FINAL REPORT

A Low-Cost, Passive Approach for Bacterial Growth and Distribution for Large-Scale Implementation of Bioaugmentation

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ACRONYMS

AOC Area of Concern

ARD anaerobic reductive dechlorination

ASTM American Society of Testing and Materials

ATSDR Agency for Toxic Substances and Disease Registry

BCI Bioremediation Consulting, Inc.

bgs below ground surface

Cal EPA California Environmental Protection Agency

CDM Camp Dresser & McKee Inc.

CERCLA Comprehensive Environmental Response, Compensation, and Liability

Act

cis-DCE cis-1,2-dichloroethene

CMT Continuous Multichannel Tubing

COC chemicals of concern COD chemical oxygen demand

CPT cone penetrometer

CSIA carbon stable isotope analysis

d days

DCE 1,2-dichloroethene

DHC Dehalococcoides species

DNA deoxyribonucleic acid

DNAPL dense, non-aqueous phase liquid

DO dissolved oxygen
DoD Department of Defense
DQO data quality objectives
GAC granular activated carbon

gpm gallons per minute

EPA Environmental Protection Agency ERH electrical resistance heating

ERSE Extended Remedial Site Evaluation

ESTCP Environmental Security Technology Certificate Program

FS Feasibility Study
ft/d feet per day
FTL field team leader
HASP health and safety plan
HDPE high density polyethylene
IR Installation Restoration

ITRC Interstate Technology Regulatory Council

L liters

LBL Lawrence Berkeley National Laboratory

m meters

MCLs maximum contaminant levels MNA monitored natural attenuation

ug/L micrograms per liter

mg/L milligrams per liter

NASA National Aeronautics and Space Administration NAVFAC ESC Naval Facilities Engineering Services Command NAVFAC SW Naval Facilities Engineering Command Southwest

NAVWPNSTA Naval Weapons Station

NFESC Naval Facilities Engineering Service Center

NPL National Priorities List
O&M operations and maintenance

OCHCA Orange County Health Care Agency

ORP oxidation/reduction potential

OSHA Occupational Safety and Health Act

PA Preliminary Assessment

PCE tetrachloroethene

PLC Programmable Logic Controller PPE personal protective equipment

qPCR quantitative polymerase chain reaction
QA/QC quality assurance/quality control
QAPP Quality Assurance Project Plan
RAB Restoration Advisory Board

RCRA Resource Conservation and Recovery Act

RDO Remedial Design Optimization RFS Revised Feasibility Study

ROD Record of Decision
RSE Removal Site Evaluation

RWQCB Regional Water Quality Control Board

SDWA Safe Drinking Water Act SOW Statement of Work TCE trichloroethene

trans-DCE trans-1,2-dichloroethene

VC vinyl chloride

VOC volatile organic compounds

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EXECUTIVE SUMMARY

Chlorinated solvents remain the most common class of contaminants at hazardous waste in the United States in general, as well as for the Department of Defense specifically. Bioremediation has emerged as a promising technology for addressing chlorinated solvents with relatively low capital costs, minimal (or no) se condary waste st reams, minimal hazard to workers and the environment, in situ contaminant de struction, low maintenance, and minimal site disturbance. However, not all contaminated sites have significant populations of the most important bacteria required for efficient bi odegradation of these contaminants, namely, *Dehalococcoides spp*. In those cases, bioaugmentation (adding a concentrated culture of the desired bacteria to a site) is becoming widely used to address potential biological limitations to degradation. While this has been demonstrated to be effective on a small scale, no rigorous full-scale demonstrations have been performed to evaluate different strategies for achieving successful growth and distribution of *Dehalococcoides spp*. bacteria to achieve site cleanup goals.

OBJECTIVES OF THE DEMONSTRATION

The over all objective of this work is to compare the cost and performance of full-scale bioaugmentation of chlorinated solvent contaminated groundwater using passive and active bacterial distribution approaches. The technical objectives for this demonstration are as follows:

- Extend bioaugmentation cost-effectively to full scale
 - Demonstrate cos t-effective ba cterial dist ribution at sca les of hundr eds, rather than tens, of feet
 - Demonstrate induction of complete dechlorination at the same scale
- Demonstrate that a low-cost, passive approach to bioaugmentation will achieve large-scale bacterial distribution and induction of complete dechlorination
- Compare and contrast effectiveness of passive and active approaches of bacterial distribution

The relative pros and cons of active recirculation and passive inject-and-drift strategies for large-scale bioaugmentation of chlorinated solvents in groundwater were evaluated in a side-by-side comparison at the Seal Beach Naval Weapons Station (NAVWPNSTA) Seal Beach Site 70 in the City of S eal B each, California. Three phases of activities were completed for each of the treatment cells, as follows:

- Phase 1 Pre-Demonstration L aboratory Investigations. Bench-scale testing was performed to demonstrate that the bi oaugmentation culture could overcome the high sulfate concentrations at the site. In addition, de oxyribonucleic acid (DNA) analysis of site groundwater samples and commercially available cultures, were used to identify "biomarkers" that provided the ability to differentiate between the injected cultures and any native *Dehalococcoides spp. (DHC)*.
- Phase 2 Tracer Test, Baseline Sampling, and "Pre-conditioning." Following treatment cell construction, a tracer test was conducted in each of the treatment cells to ve rify the groundwater hydraulics in the shallow agui fer. Baseline

sampling was then conducted to a ssess conditions, including contaminant and degradation product concentrations, redox parameters, biological activity indicators, and *DHC* concentrations. Following baseline sampling, electron donor was injected into each treatment cell to create strongly reducing conditions and remove sulfate prior to bioaugmentation.

• **Phase 3** — **Bioaugmentation and M onitoring.** This t hird and f inal phase involved injecting the dechlorinating culture into each of the two treatment cells and performing groundwater monitoring to compare with results from Phase 2.

DEMONSTRATION RESULTS

Bench-scale testing showed that complete dechlorination of TCE to ethene could be achieved even in the presence of high concentrations of sulfate, as long as sulfate-reducing conditions prevailed. Two dechlorinating cultures in microcosms with initial sulfate of 1,650 m g/L were equally successful in dechlorinating 16 mg/L TCE and 6 mg/L cis-DCE completely to ethene in 112 days with complete sulfate removal. In microcosms with much higher initial sulfate (9,270 mg/L), one of the cultures succeeded in converting all of the TCE to VC (45µM) and ethene (119 µM) in 112 days, while removing about 36 percent of the sulfate. While DNA analysis revealed low concentrations of native *DHC* at the site in a few locations, it was determined that not all of the known functional genes for dechlorination were present. Specifically, the *vcrA* gene was absent in site ground water. As this functional gene is present in commercially available dechlorination cultures, it was tentatively selected as a na ppropriate bit omarker for the bioaugmented culture pending results of DNA analysis of groundwater samples following the pre-conditioning phase.

Tracer testing performed following well installation confirmed that travel times in the two treatment cells were sufficiently short to satisfy project objectives. B aseline groundwater sampling confirmed that initial conditions were mildly reducing, with very little conversion of TCE to cis-1,2-dichloroethene (cis-DCE). It was also noted that baseline TCE conditions were quite high in both t reatment cells. In the active treatment cell, a high of 140,000 $\mu g/L$ was observed in the downgradient part of the cell, though concentrations were generally more like 5,000 to 10,000 μ g/L. In the passive cell, TCE concentrations typically ranged from 1,000 μ g/L to 3,000 μ g/L in the upgradient part of the cell, but were much higher (on the order of 50,000 to 60,000 μ g/L) in the bottom part of the middle of the passive cell. In the downgradient monitoring wells, TCE concentrations were more 10,000 to 20,000 μ g/L.

During pre-conditioning, electron donor was distributed throughout most of the passive cell, and throughout the upgr adient por tion of the active cell. Where electron donor was distributed, sulfate-reducing conditions were generally achieved, and in some locations, TCE transformation to cis-DCE was observed. However, almost no vinyl chloride was detected, and *DHC* detections were few and at very low concentrations. Most importantly for the DNA analysis of groundwater samples, no detections of the vcrA functional gene were observed, confirming its utility as a biomarker of the bioaugmentation culture.

Bioaugmentation of bot h t reatment c ells o ccurred in J anuary 2009, with t hree passive c ell injection wells receiving culture, and the two active c ell in jection and recirculation wells

receiving culture. Following bioaugmentation and during injection of one percent sodium lactate, considerable increases in numbers of DHC bacteria (ranging from > 10^6 gene copies/mL to > 10^9 gene copies/mL) and all three functional genes (tceA, bvcA, and vcrA) were observed in all wells in the upper portion of the active cell. However, electron donor distribution became less effective over time, and more frequent and higher concentration injections were required to maintain an adequate distribution and efficient DHC growth and dechlorination. Overall, conversion of TCE to ethene was proceeding effectively in the upgradient third to half of the active treatment cell, but was not observed at the monitoring well two-thirds of the way down the treatment cell axis.

In the passive treatment cell, the electron donor distribution appeared to improve over time using the original monthly injection frequency. During the post-bioaugmentation phase, TCE and DCE were mostly removed, with VC and ethene observed for the first time at injection wells PIW-2 and -3 within two weeks after inoculation in January 2009. As of October 2009, total CVOCs continue to remain low at all three injection wells. However, little to no dechlorination was observed in the upper portion of the passive cell during the post-bioaugmentation phase. While it was not conclusively demonstrated, it is speculated that inhibition of dechlorination due to the presence of othe r con taminants in this ar ea might have be en a f actor, as chloroform concentrations as high as 1,500 µ g/L and c arbon t etrachloride as high as 15,000 µg/L w ere measured in this area. In contrast, complete reductive de chlorination of T CE to ethene was observed in the central and lower portion of the passive cell. In October 2009 bi odegradation accounted for reduction of total CVOC concentrations by 72 to greater than 92 percent at central and downgradient monitoring wells compared to CVOC concentrations observed in November 2008. E thene production was observed as high as 410 μ g/L. During the post-bioaugmentation phase, DHC bacteria and functional gene (tceA and vcrA) numbers increased immediately (within 2 weeks of inoculation) at all three injection wells on the order to $>10^6$ gene copies/L, and subsequently increased to similar concentrations in the downgradient two-thirds of the cell. These concentrations were sustained through October 2009.

The growth of *DHC* was measured in each cell using DNA analysis of groundwater samples based on the total number cells at the end of the study compared to the number injected, as well as by tracking increases over time at monitoring wells. Growth was very similar in both cells, with a bout a two order of magnitude increase in cell numbers estimated in each. It was also observed that concentrations at injection wells were sustained above about 10⁶ cells/L throughout the test, and concentrations at monitoring wells increased to concentrations approximately equal to the injection wells by the end of the test. As with the first measure of growth, the two bioaugmentation strategies appeared equally effective based on this analysis.

Comparing and contrasting the distribution of *DHC* by the two bioaugmentation strategies was the key objective of this demonstration. B ased on previous studies of bacterial transport in general, and bioaugmentation specifically, groundwater velocity appeared to be one of only a few parameters than can be easily manipulated during bioremediation that might have a significant impact on transport of *DHC*. Relative distribution efficiency of passive vs. active transport was assessed by comparing travel time of injected DHC to travel of the conservative tracer (iodide) used in Phase 2 of the demonstration. The groundwater velocity in the active cell was 1 to 1.8 ft/day, and for the passive cell it was 0.22 to 0.44 ft/d, a difference of approximately a factor of 5. The tracer and *DHC* data indicated that bacterial transport was not significantly

retarded compared to groundwater flow in either the active or passive cells. In fact, arrival of *DHC* was faster than that of the conservative tracer in the majority of the passive cell monitoring wells. In the active cell, *DHC* transport velocity appeared to be approximately equal to that of the conservative tracer. These results demonstrate that *DHC* was transported more rapidly relative to groundwater flow under passive conditions than active recirculation. This is consistent with previous indications that retardation of *DHC* transport relative to a conservative tracer increases with groundwater velocity. The net result was that the passive distribution strategy provided effective distribution of *DHC* (along with complete de chlorination to ethene) over a larger portion of the treatment cell than was achieved with active recirculation.

COST ANALYSIS

Projected implementation costs f or a "typical" a pplication (not inc luding the int ensive monitoring required for a rigorous demonstration) of bioaugmentation at a 0.5-acre site using the active and passive approach were estimated based on the demonstration costs. Most of the costs are similar (e.g. start-up, general construction, monitoring, and performance assessment) because they are common to both active and passive approaches. However, the construction and O&M costs for the active approach are approximately three times as high as for the passive approach. The result is an estimated cost for the active approach of \$2.5M, compared to \$1.5M for the passive approach. The primary drivers for this cost increase are the significantly higher amount of lactate r equired, and the higher costs for construction and maintenance of r ecirculation systems. For a sitelike S eal B each, the benefits of implementing an active r ecirculation approach do not appear to be justified by the increased costs.

It should be noted, however, that some sites have conditions that would lead to more significant benefits f or r ecirculation s ystems. F or s ites with very high groundwater f low velocities, recirculation might be needed to manage residence time within the treatment zone to avoid potential off-site migration of partially chlorinated byproducts such as *cis*-DCE and VC. Such a site would also allow electron donor to be distributed over a much larger distance prior to being degraded than was possible at Seal Beach, which would also increase the benefit. On the other hand, s ites with very low groundwater velocities might make a passive system impractical because very little distribution can be achieved without enhancing the hydraulic gradient. What this demonstration indicates is that for sites that are closer to the "average" in terms of groundwater velocity, passive bioaugmentation systems are likely to be more cost-effective than active systems.

1.0 INTRODUCTION

This r eport pr ovides t he c ost a nd pe rformance da ta f or f ull-scale bi oaugmentation systems designed to transform chlorinated ethenes to e thene in groundwater. In particular, this report demonstrates the r elative pros and cons of active recirculation and passive i nject-and-drift strategies as a side-by-side comparison between the two approaches for large-scale bioaugmentation of chlorinated solvents in groundwater at the Seal Beach Naval Weapons Station (NAVWPNSTA) Site 70 in the City of Seal Beach, California. This project is sponsored by the Environmental Security Technology Certification Program (ESTCP) Project CU-0513, with additional funds provided by Naval Facilities Engineering Command Southwest (NAVFAC SW). The principal investigator for this project is Mr. Joey Trotsky from Naval Facilities – Engineering Services Command (NAVFAC ESC), and the co-principal investigator is Dr. Kent Sorenson of Camp Dresser & McKee Inc. (CDM). C DM is a demonstration partner under contract number N68711-05-C-0063.

The two full-scale bioaugmentation strategies were evaluated in treatment cells in the same chlorinated solvent source area at Site 70. Three phases of activities were completed for each of the treatment cells, as follows:

- Phase 1 Pre-Demonstration L aboratory Investigations. Bench-scale testing was performed to demonstrate that the bioaugmentation culture could overcome the high sulfate concentrations at the site. In addition, de oxyribonucleic acid (DNA) analysis of site groundwater samples and commercially available cultures, including quantitative polymerase chain reaction (qPCR), clone lib rary development, and DNAs equencing were used to identify "biomarkers" that provided the ability to differentiate between the injected cultures and any existing Dehalococcoides spp. (DHC) that may have naturally existed in the groundwater.
- Phase 2 Tracer Test, Baseline Sampling, and "Pre-conditioning." Following treatment cell construction, a tracer test was conducted in each of the treatment cells to verify the groundwater hydraulics in the shallow a quifer. Following the tracer t est, ba seline s ampling was conducted to assess ba seline c onditions including contaminant and degradation product concentrations, redox parameters, biological activity indicators, and *DHC* concentrations. Following ba seline sampling, electron donor was injected into each treatment cell to create strongly reducing conditions and remove sulfate prior to bioaugmentation.
- Phase 3 Bioaugmentation and M onitoring. This third and f inal phase involved injecting the dechlorinating culture into each of the two treatment cells and performing groundwater monitoring to compare with results from Phase 2.

The remainder of Section 1 briefly discusses background information, demonstration objectives, and regulatory drivers. Section 2 c ontains a description of the technology to be demonstrated. The performance objectives are provided in Section 3, and Section 4 gives as ite description. Section 5 o utlines the test design and results, while Section 6 provides a detailed performance assessment. Section 7 uses the demonstration data to provide a cost assessment of the technology, and Section 8 outlines implementation issues.

1.1 BACKGROUND

Chlorinated solvents are the most common class of contaminants in groundwater at hazardous waste sites in the U.S. In 1993, the Agency for Toxic Substances and Disease Registry (ATSDR) compiled a list of the top 25 c ontaminants detected at hazardous waste sites on the National Priorities List (NPL). The ATSDR ranking identified 8 of the top 20 contaminants as chlorinated solvents and their intrinsic de gradation products, including two of the top three (Pankow & Cherry, 1996). The ranking was updated by the ATSDR on their Internet site based on 1996 data with similar results. Of particular significance is the identification of trichloroethene (TCE) and tetrachloroethene (PCE) as the first and third most common contaminants at NPL sites in both surveys. Chlorinated solvents are also the most common contaminants at Department of Defense (DoD) sites. While NAVWPNSTA Site 70 is not on the NPL, it does have chlorinated solvent-contaminated groundwater.

While s ignificant pr ogress has be en made in addressing solvent sites, parties responsible for cleaning up si tes w ith chlo rinated solvents in groundwater are s till f aced w ith several technologies w ith significant cap ital costs, secondary wastes treams, the involvement of hazardous materials, and the potential for additional worker or environmental exposure. A more ideal technology would involve lower capital costs, would not generate secondary waste streams, would be non-hazardous to workers and the environment, would destroy contaminants in situ, would be low maintenance, and would minimize disturbance of the site.

Bioremediation has been identified as one of the major technologies that may be able to address this problem at chlorinated solvent sites. However, bacteria capable of complete dechlorination of chloroethenes to ethene are not always present at these sites, which can cause dechlorination to "stall" at *cis*-1,2-dichoroethene (cis-DCE). When this occurs, one mitigation strategy is to perform bi oaugmentation, which is the introduction of bacteriac apable of complete dechlorination to ethene into the affected groundwater. This process has only been successfully demonstrated at the pilot scale, however, and many issues related to full-scale implementation with important cost implications still need to be addressed.

Previous bi oaugmentation pi lot studies were conducted on the scale of tens of feet and used active recirculation for distribution of the bi oaugmentation culture. The current demonstration will complement and build on pilot testing a lready completed by NAVFAC SW at NAVWPNSTAS eal Beach, Site 40 that successfully uses a low-cost, passive approach for implementation of bioaugmentation. The purpose of this demonstration is to compare the low-cost, passive method for implementation of bioaugmentation to the active recirculation method for full-scale application at a scale of hundreds of feet or more.

1.2 OBJECTIVE OF THE DEMONSTRATION

The ove rall objective of t his w ork is to compare the cost and p erformance of full-scale bioaugmentation of c hlorinated s olvent c ontaminated groundwater u sing passive and a ctive distribution approaches. The technical objectives for this demonstration are as follows:

• Extend bioaugmentation cost-effectively to full scale

- Demonstrate cost-effective bacterial distribution at a scale of greater than one hundred feet, r ather t han tens of feet as ha s pr eviously b een demonstrated
- Demonstrate induction of complete dechlorination at the same scale
- Demonstrate that a low-cost, passive approach to bioaugmentation will achieve large-scale bacterial distribution and induction of complete dechlorination
- Compare and contrast effectiveness of passive and active approaches of bacterial distribution

Specific performance objectives for each test scenario are provided in Section 3.

1.3 REGULATORY DRIVERS

The presence of chlorinated solvents including PCE, TCE, *cis*-DCE, *trans*-1,2-dichloroethene (*trans*-DCE), and vinyl chloride (VC) in groundwater is one of the most persistent environmental problems at NPL si tes, as discus sed in Section 1.1. The Safe D rinking Water A ct (S DWA) maximum contaminant levels (MCLs) for these compounds are very low, as shown in Table 1-1, which makes cleanup of these sites difficult given that solubilities can be six orders of magnitude above the MCL.

Table 1-1. Regulatory Limits for Chlorinated Compounds

Compound	Regulatory Limit (MCL) ¹ mg/L	Solubility @ 25°C mg/L
Tetrachloroethene	0.005	150^2
Trichloroethene	0.005	$1,100^2$
cis-1,2-dichloroethene	0.07	$3,500^3$
trans-1,2-dichloroethene	0.1	$6,300^2$
Vinyl chloride	0.002	2,763 ⁴

⁴⁰ CFR 141.61

² Knox et al., 1993

³ Howard, 1990

⁴ Howard, 1989

2.0 TECHNOLOGY

The first publications de scribing field-scale bioaugmentation us ing *DHC* bacteria to treat chlorinated ethenes appeared in about 2000, so this is still a relatively new technology for full-scale field applications. This section provides a description of the underlying theory that is fundamental for technology application, an overview of the history of the development of the technology, and a brief comparison of the advantages and limitations of bioaugmentation relative to other source remediation technologies.

In general, bioaugmentation for remediation of chlorinated solvents involves addition of electron donor (biostimulation) and a bacterial culture that contains *DHC*. Different techniques are available for bioaugmentation of groundwater, and the appropriate technique depends not only on the relevant application (i.e., plume containment vs. source treatment), but a lso on the electron donor selected. Because all bioaugmentation methods require the addition of electron donor, it is important to consider the electron donor delivery method when selecting a bioaugmentation approach. Several electron donor emplacement methodologies have been used for biostimulation, including (adapted from Interstate Technology Regulatory Council (ITRC) [2005]):

- Conventional injection wells one or an etwork of wells is usually used with large volume, 1 iquid e lectron donor i njections; most a pplicable f or m oderate t o high permeability conditions
- Direct-push injection points a network of more closely spaced points is usually used with s mall vol ume, I iquid e lectron donor i njections; m ost a pplicable f or re latively homogeneous, moderate to high permeability conditions with low to medium advection to dispersion ratios
- Trenching passive trenches are usually backfilled with a large mass of solid electron donor (e.g., mulch or chitin) and/or a long-lived liquid electron donor, often mixed with sand; can be used in all permeability conditions as long as the permeability of the trench is at least as high as the formation
- Hydraulic or pneumatic fracturing either solid or liquid electron donors are emplaced during or immediately after fracturing; generally used in low permeability conditions or highly heterogeneous conditions in which low permeability zones require treatment

The c urrent de monstration f ocuses on i mplementing bot h pa ssive a nd a ctive approaches f or bioaugmentation, both of which use conventional injection wells.

2.1 TECHNOLOGY DESCRIPTION

This description of the fundamentals required for a pplication of the technology provides a n overview of bi oaugmentation for c hlorinated s olvent contaminated groundwater. First, a discussion of the basics of chlorinated ethene degradation is provided. Second, issues related to scale-up of bioaugmentation are presented. Finally, factors that can affect bacterial transport in the subsurface are discussed.

2.1.1 Chlorinated Ethene Degradation

Complete biological reductive dechlorination of PCE and TCE to ethene was first documented only 2 decades ago (Freedman and Gossett, 1989), and the pathway was observed to proceed as follows: PCE → TCE → DCE → VC → ethene. It has since been well documented (DiStefano et al., 1991; deBruin et al., 1992; DiStefano et al., 1992; Ballapragada et al., 1997; Fennell et al., 1997; C arr and Hughes, 1998) and is being used successfully to treat chlorinated ethenes in groundwater (e.g., Song et al., 2002). Complete reductive de chlorination generally has two requirements. First, redox conditions must be sufficiently reducing that reductive dechlorination of DCE and VC to ethene is thermodynamically favorable. The free energy yielded by r edox reactions v aries substantially depending upon the el ectron acceptor. During r espiration, microorganisms will preferentially use the electron acceptors yielding the greatest free energy (e.g., Bouwer, 1994). The order of preference for the most common inorganic electron acceptors is oxygen, nitrate, manganese (IV), iron (III), sulfate, and carbon dioxide (Bouwer, 1994; Cord-Ruwisch et al., 1988). Therefore, the dominant microbial community in a groundwater system is largely dependent upon the distribution of electron acceptors. While PCE and TCE reduction might oc cur unde r i ron-reducing c onditions, reduction of D CE and VC to e thene ge nerally requires a t least s ulfate r educing c onditions, o r more pr eferably m ethanogenic conditions (Semprini e t a l., 1995; S orenson, 2000; N AVFAC, 2003, ht tp://www.ert2.org/dce/tool.aspx). When e lectron donor is limited, conditions will often not be sufficiently reducing to a chieve complete dechlorination, causing it to "stall" at DCE. This can be overcome simply through the addition of a compound that acts as an electron donor, often consisting of a fermentable carbon source (Sorenson, 2003).

The second requirement for complete reductive dechlorination is a biological community capable of carrying out the reaction. It is widely accepted that bacteria capable of ana erobic reductive dechlorination are vital to biological dehalogenation processes in anoxic environments (Smidt et al., 2000). In fact, an increasing body of evidence suggests that complete biological reductive dechlorination of PCE and TCE to ethene requires the presence of a strain of the bacterium DHC (Cupples et al., 2003; He et al., 2003; Hendrickson et al., 2002). Recent advances in molecular techniques now allow scientists to characterize microbial communities, including identification of de chlorinators, m ore fully. This has I ead to the discovery of m any or ganisms c apable of dechlorinating various compounds (Holliger et al., 1999). Many of these organisms are capable of reducing PCE and TCE to DCE (Holliger et al., 1999; Drzyzga and Gottschalk, 2002), but only DHC have been found to be capable of complete dechlorination of PCE and TCE to ethene in a pur e c ulture (Maymó-Gatell e t a l., 1997; Maymó-Gatell e t a l., 1999; Maymó-Gatell a nd Zinder, 200 1). A di fferent s train, DHC strain F L2, ha s be en i mplicated f or c omplete dechlorination in a mixed culture, but it has not been isolated to date (Löffler et al., 2000). Of particular importance is that a recent study of 24 field sites in North America and Europe found that strains of this organism were present at all 21 sites that exhibited complete dechlorination to ethene, while none were found at the three sites examined where dechlorination stopped at cis-DCE (Hendrickson et al., 2002). This suggests that while DHC are relatively common and widely distributed, their absence at a site might prevent complete dechlorination. It should be noted that detection of the DHC genus does not necessarily mean that complete dechlorination of PCE or TCE will occur at a site because some strains are not capable of dechlorinating PCE and TCE. For example, s train C BDB1 grows by t he de chlorination of c hlorinated b enzenes and possibly dioxins, but cannot grow by dechlorination of PCE or TCE (Adrian et al., 2000; Bunge et al., 2003).

2.1.2 Bioaugmentation Scale-Up Issues

Bioaugmentation, the *in situ* addition of an exogenous bacterial culture containing *DHC* (in this case) to site groundwater, is gaining acceptance as a viable strategy for remediation of chlorinated solvents in groundwater, especially when these bacteria are not naturally present at a site and reductive de chlorination is found to "stall" at *cis*-DCE. Several la boratory cultures containing *DHC*, e.g., *Dehalococcoides ethenogenes* strain 195, have been shown to be capable of complete de chlorination of PCE, TCE, and DCE to ethene (Fennell et al., 2001; Maymó-Gatell et al., 1999; Maymó-Gatell et al., 1997; Richardson et al., 2002). In a ddition, several studies have demonstrated that bi oaugmentation using *DHC*-containing mixed cultures can overcome DCE stall and facilitate complete dechlorination at the field pilot scale (Ellis et al., 2000; Lendvay et al., 2003; Major et al., 2002).

While these results are very promising, the transport scale of this work has been no greater than 30 f eet. To receive regulatory and DoD end us er acceptance, cost-effective approaches for growing large volumes of *DHC*-containing cultures and distributing them across a scale of hundreds of feet or more need to be demonstrated and validated. In particular, distribution of bacteria on a large scale presents a challenge both from a subsurface transport and from a cost standpoint. The distribution of introduced cultures during bioaugmentation is generally quite limited initially both because of the adhesion of bacteria to the soil matrix and the filtering effect of soil to particles such as bacteria. Although low-adhesion strains of bacteria have been developed for bioaugmentation in some applications (Steffan et al., 1999), this is only possible with pure cultures. Because *Dehalococcoides ethenogenes* is only grown in mixed culture for bioaugmentation, its adhesion has not been manipulated.

Filtration theory has been used to model bacterial transport during injection, and predicts that soil will be an efficient filter for bacteria, reducing concentrations by several orders of magnitude within the first meter of transport from the injection well and generally limiting transport to less than 2 meters (m) from the injection location, even in the absence of sorption (Goltz et al., 2001; Martin e t a 1., 1996). During pi lot-scale d emonstrations, D. ethenogenes has be en further distributed after inoculation through forced advection (recirculation) systems (these are described in more detail in Section 2.2). While these systems have been effective at transporting bacteria approximately 10 feet in 5 weeks (Lendvay et al., 2003) or up to 30 feet in 3 months (Major et al., 2002), larger scale distribution has not been well documented. Furthermore, the use of such systems on a scale of hundreds of feet would either require many injection and extraction wells to achieve distribution on a similar time scale, or would require much higher extraction rates. Thus, the cost of scale-up could be very high. At active sites, cost increases go beyond merely the scale because recirculation pipes must be installed across roads, railroad tracks, or utilities, all of which can be problematic. A further complication is that obtaining regulatory approval to extract and reinject contaminated groundwater remains challenging at many sites. In some cases treating the extracted water is required, which eliminates many of the benefits of bioremediation.

2.1.3 Factors Affecting Bacterial Transport

The many factors that affect bacterial transport in the subsurface are widely varied and complex. Some of the physiological factors that have been implicated as influencing bacterial transport include cell size and shape, motility, cell wall type, and adsorption characteristics (Becker et al., 2004; C amesano and L ogan, 1998; Witt et al., 1999). For bi oaugmentation, i noculation fluid characteristics such as ionic strength and cell concentration have been identified as playing a role (Camesano and L ogan, 1998; Gross and L ogan, 1995), as well as flow velocity (Becker et al., 2004; C amesano a nd L ogan, 1998). Other r esearchers have sugge sted that the phys ical heterogeneity of the p orous medium is a primary factor influencing bacterial at tachment (Campbell R ehmann and W elty, 1999; F ontes et al., 1991; Renet al., 2000). Finally, heterogeneity of the attachment characteristics within a particular bacterial population has also been implicated as affecting transport (Mailloux et al., 2003; Albinger et al., 1994; Glynn et al., 1998).

With all of these factors contributing to bacterial transport, development of a rigorous model that accurately accounts for any one of the factors, let alone the interactions of several factors, would be a lofty goal. Taking flow velocity as an example, some studies have found that attachment of motile bacteria to porous media inc reased more with decreased flow rates than nonmotile bacteria (Becker et al., 2004), while others have found that it increased less (Camesano and Logan, 1998). As different bacteria were used in the studies, it is likely that some of the other factors mentioned above also played a critical role, but that there are simply too many variables to de sign a comprehensive study that can elucidate their complex interactions. Becker et al. (2004) noted that some of their results for different flow rates were "perplexing;" that is, flow rate a ffects transport behavior in ways that a re not well understood. To complicate matters further, it has be en not ed that laboratory studies of bacterial transport have not successfully predicted field-scale transport (Harvey et al., 1993).

Although the complexity of bacterial transport and the development of a general, predictive model that can be used to design bioaugmentation strategies for a wide range of bacteria is daunting, such a ge neral unde retanding is not required in the specific case of optimizing strategies f or bi oaugmentation using DHC-containing cultures f or c hlorinated s olvent remediation. Given that the focus is on only one population of bacteria, the physiological factors that affect transport are no longer variable, and an empirical approach can be used to evaluate the remaining factors. An empirical approach is further justified by the difficulty noted above in accurately representing field-scale transport phenomena at the laboratory scale. Ignoring the physiological f actors, the t ransport f actors remaining that c an be c ontrolled during bioaugmentation are reduced to flow velocity, ionic strength, and cell concentration. While low ionic strength solutions have been shown to improve bacterial transport (Gross and Logan, 1995; Fontes et al., 1991), the improvements are not a lways large, and the logistical difficulty of injecting large volumes of low ionic strength solutions at field scale in varied geologic conditions is problematic (Camesano and Logan, 1998). The degree to which bacterial dispersal can be achieved at high concentrations depends upon whether the cells exhibit "blocking" behavior or "ripening" behavior (Camesano and Logan, 1998). Blocking implies that the cells do not tend to stick to each other, so they block attachment sites, forcing other cells to flow beyond them. This behavior allows high cell concentrations to be used to enhance dispersal. Ripening implies that the cells adhere strongly to each other and tend to increase the filtering efficiency of the porous

medium, pr eventing d istribution of c ells at high c oncentrations. As it has a lready been demonstrated that injection of DHC at relatively high concentrations ($\sim 10^8$ cells/mL) can be used successfully to a chieve distribution at a scale of tens of feet, ripening does not appear to be a problem. Thus, for a given site, flow velocity appears to be one of the most important factors affecting bacterial transport that can easily be controlled during full-scale implementation.

While the fundamental issues a ffecting transport of *DHC* (or bacteria in general) are not well understood, results from a recent study at NAVWPNSTA Site 40 (see Section 2.2) suggest that a passive distribution system (low velocity) may be far more cost-effective for scale-up than an active recirculation system (high velocity). This study is designed to validate these results by measuring *DHC* transport and the resulting induction of complete dechlorination using bot h passive and active distribution approaches at full scale. The empirical approach described herein will provide information regarding a potential key control on bacterial transport at full scale, avoiding the concern of representativeness of laboratory-scale studies. It will also provide this information in a timely manner so that the results can be applied to current problems quickly, which would be very unlikely if a fundamental research approach were used.

2.2 TECHNOLOGY DEVELOPMENT

The use of active recirculation to distribute bacteria and induce complete dechlorination is well documented at the pilot scale (Ellis et al., 2000; Lendvay et al., 2003; Major et al., 2002; Hood et al., 2008), although sufficient sampling was not performed in all cases in order to provide a full assessment of bacterial growth and distribution. For example, in the Ellis et al. study (2000) at Dover Air Force Base, DHC was not analyzed in field samples. For the Bachman Road Site study (Lendvay et al., 2003), DHC analysis was performed, but it was already present in the first post-inoculation samples 35 days after inoculation. A study at Kelly Air Force Base (Major et al., 2002) was the only one for which DHC transport times could be reasonably estimated and compared to conservative transport times. Based on bromide transport data and DHC detections provided in Major et al., travel times for DHC were between 61 and 176 times longer than conservative transport. Based on the fact that VC was detected 15 days a fter inoculation, the Bachman Road Site study suggested that DHC transport time along the short flow path from the injection/inoculation wells (approximately 3.2 m eters) was only a bout 2.3 t imes greater than conservative transport times based on the reported Darcy velocity for the test area (Lendvay et al., 2003). The Cape Canaveral L C-34 project (Hood et al., 2008) had DHC bacteria al ready present in the treatment cell prior to bioaugmentation; however, post-bioaugmentation operations showed a 2-3 orders of magnitude increase in cell counts, as well as significant production of ethene. Still, quantification of transport of the added *DHC* bacteria could not be performed.

While these studies were conducted on small scales, other studies looked at bi oaugmentation using active recirculation at a larger scale. Scheutz et al., 2008 us ed a ctive recirculation for bioaugmentation at a larger scale (approximately 100 f eet be tween injection and extraction wells). This field demonstration showed distribution of electron donor more than 65 f eet from injection wells, as well as induction of dechlorination to ethene at a similar scale. However, it was determined that indigenous bacteria were capable of performing dechlorination to ethene, and that the *vcrA* gene that encodes for VC reductase was present during baseline sampling. Bioaugmentation was performed to reduce lag times for complete dechlorination; how ever, quantification of transport of introduced bacteria could not be performed.

In contrast to these recirculation systems, passive *DHC* distribution appr oach was recently demonstrated in a bioaugmentation pilot test at NAVWPNSTA Seal Beach, Site 40. Prior to the bioaugmentation phase at Site 40, b iostimulation was performed for 8 m onths to overcome the electron donor limitation at the site, which initially had sulfate concentrations of 200 to 500 milligrams per liter (mg/L) (Figure 2-1). As predicted b ased on thermodynamics and field observations (Bouwer, 1994; Semprini et al., 1995; Sorenson, 2000; NAVFAC, 2003), dechlorination of PCE to DCE occurred shortly after the onset of sulfate reduction and the removal of sulfate within about 2 to 3 months after the start of biostimulation (French et al., 2003; Rahm et al., 2006). Although conditions became methanogenic and electron donor was abundant for over 6 months, dechlorination beyond *cis*-DCE did not occur after more than a year (Figure 2-2(a)). Highly sensitive DNA analysis performed after biostimulation revealed that no *DHC* were present at the site (Rahm et al., 2006).

In April 2003, two wells (MW-40-22 and MW-40-25, see Figure 2-3), were inoculated with 20 liters (L) each of a com mercially available *DHC*-containing c ulture. Forced adve ction occurred only during brief periods when sodium lactate was periodically injected in MW-40-28, approximately 8 f eet f rom the two inoculation wells (F igure 2 -3). No ot her i njection o r extraction was performed during the test. During injection, the average hydraulic gradient in the treatment cell was 0.004, while it was approximately 0.00024 under ambient conditions. Based on these conditions, the injection durations, an average hydraulic conductivity of 97 feet per day (ft/d), and an estimated effective porosity of 0.20, the expected travel times for groundwater to move from inoculation well MW-40-22 to downgradient monitoring wells MW-40-23 (7.2 feet) and MW-40-24 (16.5 feet) are 26 d ays and 93 d ays, respectively (Table 2-1). In MW-40-23, DHC were detected using qPCR in the first post-inoculation samples analyzed from that well, some 91 days after inoculation (Figure 2-4) (see also Rahm et al., 2006). Thus, the maximum travel time for *DHC* was about 3.5 times longer than that expected for conservative transport with groundwater. The detection of VC at this location in the June 18 sample (Table 2-2), however, suggests that DHC activity may have been present much earlier, just 63 days after inoculation. In that case the travel for DHC would be only 2.4 times longer than conservative transport.

Similar to MW-40-23, *DHC* were detected in the first post-inoculation samples analyzed from MW-40-24, in this case 119 days after inoculation (Figure 2-4 and Table 2-1). VC was actually not detected at this location until the next sampling round (Table 2-2), so 119 days is likely close to the actual arrival of *DHC* at this well. The travel time for *DHC* was therefore only about 1.3 times longer than would be expected for conservative transport. Therefore, although groundwater velocities were fairly slow (0.12 ft/d without injection) in this passive system, transport of *DHC* was only slightly retarded. In addition, no lag time was observed for dechlorination activity after inoculation. VC and ethene were both observed for the first time in the inoculation wells about 1 week after inoculation (which might have been facilitated by the strongly reducing conditions already present). As noted above, VC was also detected in downgradient monitoring wells within a few weeks of the estimated arrival time of *DHC*.

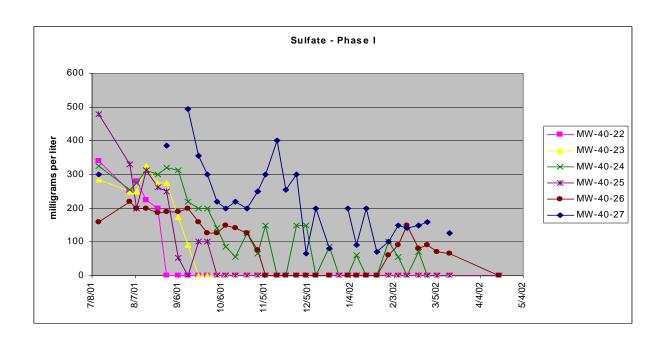


Figure 2-1. Sulfate removal at Seal Beach Site 40 wells following the start of lactate injections during biostimulation.

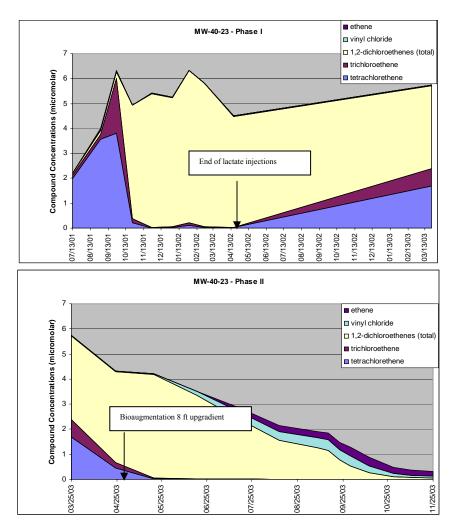


Figure 2-2. Typical dechlorination results during biostimulation at Seal Beach Site 40 including stoichiometric conversion of PCE to cis-1,2-DCE without any production of vinyl chloride or ethene, and with some rebound of PCE and TCE in the absence of lactate injections (a). Typical dechlorination results following bioaugmentation including disappearance of cis-1,2-DCE concomitant with the appearance of vinyl chloride and ethene; chloroethenes near or below MCLs after 8 months (b).

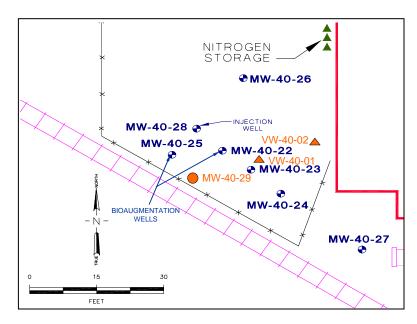


Figure 2-3. Site plan for pilot test at Seal Beach Site 40.

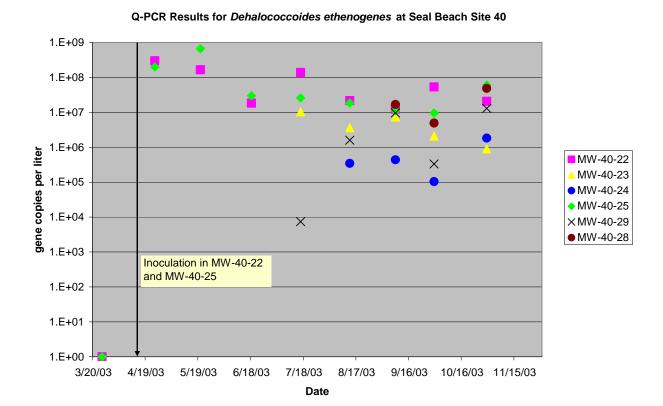


Figure 2-4. Q-PCR results for *D. ethenogenes* at Seal Beach Site 40 showing passive transport of the bacteria more than 16 ft downgradient (MW-40-24) and 8 ft upgradient (MW-40-28) in a few months.

Table 2-1. Estimated Retardation Factors for *DHC* Transport for Kelly Air Force Base Recirculation System (estimated from Major et al. (2002) and the NAVWPNSTA Site 40 passive system)

	Kelly Air	Force Base	NAVWPNSTA Site 40	
Well	B1	E1	MW-40-23	MW-40-24
Distance from inoculation point (ft)	7.9	30	7.2	16.5
Estimated conservative travel time (d)	0.17	1.2	26	93
Estimated <i>D. ethenogenes</i> travel time (d)	17 < t < 30	72 < t < 93	63 < t < 91	119
Retardation factor	100 < R < 176	61 < R < 79	2.4 < R < 3.5	1.3

Table 2-2. Chloroethene and Ethene Concentrations ($\mu g/L$) for Bioaugmentation Pilot Test at NAVWPNSTA Site 40 (initial VC and ethene were < 2 $\mu g/L$ at all wells and ethene arrived with VC in all wells).

Well	Max DCE	Final DCE	Max VC	Final VC	First VC	Max Ethene
MW-40-22	310	4	45	2.7	April 24	5
MW-40-25	390	4	62	2	April 24	8
MW-40-23	400	6	26	4.8	June 18	9
MW-40-29	410	16	30	8.3	June 18	6
MW-40-24	410	35	63	31	August 14	21

The introduced *DHC* were observed not only 16 feet downgradient from the inoculation point, but also 8 feet upgradient in the lactate injection well, in less than 4 months (Figures 2-3 and 2-4) (see also Rahm et al., 2006). Just as important, the arrival of *DHC* corresponded closely to the first appearance of VC and ethene in each of the monitoring wells (Table 2-2). Furthermore, concentrations of PCE, *cis*-DCE, and VC were all near or below MCLs throughout the treatment area in less than 8 months (Figure 2-2(b), Table 2-2). While aqueous and soil gas concentrations of degradation products only a counted for approximately 50 percent of the mass of *cis*-DCE degraded (data not shown), many months of bi ostimulation data with far larger electron donor injections demonstrated that dilution or displacement did not play a significant role in *cis*-DCE's disappearance (Figure 2-2(a)).

Table 2-1 summarizes the bacterial transport that was observed at both the Kelly Air Force Base study and NAVWPNSTA Site 40. From Table 2-1, the Kelly Air Force Base travel times suggest far greater retardation of *DHC* than was observed at Site 40. As the same culture was used in both cases, the reason for this disparity is not clear. One significant difference was the electron donor solution used. At Kelly Air Force Base, the solution consisted of a combination of a time-weighted a verage of 3.6 mM methanol (approximately 115 mg/L) and 3.6 mM acetate (and approximately 212 mg/L). At NAVWPNSTA Site 40, a 3 percent solution of sodium lactate was injected weekly for 5 w eeks, then the f requency was de creased to less t han monthly. Groundwater was methanogenic in both studies prior to bioaugmentation, so redox conditions do not appear to be a factor in the transport differences observed. Another significant difference was the use of a recirculation system at Kelly Air Force Base compared to the passive system at NAVWPNSTA Site 40. In the case of motile bacteria, at least one study has shown that they were a ctually transported more effectively under low flow or no f low conditions than under forced advection conditions (Camesano and Logan, 1998). While it is not known whether flow

conditions are an important factor for distribution of *DHC*, which are non-motile, this possibility cannot be dism issed. As discus sed above, flow conditions are the primary factor affecting transport of a given bacterium in the field that we can easily control. It is interesting that *DHC* were detected under passive conditions at NAVWPNSTA Site 40 not only downgradient, but also in MW-40-28, 8.1 feet upgradient of the inoculation well. This seems remarkable given that this transport occurred without any injections in the inoculation well to facilitate it.

The lag time prior to onset of dechlorination was insignificant in Lendvay et al. (2003), as was true at NAVWPNSTA Site 40. This raises the question of whether transport of *DHC* might be related to its growth. The two studies that had insignificant dechlorination lag periods (and presumably more rapid growth) showed *DHC* transport that was only mildly retarded relative to conservative transport, while the Kelly Air Force Base study had a significant lag period and exhibited greatly retarded transport of *DHC*. Prior to this demonstration, any potential connection between growth and transport or flow conditions and transport would have be en speculative, but this demonstration directly measured the latter at full scale. In any case, the passive distribution system was not only highly successful for destroying *cis*-DCE at the pilot scale at Site 40, but also appeared to be equal or superior to more expensive and logistically challenging recirculation systems for distributing *DHC* throughout the area of interest (Table 2-2).

The second issue for which preliminary data have been collected at NAVWPNSTA, Site 40 is the potential ease with which the *in situ* dechlorinating community can be transferred from one location to another after growth and adaptation under site conditions. The bioaugmentation pilot test w as c ompleted in the fall of 2003. In J une 2004, i n or der t o d emonstrate the proof of principle f or the c oncept of r edistributing t he *in situ DHC*-containing c ommunity f rom t he bioaugmentation a rea to ne w a reas on-site, g roundwater from M W-40-22 i n t he pi lot a rea (Figure 2-3) w as pum ped i nto a tank a nd r einjected i nto MW-40-27, l ocated do wngradient. Sodium lactate had been injected into MW-40-27 in April 2004 to create conditions that would facilitate rapid growth of the dechlorinating populations. This exercise also served to help ensure the s ustenance of t he d echlorinating populations, w hich were be ing transferred from an area where chlorinated compounds were depleted to an area where PCE concentrations were above 800 micrograms per liter (µg/L).

Figures 2-5 and 2-6 reveal the results of this "proof of principle." Figure 2-5 shows the chemical oxygen demand (COD) increasing in response to the April lactate injections. It also shows the increase in the number of gene copies per liter of groundwater measured in the well following transfer of groundwater from the pilot test area.

It is particularly interesting to not et he *DHC* DNA concentrations were a ctually an or der of magnitude higher in MW-40-27 in July 2004 than they were in MW-40-22 (the well from which groundwater containing large amounts of *DHC* DNA was transferred). This indicates that *DHC* were actively growing to large numbers in MW-40-27 within 1 month. This point is further demonstrated by the data shown in Figure 2-6. While a small increase in DCE was observed prior to injection of water from the bioaugmentation area, PCE concentrations also increased slightly.

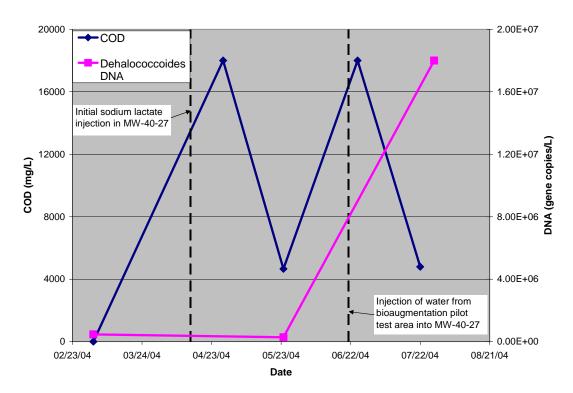


Figure 2-5. COD concentrations in well MW-40-27 at Seal Beach in response to lactate injections, and *Dehalococcoides* DNA concentrations in response to transferring groundwater from the bioaugmentation pilot test area.

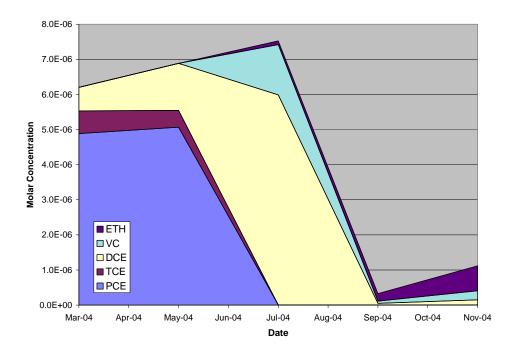


Figure 2-6. Dechlorination at well MW-40-27 following injection of water from MW-40-22. Note that initial PCE concentrations were 810 ug/L, TCE was at 85 ug/L, and DCE was 65 ug/L.

Within 1 month of the injection of MW-40-22 water in MW-40-27, however, PCE and TCE both were undetectable, VC and ethene were detected at significant levels for the first time, and the degradation products more than accounted for the mass of PCE that was degraded. These data demonstrate that once a robust, dechlorinating community is established *in situ*, it can easily be transported around a site to facilitate semi-passive distribution at a large scale without the need to pur chase large volumes of culture or construct large-scale recirculation systems. The same thing could potentially be accomplished either by waiting long periods of time for transport of the bacteria with the natural gradient, or by installing a groundwater recirculation system, both of which have been done, but this approach appears to be much more cost-effective. This is an important consideration a llowing for the potential use of a passive bacterial growth and distribution strategy.

2.3 ADVANTAGES AND LIMITATIONS OF THE TECHNOLOGY

Significant advantages of bi oaugmentation technology in general include low risk to hum an health and the environment during implementation, low secondary waste generation, minimal impacts during operations, and overall risk reduction. In addition, when applied in a source area, bioaugmentation offers the potential for complete source cleanup using one technology without a requirement for separate polishing technologies, which is a significant advantage from a cost standpoint. Source removal technologies generally do not remove all of the chlorinated solvent present, and rely on polishing technologies including *in situ* bioremediation and monitored natural attenuation to a chieve cleanup standards. *In situ* bioremediation with bioaugmentation integrates source removal and polishing, thereby facilitating attainment of cleanup goals by reducing the need for further infrastructure, treatability studies, modification of site conditions, etc. that may be required to implement a polishing technology following source removal.

Challenges for bioaugmentation can include any of the site-specific characteristics that I imit application of many remedial technologies, including complex I ithology, I ow permeability media, and high concentrations of competing electrons acceptors. In addition, this technology is probably not a pplicable for sites contaminated by large volumes of free-phase dense, non-aqueous phase liquid (DNAPL) (ITRC, 2005). Finally, the generation of methane is common at bioremediation/bioaugmentation sites, as is the temporary production of VC. Both of these can partition into the vadose zone above the water table, which can be a concern if the contamination is present in shallow groundwater underneath buildings or utility corridors.

Other t echnologies currently de monstrated a t pi lot or f ull-scale f or c hlorinated s olvent remediation i nclude thermal t echnologies (i.e., steam, electrical resistance he ating or E RH, conductive he ating), *in situ* chemical oxidation, surfactant/cosolvent floods, S VE/air s parging, and pump and treat. Steam injection and ERH both heat soil and water to volatilize chlorinated solvents f or r ecovery, while *in situ* chemical oxidation d estroys c ontaminants *in situ* using Fenton's reagent or other oxidants. Pumping and treating groundwater is currently used more for hydraulic c ontainment, or t o i nduce a gr adient, t han f or c hlorinated s olvent remediation; however, this t echnology i s f requently us ed as a ba seline f or c omparison. Table 2 -3 lists advantages and limitations of each.

Table 2-3. Advantages and Limitations of Competing Technologies

Technology	Advantages	Limitations
Bioremediation/ Bioaugmentation	All treatment performed in situ; low infrastructure and energy requirements; no secondary waste produced; costs moderate.	Relatively slower; requires longer monitoring period; not applicable for large volumes of free-phase DNAPL; production of methane and VC must be considered
Thermal (steam, ERH, conductive heating)	Relatively rapid source reduction possible; can be used for large volumes of free-phase DNAPL	Energy intensive, expensive; high secondary waste production
In situ chemical oxidation	Source reduction might be more rapid than bioremediation, though this is not well-documented; very little secondary waste produced.	Carbonates and organics compete for hydroxyl radical, oxidant is quickly consumed, limiting distribution in the subsurface; rebound of contaminants common; not applicable for large volumes of free-phase DNAPL
Surfactant/cosolvent flooding	Can be used for large volumes of free-phase DNAPL	Requires uniform, moderate to high permeability; high secondary waste production; only applicable for source areas; usually expensive and requires polishing
SVE/air sparging	SVE effective for vadose zone, short-term costs moderate.	Ineffective for source removal; air sparging requires intensive research at pilot scale; typically requires offgas treatment at the surface
Pump and treat	Effective for hydraulic containment during remediation	Ineffective for source removal; difficult to terminate operations; expensive

In addition to the general advantages and limitations for bioaugmentation discussed above, each bioaugmentation a pproach be ing t ested in this demonstration has its own a dvantages and limitations. For the active recirculation for bioaugmentation, the most significant advantage is that it provides the most control over am endment distribution because the gradients can be manipulated. Other advantages include:

- The ability to achieve fastest initial donor distribution, which can lead to more rapid onset of reducing conditions
- Can achieve larger distribution from an individual injection point (i.e. larger radius of influence during injection)
- Ability to add large amounts of amendments over a relatively short timeframe

The most significant disadvantage for active recirculation is that it generally has the highest capital costs and O&M requirements of any approach. Continual system monitoring, either by automated i nstrumentation, or by o nsite staff, is needed to ensure upset conditions are not encountered and that all above ground equipment is operating as designed. In addition, logistical constraints at active facilities may impact placement of above ground infrastructure.

The primary advantage to passive approach for bioaugmentation is that it is a flexible approach that a llows f or f requent a pplications of e lectron donor, while ke eping the operational requirements (and costs) low. Other advantages include:

- Ability to distribute and maintain high concentrations of electron donor to a large radius of influence from individual injection points
- Ability to perform frequent (i.e., monthly to quarterly) a mendment injections cost effectively (on smaller scales)
- Large areas can be treated effectively with multiple injection points
- Minimal O&M and capital requirements compared to active recirculation.

The main disadvantage for the passive approach is because the primary distribution mechanism is a mbient groundwater flow; the success of this injection technique is highly dependent on subsurface conditions at the site. If ambient groundwater is too slow, then the area treated using this approach may be limited. In a ddition, the time and number of injections required be fore reducing conditions are achieved can be significantly longer compared to an active recirculation system. Also, individual injections can take multiple days depending on subsurface conditions.

3.0 PERFORMANCE OBJECTIVES

This demonstration complemented work completed under the ESTCP project "Bioaugmentation for Chlorinated Solvent Remediation: Microbial Transport, Growth, Survival and Dechlorinating Activity" (ER-0315). It a lso built upon pi lot t esting completed by NAVFAC SW at NAVWPNSTA Site 40 that successfully used a low-cost, passive approach for implementation of bioaugmentation. As described in Section 1, the technical objectives for this project are as follows:

- Extend bioaugmentation cost-effectively to full scale
 - Demonstrate cos t-effective ba cterial di stribution a t s cales of hundr eds, rather than tens, of feet
 - Demonstrate induction of complete dechlorination at the same scale
- Demonstrate that a low-cost, passive approach to bioaugmentation will achieve large-scale bacterial distribution and induction of complete dechlorination
- Compare and contrast t he ef fectiveness of passive and active approaches of bacterial distribution

The critical performance elements to measure were the results of the Phase 1 laboratory studies, the effects of the Phase 2 bi ostimulation/pre-conditioning, and the distribution of bacteria and extent of dechlorination in each of the treatment cells during Phase 3. Thus, the parameters to be monitored i nclude DHC cell c ounts, c hloroethenes and m etabolites, e lectron donor and fermentation products, bioactivity and redox indicators, and cost. The performance criteria are identified specifically in Table 3-1. These performance objectives were derived from those that were presented in the ER-0513 Demonstration Plan.

3.1 PHASE 1 PERFORMANCE O BJECTIVES – BENCH S CALE T ESTING AND BIOAUGMENTATION CULTURE SELECTION

Phase 1 of the E R-0513 project comprised conducting laboratory studies to confirm that dechlorination could be stimulated in the high sulfate environment present at NAVWPNSTA Site 70, and to select a bioaugmentation culture for the demonstration. These objectives are described further below.

3.1.1 Demonstration of Dechlorination using Site Groundwater

Site 70 w as known to have sulfate and chloride concentrations in excess of 1,000 mg/L throughout the source area, with concentrations as high as 8,000 mg/L or more in some areas due to past chemical oxidation activities. Sulfate-reducing bacteria can compete with dechlorinators for available electron donor, and high sulfate concentrations have been shown to inhibit complete dechlorination when the sulfate cannot be removed. For this reason, ESTCP requested be neh-scale testing be performed to evaluate a commercially available bioaugmentation culture for its ability to overcome the high sulfate concentrations and dechlorinate TCE to ethene.

Table 3-1. Technology Demonstration Performance Objectives

Project Phase	Performance Objective	Data Requirements	Success Criteria	Results
Quantitative Performance Ob	jectives			
Phase 1: Demonstrate that selected bioaugmentation culture can overcome high sulfate conditions and perform dechlorination to ethene; select a	Demonstrate that at least one commercially available bioaugmentation culture can carry out complete dechlorination in the presence of high sulfate concentrations.	Electron donor, sulfate, chloroethene, and dissolved gas concentrations in bench-scale study	Production of ethene at concentrations at least 2X detection in bench study using site groundwater samples, reduction of 95% TCE	Successful – see Section 6.1.1
bioaugmentation culture that contains <i>DHC</i> that can be distinguished from indigenous <i>DHC</i>	Determine if <i>DHC</i> are present onsite; if so select a culture that contains a <i>DHC</i> strain or functional gene not present naturally at site.	qPCR results; DNA sequencing results	Identification of a biomarker that is present in bioaugmentation culture(s) but not in native strains of <i>DHC</i>	Successful – see Section 6.1.2
Phase 2: Determine baseline conditions and pre-condition treatment cells	Demonstrate that the layout and residence time of each treatment cell are such that demonstration performance can be meaningfully evaluated in a sufficient time.	Tracer compound (iodide) concentrations over time, groundwater velocity and direction, residence time	Construct treatment cells such that travel time from injection wells to monitoring wells is 6 months or less	Successful – see Section 6.2.1
	Demonstrate that electron donor can be adequately distributed to remove sulfate from the system and create strongly reducing conditions in both treatment cells.	Electron donor, sulfate, ferrous iron, and methane data to verify that whey injections have created strongly reducing conditions	Sulfate reducing conditions achieved at monitoring wells nearest to injection locations	Partially Successful – see Section 6.2.2
Phase 3: Determine full-scale effectiveness of bacterial distribution using passive and active circulation systems	Determine bacterial growth and distribution throughout the treatment cells using both bioaugmentation scenarios.	qPCR analysis, iodide tracer	Collect data that allow for quantitative assessment of tracer and bacterial transport time, and growth of bacteria over time	Successful – see Section 6.3.1
	Determine extent of dechlorination in both treatment cells during the test period	Chloroethene and dissolved gas concentrations; stable carbon isotope analysis	Achieve full dechlorination to ethene using both approaches – detection of ethene at greater than 2x detection limit at greater than or equal to 2/3 of the monitoring wells in a given treatment cell	Partially successful – see Section 6.3.2
Qualitative Performance Obje				T = -
	Determine ease of use for both active and passive approaches	Feedback from field personnel; injection and operational logs	Quantify operational requirements for each approach	Successful – see Section 6.4

The microcosm tests were conducted using site groundwater. Two mixed cultures of *DHC* that were most likely to tolerate high concentrations of sulfate and chloride were used in these tests. Whey was used as the electron donor, and live microcosms received trace nutrient amendments (e.g., NH4, P O4, ye ast e xtract, a nd vi tamin B 12). The test for each well consisted of three microcosm bottles: 1) killed control; 2) whey, trace amendments, and bioaugmentation culture #1; and 3) whey, trace amendments, and bioaugmentation culture #2. The tests were conducted for approximately 3-4 months. Data collected during the lab study included monthly sampling for sulfate, electron donor, chlorinated compounds, ethene, ethane, and methane.

The success criterion for this performance objective was production of ethene at concentrations at least 2X detection, and reduction of TCE by at least 95 percent in the microcosms. The results of the study showed that dechlorination of TCE to ethene was achieved in less than 4 months, with nearly complete removal of TCE. Therefore, this performance objective was met. The full discussion of the results related to this performance objective is presented in Section 6.1.1.

3.1.2 Select Bioaugmentation Culture with Reliable Biomarker

Another c oncern for implementation of the de monstration is that the site might have a lready contained *DHC* prior to the de monstration, which would make tracking of the introduced bacteria difficult. In order to address this concern, samples of site groundwater were collected from MW-70-27 and EW-70-01 and analyzed for *DHC* DNA. In addition, three commercially available bioaugmentation cultures were screened and DNA was sequenced in order to select a bioaugmentation culture that could be reliably distinguished from any indigenous species.

The success c riterion for this objective was i dentification of a bi omarker that is present in bioaugmentation culture(s) but not in native *DHC*. The results from the DNA study showed that the functional gene *vcrA* was not present at the site, but was present in a commercially available bioaugmentation culture. Therefore, this performance objective was met. The full discussion of the results related to this performance objective is presented in Section 6.1.2.

3.2 PHASE 2 PERFORMANCE OBJECTIVES – BASELINE C ONDITIONS AND PRE-CONDITIONING

The pur pose of P hase 2 of the ER-0513 project was to determine groundwater hydraulic conditions and baseline contaminant distribution, *DHC* distribution, and geochemical concentrations prior to beginning the biostimulation and bioaugmentation in each treatment cell. Performance objectives were established related to demonstrating that the treatment cell layout was such that meaningful results could be obtained during the timeframe of the project, and related to establishing appropriate conditions prior to conducting bioaugmentation. These objectives are discussed further below.

3.2.1 Treatment Cell Construction and Residence Time

Due to the slow ambient groundwater velocity in the Site 70 source area, ESTCP was concerned that ef fects of e lectron donor i njections a nd bi oaugmentation w ould not be observed a t monitoring wells within the timeframe of the demonstration, at least for the passive cell. In

addition, historical data that were available for the site did not provide conclusive information regarding groundwater flow magnitude and direction in the Upper Fines unit (see Section 4.2.2) on the scale of the source area. In order to verify that meaningful results could be obtained using the propo sed treatment cell layout, a tracer test was conducted to verify the groundwater hydraulic conditions in the treatment cells. Data collected in support of this objective were multiple iodide tracer samples collected from active cell and passive cell monitoring wells.

The success criterion for this objective was to construct the treatment cells such that travel time from injection wells to monitoring wells was 6 months or less. The results of the tracer test showed arrival in some wells in less than 1 month, and subsequent sampling for tracer indicated that travel times to most monitoring wells were less than 4 months. These results were documented in a memo to E STCP dated June 6, 2008 (see A ppendix B). Therefore, this performance objective was met. The full discussion of the results related to this performance objective is presented in Section 6.2.1.

3.2.2 Pre-Conditioning Results

Baseline sampling w as c onducted t o a ssess baseline c onditions i ncluding c ontaminant a nd degradation product concentrations, redox parameters, and biological activity indicators (refer to Section 5.2 for complete baseline sampling results). In summary, the baseline results confirmed the pre-demonstration conditions in the source area, namely that conditions were anaerobic but mildly reducing, with very high sulfate concentrations and very limited dechlorination to *cis*-DCE in some areas. Because these conditions were not ideal for bioaugmentation, electron donor additions were performed to "pre-condition" the aquifer to reduce sulfate concentrations and to drive r edox c onditions more s trongly r educing. Data c ollected i n s upport of t his objective included r edox-sensitive pa rameters (specifically sulfate, ferrous iron, and methane), electron donor (as COD), volatile organic compounds (VOCs), and *DHC* using qPCR.

The suc cess criterion for this objective was to create at least sulfate-reducing conditions at monitoring wells nearest to injection locations, such that the bioaugmentation culture would have a favorable environment following inoculation. Results showed that redox conditions nearest the injection locations were sulfate reducing to methanogenic in both treatment cells following the pre-conditioning phase. These results were documented in a memo to ESTCP dated December 28, 2008 (see Appendix B). Therefore, this performance objective was met. The full discussion of the results related to this performance objective is presented in Section 6.2.2.

3.3 PHASE 3 PERFORMANCE OBJECTIVES – BIOAUGMENTATION RESULTS

The purpose of Phase 3 of the ER-0513 project was to demonstrate full-scale bioaugmentation and dechlorination using both the active and passive approaches. Phase 3 of the ER-0513 project began with inoculation of both treatment cells. Performance objectives were established related to collection of data that would allow for quantification of bacterial distribution and growth, and assessment of the extent of dechlorination. These objectives are discussed further below.

3.3.1 Bacterial Growth and Distribution

The first Phase 3 objective was to assess and quantify bacterial growth and distribution in both treatment cells. Bacterial distribution was assessed by analyzing the first arrival of *DHC* bacteria (as measured by qPCR analysis) at a giving monitoring location following i noculation. This travel time was then compared to the travel time for ambient groundwater, as determined from the tracer test. Bacterial grow th was then assessed by analyzing the increase of *DHC* and functional gene counts at a given location once first arrival had been established. Data collected in support of this objective included concentrations of *DHC* using qPCR and iodide tracer.

The success criterion for this objective was to collect data that allow for quantitative assessment of tracer and bacterial transport time, and growth of bacteria over time. No specific criteria were set in terms of bacterial transport times or cell counts. Therefore, this performance goal was met. The full discussion of the results related to this performance objective is presented in Section 6.3.1.

3.3.2 Extent of Dechlorination

The second Phase 3 objective was to assess and quantify the extent of dechlorination using both the active and passive bioaugmentation approaches. In the ER-0513 work plan, decision rules were defined for this performance objective based on trends observed in monitoring data, as shown in Table 3-2:

	Redox Conditions	Chloroethenes	Ethene	qPCR
Favorable trends	Sulfate decreasing or absent; Methane detected	Decreasing or not detected	Increasing or molar equivalent to initial TCE	DHC bacteria detected
Unfavorable trends	Sulfate present and not decreasing; no methane detected	Stable or increasing	Not detected	No <i>DHC</i> bacteria detected

Table 3-2. Decision Rules for Dechlorination Performance Objective

Decision Rule 1: If the passive treatment cell shows all of the favorable trends in Table 3-2 at >2/3 of all monitoring wells, then it will be determined that full-scale bi oaugmentation was successfully implemented using the passive approach. If less than 1/2 of all monitoring wells in the passive cell show all favorable trends in Table 3-2, then it will be determined that full-scale bioaugmentation was not successfully implemented using the passive approach. If more than 1/2 but less than 2/3 of all monitoring wells show favorable trends, then further evaluation will be required.

Decision Rule 2: If the active recirculation treatment cell shows all of the favorable trends in Table 3-2 over a distance of greater than or equal to 75 feet from the reinjection wells, then it will be determined that full-scale bioaugmentation was successfully implemented using the active recirculation approach. If the active recirculation treatment cell does not show all of the favorable trends in Table 3-2 over a distance of at least 50 feet from the reinjection wells, then it will be determined that full-scale bioaugmentation was not successfully implemented using the

active recirculation approach. All other combinations of potential outcomes will require further evaluation

A third decision was identified in the D emonstration P lan, to determine whether, and to the extent possible, under what conditions the passive approach is more technically effective and cost effective than the active recirculation approach. Decision #3 is based on the outcomes of Decisions 1 and 2, as well as on cost. Because of the multiple combinations of outcomes, and because of the fact that Decision Rules 1 and 2 are qualitative and are based on trends rather than explicit action 1 evels, no decision rule was presented for Decision #3. However, a no verall evaluation was made considering all available data in order to determine whether the passive approach was more technically effective and more cost effective than the active approach. This discussion is presented in Section 6.

Based on these decision rules, data collected in support of this performance objective include chloroethene and dissolved gas concentrations; stable carbon isotope analysis, redox sensitive parameters, and DHC using qPCR. The success criterion for this performance objective was to achieve full dechlorination to ethene using both approaches, as indicated by detection of ethene at greater than 2X detection limit at greater than or equal to 2/3 of the monitoring wells in a given treatment cell. Based on data collected during Phase 3, this performance objective was partially met. The full discussion of the results related to this performance objective is presented in Section 6.3.2.

3.4 QUALITATIVE PERFORMANCE OBJECTIVES

One qualitative performance objective was established for the ER-0513 project. This objective was to assess the ease of use for both passive and active approaches. This includes operational time r equired in the field, time spent conducting maintenance and repair a ctivities, and the amount of training required to operate each system. Data collected in support of this objective include feedback from field personnel; injection and operational logs, and the field team leader logbook.

The success criterion for this performance objective was to quantify the operational requirements for each approach. Data collected during the course of the ER-0513 demonstration did allow for an assessment of the ease of use of both approaches. Therefore, this performance goal was met. The full discussion of the results related to this performance objective is presented in Section 6.4.

4.0 SITE DESCRIPTION

This site d escription i neludes a discussion of t he s ite l ocation and hi story, ge ology and hydrogeology, ge ochemistry, and contaminant di stribution. This i neludes site background conditions at the outset of the demonstration project, not i neluding baseline characterization activities. Results of baseline sampling are provided in Section 5.2.

4.1 SITE LOCATION AND HISTORY

NAVWPNSTA IR S ite 70 w as the former N ational A eronautics and S pace A dministration (NASA) Research Testing and Evaluation Area, a rocket engine test facility located just south of Westminster Boulevard and east of Seal Beach Boulevard in Seal Beach, California (Figure 4-1). Site 70 encompasses approximately 40 acres on the northwestern quadrant of the NAVWPNSTA Seal Beach. Site 70 includes seven office and production buildings, asphalt-paved parking areas, several aboveground storage tanks, and distribution pipelines.

Past operations at the facilities reportedly included the use of dilute acids, chlorinated solvents including TCE, phenolic compounds, petroleum oils, sodium dichromate containing hexavalent chromium (C r⁶⁺), detergents, paint waste containing metals, volatile organics, and machine lubricating oil (Naval We apons S tation S eal B each, 2005). Currently these facilities are being used for industrial operations, storage, communications research, and office space.

4.2 GEOLOGY AND HYDROGEOLOGY

4.2.1 Regional Geology

Most of NAVWPNSTA Seal Beach slopes evenly from approximately 20 feet above sea level in the nor thwestern part of the facility to sea level at the tidal flats of the Seal Beach National Wildlife Refuge in the southeast. NAVWPNSTA Seal Beach is located on the Los Angeles-Orange County coastal plain and is underlain by approximately 20,000 feet of alluvial deposits. Recent age alluvial and coastal deposits overlay the NAVWPNSTA Seal Beach area.

4.2.2 Site-Specific Geology

The most recent characterization events at the site were conducted as a part of Remedial Design Optimization (RDO) a ctivities in 2005 by GeoSyntec C onsultants (GeoSyntec C onsultants, 2006). The RDO included cone penetrometer (CPT) soil and groundwater sampling within the Site 70 source area, as well as other characterization and testing activities in the downgradient plume area. Based on boring logs and site geologic models (GeoSyntec Consultants, 2006), the following hydrostratigraphic units, in order of increasing depth, have been characterized beneath NAVWPNSTA IR Site 70:

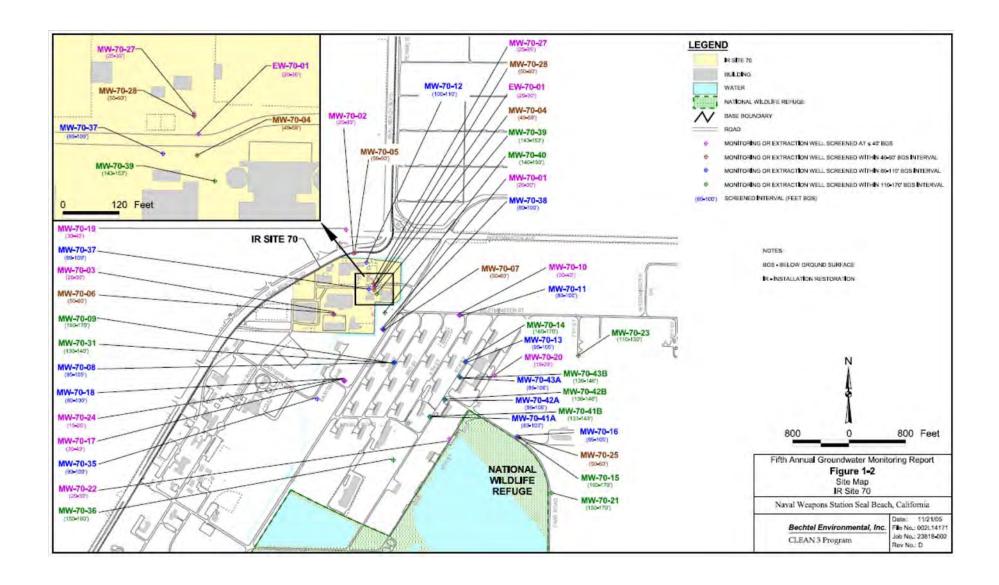




Figure 4-1
BECHTEL SITE LOCATION MAP
NAVWPNSTA SEAL BEACH, SITE 70
SEAL BEACH, CALIFORNIA

- Upper Fi nes U nit. This unit extends f rom ground s urface t o a pproximately 60 feet below ground surface (bgs) and comprises three zones: a shallow zone of surficial so ils and recent clayey sed iments; an intermediate zone of interbedded silts, clays, and sandy silts and clays including a semi-perched zone; and a lower zone of interbedded silts, clays, and fine to coarse-grained, silty to clayey sands. Based on CPT boring logs from the RDO activities, fine to medium grained sands are present from approximately 20 to 30 feet bgs in the source area. These sands are underlain by a clay unit to about 40 feet bgs.
- **First Sand Unit.** This unit extends from a pproximately 60 to 105 f eet bgs. It consists of poorly-graded fine-grained s and s ilty s ands. A coarse sand/fine gravel layer is present between 80 and 95 feet bgs in some areas.
- **Shell Horizon.** The shell horizon extends from approximately 105 to 135 feet bgs and comprises interbedded clays, silts, sands, and gravels below the source area transitioning to mainly f ine-grained s and t o t he s outheast. T his unit w as subdivided into two zones: interbedded clays and fine-grained sands.
- **Second Sa nd.** This u nit is similar to the F irst S and unit a nd extends f rom approximately 135 to 170 feet bgs.
- **Deep Clay Unit.** This unit extends from approximately 170 to 190 feet bgs and appears to be a continuous unit throughout the entire area of Site 70.
- **Deep San d U nit.** This unit is encountered at approximately 190 f eet bgs and appears to be similar in character to the First and Second Sands.

It should be noted the site specific geology presented above differs from what is described in the Final Extended Removal Site Evaluation Report (Bechtel Environmental, Inc., 1999) in that the Upper Fines Unit is separated into three separate units – the Surficial Soils, Shallow Clay Unit, and the Interbedded Unit.

4.2.3 Hydrogeology

The principal source of the deposited alluvium referenced above is the San Gabriel River, which cuts t hrough t he c oastal pl ain c reating t he A lamitos a nd S unset G aps. G roundwater f lows preferentially through the gaps due to the higher permeability of the alluvial fill within them. Regional gr oundwater f low i s a lso i nfluenced by t he L os A lamitos i njection b arrier, t idal influences, groundwater production wells, and manmade recharge basins (Jacobs Engineering Group, 1994).

Groundwater occurrence has been described as semi-perched and unconfined in the fine grained silt and silty sand that ge nerally comprises the upper 60 feet of the R ecent A ge de posits. Confined f reshwater z ones have been i dentified a t depths of 75 a nd 200 f eet bgs a t NAVWPNSTA Seal Beach and at depths of 250 to 1,000 feet bgs beneath NAVWPNSTA Seal Beach and neighboring cities (Jacobs Engineering Group, 1994).

This demonstration was conducted in the contaminant source area in the Upper Fines Unit, from approximately 15 f eet bgs to 35 feet bgs. The water table in the source area was historically

present at 5 to 12 feet. Hydraulic conductivity was not directly measured during the RDO. Based on hi storical da ta, the estimated conductivity in the U pper F ines U nit is 10 ft/d (Bechtel Environmental, Inc., 1999).

Sitewide historical hydraulic gradients in the Upper Fines Unit range from 0.0002 to 0.0011. However, hydraulic gradients w ithin the contaminant source are a from 0-40 feet bgs are confounding, as described in a recent groundwater monitoring report for Site 70 (Bechtel Environmental, Inc., 2005). Groundwater level data from July 2005 are shown as Figure 4-2, which is taken from (Bechtel Environmental, Inc., 2005). Figure 4-2 shows that groundwater flows generally northwest to southeast and culminates in a southwest-to-northeast trough in the general area of EW-70-01. The occurrence of the trough is attributed to an old stream drainage system that flowed through IR Site 70 (Bechtel Environmental, Inc., 2005). Groundwater flow in areas northwest of this trough is to the southeast into the trough, which is consistent with flow directions in deeper aquifer zones at Site 70. However, in areas that are southeast of the trough, groundwater actually flows northwest into the trough, with a gradient within the same range as the overall gradients for the Upper Fines Unit. Once groundwater reaches the trough, it appears to flow to the southwest (Bechtel Environmental, Inc., 2001), although the resolution of water level measurements in the source area may not be sufficient to fully characterize the flow direction

Also, quarterly water level data collected during 2004 and 2005 s how that water levels vary seasonally at Site 70 by nearly 7 feet. However, the occurrence of the trough near the Site 70 source area was observed in all quarters of monitoring, although its inferred location was slightly further s outheast during the D ecember 2004 s ampling r ound (Bechtel E nvironmental, Inc., 2005).

It is important to note that while this trough was observed during multiple groundwater sampling rounds, the number of data points used to create the historical groundwater elevation maps is not sufficient to elucidate detailed hydraulic gradients on the scale of source area (i.e., 200-400 feet). For example, the site-wide elevation maps from (Bechtel Environmental, Inc., 2005) show the entire s ource a rea as having the same groundwater elevation, which would imply that no groundwater flows through the source area (Figure 4-2). However, the gradient between wells EW-70-01 and MW-70-27 ranges from 0.0012 to 0.0026, with the flow direction toward MW-70-27, suggesting that the location of the trough may be closer to MW-70-27 than to EW-70-01, as suggested in Bechtel Environmental, Inc. (2005).

4.3 GEOCHEMISTRY

Redox c onditions in the source a rea, as measured in July 2005 (Bechtel Environmental, Inc., 2005), were mildly reducing, with oxidation/reduction potential (ORP) ranging from 56 to 179 mV. Dissolved oxygen (DO) was less than 0.5 mg/L, and some ferrous iron was detected in source area well EW-70-01, at 1.25 mg/L. One unique attribute of Site 70 is very high levels of sulfate in the source area. In source area well MW-70-27, sulfate was 7,650 mg/L; how ever, approximately 50 feet away at EW-70-01, sulfate was 1,150 mg/L, indicating that the very high concentrations are 1 ocalized around M W-70-27. C onsistent with the high I evels of sulfate, methane was not detected above 110 μ g/L.

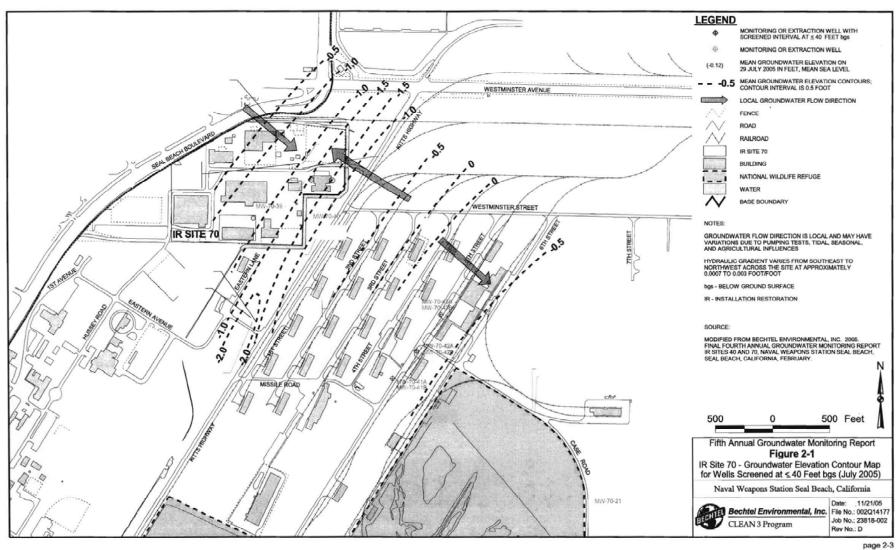




Figure 4-2 **BECHTEL GROUNDWATER ELEVATION MAP NAVWPNSTA SEAL BEACH, SITE 70 SEAL BEACH, CALIFORNIA** Chloride is also high in the source area, with MW-70-27 having a concentration of 3,920 mg/L. As with the sulfate, chloride decreases significantly from this well to EW-70-01, which had a concentration of 577 mg/L. Total or ganic carbon is 1 ow t hroughout the a quifer, with concentrations ranging from 0.5 mg/L to 14.8 m g/L. This is consistent with the limited dechlorination that has occurred intrinsically at this site (see Section 4.4). Alkalinity in the source area is 500-660 mg/L as CaCO₃, indicating that the aquifer has a reasonable buffering capacity.

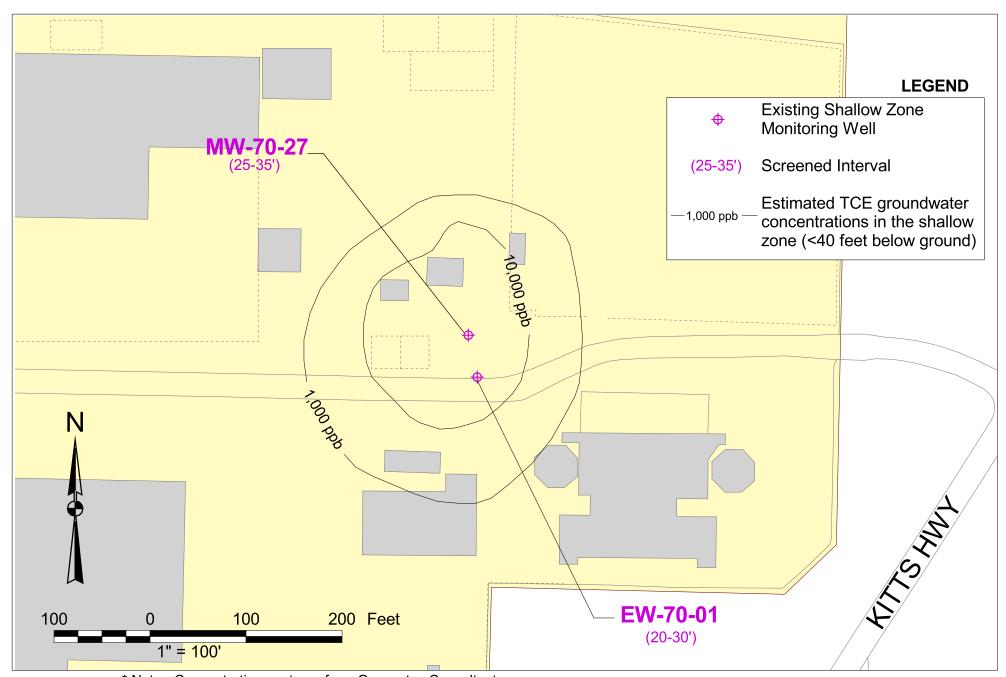
Overall, it appears as though the chemical oxidation activities that were conducted near MW-70-27 have significantly increased sulfate and chloride concentrations locally, and have created less reducing (although still anaerobic) conditions compared to the rest of the source area. This is confirmed by the fact that none of the source area R DO groundwater samples had sulfate or chloride concentrations of more than 1000 mg/L. Because of this, the geochemistry of EW-70-01 is thought to be more representative of the overall conditions in the source area.

4.4 CONTAMINANT DISTRIBUTION

The groundwater plume at Site 70 contains primarily TCE and other VOCs such as PCE, DCE, VC, chloroform, and others (Bechtel Environmental, Inc., 2005). The plume is estimated to be approximately 2,400 f eet long by 2,000 feet wide and approximately 195 f eet deep. There are two parts to the VOC plume: a small, high concentration source zone and a large area consisting of lower concentration VOCs in the dissolved-phase. The location for this demonstration is the shallow source zone, and the estimated extent of TCE in the source zone is shown in Figure 4-3.

Wells MW-70-27 and EW-70-01 are the only Upper Fines Unit permanent monitoring wells in the source area. These wells are completed from 25-35 feet bgs (MW-70-27) and from 20 to 30 feet bgs (EW-70-01). Each of these wells has very high levels of TCE, as concentrations in July 2005 were 130 mg/L and 53 mg/L, respectively. Concentrations of other chloroethenes are much lower in MW-70-27, where cis-DCE was 670 μ g/L, and VC and ethene were each less than 25 μ g/L. The most significant concentrations of daughter products were measured at EW-70-01, which had cis-DCE at 27 mg/L, while VC was 720 μ g/L.

The RDO involved collection of groundwater samples through temporary CPT wells throughout the source area. However, these samples were all collected at depths of 45-60 feet bgs, which represent the low er part of the Upper F ines U nit. The highest TCE concentration measured during the RDO sampling was 4 mg/L. This, combined with the data from MW-70-27 and EW-70-01, indicates that the high contaminant concentrations are limited to depths shallower than 40 feet bgs, which is the target zone for this demonstration. Overall within the source area, while some limited dechlorination has occurred, the majority of contamination in the source area is present as TCE.





* Note - Concentration contours from Geosyntec Consultants 2005 Draft Technical Memorandum.
Contours are based on September 2005 permanent monitoring wells groundwater concentrations including MW-70-27 and EW-70-01.

5.0 TEST DESIGN

This section provides the detailed description of the system design and testing conducted during the demonstration. This includes the conceptual design, treatability studies, system installation, baseline c haracterization, bi oaugmentation, a nd m onitoring. The s ampling a nd a nalysis is described in Section 5.7. The results of these activities are presented throughout this section, with the r esults of all P hase 3 activities being presented in Section 5.8. D iscussion and interpretation of the key results is provided in Section 6.

5.1 CONCEPTUAL EXPERIMENTAL DESIGN

The overall experimental design is based on the performance objectives presented in Section 3. The de sign comprised two independent treatment c ells to test the passive and active bioaugmentation approaches in a side-by-side comparison. The passive treatment cell consists of three injection wells, three multilevel (Continuous Multichannel Tubing [CMT]) monitoring wells, and six standard monitoring wells. The active recirculation cell consists of two injection wells, two extraction/recirculation wells, three multilevel (CMT) monitoring wells, and three standard monitoring wells.

The design was performed in three phases as described below:

Phase 1 – Pre-Demonstration Laboratory Investigations. Bench-scale testing was performed to demonstrate that the bioaugmentation culture could overcome the high sulfate concentrations at the site. DNA analysis of site groundwater samples and commercially available cultures, including qPCR, clonel ibrary development, and DNA sequencing were used to identify "biomarkers" that provided the ability to differentiate between the injected cultures and any existing *DHC* that may have naturally existed in the groundwater.

Phase 2 - Tracer Test, Baseline Sampling, and "Pre-Conditioning". Following treatment cell construction, a tracer test was conducted in each of the treatment cells to verify the groundwater hydraulics in the shallow aquifer. Following the tracer test, baseline sampling was conducted to assess baseline conditions including contaminant and degradation product concentrations, redox parameters, biolog ical act ivity indicators, and *DHC* concentrations. Following baseline sampling, electron donor was injected into each treatment cell to create strongly reducing conditions and remove sulfate prior to bioaugmentation.

Phase 3 – Bioaugmentation and Monitoring. This third and final phase involved injecting the dechlorinating c ulture into e ach of t he t wo t reatment c ells a nd pe rforming groundwater monitoring to compare with results from Phase 2.

5.2 BASELINE CHARACTERIZATION

The objectives of the baseline characterization were to determine groundwater hydraulic conditions and baseline contaminant distribution, *DHC* distribution, and geochemical concentrations prior to beginning the biostimulation and bioaugmentation in each treatment cell. In order to perform the baseline characterization, the active recirculation system and select

monitoring wells were installed prior to baseline activities. The remaining wells were installed based on o bserved w ater levels during a mbient and p umping c onditions. D etails of t he recirculation system and well installations are provided in Section 5.4.

A tracer test was then conducted in the active cell to verify the groundwater hydraulic conditions in the treatment cells. In order to create similar conditions to the demonstration, the recirculation system was started 5 days prior to starting the tracer test and continued operating during baseline sampling. Following the tracer test, the additional wells were installed and baseline sampling was conducted to a ssess baseline conditions including contaminant and degradation product concentrations, redox parameters, and bi ological activity indicators. A summary of these activities is provided below.

5.2.1 Installation Activities

Well installation was not performed in one mobilization because the groundwater flow patterns needed to be understood with the active cell recirculation system running. Once the groundwater flow pattern under pumping conditions was understood, the most appropriate cell orientation was determined for the passive cell. This phased approach for treatment cell construction allowed for the opportunity to assess groundwater flow direction in the area of the planned passive cell wells before installing the remaining ten wells. This helped avoid a scenario in which the entire passive treatment cell was in stalled, only to find out that groundwater did not flow parallel to the treatment cell axis.

5.2.1.1 Active Cell Well Installation

Injection, extraction, and monitoring wells for the active cell were installed in September and October 2007, along with two of the passive cell monitoring wells. The active cell recirculation system i tself w as cons tructed, i nstalled, and tested in March and A pril 2008. The system operated by extracting groundwater from wells AEW-1 and AEW-2 into a 275 gallon surge tank; the surge tank water was reinjected i nto A IW-1 and A IW-2, which is a distance of 100 feet upgradient from the extraction wells (refer to Figure 5-1 for well locations). Photos of the recirculation system a re included in A ppendix C. Once the system was functional, it was operated for several days, and water levels were measured in active cell monitoring wells, and in the two existing passive cell monitoring wells, in order to determine the groundwater flow direction in the area of the proposed passive cell wells. Synoptic water level data were collected in several wells using transducers, and in other wells by taking water levels using a water level meter.

Following a tracer study with the active cell running, the location of the passive treatment cell was modified to reflect the groundwater flow direction under pumping conditions. A more detailed description of the active cell tracer study is provided in Section 5.3.2 and in Appendix B. The groundwater flow direction was different than assumed based on data available at the time the ESTCP Demonstration Plan was submitted. The final well construction locations and details are shown in Table 5-1.

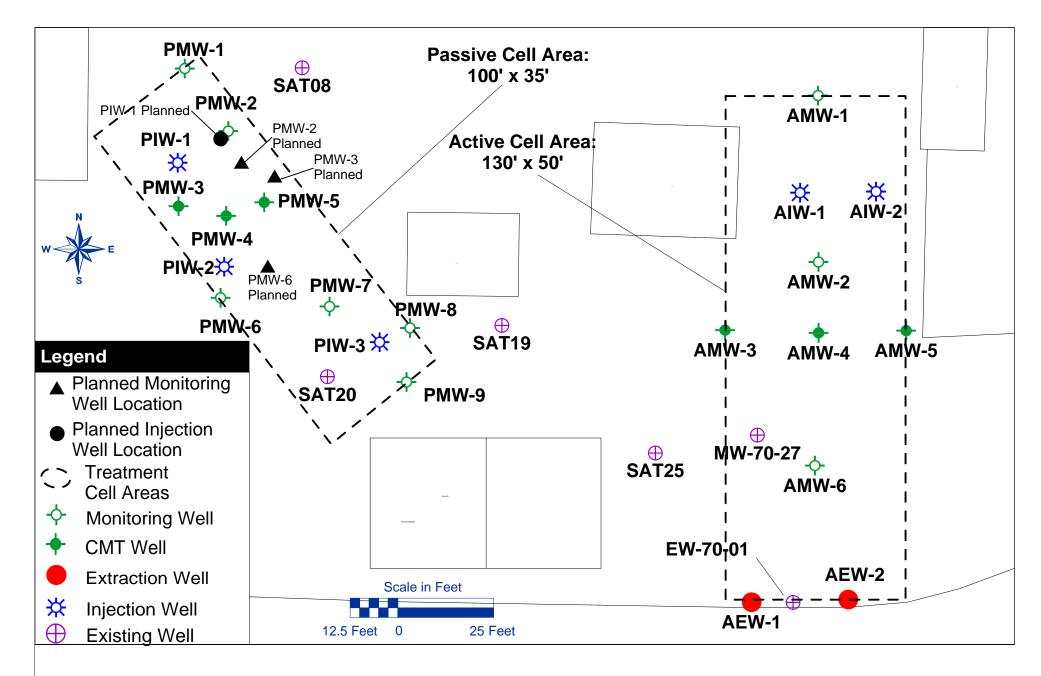


FIGURE 5-1 WELL LOCATION MAP ER-0513 FINAL REPORT SEAL BEACH NAVAL WEAPONS STATION, SEAL BEACH, CA



Table 5-1. Well Construction Summary

Well ID	Well Type	Easting	Northing	Surface Elevation	Construction	Well	Screen Interval	Total Depth
Well ID		UTM (feet)	NVI) QQ (foot		Construction	Diameter	ft bgs	ft bgs
Passive Ce	ell							
PMW-1	Monitoring	6005842.00	2224265.34	11.22	PVC	4-inch	15-35	35.3
PMW-2	Monitoring	6005853.47	2224248.64	11.69	PVC	4-inch	15-35	35.5
PMW-3	CMT - Zone 1	6005840.36	2224229.27	11.50	CMT	1.7-inch	34-35	36
	CMT - Zone 2						26-27	
	CMT - Zone 3						22-23	
	CMT - Zone 4						16-17	
PMW-4	CMT - Zone 1	6005852.88	2224226.69	11.43	CMT	1.7-inch	33.5-34.5	36
	CMT - Zone 2						30-31	
	CMT - Zone 3						26.5-27.5	
	CMT - Zone 4						22.5-23.5	
	CMT - Zone 5						15.5-16.5	
PMW-5	CMT - Zone 1	6005862.70	2224230.18	11.46	CMT	1.7-inch	33.5-34.5	35.9
	CMT - Zone 2						27-28	
	CMT - Zone 3						23-24	
	CMT - Zone 4						17-18	
PMW-6	Monitoring	6005851.40	2224205.28	11.20	PVC	4-inch	15-35	35.5
PMW-7	Monitoring	6005879.93	2224203.13	11.32	PVC	4-inch	15-35	35.5
PMW-8	Monitoring	6005900.79	2224197.53	11.23	PVC	4-inch	15-35	35.5
PMW-9	Monitoring	6005899.76	2224183.81	10.88	PVC	4-inch	15-35	35.5
PIW-1	Injection	6005840.29	2224240.54	11.63	PVC	4-inch	15-35	35.5
PIW-2	Injection	6005852.43	2224213.43	11.22	PVC	4-inch	15-35	35.5
PIW-3	Injection	6005892.86	2224193.80	11.27	PVC	4-inch	15-35	35.5
Active Cel								
AMW-1	Monitoring	6006006.90	2224258.32	10.60	PVC	4-inch	15-35	36.5
AMW-2	Monitoring	6006007.00	2224215.03	10.65	PVC	4-inch	15-35	36
AMW-3	CMT - Zone 1	6005982.69	2224197.30	10.53	CMT	1.7-inch	33-34	36.5
	CMT - Zone 2						28-29	
	CMT - Zone 3						24-25	
	CMT - Zone 4						17-18	
AMW-4	CMT - Zone 1	6006007.03	2224196.59	10.30	CMT	1.7-inch	33-34	36
	CMT - Zone 2						28-29	
	CMT - Zone 3						24-25	
	CMT - Zone 4						18-19	
AMW-5	CMT - Zone 1	6006029.84	2224197.17	9.83	CMT	1.7-inch	33-34	36.4
	CMT - Zone 2						28-29	
	CMT - Zone 3						24-25	
	CMT - Zone 4			40 :-			18-19	
AMW-6	Monitoring	6006006.08	2224162.08	10.17	PVC	4-inch	15.5-35.5	35.5
AIW-1	Injection	6006002.22	2224233.05	11.01	PVC	4-inch	15-35	35
AIW-2	Injection	6006022.00	2224233.33	9.88	PVC	4-inch	15-35	35.6
AEW-1	Extraction	6005989.60	2224126.55	9.15	PVC	4-inch	15-35	35
AEW-2	Extraction	6006014.76	2224127.23	8.79	PVC	4-inch	15-35	35.3

CMT - Solinst® Continuous Multichannel Tubing System

UTM - Universal Transverse Mercator

NVD88 - National Vertical Datum 1988

ft bgs - feet below ground surface

AMSL – above mean sea level

5.2.1.2 Passive Cell Well Installation

In order to account for the more southerly flow direction under pumping conditions, placement of some of the passive cell wells was adjusted slightly from the original planned locations. These adjustments we ere made considering interpreted groundwater flow directions as well as accounting for the many underground utilities in the area. The planned and actual locations are presented in Figure 5-1. The most significant change was moving C MT well P MW-3 to be

directly south of PIW-1. The CMT wells are used to measure multiple depths from a single well using individual channels screened at discrete intervals. Also, well PMW-2 was moved from its planned location on the treatment cell axis to a location northeast of PIW-1. Finally, wells PIW-2 and PMW-6 were moved a few feet to the west of their planned locations in order to a void utilities.

The r emaining t en p assive c ell w ells (four m onitoring w ells, t hree injection w ells, and t hree CMT w ells) were installed in April 2008 following the tracer test. After installation of the remaining passive c ell wells, a new round of water level measurements was collected under pumping conditions.

5.2.1.3 CMT Well Installation

The E STCP D emonstration P lan called for three sam ple ports in each CMT well. During installation of both the active and passive cell CMT wells, four sample ports were completed in all CMT wells except PMW-4, which has five sample ports. This was done in order to account for the possibility that some ports would not produce enough water for sampling.

5.3 BASELINE SAMPLING

Baseline sampling was completed in April 2008, after the active cell recirculation system was operating. In the active cell, this included sampling the three standard monitoring wells, all ports in the three CMT wells, and the water being produced from the extraction wells (refer to Figure 5-1 for well locations). Baseline sampling for the passive cell included sampling the six standard monitoring wells, all p orts in the three C MT wells, and the three injection wells. Analytes sampled i ncluded VOCs, dissolved gases (ethene/ethane/methane), a nions (sulfate, c hloride, nitrate/nitrite), alkalinity, COD, DNA samples, compound-specific isotope analysis, and iodide tracer (for background measurements). A summary of the analyses performed in each monitoring well is provided in Section 5.6.

During the baseline sampling events, it was determined that the uppermost port in each active cell CMT well did not produce su fficient water to complete a full set of samples. However, because extra ports were installed in each well, data are available from multiple depths in each CMT well.

5.3.1 Baseline Sampling Results

Results of baseline sampling are summarized here and are presented in Table H-1 for the passive cell and Table I-1 for the active cell. For the active treatment cell, concentrations were generally around 1,00 0 t o 3,000 μ g/L for TCE, w ith other contaminants present at 1 ow 1 evels, but concentrations increased significantly at the southern end of the cell. The highest concentration measured anywhere in the ESTCP demonstration area was 140,000 μ g/L at well AMW-6. This is adjacent to a previous chemical oxi dation pilot test and was known to be the highest concentration area within the source. The sample collected from the water being extracted from wells AEW-1 and AEW-2 had a TCE concentration of 10,000 μ g/L.

For the passive cell, TCE concentrations were around 1,000 μ g/L at each end of the treatment cell (wells PMW-1 and PMW-9). However, TCE concentrations were much higher in the center

of the passive cell (15,000 μ g/L to 63,000 μ g/L). Concentrations of other VOC contaminants were low in all passive cell wells.

Vertically discrete samples of contaminants in upper zones of the CMT wells in the active cell generally had low levels of contaminants and also produced very little water when purged. TCE concentrations were approximately 600 to 1,800 μ g/L in middle to lower zones. For the passive cell, TCE concentrations are generally an order of magnitude higher than the active cell; upper zones had TCE concentrations of 1,000 to 10,000 μ g/L, while middle and lower zones had TCE as high as 63,000 μ g/L.

Results for other parameters show that the aquifer was generally mildly reducing with low levels of available carbon. DO was less than 1 mg/L and ferrous iron was generally less than 0.1 mg/L at al 11 ocations. Sulfate was very hi gh a t t his s ite, w ith c oncentrations r anging f rom approximately 1,600 mg/L to as high as 8,700 mg/L near the area where the chemical oxidation pilot t est w as c onducted. Methane w as de tected at som e w ells up t o 230 μ g/L, w hile C OD ranged from non-detect to 100 m g/L. Overall, the pH was near neutral, and ORP ranged from -150 t o +300 m V. The only exception to these general trends was well P MW-9, w hich had relatively high concentrations of methane of 2.8 mg/L, and somewhat depressed sulfate of 1,100 mg/L. While TC E was lower a t t his loc ation than others in the passive c ell, very low concentrations of reductive d aughter products were present, and C OD w as low as well (16 mg/L). This suggests that while redox conditions may have been approaching methanogenesis at this location, little dechlorination was occurring.

Finally, the baseline compound-specific isotope analyses results show that the TCE present near the active extraction wells was "heavier" than in other places. This implies that a mechanism which results in fractionation of TCE (i.e., preferential transformation of the TCE molecules with the "lighter" carbon-12 isotope) is or was active in the past in this area. This is consistent with the fact that this area of the site is near the former chemical oxidation pilot test, because chemical oxidation is known to cause fractionation of TCE, similar to what biodegradation causes. Thus, it appears that the effects of the chemical oxidation are still evident in the isotope signatures at this monitoring I ocation. This was no t expected to affect d ata interpretation for the E R-0513 demonstration be cause future bi odegradation would cause further f ractionation of TCE, and would also produce daughter products, whose isotope signatures could then be monitored over time.

5.3.2 Active Cell Tracer Test

In order to verify the groundwater velocities estimated based on existing data, a tracer test was conducted in the active cell using an iodide tracer. The purpose of the tracer test was to determine hydraulic properties of the active cell and its effect on hydraulics in the passive cell, and to measure the first arrival of tracer at the nearest monitoring locations, which represents the earliest expected arrival of injected bacteria and donor. In order to determine the hydraulic properties of the treatment cells, peak b reakthrough had to be measured in at least one monitoring well for each treatment cell.

Approximately 500 ga llons of potassium iodide was injected into the active cell on April 10, 2008. The a verage c oncentration of i odide in the injected solution was a pproximately

13,100 mg/L. Samples for iodi de tra cer w ere c ollected once per day from well A MW-2 for approximately 4 weeks. Periodic CMT monitoring was then performed for seven weeks after the tracer injection.

A detailed summary of the active cell tracer study is provided in Appendix B, including tracer breakthrough curves for the active cell tracer test. Tracer breakthrough was observed in AMW-2 (18 feet from injection wells) within 2 weeks. Breakthrough was observed at AMW-4 Zone 2 (screened 28 feet bgs) within approximately 2.5 weeks, Zone 1 (33 feet bgs) within 3 weeks, and Zone 3 (24 feet bgs) within 4 weeks. In addition, tracer breakthrough occurred in AMW-5 Zone 2 and AMW-3 Zone 3 in approximately 5 weeks, and tracer was eventually detected in the other ports in these CMT wells. These results showed that the deeper zones are more transmissive, which is also where the higher contaminant concentrations are found in these wells. The long tail on the AMW-2 tracer breakthrough curve is likely the result of different tracer arrival times in the various lithologic units.

A prel iminary analysis of the tracer test data was performed in order to estimate a quifer properties for the purpose of calculating potential ranges of travel times within the passive cell. The model used was developed for an instantaneous point source (Baetsle, 1969). The analytical equation is found in Domenico and Schwartz (1990, p. 650). A hydraulic conductivity of 10 ft/d was assumed as a starting point based on a pumping test performed in the source area at the site several years ago. An effective porosity of 0.20 was assumed based on CDM's experience with this soil type. A longitudinal dispersivity value equivalent to approximately 10 percent of the scale of the cell was assumed, and the transverse dispersivity was assumed to be 10 percent of the longitudinal. The hydr aulic gradient us ed was 0.04 b ased on water level measurements during pumping. The final variable in this model is distance from the axis (or centerline) of transport. Given the two injection wells in the active cell, this analytical model does not perfectly represent the real system, and the distance from the axis has a questionable meaning. Also, solutions using this model will be non-unique as multiple combinations of the conductivity, effective p orosity, and distance f rom t he center line can produce very s imilar r esults. Nevertheless, it is believed that this approach is useful to estimate aguifer properties reasonably, especially given the fact that the hydraulic conductivity has previously be en measured by a multiple well pumping test at the site.

Using t his approach, inverse m odeling w as performed to e stimate a r ange of hydr aulic conductivities based on matching model predictions to measured iodide breakthrough at several of the monitoring locations. For the three active cell monitoring locations shown, the hydraulic conductivity ranged from 5 to 10 ft/d. Thus, the tracer test data could be reasonably matched using hydraulic property values consistent with the soil type and previous hydraulic testing at the site.

Based on the estimated values of parameters determined by the tracer test as listed above, travel times from passive cell injection wells to passive cell monitoring wells were estimated. The most significant f actor af fecting the travel time is the injection event itself. The target in jection volume of 1,000 gallons per well is based on achieving a radius of influence of 5 feet. Therefore, it was assumed that the injected substrate would be distributed 5 feet from the injection point at time zero. Given the range of hydraulic conductivities that were estimated based on the tracer

test, along with the measured groundwater elevations, groundwater velocity in the passive cell was expected to be approximately 4-8 feet/month, or 45-90 feet/year. This is well within the range of ambient groundwater velocity at other sites where bioremediation and bioaugmentation have been successful, and is in fact two to four times higher than what was originally assumed in the ER-0513 ESTCP Demonstration Plan.

The transport during i njection c ombined with advection under a mbient c onditions r esults in travel times from injection wells P IW-1 and P IW-3 to their corresponding monitoring wells ranging from 1 to 3 months, a ssuming a hydraulic conductivity of 10 ft/d. Even if the low estimate of 5 ft/d for conductivity were assumed, travel times from PIW-1 and PIW-3 range from 2 to 5 months. Well P IW-2 has a monitoring well located 8 f eet away (PMW-6), and another monitoring well located 29 f eet away (PMW-7). Depending on the local flow direction in this area, travel times to PMW-6 could be less than one month, while travel times to PMW-7 could be 3 to 7 months. These travel times were deemed acceptable for the demonstration, and the data indicated that travel times were less than predicted (refer to Sections 5.8 and 6.3).

5.4 TREATABILITY AND LABORATORY STUDY RESULTS

The objectives of Phase 1 were to demonstrate that a com mercially available bioaugmentation culture is a ble to perform complete dechlorination under high sulfate conditions, and a lso to choose a culture that can be differentiated from naturally existing bacteria in the groundwater at the site. These objectives were successfully met by performing bench-scale studies of the groundwater and analyzing the existing cultures in the groundwater using qPCR, clone library development, and DNA sequencing.

5.4.1. Bench-Scale Study

Site 70 is known to have sulfate and chloride concentrations in excess of 1,000 m g/L in the source area, likely due to past chemical oxi dation activities. Sulfate-reducing bacteria can compete with dechlorinators for available electron donor, and high sulfate concentrations have been shown to inhibit complete dechlorination when the sulfate cannot be removed. For this reason, ESTCP requested bench-scale testing be performed to evaluate a commercially available bioaugmentation culture for its a bility to over come the high sulfate concentrations and dechlorinate TCE all the way to ethene.

Microcosm Study Setup

The pur pose of the microcosm t est w as to de termine w hether either of two bi oaugmentation cultures could achieve dechlorination in well samples from the NAVWPNSTA Site. The tests were performed by B ioremediation C onsulting, Inc. (BCI) and the full report is provided as Appendix D.

CDM selected two wells for testing: (1) E W-70-01, which had a high chloride content of 2,200 mg/L and high sulfate content of 1,650 mg/L, and (2) MW-70-27, which had high chloride of 4,400 mg/L and extremely high sulfate of 9,300 mg/L. Both wells contained total chlorinated ethene concentrations of less than 30 mg/L.

Two *DHC* cultures were used for testing: Culture "S" (a T CE-degrader) and C ulture "B" (a mixed chloroethene-degrader), both of which had capabilities with high chloride concentrations. Both cultures were augmented with a sulfate-reducing culture active at high sulfate concentrations.

Anaerobic microcosms were constructed to test each culture with each groundwater sample, using whey as an electron donor (food source), and adding small amounts of nutrients needed by bacteria (ammonia and phosphate), as well as ye ast extract and vitamin B12. Killed control microcosms were also constructed for each well sample. Microcosms were monitored by removing smalls amples and analyzing for chlorinated or ganics and eithene by gas chromatography, and organic acids and sulfate by capillary ion electrophoresis.

Results and Conclusions

For EW-70-01, which contained 1,650 mg/L sulfate and 2,200 mg/L chloride, BCI Cultures "S" and "B" were e qually s uccessful i n de chlorinating 16 mg/L T CE a nd 6 m g/L cis-DCE completely to e thene in 112 days. Ethene was measured as high as 177 μ M, and sulfate was reduced to non-detect using both cultures. Figures showing results from the study are included in Appendix D.

For M W-70-27, which contained very high sulfate of 9,270 mg/L and very high chloride of 4,350 mg/L, Culture "S" succeeded in converting all of the TCE to VC ($45\mu M$) and ethene (119 μM) in 112 days (see Appendix D). Sulfate was reduced by 36 percent to 6,020 mg/L during this time. Culture "B" was a ble t o d egrade a ll of t he T CE pr esent in t he m icrocosm, but dechlorination only proceeded to cis-DCE and V C, with trace a mounts of e thene produced. Sulfate was reduced by 35 percent to 5,990 mg/L during this time. Based on these results, it was concluded that c omplete de chlorination to e thene was a chievable in the presence of the high sulfate concentrations at the site.

5.4.2 DNA Sequencing Study

Another concern for implementation of the demonstration was that the site might already contain *D. ethenogenes* or other *DHC* that would make tracking of the introduced bacteria difficult. In order to address this concern, samples of site groundwater were collected from MW-70-27 and EW-70-01 and a nalyzed for *DHC* DNA. The DNA was a mplified using specific primers for *DHC*, then the amplified DNA was inserted into clones, from which the DNA was later extracted and sequenced. Up to 20 c lones were analyzed in this clone library, allowing determination of the *DHC* strains that are present at the site. Results from this study are provided in Appendix D.

The results from the 16S rRNA clone library GenBank analysis suggest that most of the *DHC* identified in the NAVWPNSTA Site 70 and bioaugmentation clone libraries were most closely related to *Dehalococcoides ethenogenes* strain 195, or *Dehalococcoides* species TM-EtOH with greater than 98-99 percent sequence similarity. These data illustrate that the *DHC* 16S rRNA rRNA sequences are highly similar, and while there are so me regions be tween different sequences that are significantly different, it would be difficult to distinguish between the observed sequences found within the different bioaugmentation cultures and those indigenous to the NAVWPNSTA Site 70 by 16S rRNA molecular analysis alone.

Baseline qPCR analysis showed that indigenous DHC were only detected at low levels at two monitoring locations – the active extraction wells had 448 ± 75 cells/L, and the passive cell well PMW-3 had 110 ± 28 cells/L. These cell counts are just above the minimum quantification level for the qPCR analysis, and a re four to six or ders of magnitude lower than what is typically observed following bioaugmentation.

While results from the 16S rRNA clone library analysis did not provide a clear biomarker for any of the commercially a vailable bi oaugmentation cultures, qPCR analysis i ndicated that the functional reductase gene *vcrA* was not present at NAVWPNSTA Site 70, but was present in high concentrations in bioaugmentation cultures. In order to determine if there were significant differences between the *vcrA* gene sequences present within the bioaugmentation cultures, clone libraries were constructed using *vcrA*-specific PCR primers. The NAVWPNSTA Site 70 sample did not amplify, confirming that the *vcrA* gene was not detected using either the qPCR or PCR protocols d escribed. The B CI bi oaugmentation culture, how ever, did not a mplify e ither. Therefore, only the Shaw SDC-9TM and KB-1TM cultures had clone libraries constructed for the *vcrA* gene. The *vcrA* clone library DNA data would have been used to design a biomarker if the standard qPCR analysis for *vcrA* was not sufficient.

5.5 DESIGN AND LAYOUT OF TECHNOLOGY COMPONENTS

The demonstration a rea was designed to include two independent cells, one utilizing a recirculation system (active cell), and one relying on passive distribution of the introduced culture. The primary technology components of this demonstration included groundwater wells (injection, extraction, and monitoring wells), a gravity fed electron donor delivery system, a groundwater recirculation system, and a bacteria injection system.

5.5.1 Well Layout and Cell Placement

Two treatment cells were installed at NAVWPNSTA Site 70, one for the passive distribution system and one for the active distribution system (Figure 5-1). The treatment cells were based on the following criteria:

- Both cells should be located within the source area or the high concentration area surrounding the source area (i.e., TCE concentrations greater than 1,000 ppb).
- The cells should be located such that hydraulic autonomy could be maintained between the passive and active cells; therefore the extraction wells in the active cell do not capture significant volumes of groundwater from the passive cell during the duration of the demonstration.
- The well layout within each cell must allow for meaningful results to be observed within the 12-month duration of Phase 3 bioaugmentation activities.
- Both cells should be oriented generally in the direction of groundwater flow

These criteria were met by the treatment cell layouts based on tracer test results and phased treatment cell construction, as described in Section 5.2

For the active treatment cell, the overall dimensions are 130 feet by 50 feet. A pair of extraction wells a nd a pa ir of in jection wells were installed 105 feet a part, with the spacing be tween extraction wells and injection wells 25 feet and 20 feet, respectively. The final active cell well screened depth intervals and CMT sampling depths are shown in Table 5-1. During drilling, soil lithology was recorded based on the Unified Soil Classification system (ASTM-D 2488-93) for all boreholes. The soil boring / well construction logs for each well are provided in Appendix E. All wells, including CMT wells, were developed to comply with California Division of Water Resources Water well standards. A summary of the development of each well is also provided in Appendix E.

For the passive cell, the overall dimensions are 100 feet by 35 feet (Figure 5-1). Within this area, three injection wells are located along the axis of the treatment cell at a spacing of 35-45 feet. A total of six standard monitoring wells are located in the passive treatment cell. Three of these wells are located along the axis of the treatment cell and are spaced between 12 and 17 feet from the injection wells. The other three monitoring wells are located just off-axis, at a distance of about 7 to 9 feet from each of the three injection wells. The passive cell also has a transect of three CMT wells placed halfway between the first and second injection wells. The CMT wells are spaced approximately 17.5 feet laterally and were completed at three discrete sampling depths based on observed field conditions. Many of the proposed well locations were moved because of above ground and utility obstructions. The final passive cell screened depth intervals are provided in Table 5-1.

5.5.2 Standard Well Installation

Four different types of wells were installed for this demonstration: injection wells, extraction wells, Solinst® CMT monitoring wells, and standard monitoring wells. Except for CMT wells, all wells were completed with approximately 20 feet of 4-inch diameter schedule 40 PVC, wire wrapped 0.05 slot screen, and 4-inch schedule 40 PVC riser pipe installed from the top of screen approximately to ground surface. One foot of a ppropriately sized silicals and filter pack was added to the annular space beneath the bottom of the well. Well installation details are provided in Table 5-1 and well construction diagrams in Appendix E. The annular space surrounding the screen was backfilled with the silical sand filter pack to a depth of approximately 3 feet above the well screen and capped with a bentonite seal to at least 2 feet bgs. The remainder of the annular space was filled with concrete to ground surface and if necessary, widened into a 24-inch by 24-inch concrete pad at the surface (if the surface was not already concrete). All wells were flush mounted with bolted manhole covers and locking caps.

5.5.3 CMT Monitoring Well Installation

Three CMT monitoring wells were installed in each treatment cell as shown in Figure 5-1. The wells were aligned perpendicular to flow in each cell to evaluate three-dimensional transport. The CMT wells are 1.7-inch diameter and each has a minimum of four sampling ports as detailed in Table 5-1. Well construction diagrams for CMT wells are provided in Appendix E.

5.5.4 Passive Cell Electron Donor Distribution System

A gravity-fed e lectron donor distribution system was constructed to deliver a sodium la ctate (electron donor) solution to all three of the passive cell injection wells simultaneously during discrete injection events. A process flow diagram is provided as Figure 5-2.

Make-up water for the passive cell injections was from a potable water source available onsite. The potable water was fed through a proportional flow mixer, which de livers lactate to the injection line at a concentration that is in proportion to the water flow rate.

The di luted la ctate s olution was transferred to a manifold capable of injecting into all three passive cell wells simultaneously. Each line of the manifold included a metered valve with a totalizer, and the manifold itself was mounted on plywood or similar board. Reinforced flex hose was used to convey the dilute lactate solution to the injection wells. These hoses were lowered in the well and placed near the middle of the well s creen, and injections were performed under gravity flow (i.e., not under pressure).

5.5.5 Active Cell Recirculation System

For the active cell, a recirculation system was constructed to extract and re-inject groundwater continually (i.e., 24 hours per day, 7 days per week) across an area of approximately 130 feet. The system was designed to be capable of pumping total groundwater flows in the range of 0.5 – 5 gallons per minute (gpm) from each of two extraction wells (1-10 gpm total). To periodically pulse lactate into the recirculation line, a second proportional feed mixer was installed for use only when lactate injections were required. Instrumentation and controls were provided such that the system can run without an operator onsite, except for periodic inspections and maintenance. Below is a brief description of the operating requirements and parameters for the active treatment cell. A process flow and instrumentation diagram is provided as Figure 5-3.

The system was designed to extract groundwater from each of the two extraction wells using environmental duty submersible pumps and pump it into a double walled surge tank. The pumps were controlled by two float switches. The high level switch LSH-100 initiates the pumps' run operation (Figure 5-3). When the groundwater level drops below LSL-100, the pumps would stop. The pumps' operation was interlocked with Hi-Hi level switch LSHH-200 in the surge tank. If the Hi-Hi level was reached in the surge tank the extraction pumps stopped. The level switch locations in the extraction well were modified after low groundwater levels caused the pumps to cycle during a period of low precipitation.

Extracted groundwater was conveyed to the treatment skid, which consisted of the surge tank, transfer pump, manifold, and electronics. Each extraction well was plumbed independently back to the treatment skid where they were combined prior to discharge into the surge tank. Each leg contained a check valve to prevent extracted groundwater from flowing back into the well. Each leg also included a pressure gauge, a totalizing flow meter and a gate valve.

The surge tank included two level switches to control the injection pump. Level switch LSH-200 initiated the pumps' run operation. When the water level dropped to below LSL-200 the pumps would stop. The pump operation was interlocked with the Hi-Hi level switch LSHH-300 located

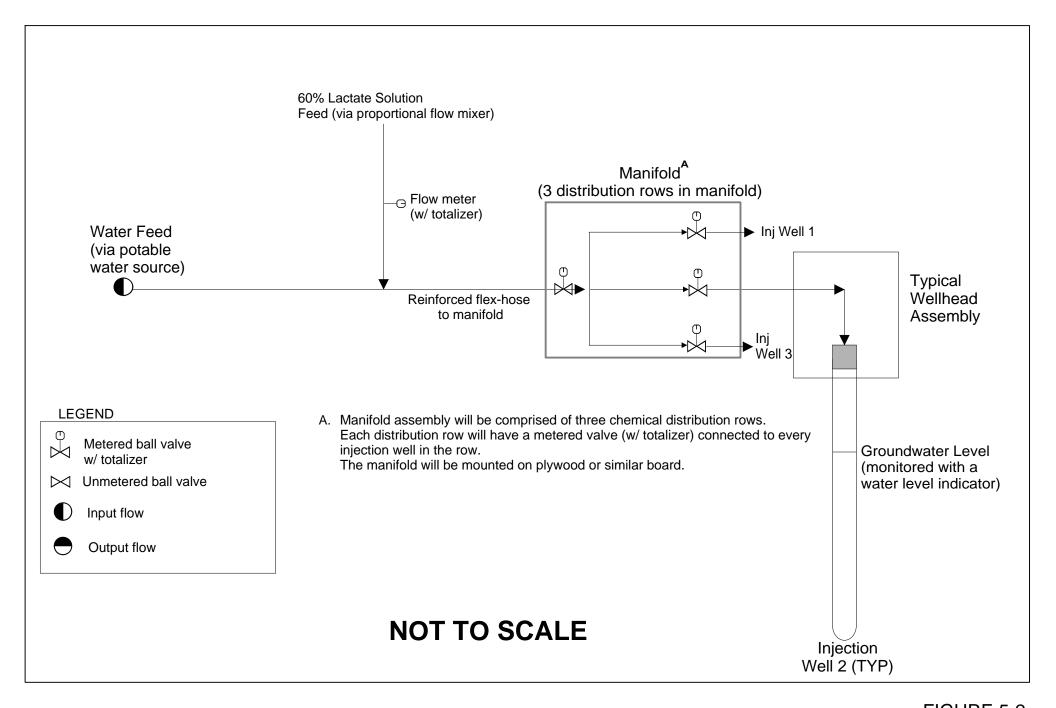
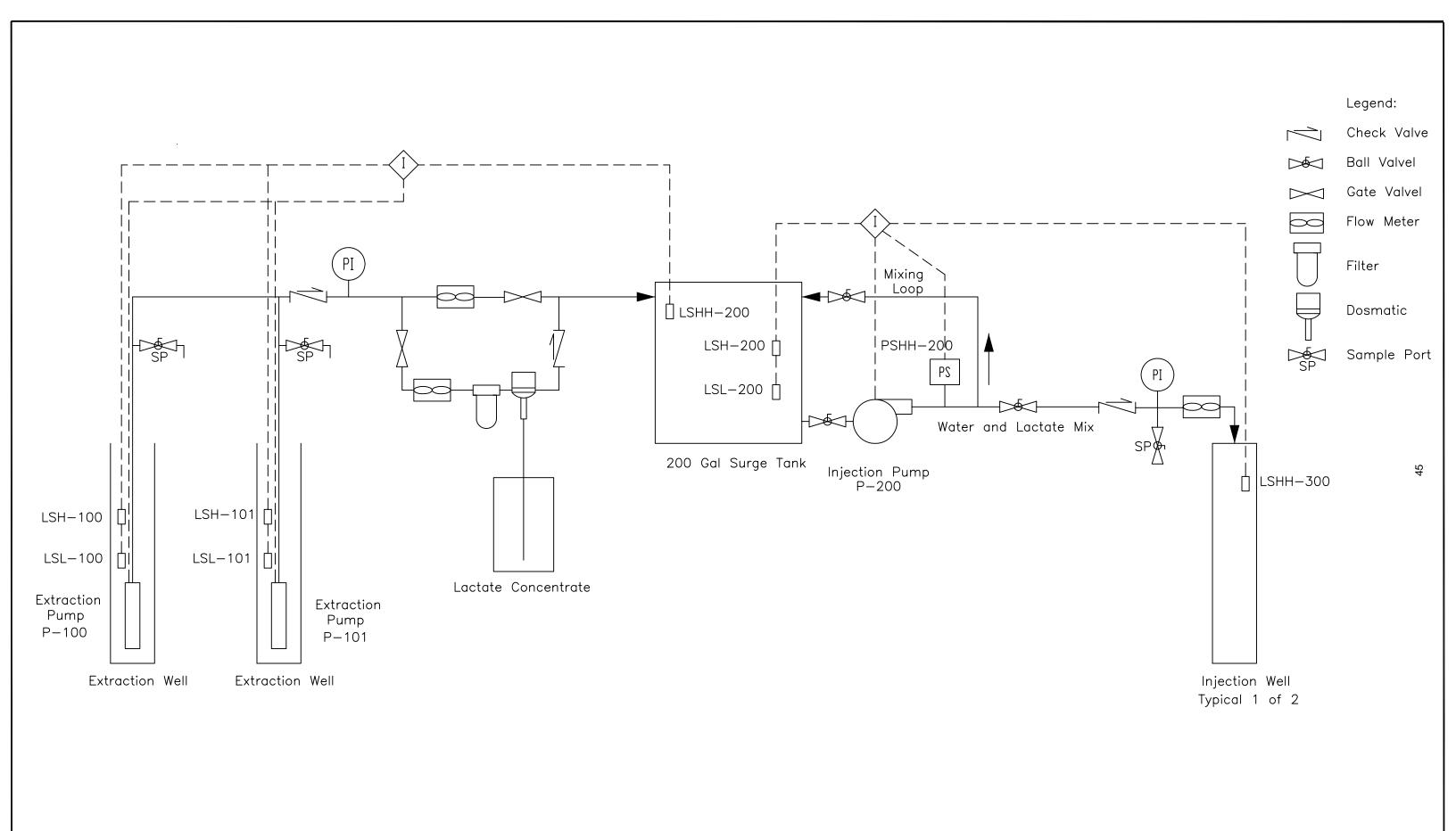




FIGURE 5-2 PROCESS FLOW DIAGRAM - SUBSTRATE DELIVERY FINAL REPORT SEAL BEACH NAVAL WEAPONS STATION, SEAL BEACH, CA



REV.	DATE	DESCRIPTION	BY	CKD.	APP.	DATE:
						APPROVED BY:
						CHECKED BY:
						DRAWN BY: DAG
						DESIGNED BY: DAG

CDW Camp Dresser & McKee Inc.



Navy Bio-Augmentation Site

Seal Beach, CA

Figure 5-3 Active Cell Test Area Injection System P&ID

CADD FILE:	CADD DATE:					
FIG 1	8/15/07					
SCALE						
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	-	SHEET 1 of 1				

in the injection well. When the Hi-Hi level was reached in injection well the injection pumps would s top. The pum p w as a lso i nterlocked with t he H i-Hi pres sure sw itch l ocated on t he discharge. The Hi -Hi pressure all arm would be a result of either c logging of the discharge totalizing meter or the well screen of the injection well. The Hi-Hi pressure alarm was never tripped during operation of the recirculation system.

The disch arge was plu mbed to allow recirculation of the water and donor mixture prior to injection if needed. The discharge line was also equipped with a check valve, pressure gauge, and totalizing flow meter.

All processes were controlled by an Idec brand Programmable Logic Controller (PLC). The PLC allowed for field modifications of the process without the need to rewire the control panel. A wireless telemetry unit was later added to notify the operator of any operational alarms.

Extracted groundwater from the 275-gallon surge tank was pumped to the injection wells during normal operations. During a nelectron donor injection event the extracted groundwater was diverted to a standalone proportional inline mixer, where lactate was added. The lactate-amended water was then conveyed to the injection wells.

The equi pment area was located between the extraction and injection wells based on site constraints. The signal cable between the equipment area and the piezometer level switches was placed in a conduit. Double walled piping was used to convey extracted water to the lactate injection system and to the injection wells.

Submersible pumps were installed 6 inches from the bottom of each extraction well, and piping was installed between the extraction wells and injection wells in the recirculation cell. The extraction wells each transfer to one central vault, which housed all of the controls, sampling ports, flow meters, and check valves for both extraction wells. Piping was then run from the vault to the reinjection wells. The vault, all transfer piping and wiring was installed below ground to minimize impacts to normal operations at the site. Because of low traffic in the area, a shallow (6-inch) trench was dug to install the piping and wiring within a PVC conduit. Once the piping was installed, the trench was covered with new asphalt. All transfer piping between the extraction wells and the injection wells was constructed with high density polyethylene (HDPE), double-lined piping.

5.5.6 Bacteria Distribution System

The bacteria distribution system was designed to inject the desired bacteria directly into each injection well at the wellhead. The bacteria were provided in 20-L pressurized vessels. Pressurized argon was used to evacuate the headspace in each well and to fill the vessel as the bacteria were removed. The well headspace was then evacuated by lowering Teflon tubing to just above the water table and injecting a comparative volume of argon into the well.

Immediately following evacuation, 20 L of bacteria was injected into the subsurface using Teflon tubing. The tubing was installed approximately to the center of the well screen. Figure 5-4 shows a typical bacterial injection setup.

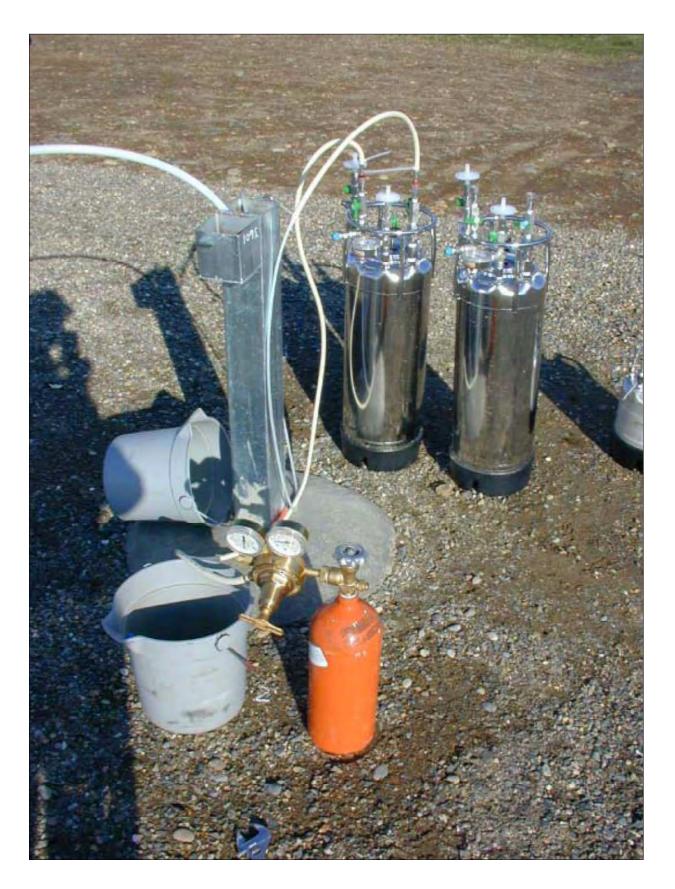




Figure 5-4
BACTERIA INJECTION SYSTEM EXAMPLE
ER-0513 FINAL REPORT
SEAL BEACH NAVAL WEAPONS STATION, SEAL BEACH, CA

5.6 FIELD TESTING

Field a ctivities during this demonstration included system startup, pre-conditioning (Phase 2), Bioaugmentation (Phase 3), and system shut down. A dditional a ctivities included temporary shutdown of the recirculation system and modification of lactate injections. This section includes details of these activities performed during the demonstration.

5.6.1 System Start-up

Once the wells for the active cell were installed and the active cell recirculation system itself was constructed and i nstalled, the recirculation system was tested in M arch and A pril 2008. As described above, the system operated by extracting groundwater from wells AEW-1 and AEW-2 into a 275-gallon surge tank; the surge tank water was reinjected into AIW-1 and AIW-2, which is a distance of 100 f eet upg radient from the extraction wells (refer to Figure 5-1 for well locations). Once the system was functional, it was operated for several days, and water levels were measured in active cell monitoring wells, and in the two existing passive cell monitoring wells, in order to determine the groundwater flow direction in the area of the proposed passive cell wells. Synoptic water level data were collected in several wells using transducers, and in other wells by taking water levels using a water level meter.

Prior to installing the passive cell wells, whose locations were being determined in part based on potential effects from the recirculation system, one round of lactate injections was performed in the active cell in April 2008. The injections were performed by "pulsing" the lactate into the recirculation line to the two injection wells. Approximately 950 gallons was injected at a weight concentration of 2.5 percent (i.e., 25,000 mg/L). The full lactate injection summary is provided in Table 5-2. During startup the observed flow rates from each of the extraction wells were less than anticipated (approximately 0.7-0.8 gp m), and because of this, the concentration of lactate was increased to 2.5 percent from 1 percent during injection.

5.6.2 Pre-conditioning

Once the system was determined to be performing as designed and the additional passive cell wells were installed, "pre-conditioning" of the treatment cells was performed, consisting of lactate injections sufficient to remove sulfate and create strongly reducing conditions. At each well, lactate was injected every 4 weeks into the passive injection wells and quarterly into active injection wells. Pre-conditioning started much sooner in the active treatment cell, though it was completed in January 2009 f or both cells. Approximately 50 ga llons of sodium lactate stock solution was injected into each cell during each injection event. The lactate injection summary is provided in Table 5-2. A more detailed injection summary is provided as Appendix F.

Groundwater s ampling w as pe rformed during pr e-conditioning to m onitor the subsurface conditions prior to initiating bioaugmentation. Groundwater samples were collected according to the s ampling s chedule s hown i n S ection 5. 7.

Table 5-2. Lactate Injection Summary

Well ID	Injection Date Range	Volume Water Injected (gallons)	Volume 60% Sodium Lactate Injected (gallons)	Sodium Lactate Conc. (%)	Volume Lactate Injected ¹ (gallons)	Lactate Injection Conc. (%)	Total Injection Time (Hours)	Average Lactate Injection Flowrate (gpm)
PASSIVE CELL								
Pre-Condi	tioning Totals (Phase	2)						
PIW-1	8/7/08-1/12/09	4,011	67	1.7%	32	0.8%	64.6	1.0
PIW-2	8/7/08-1/12/09	4,156	67	1.6%	32	0.8%	59.5	1.2
PIW-3	8/7/08-1/12/09	4,151	67	1.6%	32	0.8%	59.5	1.2
TOTAL	8/7/08-1/12/09	12,319	201	1.6%	96	0.8%	64.6	3.2
	gmentation Totals (P	hase 3)						
PIW-1	1/13/09-10/31/09	8,481	143	1.7%	69	0.8%	143.4	1.0
PIW-2	1/13/09-10/31/09	8,519	143	1.7%	69	0.8%	143.4	1.0
PIW-3	1/13/09-10/31/09	8,549	144	1.7%	69	0.8%	143.4	1.0
TOTAL	1/13/09-10/31/09	25,549	430	1.7%	206	0.8%	143.4	3.0
OVERALI								
PIW-1	8/7/08-10/31/09	12,492	209	1.7%	101	0.8%	208.0	1.0
PIW-2	8/7/08-10/31/09	12,675	211	1.7%	101	0.8%	202.9	1.0
PIW-3	8/7/08-10/31/09	12,701	211	1.7%	101	0.8%	202.9	1.0
TOTAL	8/7/08-10/31/09	37,868	631	1.7%	303	0.8%	208.0	3.0
			A	CTIVE CELL				
	tioning Totals (Phase							
AIW-1	4/23/08-1/12/09	2,343	96	4.1%	46	2.0%	60.5	0.6
AIW-2	4/23/08-1/12/09	2,507	101	4.0%	49	1.9%	60.5	0.7
TOTAL	4/23/08-1/12/09	4,850	198	4.1%	95	2.0%	60.5	1.3
	gmentation Totals (P							
AIW-1	1/13/09-10/31/09	15,389	547	3.6%	262	1.7%	312.9	0.8
AIW-2	1/13/09-10/31/09	14,375	504	3.5%	242	1.7%	312.9	0.8
TOTAL	1/13/09-10/31/09	29,764	1,061	3.6%	504	1.7%	312.9	1.6
OVERALI								
AIW-1	4/23/08-10/31/09	17,732	643	3.6%	309	1.7%	373.4	0.8
AIW-2	4/23/08-10/31/09	16,882	605	3.6%	290	1.7%	373.4	0.8
TOTAL	4/23/08-10/31/09	34,614	1,258	3.6%	599	1.7%	433.9	1.3

¹ 60% Sodium Lactate contains approximately 48% bioavailable lactate.

5.6.3 Temporary System Shutdown

The r ecirculation s ystem w as s hut down t emporarily to add a dditional c ontrols i ncluding a secondary overflow tank and an autodialer in late 2008. Therefore, the recirculation system was not operating between October 2008 and January 2009. The system was re-started approximately one week before beginning Phase 3 – Bioaugmentation.

5.6.4 Bioaugmentation

Once the p re-conditioning phase was completed, both the passive and active cells were inoculated with the SDC-9TM *DHC* culture in January 2009. The inoculation was performed by first in jecting 90 percent of the monthly electron donor volume into each cell, followed by inoculation, and finally by "flushing" the wells with anoxic water.

To do this, I actate injections i nto the passive and active cells were performed the week of January 5, 2009. In the passive cell, approximately 953 gallons of 1 percent lactate solution were injected into wells PIW-1, PIW-2, and PIW-3.

A lactate i njection was also performed into the active cell the week of January 5, 2009. Approximately 2,975 g allons of 1 percent to 1.5 percent lactate was injected into wells AIW-1 and AIW-2 by feeding lactate into the recirculation water.

Following the initial lactate injections, each cell was inoculated with approximately 100 L of SDC-9TM. The inoculation was performed by injecting proportional amounts of culture into each injection well (50 L per well in active cell, 33 L per well in passive cell) with argon as a carrier gas to ensure the culture did not come in contact with air.

Once the wells were inoculated, the final 10 percent of lactate-amended water for the injection was added to each injection well (i.e., 100 ga llons per well). This lactate solution was mixed approximately 72 hours before injecting to ensure that the water was anoxic.

5.6.4.1 Lactate Injection Modifications

Following the bioaugmentation, lactate injections were continued for 8 months. However, the injection strategy was modified in the active cell. Because carbon distribution was less than anticipated in the active cell, the pulsing strategy was modified to weekly from monthly. Although the frequency of injections was increased, the volume was decreased to approximately 12.5 gallons of stock lactate per event such that the monthly lactate mass injected did not change.

In June 2009, the active cell lactate injection strategy was modified again based on continued low carbon distribution throughout the active cell. The lactate concentration during each weekly injection was increased such that 50 gallons of stock sodium lactate were injected per event.

5.6.4.2 Groundwater Sampling

Groundwater sampling was performed following bioaugmentation to monitor the contaminant destruction, electron donor distribution, and bacterial distribution and activity. Groundwater samples were collected according to the sampling schedule shown in Section 5.7.

5.6.5 System Shut-down

In October 2009, the recirculation system was shut down. Once it was determined in March 2010 that no additional data would be collected, the system was decommissioned in April 2010, and all equipment was removed from the site.

5.7 SAMPLING METHODS

Groundwater sampling was performed in each of the three phases of the demonstration to collect data sets that would achieve the project objectives. Phase 1 included one round of baseline sampling, and Phase 2 included three rounds of sampling. Following bioaugmentation, eight rounds of sampling were performed.

5.7.1 Sampling Summary

Samples were collected as shown in Table 5-3 during the demonstration. All injection wells and monitoring wells (including CMT wells) were sampled in the passive cell during each event, and the combined effluent from the two extraction wells and all monitoring wells (including CMT wells) were sampled in the active cell during each event. Not all analyses were performed during each event, as specified in Table 5-3. Not all screened intervals in the CMT wells were sampled during each event by design. Additionally, because the depth to water varied during the course of the demonstration, the amount of intervals sampled had to be modified if certain intervals were dry. A detailed summary of the samples collected is provided in Appendix G.

5.7.2 Analytical Methods

Analytical t echniques f or t his d emonstration i ncluded s tandard E PA methods f or V OCs, ethene/ethane/methane, anions, COD, and alkalinity, as well as a ccepted field measurements using water quality instruments and field test kits. Two innovative a nalytical techniques f or which no s tandard E PA methods exist are included in this demonstration, both of which are important for assessing the demonstration's performance. A summary of the analytical methods used is provided in Table 5-4.

The two innovative analytical techniques used during this demonstration were qPCR and carbon stable isotope analysis (CSIA). As discussed above, these techniques do not have standard EPA methods, although the methods have been published. The actual analytical method is published for qPCR by Rahm et al. (2006) and for CSIA by Song et al. (2002).

qPCR

The most crucial of these methods is qPCR, which was used to track the growth and distribution of the introduced bacteria. Initial detections of bacteria at a given well were used to calculate bacterial transport times, which were used to infer whether differences in the bioaugmentation strategies im pacted distribution.

Table 5-3. Monitoring Summary

Sampling Round	Sampling Date	Number Recirculation Cell Well Samples			Number Passive Cell Well Samples			- Total	Number of
	Sampling Date	Extraction	Monitoring	CMT ¹	Injection	Monitoring	CMT ¹	Total	QA/QC samples ²
Baseline Sampling (C)	April-08	2	3	9	3	6	9	32	4
Pre-conditioning – Month 1	May-08	2	3	3	3	6	3	20	2
Pre-conditioning – Month 2	September-08	2	3	3	3	6	3	20	2
Pre-conditioning – Month 3 (C)	November-08	2	3	9	3	6	9	32	4
Bioaugmentation sampling – Month 4	January-09	2	3	3	3	6	3	20	2
Bioaugmentation sampling – Month 5 (C)	February-09	2	3	3	3	6	3	20	2
Bioaugmentation sampling – Month 6	March-09	2	3	3	3	6	3	20	2
Bioaugmentation sampling – Month 7 (C)	April-09	2	3	9	3	6	9	32	4
Bioaugmentation sampling – Month 8	May-09	2	3	3	3	6	3	20	2
Bioaugmentation sampling – Month 9	June-09	2	3	3	3	6	3	20	2
Bioaugmentation sampling – Month 10 (C)	October-09	2	3	9	3	6	9	32	4
Bioaugmentation sampling – Month 13 (C)	December-07	2	3	3	3	6	3	20	2
Bioaugmentation sampling – Month 16 (C)	N/S	N/S Month 16 Sampling Event not required based on meeting demonstr				emonstra	tion obje	ectives.	
Totals			· T11 7 A					320	36

All samples were analyzed for the following parameters (analysis details shown in Table 5-4):

- Field parameters
 - o Conductivity, pH, Temperature, Dissolved Oxygen, Oxidation Reduction Potential, Turbidity, Ferrous iron, and iodide tracer
- Lab parameters (Method ID)
 - o VOCs (8260B), Dissolved Gases Methane, Ethane, and Ethene (RSK 175), Anions (353.2), Alkalinity (310.1), DNA Analysis (qPCR), Chemical Oxygen Demand COD (410.4)
- **(C)** All samples collected during the Baseline, Month 3, Month 5, Month 7, Month 10, Month 13, and Month 16 sampling periods were analyzed for stable carbon isotopes.
- 1 Only one depth sampled from each CMT well during months 1, 2, 4, 5, 6, 8, 9 and 13. Up to 3 depths sampled in other sampling periods, depending on observed water levels.
- 2 Approximately 10% of all samples were collected for QA/QC during the monitoring period.

Table 5-4. Sample Collection and Analysis Summary

Analytes	Sample container size and type	Preservative	Analytical Method	Holding time	Comments
Field laboratory analyses [priority]					
Ferrous Iron [1]	One 125-mL HDPE	4°C	Hach Method 8146	30 minutes	Must be analyzed immediately; no headspace
Tracer - Iodide [2]	One 125-mL HDPE	4°C	Ion specific Electrode	4 hrs	
Off-site laboratory analyses					
VOCs	Two glass 40-mL VOA vials	4°C	SW-846 8260B	7 days	No headspace
DNA Sequencing	One 1-L HDPE	4°C	qPCR	3 days	No headspace
Stable Carbon Isotopes	One 1-L HDPE	4°C	GC-IRMS	7 days	No headspace
Ethene/ethane/methane	Three glass 40-mL VOA vials	4°C	RSK-175 (or equivalent)	7 days	No headspace
Chloride	One 250-mL HDPE	4°C	EPA 325.3	28 days	
Chemical Oxygen Demand	One 50-mL HDPE	H ₂ SO ₄ / 4°C	EPA 410.4	28 days	
Alkalinity	One 250-mL HDPE	4°C	EPA 310.1	14 days	
Nitrate	One 250-mL HDPE	4°C	EPA 300.0	48 hours	See below
Nitrite/Nitrate	One 250-ml HDPE	H ₂ SO ₄ / 4°C	EPA 353.2	14 days	Added because 48-hour hold time not always achievable for Nitrate analysis
Sulfate	One 250-mL HDPE	4°C	EPA 375.4	28 days	

qPCR = quantitative polymerase chain reduction HDPE = high-density polyethylene VOA = volatile-organic analysis

The DNA extractions and qPCR analyses were performed by North Wind, Inc. because of their specialized expertise in c lone l ibrary de velopment, D NAs equencing, and qPCR method development.

CSIA

The second innovative analytical technique was CSIA for TCE, *cis*-1,2-DCE, VC, and ethene. Following the analysis, stable carbon isotope ratios for each compound were determined to evaluate degradation patterns and the extent of dechlorination of parent compounds. Stable carbon isotope ratios are described in terms of δ^{13} C, which is defined by the following equation:

$$\delta^{13}$$
C = ((R_{sample}/R_{standard}) - 1) x 1,000

where:

 δ = delta notation of stable isotope ratio

 ^{13}C = carbon-13

R = concentration of carbon-13/concentration of carbon-12

Thus, if the sample has a lower ratio of carbon-13 to carbon-12 than the ratio of the reference standard, δ^{13} C is negative. If the sample has a higher ratio, then δ^{13} C is positive. Stronger molecular bonds are formed by c arbon-13 than by c arbon-12. When dechlorination starts, the weaker-bonded c arbon-12 i sotopes t end t o b e transformed m ore qui ckly, resulting i n t he enrichment of carbon-13 in the residual reactant (e.g., cis-1,2-DCE that is being transformed to VC). This causes δ^{13} C to increase for *cis*-1,2-DCE. On the other hand, the amount of carbon-12 in the product (in this case, VC and ethene) is initially higher, causing δ^{13} C to be more negative. However, if a finite amount of reactant is present and the reaction proceeds to completion, then δ^{13} C of the product(s) will equal that of the initial reactant (Song et. al, 2002). In other words, when dechlorination starts, the δ^{13} C of the newly formed vinyl chloride and ethene will initially be much "lighter" (more negative) than baseline samples of cis-1,2-DCE (because of a higher amount of carbon-12 in the newly formed compounds than in the original cis-1,2-DCE). The cis-1,2-DCE's δ^{13} C will, in turn, become "heavier" (less negative) than baseline (because of a higher amount of carbon-13 than carbon-12 in the remaining cis-1,2-DCE) as it is dechlorinated. As the cis-1,2-DCE is completely dechlorinated, the δ^{13} C in the degradation products will approach and eventually equal that of the original *cis*-1,2-DCE.

The CSIA was performed by Lawrence Berkeley National Laboratory (LBL). The Center for Isotope Geochemistry stable isotope laboratory at LBL conducts basic and applied geochemical research using the isotope ratios of light elements including hydrogen, carbon, nitrogen, oxygen and chlorine. Results are included in Appendix H for the active cell and Appendix I for the passive cell.

Field Analyses

Field analyses for ferrous iron were performed as per the test kit manufacturer's instructions. Field analyses for DO, ORP, temperature, pH, and specific conductivity were performed as per the water quality meter manufacturer's instructions. Analysis for iodide tracers was performed per the ion specific electrode manufacturer's instructions.

5.7.3 Quality Control

Laboratory quality assurance (QA) for the onsite field analyses included analysis of blanks and duplicates. O ffsite la boratory quality a ssurance r equirements were defined in the laboratory SOW. Frequencies for QA analyses are specified in Table 5-5. Further details are provided in Appendix G, which addresses the appropriate sections of the Quality Assurance Project Plan for this demonstration. Also included in Appendix G is a description of the calibration procedures performed for all equipment not o perated by a contract laboratory. For all equipment us ed outside the contract laboratory, calibration procedures were performed as per the manufacturer guidelines. Sample documentation procedures are also detailed in Appendix G.

All data, checklists, photographs, and calibration logs generated during the demonstration were included as part of the project file. These data and reports will be maintained by CDM.

Sample TypeFrequencyCommentsField Duplicate1 per 20 samplesaAll samplesField blank1 per 20 samplesaAll samplesTrip blank1 per sample coolerFor off-site VOCs and ethene/ethane/methane samples only.

Table 5-5. Field QA frequency for Groundwater Monitoring

5.7.4 Decontamination Procedures

Any residuals that were generated during drilling and during the technology demonstration were handled and disposed in an appropriate manner. Residuals generated from this work included water during drilling, well de velopment, and e quipment de contamination; pur ge water from sampling; drill cuttings; field test kit wastes; sampling equipment de contamination wastes; and personal protective equipment (PPE).

Water generated during the demonstration was stored temporarily in a storage tank and then sent to an appropriate disposal facility for disposal. Soil generated during well installation was stored in a covered bin onsite

All solid waste and RCRA waste was disposed offsite. The Generator EPA ID number for this site is CA0170024491.

5.8 SAMPLING RESULTS

This se ction summarizes the sam pling results from the act ivities specified in Section 5.6. Specifically, an analysis of the concentration trends for five main parameters is provided in this section. In order for complete reductive dechlorination of TCE to ethene to occur biologically, electron donor must be adequately distributed, redox conditions must be sufficiently reducing, pH should be in the appropriate range, and appropriate microbial populations must be present and active. The performance of the active and passive cells was therefore evaluated based on the success of electron do nor injections, extent of electron do nor distribution, changes in redox

a: 1 sample for all analytes per day if number of monitoring locations is <20.

conditions, extent and rate of dechlorination, and changes in the microbial population within the aquifer of the active and passive cells.

5.8.1 Active Cell

Trends for the five parameters of interest in the active cell are presented in this section.

5.8.1.1 Electron Donor Distribution

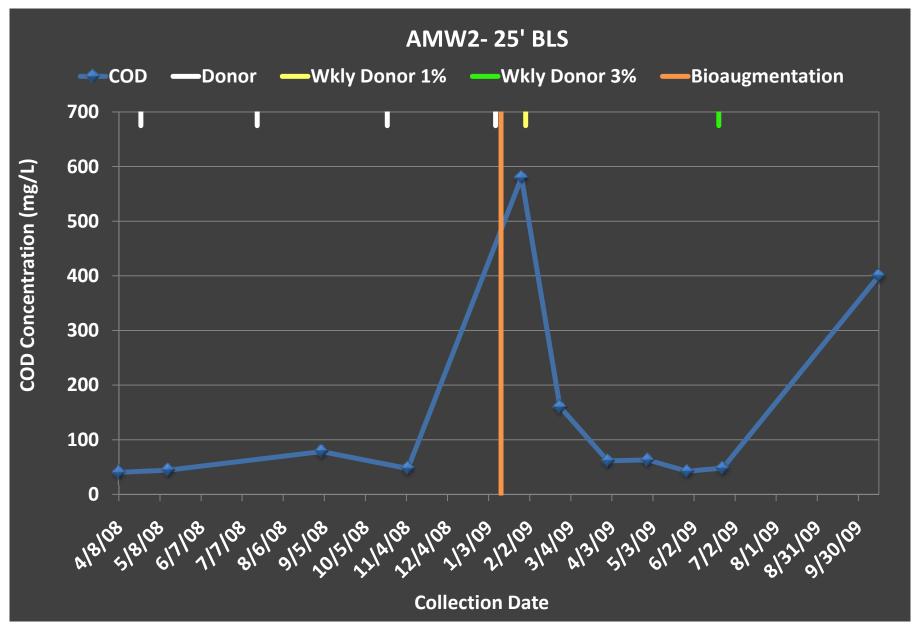
COD was measured to indicate the amount of available electron donor in the groundwater. COD is an important metric, as it represents the carbon and energy available to dechlorinating bacteria. Complete COD results for the active cell are included in Table H-1 and figures showing the key COD concentration trends are presented in Appendix H.

The baseline sampling event (April 2008) showed COD concentrations ranging between 28 and 60 mg/L in active cell wells. During the pre-conditioning phase (April 2008 to January 2009) when quarterly pulsed electron donor injections (1,000 gallons of 2 percent (v/v) of sodium lactate solution) were performed, COD concentrations were observed to increase slightly (i.e., near 2X background concentrations) only at wells AMW-2 (78 mg/L in September 2008) and AMW-4 (Z1) (120 mg/L in May 2008). The concentrations at all other wells and zones remained near ba ckground. The quarterly injections were able to achieve some increase in COD concentrations compared to baseline and the electron donor distribution was observed approximately 36 feet downgradient (well AMW-4) of the injection wells within the active cell.

To a chieve better electron donor distribution and increase the COD concentrations within the active cell, the injection strategy was modified to include weekly electron donor injections with approximately 750 gallons of 1 percent (wt/wt) sodium lactate solution between January 26 and June 9, 2009. D uring this period, a slight increase in COD concentrations near 2X background was observed at the monitoring wells (AMW-4 (Z1) - 85 mg/L in February 2009, AMW-5 (Z1 [83 mg/L in June 2009] and Z2 [70 mg/L in April 2009]) and the upgradient well (AMW-1 – 120 mg/L in February 2009). O nly well A MW-2 showed C OD concentrations of a few hundred mg/L, which peaked in January 2009 (580 mg/L), but then decreased and was observed near background by May 2009 (Figure 5-5). The concentrations at all other wells and zones remained near background. The donor distribution was still approximately 36 feet downgradient but now included well A MW-5 and effects were a lso observed a pproximately 25 f eet upgradient (AMW-1) of the injection wells within the active cell.

To further improve electron donor distribution and increase the COD concentrations within the active cell, the injection strategy was modified again to include weekly electron donor injections using a pproximately 1,000 ga llons of 3 percent sodium lactate solution be tween J une 10 a nd October 2, 2009. E levated C OD c oncentrations in the range of a few hundred mg/L were observed in October 2 009 at a number of monitoring wells including A MW-2 (400 m g/L), AMW-3 (Z2) (540 m g/L), AMW-4 (Z1 [420 mg/L]), A MW-5 (Z2) (350 m g/L), and the upgradient well, AMW-1 (180 mg/L). The donor distribution was now greater than 36 f eet downgradient of the injection wells and also included wells AMW-3 (Z2) and AMW-4 (Z2) but still failed to reach well A MW-6 located a pproximately 72 f eet downgradient of the injection well. C ontinued e ffects of donor distribution were a lso observed a pproximately 25 f eet upgradient of the injection well within the active cell. The majority of the COD increases were

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observed in zones 1 and 2 of the CMT wells. It should be noted that Zone 3 of the CMT well AMW-5 was never sampled during the pilot test due to a lack of water at that location. Furthest downgradient wells AMW-6 and AEW did not show elevated COD concentrations throughout the pilot test, indicating that donor was not distributed at these wells.

5.8.1.2 Redox conditions

Redox conditions are frequently monitored by m easuring the ORP. It is a simple indicator of redox conditions and can be easily measured on site during the field activities. However, it is not the most accurate parameter in assessing the actual redox conditions, and if considered alone can sometimes be misleading. Thus, it is also required to monitor concentrations of certain inorganic electron acceptors in addition to ORP to assess the redox conditions at a si te accurately. ORP measurements and concentrations of inorganic electron acceptors (DO, ni trate, f errous iron, sulfate and methane) for the active cell are included in Table H-1 and figures showing the key changes in electron acceptors are presented in Appendix H.

ORP

ORP is measured in a flow-through cell during sampling. Generally, ORP measurements that are slightly positive i ndicate mildly r educing c onditions. R eductive de chlorination is generally possible with ORP values less than approximately +50 millivolts (mV), but more negative ORP measurements (less than -100 mV) indicate strongly reducing conditions that are favorable for complete reductive dechlorination (EPA, 1998).

ORP values at all the active cell wells were mostly high and ranged from 73 mV to 443 mV during the baseline sampling event, with the exception of wells AMW-5 (Z1 [-83 mV] and Z2 [15 mV]). Following electron donor injections ORP values were reduced at all the active cell wells. As of O ctober 2009 the ORP values were observed below 50 mV at all the active cell wells, and were observed below -100 mV at wells AMW-1, AMW-2, AMW-3 (Z2), AMW-4 (Z1 and Z2), and AMW-5 (Z1 and Z2). Overall, the ORP values at the monitoring wells were in the appropriate range for dechlorination and indicate establishment and sustenance of moderate to strongly reducing conditions within the active cell.

Electron Acceptors and Reduced Products

As discussed above, the aqueous concentrations of inorganic electron acceptors and their reduced products are a more reliable indicator of reducing conditions in the groundwater than ORP. The redox conditions typically progress from aerobic \rightarrow nitrate reducing \rightarrow iron reducing \rightarrow sulfate reducing \rightarrow methanogenic following addition of a sufficient supply of electron donor. Decreases in concentrations of DO, nitrate, and sulfate, and increases in ferrous iron and methane indicate that conditions are becoming favorable for dechlorination.

Dissolved Oxygen

Low DO concentrations are required for reductive dechlorination to occur; generally DO concentrations less than 0.5 m g/L are best for reductive dechlorination, whereas higher DO concentrations (generally greater than 1 mg/L) are harmful (EPA, 1998). DO was not a r eliable r edox i ndicator during t his demonstration, l ikely be cause of e quipment problems, and so it is not discussed here.

Nitrate Reduction

Nitrate concentrations of less than 1 mg/L are considered appropriate for dechlorination (EPA, 1998). The baseline sampling events howed nitrate concentrations less than 1 mg/L at all the active cell wells. The already low nitrate concentrations were reduced and observed near or below detection limit at all the active cell wells during the pilot test. Overall, the results indicate that nitrate reduction was not an important process within the active cell due to the lack of nitrate available.

Iron Reduction

Ferrous iron is the product of ferric iron reduction. Ferrous iron concentrations of near or greater t han 1 mg/L a re c onsidered indicative of iron -reducing c onditions t hat c ould support dechlorination (EPA, 1998). The baseline sampling event showed ferrous iron concentrations of 1 ess t han 0.25 mg/L at al 1 t he act ive cel 1 w ells. Ferrous i ron concentrations increased at all the active cell following donor distribution except wells AMW-6 and AEW. At well AMW-2 the ferrous iron concentration were near or above 3 mg/L be tween S eptember 2008 a nd June 2009 but were reduced to be low detection limit in October 2009. The blackish water observed during this sampling event indicates that the decrease in ferrous iron concentration may be due to the production of reduced iron sulfide minerals (ferrous iron reacts with sulfide, which is formed from sulfate reduction). This has been observed at sites where ferrous iron is not available in dissolved form under intrinsic conditions and sulfate is present in large a mounts. As of October 2009, elevated ferrous i ron concentrations of n ear or a bove 3 m g/L w ere observed a t wells AMW-3 (Z2 and Z3), AMW-4 (Z1 and Z2), and AMW-5 (Z2). Increases in ferrous iron concentrations were also observed at wells A MW-3 (Z1) (February 2009) and AMW-4 (Z3) (April and June 2009), but the concentrations were not sustained. At the upgradient well AMW-1 the ferrous i ron concentrations varied and depended on the donor di stribution. A s of O ctober 2009 e levated f errous i ron c oncentration (above 3 mg/L) were observed at well AMW-1. Overall, the results indicate that iron reducing conditions were established at the wells in the upper portion of the active cell.

Sulfate Reduction

Optimal dechlorination rates are typically supported by sulfate concentrations of less than 20 mg/L (EPA 1998). However, as shown in Section 5.4, dechlorination c an oc cur at sulfate concentrations higher than this at sites where initial sulfate is greater than 500-1,000 m g/L. B ecause of t his, t he more i mportant i ndicator of a ppropriate redox conditions i s dow nward t rends in s ulfate c oncentrations, w hich i ndicate t hat s ulfate reduction is occurring.

Baseline sulfate concentrations were above 3,000 mg/L in all the active cell wells except well AEW. Near the injection wells, sulfate was above 7,000 mg/L; closer to extraction wells AEW the sulfate concentration was 1,600 mg/L. Following donor injections sulfate concentrations decreased considerably at all the wells in the upper portion of the active cell: AMW-1, AMW-2, and all three zones of CMT wells except well AMW-5 zones 1 and 3 (no data c ollected). As of O ctober 20 09 s ulfate reductions in the range of 62 percent to 98 percent were achieved at wells AMW-2, AMW-3 (Z1 to Z3), AMW-4 (Z1 to Z3), and AMW-5 (Z2) depending on the extent of donor distribution. At the upgradient well AMW-1 the sulfate concentrations varied and depended on the donor distribution.

Compared to baseline, 57% removal of sulfate was observed at well AMW-1 in October 2009. Some increase in sulfate concentrations was observed at wells AMW-6 and AEW during the pilot test indicating the breakthrough of water from upgradient at these wells. Overall, the results indicate that sulfate reducing conditions were established at the wells in the upper portion of the active cell.

Methanogenesis

Methanogenesis, the production of methane from carbon dioxide, is the most favorable redox c ondition f or complete dechlorination. Met hanogenesis r esults i n i ncreased concentrations of methane. D uring t he ba seline sampling e vent, l ow methane concentrations (less than 0.15 mg/L) were observed at all the active cell wells. Methane concentrations r emained near ba seline (less t han 0.15 mg/L) at t he active cell wells throughout the operation of the pilot test indicating that strongly methanogenic conditions were not observed at any well within the active cell.

Redox Summary

Based on the results discussed in this section, it can be concluded that redox conditions shifted in accordance with the electron donor distribution, and as of October 2009, sulfate reducing to methanogenic conditions were established within the active cell except in the furthest downgradient locations, AMW-6 and the AEW wells. An example of redox conditions is included in Figure 5-6 for AMW-4 Zone 1.

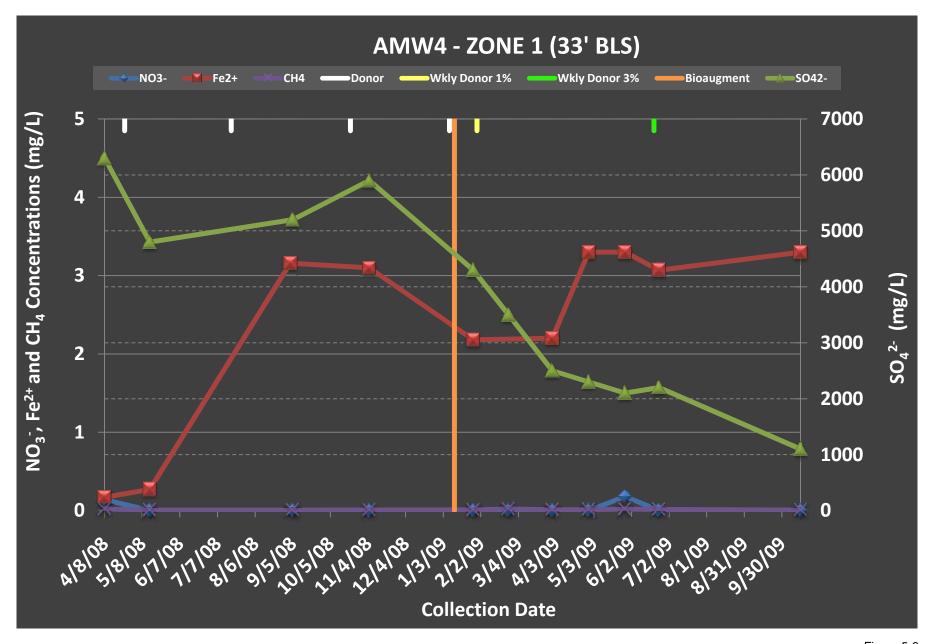
5.8.1.3 VOC Concentrations

The concentrations of electron donor and redox conditions only indicate whether conditions are favorable for reductive dechlorination to progress at a site. The concentrations of chloroethenes and ethene need to be monitored as direct evidence. Complete results for VOC concentrations for the active cell are presented in Table H-1 and figures showing the key VOC concentration trends are presented in Appendix H.

Baseline c onditions (April 2008) were c haracterized by hi gh c hloroethene co ncentrations observed at the active cell wells, primarily consisting of TCE concentrations ranging between 96 $\mu g/L$ and 10,000 μ g/L and DCE c oncentrations ranging be tween 5 $\mu g/L$ and 660 μ g/L. Exceptions were we lls AM W-6, which e xhibited a much hi gher TCE c oncentration of 140,000 $\mu g/L$, and A EW, which exhibited a D CE c oncentration of 1,9 00 $\mu g/L$. A l ow concentration of VC was detected only at well AEW (48 $\mu g/L$), whereas ethene was not detected at any of the active cell location during the baseline sampling event.

Following electron donor injections, an increase in TCE and total chloroethene concentrations was not ed at all the wells sampled. This was likely caused by de sorption and/or enhanced dissolution from a residual TCE source, and also due to the fact that TCE concentrations near the extraction well were higher than those near the injection wells at the start of recirculation. The concentrations of total chloroethenes increased by a factor ranging from nearly 4X at well AMW-2 to greater than 39X at well AMW-4 (Z2) in April 2009 when compared to the baseline concentrations.

During the pr e-conditioning pha se (April t o N ovember 2008), a dramatic inc rease in DCE concentrations ranging from 650 μ g/L to 8,400 μ g/L in November 2008 was observed at the



wells loc ated in the upper half of the a ctive c ell (A MW-1, A MW-2, A MW-3 (Z1 t o Z 3), AMW-4 (Z1 to Z3), and AMW-5 (Z1 to Z3). The increase in DCE concentrations indicated that degradation of the TCE was occurring. However, very little increase in VC concentration was observed during pre-conditioning, (detected at 9 μ g/L to 35 μ g/L at wells A MW-2, AM W-3 (Z1), AMW-4 (Z1), and AMW-5 (Z2)) and no e thene production was observed during the preconditioning phase. These results suggested the necessity for bioaugmentation for dechlorination to progress within the active cell.

Following b ioaugmentation and the change to weekly lactate injections, further progress in dechlorination was observed rapidly in the upper half of the active cell with increases in removal of TCE and conversion largely to VC and some ethene. The highest ethene concentration of 200 μ g/L was observed at well AMW-3 (Z1) during June 2009. Well AMW-5 (Z3) could only be monitored in April 2009 and the presence of large concentrations of DCE (>4200 μ g/L) and some VC (170 μ g/L) indicated that this well was also being impacted by the injections. At the upgradient well AMW-1, good progress in dechlorination was observed following donor injections and bioaugmentation but the TCE concentrations started rebounding between April and June 2009 due to limited electron donor availability. Based on the lack of complete conversion of TCE to ethene in the upper part of the active cell, combined with less favorable conditions observed at well AMW-1, the electron donor injection strategy was modified again by increasing the volume and concentration of weekly electron donor injections.

Complete reductive dechlorination of TCE to ethene was observed in the upper half of the active cell following the increase in electron donor volume and concentration that began in June 2009. As of October 2009, TCE degradation ranging from 85 percent to 99.7 percent was achieved in the upper p ortion of the active cell. In addition, large increases in VC concentrations ranging from 510 μ g/L to 6, 000 μ g/L, and significant ethene production ranging from 47 μ g/L to 1,500 μ g/L at wells AMW-1 and AMW-2, and all three zones of the three CMT wells, indicated that complete dechlorination was achieved. Zone 2 of the CMT wells appeared to be the most impacted with much higher ethene production observed, followed by zone 1 and then zone 3.

At well AMW-6, TCE concentrations decreased by 79 percent, DCE concentrations increased by 642 percent, and dramatic increases in VC c oncentration from be low de tection limit to 4,900 μ g/L were observed in October 2009. Because little change was observed in the COD and redox data at AMW-6, these VOC results suggest that the shift in VOC concentration is a result of biodegradation occurring upgradient and degradation products being transported to this well. Similarly, at well A EW, TCE co ncentrations increased by 50 percent, D CE c oncentrations increased by 15 percent, and a large increase in VC concentration from 48 μ g/L to 510 μ g/L was observed in October 2009.

Once complete reductive dechlorination of TCE to ethene was achieved, a loss of chloroethene mass balance was observed at all the wells located in the upper half of the active cell. This phenomenon has been observed at other sites with similar conditions, namely shallow, relatively "thin" contaminated aquifers (e.g., F rench et al, 2003). This result can at least partially be attributed to the volatilization of VC and ethene to the vadose zone.

In summary, complete reductive dechlorination of TCE to ethene was achieved only in the upper half of the active cell (greater than 36 feet downgradient and approximately 25 feet upgradient of the injection wells) as a function of electron donor distribution. An example of this is included in Figure 5 -7 f or A MW-1. Vertical di stribution of e lectron donor a ppears e ffective w ith dechlorination of TCE to ethene being observed in all three zones of all the three CMT wells. However, Zone 2 of the CMT wells was impacted the most, followed by Zone 1 and Zone 3. It should also be noted that the considerable production of ethene occurred in the presence of high sulfate concentration and minimal methane production which confirmed that complete reductive dechlorination could be achieved in the presence of high sulfate concentrations.

CSIA da ta for the a ctive c ell ge nerally were c onsistent with the C VOC da ta, in that they suggested degradation to VC and ethene was occurring. An example CSIA chart is included as Figure 5-8 for A MW-2. T his chart shows a very "heavy" signature (less negative) for TCE, indicating that it has been substantially degraded. A lso, c-DCE and VC also become he avier during the course of the demonstration, indicating degradation is occurring. Ethene was detected at this location, but not in high enough concentrations to be able to perform an isotope analysis. The rest of the active cell CSIA data are included in Appendix H.

5.8.1.4 Biological Indicators

Dechlorinating bacteria, pH, and alkalinity can serve as indirect lines of evidence for occurrence of bi ological a ctivity within the aquifer. In particular, increase in numbers (i.e., gr owth) of dechlorinating bacteria suggests the occurrence of bi odegradation of VOCs within the aquifer. These parameters are discussed below.

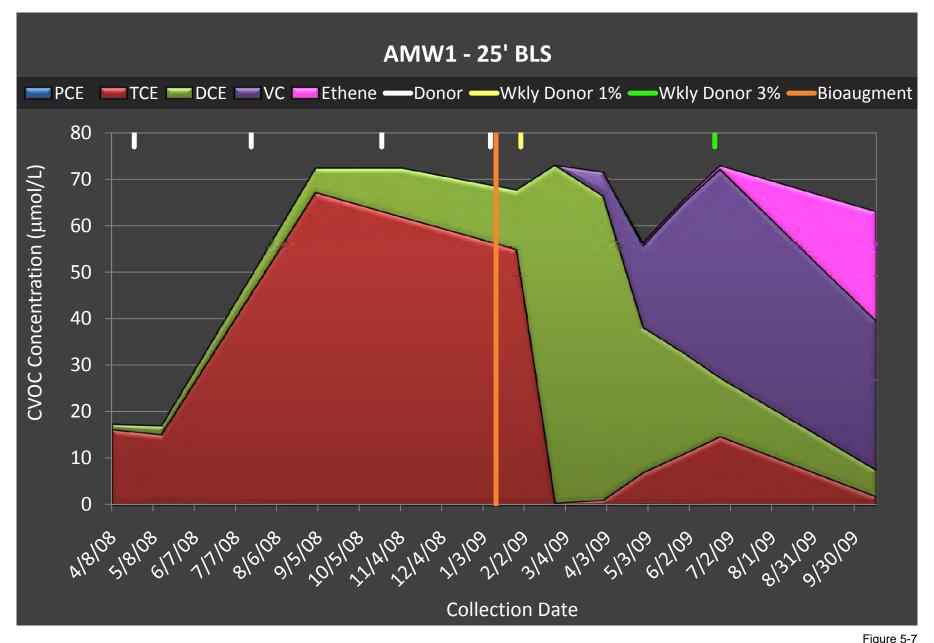
Dechlorinating Bacteria

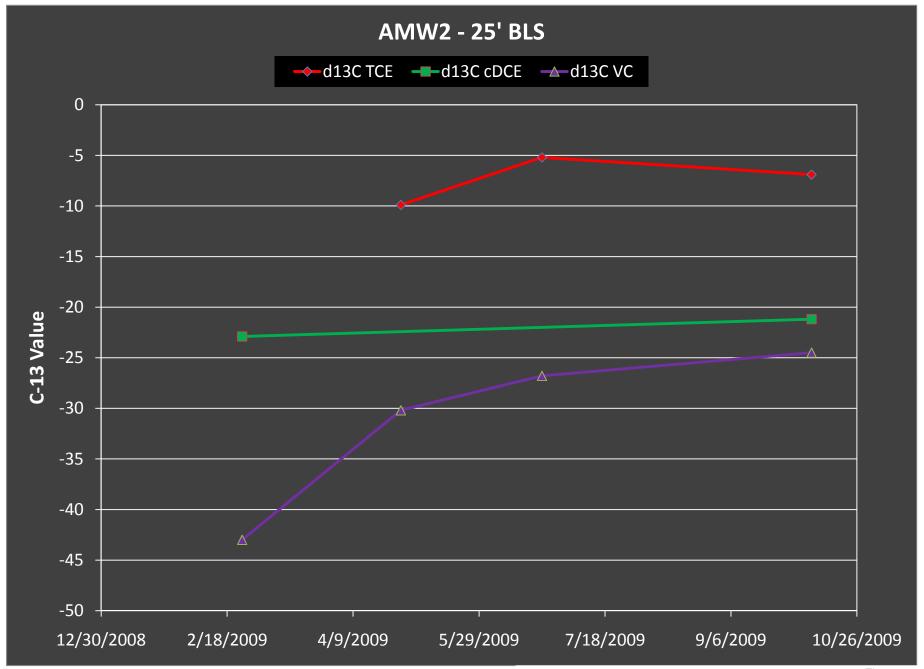
DNA sa mpling was p erformed at t he ac tive cel l w ells t o evaluate t he pr esence of the dechlorinating ba cteria *DHC* prior t o bi oaugmentation, and more importantly, the success of bioaugmentation f ollowing inoculation of bacteria. Complete r esults f or t he a ctive cell ar e provided in Table H-1 and figures showing the key DNA trends are presented in Appendix H.

During the baseline sampling event, low numbers of DHC (16S rRNA and functional genes tceA and bvcA) on the order of 10^2 gene copies/L were observed only at well AEW within the active cell. The functional gene vcrA was not observed at any well within the active cell.

During the pre-conditioning phase low numbers of *DHC* bacteria (16S rRNA and/or functional genes *tceA* and *bvcA*) on the order of 10² gene copies/L to 10⁴ gene copies/L were observed at wells AMW-2 (November 2008), AMW-4 (Z1) (November 2008), AMW-5 (Z1) (May 2008 and September 2008), AMW-6 (September 2008), and continued to be observed at well AEW (May through N ovember 2008). H owever, the functional gene *vcrA* was not observed at any well within the active cell. The DNA results suggested the need for bioaugmentation within the active cell, and also confirmed that the *vcrA* gene could be used as a biomarker for the introduced culture (refer to Section 6.3.1).

Following b ioaugmentation and during i njection of one percent sodium lactate, considerable increases in numbers of DHC bacteria (ranging from > 10⁶ gene copies/mL to > 10⁹ gene copies/mL) and all three functional genes (tceA, bvcA, and vcrA) were observed in all wells in the upper portion of the active cell: AMW-1, AMW-2, and all zones of CMT wells (AMW-3, -4,





and -5). However, a decline in *DHC* populations (for example, 2 order of magnitude decrease at well AMW-4 (Z1) in June 2009) was soon observed within the upper half of the active cell. Low numbers of *DHC* bacteria were observed at well AMW-6 and AEW indicating that these wells were not being impacted. The decline in numbers of *DHC* bacteria combined with the COD and VOC data indicated that donor injection strategy needed to be optimized further in order sustain and advance reductive dechlorination within the active cell.

The inc rease in w eekly e lectron d onor in jection concentration f rom 1 percent to 3 percent sodium l actate so lution r esulted in i ncreases o r sus tenance of the numbers of *DHC* bacteria and/or the functional genes in the upper portion of the active cell: AMW-1, A MW-2, and all zones of CMT wells (AMW-3, -4 and -5 except well A MW-3 (Z3) and A MW-5 (Z1)). Low numbers of *DHC* bacteria observed at well A MW-6 and A EW indicated that these wells were still not being impacted by the remedy. An example of the *DHC* population trends is presented in Figure 5-9 for AMW-1.

Overall, the dechlorination t rends t hroughout t he de monstration, t he complete c onversion of TCE to ethene only a fter bi oaugmentation, and the DNA results indicate that bioaugmentation was successful for the upper half of the active cell. B ecause low levels of *DHC* were detected prior to bioaugmentation (specifically the bvc A and tceA genes), it is possible that some of the *DHC* present in the active cell w as f rom growth of indigenous bacteria. However, the bioaugmentation culture also contained these functional genes, so it is also possible that majority of DHC was from the added culture. While it is not clear exactly whether all of the *DHC* present in the active cell were from the added culture, the most important point is that the vcrA results indicate that *DHC* bacteria that were added during bi oaugmentation were transported to monitoring wells throughout the upper half of the active treatment cell. In addition, vertical distribution of *DHC* appeared to be effective, with complete dechlorination to ethene and *DHC* bacteria observed in all 3 zones of all CMT wells.

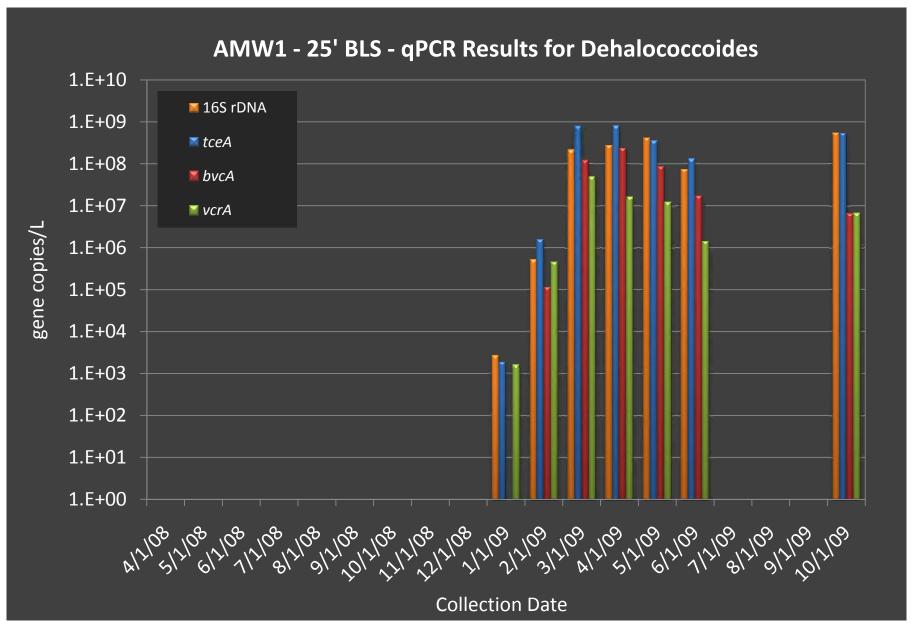
pH

While pH is not an indicator of reducing conditions or dechlorination, it can indicate whether aquifer geochemistry is favorable for biological activity. pH levels in the appropriate range (5.0 < pH < 9.0) provide verification that the progress of dechlorination (i.e., survival and performance of the *DHC* bacteria) within the pi lot test a rea is not be ing hindered (EPA 1998). Complete results of pH measurements are included in Table H-1.

pH levels were observed to decrease slightly following electron donor in jection, particularly following weekly electron injections of 1,000 gallons of 3 percent sodium lactate solution. But as of O ctober 2009, pH levels were greater than 5.2 within the active cell. This indicates that appropriate pH levels have been maintained within the active cell area, and that the aquifer has sufficient buffering capacity.

Alkalinity

Alkalinity is an indicator of microbial respiration because carbon dioxide production increases bicarbonate at typical groundwater pH levels. Alkalinity is also increased by the fermentation of injected electron donor, providing an indication of whether electron donor utilization is occurring in the treatment area. Complete results for alkalinity are presented in Table H-1.



During the baseline sampling event, alkalinity values ranging between 450 mg/L and 860 mg/L were observed. Alkalinity values were observed to increase at all the wells in the upper portion of the active cell. As of October 2009, alkalinity values ranging between 870 and 1,900 mg/L were observed. Alkalinity values mostly remained near background at wells AMW-6 and AEW. The elevated alkalinity values observed at the wells in the upper portion of the active cell indicate the presence of biological activity (specifically electron donor utilization) within the active cell.

5.8.2 Passive Cell

5.8.2.1 Electron Donor Distribution

Complete COD results for the passive cell are included in Table I-1, and figures showing the key COD c oncentration t rends a re pr esented in A ppendix I. The baseline sampling event (April 2008) showed COD concentrations ranging between 16 and 100 mg/L in the passive cell wells. During the pre-conditioning phase, the COD concentrations were observed to increase significantly only in the central and lower portion of the passive cell with concentrations above 1,000 mg/L observed at injection wells PIW-2 and PIW-3 and monitoring wells PMW-7 and PMW-8. Much lower COD values were observed in the upper portion of the passive cell with concentrations near or above 100 mg/L observed at injection well PIW-1 (September and November 2008) and wells PMW-2 and PMW-3 (Z1) (November 2008). COD concentrations at the upgradient well PMW-1 and all other monitoring wells and zones remained near background during the pre-conditioning phase.

At inj ection w ells P IW-2 a nd P IW-3 a nd m onitoring wells P MW-7 a nd P MW-8, COD concentrations continued to remain above 1,000 mg/L during the post-bioaugmentation phase, with the exception of well PIW-2 where concentrations decreased in October 2009 to 920 mg/L. At well PMW-6 COD concentrations increased above 1,000 mg/L but were observed to decrease in October 2009 (400 mg/L) whereas COD concentrations at well PMW-9 showed an increase in COD concentrations in the range of a few hundred mg/L. Thus, in the central and lower portion of the passive cell, the extent of electron donor distribution was expanded to include wells PMW-6 and PMW-9 during the post-bioaugmentation phase.

During the post-bioaugmentation phase, COD concentrations at upgradient well PMW-1 and injection well P IW-1 remained near baseline except in June and O ctober 2009 when small increases in concentration to about 60 mg/L and 45 mg/L (near 2X baseline), respectively, were observed. Samples were collected from two different depth intervals (25 feet and 35 feet bgs) at injection well PIW-1 in March 2009 to better understand the distribution of electron donor at this well. But very similar concentrations (28 mg/L at 25 feet bgs and 30 mg/L at 35 feet bgs) were observed, leaving the reason for the significant difference in COD concentrations be tween injection wells PIW-2 and PIW-3 and injection well PIW-1 unknown. At well PMW-2 the COD concentration remained near baseline, except in October 2009 (410 mg/L). COD concentrations at the CMT wells were observed to increase above 1,000 mg/L at wells PMW-3 (Z2 and Z3), PMW-4 (Z3), and PMW-5 (Z2) and in the range of a few hundred mg/L in all other zones. Zone 2 of the CMT wells appeared to be the most impacted with higher COD values followed by zone 3 and then zone 1. Thus in the upper portion of the passive cell, the extent of electron donor distribution was expanded to include a ll three C MT wells during the post-bioaugmentation phase.

In ge neral, COD c oncentrations increased and resulted in good donor distribution within the treatment z one of the passive cell extending approximately 22 f eet downgradient and 15 feet cross-gradient of the injection wells. Effects of donor injections were observed a few months earlier in the central and lower portion of the compared to the upper portion of the passive cell. Vertical distribution appeared effective, with the impact of donor observed more in zones 2 and 3 compared to zone 1 of CMT wells. Overall, the results suggest that electron donor can be easily injected using slug injections and effectively distributed to at least 22 f eet downgradient using the passive injection approach at the Site. Figure 5-10 shows an example COD concentration trend for PMW-7.

5.8.2.2 Redox conditions

ORP measurements and concentrations of inorganic electron acceptors (DO, nitrate, ferrous iron, sulfate and methane) for the passive cell are included in Table I-1 and figures showing the key redox conditions trends are presented in Appendix I.

ORP

ORP values at all the passive cell wells were mostly high and ranged from -60 mV to 484 mV during the baseline sampling event. Following electron donor injections, ORP values were observed to decrease as a function of electron donor distribution and reached the appropriate range at all the passive cell wells. As of October 2009 the ORP values were observed below 50 mV at all the passive cell wells, and were observed near or below -100 mV at all three injection wells and monitoring wells PMW-2, PMW-3 (Z3), PMW-6, PMW-7, PMW-8, and PMW-9. Overall, the ORP values at injection and monitoring wells were in the appropriate range for dechlorination and suggest establishment and sustenance of moderate to strongly reducing conditions within the passive cell.

Electron Acceptors and Reduced Products

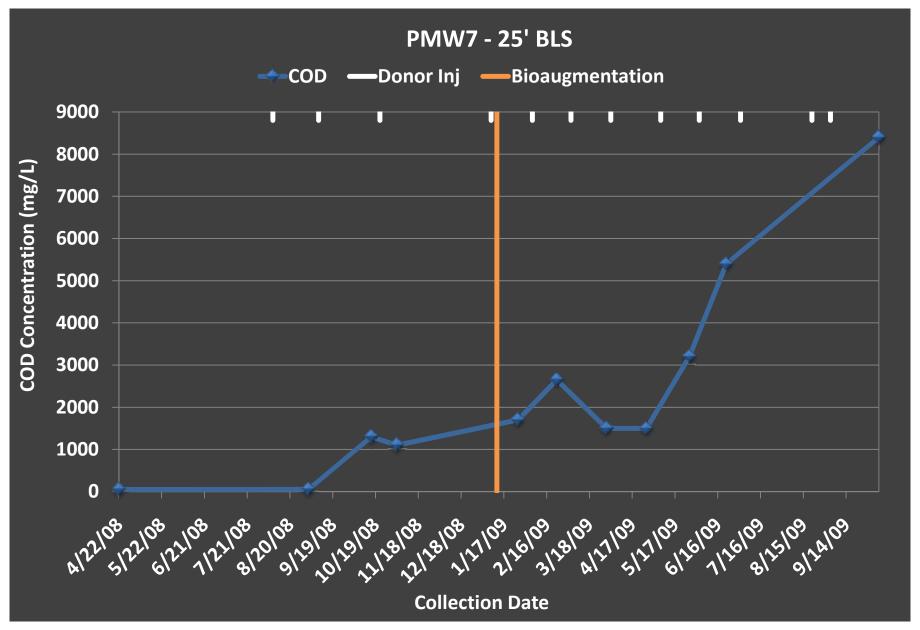
The c hanges in the concentrations of various electron a cceptors and their reduced products throughout the pilot test within the passive cell are discussed below.

Dissolved Oxygen

DO was not a r eliable r edox indicator during t his de monstration, likely because of equipment problems, and is not discussed here.

Nitrate Reduction

The baseline sampling event showed nitrate concentrations less than $1 \, \text{mg/L}$ at all the passive cell wells. The already low nitrate concentrations were reduced and observed mostly near or below detection limit at all the passive cell wells during the pilot test with the exception of the upgradient well PMW-1 which showed nitrate concentrations near baseline. Overall, the results indicate that nitrate reduction was not an important process within the passive cell due to the low initial nitrate concentrations.



Iron Reduction

The baseline sampling event showed ferrous iron concentrations below the detection limit at all the passive cell wells except wells PMW-4 (Z5) (0.53 mg/L), PMW-5 (Z3) (0.015 mg/L), a nd P MW-5 (Z4) (0.76 m g/L). E levated f errous i ron c oncentrations were observed at all the passive cell wells except the upgradient well PMW-1 and injection well PIW-1 following electron donor injection. Elevated ferrous iron concentrations were observed at zones 1 to 3 of the three CMT wells. Following the initial increase, ferrous iron concentrations were found to decrease at some of the wells over time (PIW-2, PMW-2, and PMW-6 through PMW-9). The blackish water observed during sampling event at these wells suggests the production of reduced iron sulfide minerals that explains the decrease in aque ous f errous iron. O verall, the r esults indicate that iron r educing conditions were established at most of the wells within the passive cell.

Sulfate Reduction

Baseline sulfate concentrations were above 1,000 mg/L in all the passive cell wells and ranged from 1,100 mg/L to 5,800 mg/L. Baseline sulfate concentrations were generally higher in zones 2 and 3 of CMT wells (3,900 mg/L to 5,800 mg/L) compared to zone 1 (2,000 m g/L). F ollowing e lectron donor injections, sulfate con centrations de creased significantly at injection wells PIW-2 and PIW-3 and monitoring wells PMW-2, PMW-7, and PMW-8 with removal ranging between 76 percent and 100 percent in October 2009. Significant decreases in sulfate concentrations were also observed at wells PMW-3 (Z2) and PMW-6 with removal of greater than 75 percent in June 2009, how ever; rebound in sulfate c oncentrations were observed at these w ells in October 20 09. Little s ulfate reduction was observed at wells PMW-4 (Z3) (30 percent removal) and PMW-5 (Z2) (19 percent removal) in October 2009. At well PMW-9 sulfate reduction was observed, but the sulfate concentrations varied. At upgradient well PMW-1 and other zones of the CMT wells, sulfate concentrations were observed near baseline. Overall, sulfate reduction was observed at most of the wells in the central and lower portion of the cell, and at wells PMW-2 and PMW-3 (Z2) in the upper portion of the passive cell.

Methanogenesis

During the baseline sampling event, low methane concentrations (less than 0.5 mg/L) were observed at all the passive cell wells with the exception of wells PIW-1 (2.3 mg/L) and PMW-9 (2.8 mg/L). A significant increase in methane concentration (greater than 0.5 mg/L) was observed as a result of lactate injections in all three injection wells and monitoring wells PMW-2, PMW-3 (Z2), PMW-4 (Z1), PMW-5 (Z1), and PMW-6 through PMW-9. However, with the lack of sulfate reduction observed at wells PMW-4 (Z1) and PMW-5 (Z1), methane might not have been generated locally at these wells but transported from upgradient. At all other wells and zones methane concentrations were observed near baseline.

At the injection wells, methane production was observed almost 9 months (January 2009) after beginning monthly donor injections, and at most of the monitoring wells, increases in methane c oncentration were observed about 13 months (May 2009) to 17 months (October 2009) a fter beginning monthly donor injections. At all the above mentioned wells except injection well PIW-3, methane production was observed in the presence of high sulfate concentrations indicating that all sulfate present does not need to be reduced

before methanogenic conditions are established. Overall, methanogenic conditions were observed at most of the wells in the central and lower portion of the passive treatment cell and at wells PIW-1 and PMW-2 in the upper portion of the passive cell.

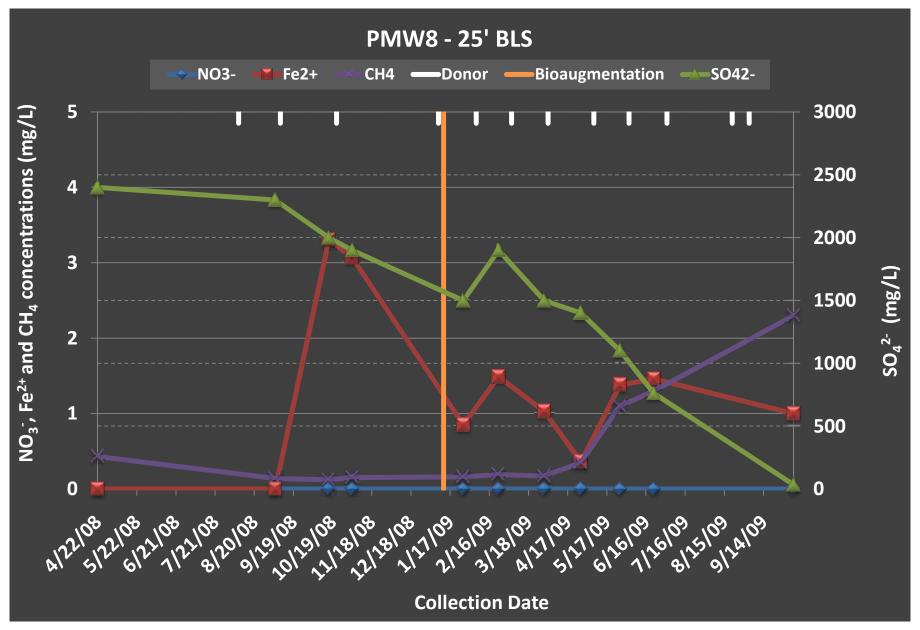
Redox Summary

Based on the results discussed in this section, it can be concluded that redox conditions shifted in a ccordance with the electron donor distribution, and as of O ctober 2009, moderate to strongly reducing conditions had been established within the passive cell. Methanogenic conditions appeared to be established in the upper portion (wells PIW-1 and P MW-2) and in the central and lower portions (wells P IW-2, P IW-3, P MW-6 through P MW-9) of the passive cell. I ron reducing conditions with little to no sulfate reduction appeared to be established within zones 1 to 3 of all three CMT wells except at well P MW-3 (Z2) where sulfate reducing conditions were achieved. It should be noted that unlike the active cell, no effects of donor injections were observed at the upgradient well, PMW-1, of the passive cell. Typical electron acceptor concentrations are presented in Figure 5-11 for PMW-8.

5.8.2.3 VOC Concentrations

VOC results for the passive cell are presented in Table I-1 and figures showing the key VOC concentration trends are presented in A ppendix I. Baseline groundwater contamination (April 2008) was characterized by high chloroethene concentrations primarily consisting of TCE at the passive cell wells. In the upper portion of the passive cell, TCE concentrations on the order of 1,100 μ g/L to 2,600 μ g/L were observed at wells PMW-1 and PMW-2, whereas injection well PIW-1 showed very low total CVOC concentration of 64 μ g/L. Zone 1 of all three CMT wells (PMW-3 to PMW-5) was characterized by very high chloroethene concentrations on the order of 50,000 to 60,000 μ g/L. Zones 2 and 3 of the three CMT wells consisted of concentrations of nearly 5,00 0 to 17,00 0 μ g/L. I n the central and lower portion of the passive cell, TCE concentrations were approximately 10,000 to 20,000 μ g/L (wells P IW-2, P IW-3, P MW-6, PMW-7, and PMW-8), with the exception of well PMW-9 where a TCE concentration of 840 μ g/L was observed. DCE concentrations were very low (below detection limit to 120 μ g/L) and VC and ethene were not detected in any passive cell well.

During the pre-conditioning phase (April to November 2008), TCE concentrations decreased by >97 percent at the injection wells PIW-2 and PIW-3 without a corresponding increase in the degradation products. At injection well PIW-1, chloroethene concentrations continued to remain low. During the first sampling event following electron do nor injections (September 2008) a slight increase in T CE and t otal chloroethene concentrations was noted at a few of the monitoring wells (PMW-2, PMW-3 (Z1), PMW-6, and PMW-9), including the upgradient well PMW-1, with the increase in total CVOCs ranging from 1.1 X to 2.4 X baseline. Following the initial increase, TCE concentrations decreased at all the wells within the passive cell except at well PMW-3 (Z1), and by November 2008 T CE decreases ranging between 11 percent and 53 percent were observed. However, no notable increase in any of the degradation products (DCE, VC, or ethene) was observed at these wells.



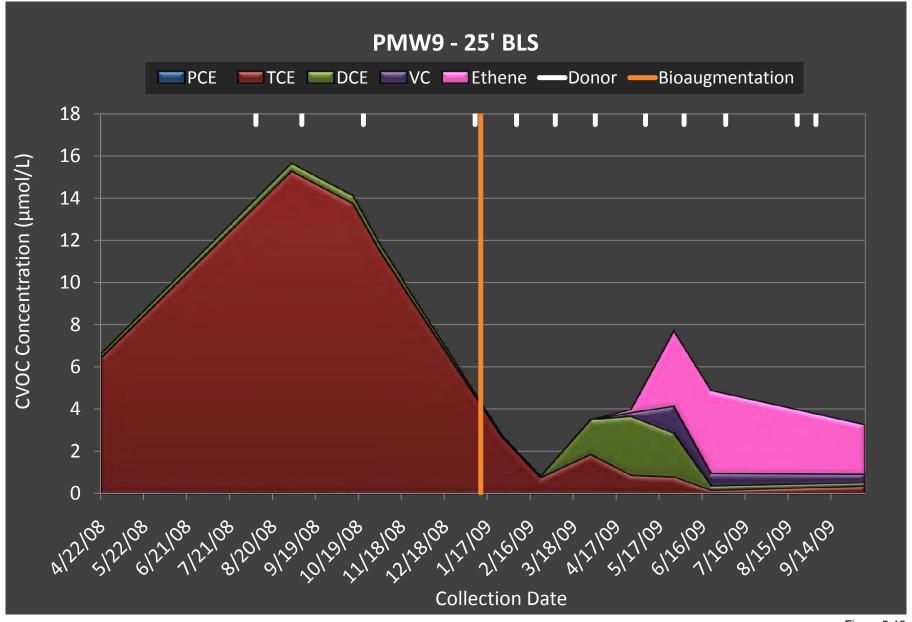
During the post-bioaugmentation phase, TCE and DCE were mostly removed, with VC and ethene observed for the first time at injection wells PIW-2 and -3 within two weeks after inoculation, in January 2009. Complete conversion of TCE to ethene was also observed at injection well PIW-1. As of October 2009, total CVOCS continue to remain low at all three injection wells. The concentrations at the upgradient well PMW-1 remained unchanged during the post-bioaugmentation phase.

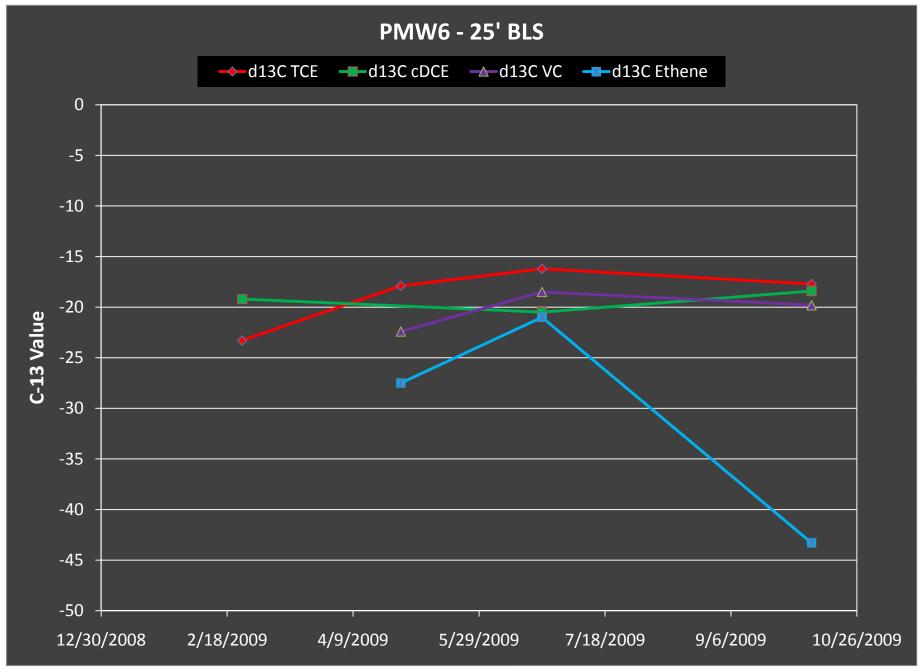
Little to no dechlorination was observed in the upper portion of the passive cell during the post-bioaugmentation phase. At well PMW-2 the TCE concentration decreased through March 2009 followed by r ebound i n T CE c oncentration ob served be tween A pril and J une 2009, but the concentration decreased aga in in October 2009 . The decrease i n TCE conc entration at well PMW-2 was not a companied by a corresponding increase in the degradation products. As of October 2009, C VOC concentrations r emain unchanged a t C MT well P MW-3 (Z1) and a decrease in TCE and increase in DCE with little VC production was observed at wells PMW-4 (Z1) and PMW-5 (Z1). As of October 2009, T CE r emoval greater than 44 percent and DCE concentrations greater than 10,000 μ g/L were observed at wells PMW-4 (Z1) and PMW-5 (Z1), and a VC concentration of 490 μ g/L was observed at well PMW-5 (Z1). At well PMW-5 (Z2) some DCE production was observed in October 2009 (220 μ g/L). At all other zones of the CMT wells the total C VOC concentrations varied but primarily consisted of T CE and no biodegradation was observed.

Complete r eductive de chlorination of TCE to e thene was observed in the c entral and lower portion of the passive cell as shown by the VOC results at wells PMW-6 through PMW-9. In October 2009 biodegradation accounted for reduction of total CVOC concentrations by greater than 92 percent at wells PMW-7 through PMW-9 and nearly 72 percent at well PMW-6 compared to C VOC concentrations observed in November 2008 , immediately be fore bioaugmentation. Ethene production was observed as high as 410 $\mu g/L$ at wells PMW-6 through PMW-9.

In summary, the VOC data indicate that complete reductive dechlorination was achieved in the central a nd low er por tions (a round injection wells P IW-2 a nd P IW-3) of the passive cell. However, complete reductive dechlorination was not observed in the upper portion of the passive cell (a round injection well P IW-1) a lthough effective electron donor distribution and redox conditions appropriate for de chlorination were a chieved. CVOC molar concentrations a represented in Figure 5-12 for PMW-9.

CSIA data for the passive cell generally were consistent with the CVOC data, in that they suggested degradation to VC and ethene was occurring near PIW-2 and PIW-3, but not in the vicinity of PIW-1. An example CSIA chart is included as Figure 5-13 for PMW-6. This chart shows that TCE, c-DCE, and V C be come he avier during the course of the demonstration, indicating degradation is occurring. Ethene was much "lighter" during the last sampling event compared to the previous two. The rest of the active cell CSIA data are included in Appendix H.





5.8.2.3 Biological Indicators

Changes in numbers of dechlorinating bacteria and values of pH and alkalinity are discussed below.

Dechlorinating Bacteria

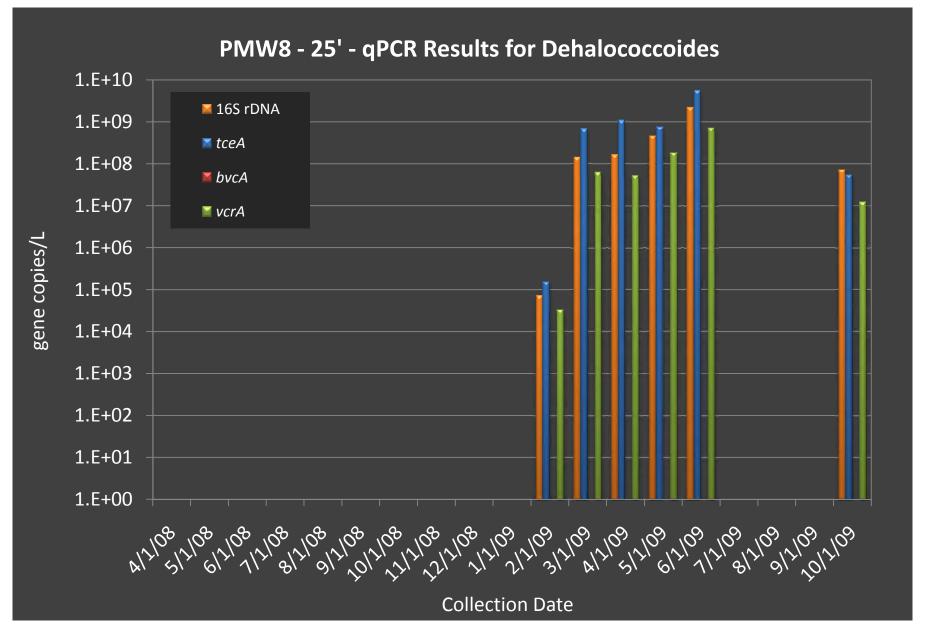
DNA results for the passive cell are provided in Table I-1 and figures showing the key DNA number trends are presented in Appendix I. During the baseline sampling event, DHC bacteria numbers were below detection limit at all the wells except well PMW-3 (Z1) which showed low numbers of DHC bacteria along with functional gene tceA (>10³ gene copies/L).

During the pre-conditioning phase low numbers of DHC (16S r RNA and/or functional genes tceA and bvcA) were observed at a few wells (PMW-1, PMW-6, and PMW-7) ranging between $>10^1$ gene copies/L and $>10^3$ gene copies/L in September 2008. The functional gene vcrA was not detected in any well. The presence of dechlorinating bacteria in such low numbers at only a few wells, along with the absence of degradation products within the passive cell confirmed the need for bioaugmentation for reductive dechlorination to progress within the passive cell.

During the post-bioaugmentation p hase, DHC bacteria and functional gene (tceA and vcrA) numbers increased immediately (within 2 weeks of inoculation) at all three injection wells on the order of >10⁶ gene copies/L. As of October 2009, the numbers were observed to decrease by one to two or ders of magnitude at the injection wells, suggesting that in the a bsence of high chloroethene concentrations, the DHC bacteria number might be decreasing. The functional gene bvcA was only detected in low numbers at well PIW-1 (May 2008).

In the upper portion of the passive cell, low detections of DHC bacteria and functional genes ranging be tween $> 10^1$ gene copies/L and $> 10^3$ gene copies/L were observed at the upgradient well PMW-1 and monitoring well PMW-2. In zone 1 of the CMT wells DHC bacteria and tceA gene numbers increased and were detected on the order of $> 10^6$ gene copies/L and vcrA gene numbers were detected in the order of $> 10^5$ gene copies/L and were sustained as of O ctober 2009. Z ones 2 and 3 of the CMT wells except well PMW-5 (Z3) also showed increases in numbers of DHC bacteria and functional genes tceA and vcrA, but the numbers were lower compared to Zone 1 of the CMT wells. In the central and lower portion of the passive cell (wells PMW-6 through PMW-9) DHC bacteria and functional gene (tceA and vcrA) numbers increased on the order of $> 10^6$ gene copies/L and were sustained as of October 2009.

Overall, the D NA re sults c ombined with the V OC datas uggest that bi oaugmentation was successful; i.e., dechlorinating bacteria were successfully distributed and maintained, and complete reductive dechlorination was achieved in the central and lower portion of the passive cell. This is shown in Figure 5-14 for PMW-8. However, the DNA data combined with the COD datas uggests that electron donor was distributed at higher concentrations in the upper zones (zones 2 and 3) of the CMT wells, whereas the bioaugmented culture was distributed (or at least survived) to a greater degree in zone 1 of the CMT wells. This discrepancy in distribution of electron do nor and bi oaugmented culture might be the reason that I imited to no progress in reductive dechlorination was observed in the upper portion of the passive cell. The cause of this difference is unclear.



pH

Results of pH measurements are included in Table I-1. Significant pH impacts were not observed in any of the passive cell wells, and remained in the appropriate range (5.0<pH<9.0) during the pilot test, indicating that the aquifer has sufficient buffering capacity.

Alkalinity

Results for alkalinity are presented in Table I-1. During the baseline sampling event, alkalinity values ranging between 530 mg/L to 1,100 mg/L were observed at all the passive cell wells with the exception of well PMW-1 (1,400 mg/L), PIW-1 (1,900 mg/L), and zone 1 of the CMT wells (220 mg/L to 360 mg/L). Alkalinity values were observed to increase at all the wells within the passive cell except the upgradient well PMW-1 and injection well PIW-1. As of October 2009, the increased alkalinity values ranged from 1,400 to 5,200 mg/L at most of the wells with the exception of z one 1 of the CMT wells. In z one 1 of the CMT wells the increased alkalinity values ranged from 630 mg/L to 650 mg/L in October 2009 except well PMW-3 (Z1), where the alkalinity value peaked in February 2009 (680 mg/L), but was reduced to baseline by O ctober 2009. The el evated alkalinity values observed at most of the wells within the passive cell compared to the near background values observed at the upgradient well PMW-1 and the injection well PIW-1 indicate significant electron donor utilization within the passive cell.

6.0 PERFORMANCE ASSESSMENT

In the previous section, the test design and results were presented, including the data collected for be nch s cale t esting, b ioaugmentation c ulture selection, pre-conditioning, a nd bioaugmentation. In this section, the implications of those data are discussed in the context of the project performance objectives.

6.1 PHASE 1 PERFORMANCE O BJECTIVES – BENCH'S CALE T ESTING AND BIOAUGMENTATION CULTURE SELECTION

The purpose of the Phase 1 of the ER-0513 project was to conduct laboratory studies to confirm that dechlorination could be stimulated in the high sulfate environment present at NAVWPNSTA Site 70, and to select a bioaugmentation culture for the demonstration. These objectives were described in Section 3. The sections below assess performance of the demonstration activities in achieving these objectives.

6.1.1 Demonstration of Dechlorination using Site Groundwater

Section 5.4 and Appendix D present the results of the microcosm studies conducted as a part of Phase 1 de monstration a ctivities. Two s ets of microcosms were run, one with groundwater collected from existing well EW-70-01, and one with groundwater collected from MW-70-27. The success criterion for this performance objective was production of ethene at concentrations at least 2X detection, and reduction of TCE by at least 95% in the microcosms.

The r esults of the lab s tudy s howed that T CE was completely r emoved under a ll c onditions investigated, which exceeded the goal of achieving at least 95% reduction of TCE. Microcosms from EW-70-01 s howed that all C VOCs were converted to ethene with complete reduction of 1,650 mg/L sulfate. Microcosms from MW-70-27 showed that dechlorination of TCE to VC and ethene was a chieved in less than four months using one of the two cultures, while TCE was converted to cis-DCE and VC using the other culture tested. These results show that three of the four conditions tested met the criteria of production of ethene of at least twice the detection limit. Based on these results, this performance objective was met.

6.1.2 Select Bioaugmentation Culture with Reliable Biomarker

Section 5.4 and Appendix E present the results of the DNA studies that were conducted as a part of P hase 1 demonstration a ctivities. D uring t he D NA s tudy, s everal m ethods w ere us ed to evaluate *DHC*, including quantitative PCR analysis and clone library analysis to evaluate various genes including the 16S rRNA gene, and functional reductase genes *vcrA*, *bvcA* and *tceA*. These analyses were performed for the 16S rRNA gene of NAVWPNSTA Site 70 indigenous *DHC* and three bioaugmentation cultures. The DNA study also included *vcrA* gene sequence analysis of the SDC-9TM and KB-1TM bioaugmentation cultures. The success criterion for this objective was identification of a biomarker that is present in bioaugmentation culture(s) but not in native strains of *DHC*.

The results from the DNA study showed that the functional gene *vcrA* was not present at the site, but w as pr esent i n bot h t he SDC-9TM a nd K B-1TM commercially available bioaugmentation culture. In addition, DNA sequence information was obtained for the *vcrA* gene in both cultures

for the purpose of designing a new biomarker in the event that *vcrA* was detected at the end of the pre-conditioning phase. B ased on the fact that the SDC-9TM culture had been demonstrated to perform better in the presence of co-contaminants detected at Site 70 compared to KB-1TM (i.e. c hloroform), t he S DC-9TM culture w as s elected f or t he de monstration. T herefore, t his performance objective was met.

6.2 PHASE 2 PERFORMANCE O BJECTIVES – BASELINE CO NDITIONS A ND PRE-CONDITIONING

The pur pose of P hase 2 of the ER-0513 project was to determine groundwater hydraulic conditions and baseline contaminant distribution, *DHC* distribution, and geochemical concentrations prior to beginning the biostimulation and bioaugmentation in each treatment cell. Performance objectives were established related to demonstrating that the treatment cell layout was such that meaningful results could be obtained during the timeframe of the project, and related to establishing a ppropriate conditions prior to conducting bioaugmentation. These objectives are discussed further below.

6.2.1 Treatment Cell Construction and Residence Time

Due to the slow ambient groundwater velocity in the Site 70 source area, ESTCP was concerned that e ffects of e lectron donor i njections a nd bi oaugmentation w ould not be observed a t monitoring wells within the timeframe of the demonstration, at least for the passive cell. In addition, historical data that were available for the site did not provide conclusive information regarding groundwater flow magnitude and direction in the Upper Fines unit on the scale of the source area. In order to verify that meaningful results could be obtained using the proposed treatment cell layout, a tracer test was conducted to verify the groundwater hydraulic conditions in the treatment cells. Data collected in support of this objective included multiple samples collected from active cell and passive cell monitoring wells and analyzed for iodide tracer.

The success criterion for this objective was to construct the treatment cells such that travel time from injection wells to monitoring wells was 6 months or less. In the active cell, arrival of tracer occurred within 6 weeks of injection for AMW-1 through AMW-5, including at the two deepest zones of all of the CMT wells. Tracer was not observed at well AMW-6 (75 ft from injection wells) during the time it was sampled (this well also turned out to be too far from the injection wells for any effects of bioaugmentation or electron donor injection to be observed).

For the passive cell, a tracer test was conducted in order to confirm the results of the active cell tracer test. Because this test was merely to confirm approximate travel times predicted from the active c ell tracer test, the f requent s ampling that w ould be re quired to quantify hydraulic parameters was not performed. Rather, samples for tracer were collected 3 weeks and 5 weeks following i njection, a nd t hen dur ing pl anned p re-conditioning s ampling e vents, w hich w ere conducted monthly from S eptember t hrough November. 1,000 ga llons of iodi de tracer w ere injected into PIW-1 on 8/7/08 at a concentration of approximately 13,000 mg/L as iodide. Tracer arrival was observed within 4 weeks at the deepest interval in PMW-4 (center CMT well located 17 ft downgradient), at the deepest zones of PMW-3 and -5, and at cross-gradient well PMW-2 within 7 weeks. By the end of the passive cell tracer monitoring period of 3.5 months, tracer was measured at PMW-2 through PMW-5, including at the two deepest zones of all CMT wells.

Overall, the results of the tracer test showed arrival in some wells in less than one month in both treatment cells, and subsequent sampling for tracer indicated that travel times to all monitoring wells that were installed near the tracer injection wells were less than 4 months. These tracer results show that meaningful data would be obtained within the 12 month planned duration of the demonstration. The groundwater velocities that were predicted for the passive cell based on the active cell tracer test were achi eved during the demonstration. Therefore, this performance objective was met. In fact, as discussed below in Section 6.3, results were obtained faster than originally planned, such that the demonstration objectives were all met within a 9 month period.

6.2.2 Pre-Conditioning Results

Sampling was conducted to assess baseline conditions including contaminant and degradation product concentrations, redox parameters, and biological activity indicators (refer to Section 5.2 for complete baseline sampling results). In summary, the baseline results confirmed the predemonstration conditions in the source area; namely, that conditions were anaerobic but mildly reducing, with very high sulfate concentrations and very limited dechlorination to cis-DCE in some a reas. Because these conditions were not ideal for bioaugmentation, electron donor additions were performed to "pre-condition" the aquifer to reduce sulfate concentrations and to drive redox conditions more strongly reducing.

The suc cess cr iterion for this objective was to create at least sulfate-reducing conditions at monitoring wells nearest to injection locations, such that the bioaugmentation culture would have a favorable environment following inoculation. Results were presented in Section 5.7 and in a memo to ESTCP dated 12-28-2008 (see Appendix B). After three lactate injections into the active cel l, r esults i ndicated that appr opriate conditions were a chieved for suc cessful bioaugmentation, particularly in wells near the reinjection locations. Ferrous iron increased to above 0.5 m g/L in all wells except A MW-6 and upgradient well A MW-1. A lso, sulfate concentrations decreased more than 10% except in AMW-6 and the extraction wells. While COD concentrations did not increase above 60 mg/L in any active cell well, the significantly increased cis-DCE concentration at AMW-2 and other wells indicated that partial dechlorination was already occurring near the injection wells.

After thr ee pa ssive c ell in jections, results indicated that c onditions w ere be coming more reducing, with the most positive results observed near the injection wells. At these wells, ferrous iron increased to above 0.5 m g/L and sulfate decreased more than 10% except in PMW-2 and PMW-6. COD increased significantly at wells near the injection points also, and significant COD still remained at two of the three injection wells.

Another key result from the post-preconditioning sampling event was that the *vcrA* functional gene was not detected at any location in either the active or passive cell, despite the fact that low concentrations of *DHC* did appear following the biostimulation phase. These results confirmed that the *vcrA* gene could be used to track the bioaugmentation culture.

Overall, t he pos t-preconditioning r esults indicated t hat s ufficient e lectron donor was be ing supplied f or bi oaugmentation, and that r edox conditions nearest the injection l ocations were

sulfate reducing to methanogenic in both treatment cells following the pre-conditioning phase. Therefore, this performance objective was met.

6.3 PHASE 3 PERFORMANCE OBJECTIVES – BIOAUGMENTATION RESULTS

The purpose of Phase 3 of the ER-0513 project was to demonstrate full-scale bioaugmentation and dechlorination using both the active and passive approaches. Phase 3 of the ER-0513 project began with inoculation of both treatment cells. Performance objectives were established related to collection of data that would allow for quantification of bacterial distribution and growth, and assessment of the extent of dechlorination. These objectives are discussed further below.

6.3.1 Bacterial Growth and Distribution

The first Phase 3 objective was to assess and quantify bacterial growth and distribution in both treatment cells. Bacterial distribution was assessed by analyzing the first arrival of *DHC* bacteria (as measured by qPCR analysis) at a giving monitoring location following i noculation. This travel time was then compared to the travel time for ambient groundwater, as determined from the tracer test. Bacterial growth was then assessed by analyzing the increase of *DHC* and functional gene counts at a given location once first arrival had been established. The success criterion for this objective was to collect data that allow for quantitative assessment of tracer and bacterial transport time, and growth of bacteria over time. No specific criteria were set in terms of bacterial transport times or cell counts. Therefore, this performance objective was met. The subsections be low quantify the arrival of tracer and bi oaugmentation culture based on *vcrA* analysis.

In general, the distribution of *DHC* bacteria was effective in both the active and passive cells. As shown in Figure 6-1, *DHC* concentrations exceeded 10⁸ cells/L in both cells based on analysis of the 16S rRNA gene. In the active cell, the high *DHC* concentrations extended greater than about 30 ft downgradient from the injections wells. In the passive cell, the high concentrations were distributed t hroughout t he dow ngradient tw o-thirds of the cell. Perhaps more im portantly, concentrations of the *vcrA* gene, while som ewhat lower than 16S r RNA gene measurements, indicated that the high *DHC* concentrations were representative of the bioaugmentation culture (Figure 6-2). The next two subsections discuss the speed at which the bacteria were distributed relative to groundwater velocity in the two cells.

6.3.1.1 Active Cell Distribution

Table 6-1 shows details for tracer arrival and first detection of *DHC* for the active treatment cell. Data are presented only for wells that were sampled monthly for *DHC* bacteria. While tracer samples were collected more frequently for the active cell CMT wells, *DHC* data were collected monthly from the deepest CMT port (Zone 1), and approximately quarterly from all other CMT ports. Because of this, the analysis of tracer and *DHC* arrival was only performed for Zone 1 of the C MT wells. A lso, t racer d ata were not collected f requently enough a t upgr adient well AMW-1 to perform the analysis. For the active cell, tracer injection was performed on 4/10/08, and bioaugmentation was performed on 1/12/09.

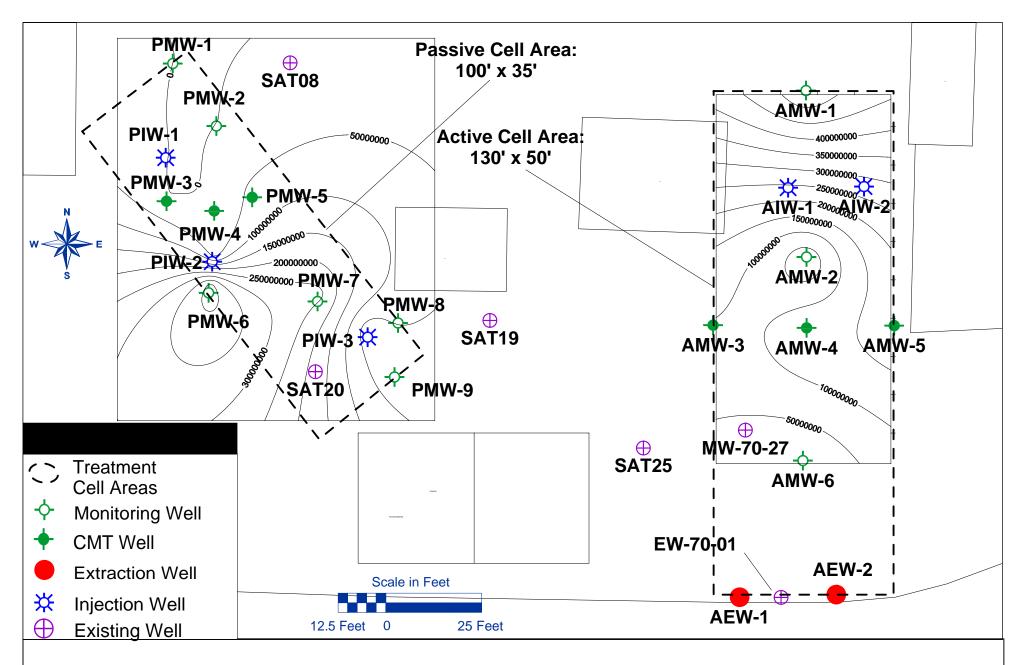


FIGURE 6-1 16S rDNA RESULTS, OCTOBER 2009 ER-0513 FINAL REPORT SEAL BEACH NAVAL WEAPONS STATION, SEAL BEACH, CA

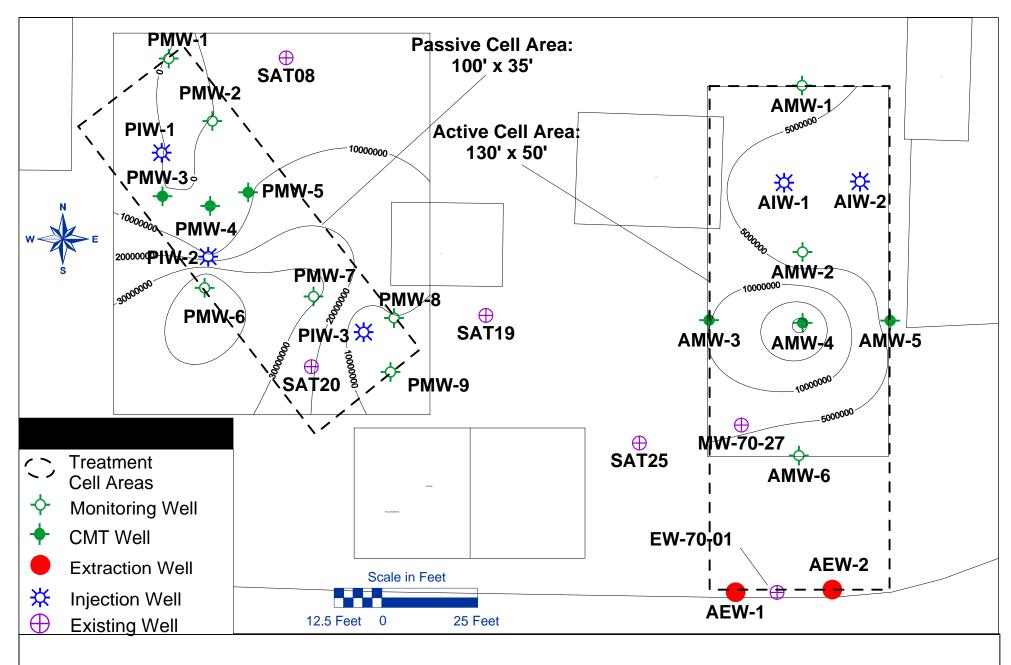


FIGURE 6-2 vcrA GENE rDNA RESULTS, OCTOBER 2009 ER-0513 FINAL REPORT SEAL BEACH NAVAL WEAPONS STATION, SEAL BEACH, CA

Table 6-1. Active Cell Tracer Test Data

Well	Distances from Nearest Injection Well (ft)	Tracer First Arrival Date	Travel Time based on tracer first arrival (days)	Velocity based on tracer first arrival (ft/day)	Tracer Peak Arrival Date	Travel Time based on tracer peak arrival (days)	Velocity based on tracer peak arrival (ft/day)	Date of First Arrival of Bacteria	First Arrival of Bacteria (days)	"Velocity" of Bacteria (ft/d)	Retardation of Bacteria - Based on tracer peak arrival	Retardation of Bacteria - Based on tracer first arrival
AMW-2	18.0	4/16/2008	6	3.00	4/24/2008	14	1.29	1/29/2009	17	1.06	1.21	2.83
AMW-3 Z1	36.0	5/19/2008	39	0.92	6/2/2008	53	0.68	2/24/2009	43	0.84	0.81	1.10
AMW-4 Z1	36.0	4/25/2008	15	2.40	5/9/2008	29	1.24	1/29/2009	17	2.12	0.59	1.13
AMW-5 Z1	36.0	5/19/2008	39	0.92	6/2/2008	53	0.68	2/24/2009	43	0.84	0.81	1.10
			Average	1.81		Average	0.97		·	Average	0.86	1.54
Distances from	Distances from AMW-1/2									Std Dev	0.26	0.86

Tracer Injection performed on 4/10/08

Bioaugmentation performed on 1/12/09

The distances from injection wells presented in Table 6-1 are south from AIW-1 and AIW-2 (refer to Figure 5-1 for well locations). No corrections in distance are made for the fact that AMW-3 and AMW-4 are slightly off the axis of the treatment cell. The first tracer arrival was the first measured i odide concentration a bove 4 mg/L, which was the highest i odide reading during baseline sampling (before tracer was injected). The peak tracer arrival was the date of the maximum concentration of tracer at those locations where it was detected.

For *DHC* data, the date of first arrival represents the first detection of *DHC* as indicated by a *vcrA* concentration that was greater than the reporting limit. *vcrA* was used rather than the 16s rRNA because *vcrA* was determined to be the best biomarker for the bioaugmentation culture; based on the results of the pre-conditioning phase, it was possible that *DHC* increases as measured by the 16s rRNA results could occur from biostimulation alone.

The retardation of bacteria was initially calculated based on the velocity derived from the peak tracer arrival, and the first arrival of *DHC* bacteria. The peak tracer arrival was used because it represents t he a verage linear gr oundwater velocity (i.e., Darcy velocity divided by effective porosity). However, from Table 6-1, *DHC* arrival was faster than peak tracer arrival for 3 of the 4 wells for which the analysis was performed. The average retardation using this method was 0.86, with a standard deviation of 0.26.

The travel time of first arrival of *DHC* was also compared to the first arrival of tracer. From Table 6-1, the average retardation of *DHC* using this method was 1.54, with a standard deviation of 0.86. The arrival of *DHC* was nearly 3 times longer than first tracer arrival for well AMW-2, but was only a few days longer for all 3 CMT wells.

The apparently very low retardation of *DHC* as shown by the CMT well results could be a result of several factors other than truly having such low retardation. The first possible factor was sampling methods. The CMT is able to target discrete zones and could detect arrival of DHC faster than a conventional well (AMW-2), which would be subject to dilution. However, such dilution would also have a ffected the tracer's ampling, and therefore would not cause "false negatives" for *DHC* but not for tracer. Another possible reason for the minimal *DHC* retardation at the C MT wells relative to AMW-2 is that actual growth of DHC bacteria was a more significant factor in distribution to the CMT wells compared to AMW-2. This could explain the minimal retardation seen at AMW-3 and AMW-5, because DHC arrival was detected at 43 days, which is sufficient time for DHC to grow in situ. However, DHC was detected at both AMW-2 and AMW-4 at the first sampling event (17 days following bioaugmentation), during which time significant growth of DHC is unlikely. The final factor that could have contributed to the minimal DHC retardation is the sampling frequency for tracer didn't a llow for a precise assessment of first arrival, and that tracer actually arrived sooner than it was detected. However, all three CMT wells were sampled 3 days before first arrival of tracer occurred, and in all cases the iodide concentration was less than the baseline concentrations. Because of this, the earliest that tracer could have arrived at these wells was two days earlier, which would have only had a minimal impact on the *DHC* retardation factor.

Overall, the re sults from the a ctive c ell indicate that m inimal re tardation of *DHC* bacteria occurred compared to transport of conservative tracer. In terms of actual velocity, based on the

distance from injection to monitoring wells, the average *DHC* "velocity" was 1.21 ft/day. During the active cell tracer test, the groundwater velocity was estimated to be 1-2 ft/day. Based on the actual tracer arrival, groundwater velocity using first arrival of tracer was 1.81 ft/d, while peak arrival yields a velocity of 0.97 ft/d. This implies that the *DHC* "velocity" was approximately the same as the actual groundwater velocity. Work published previous to this demonstration, suggested that retardation factors of *DHC* under forced advection could be as high as 60-200 (Major et al, 2002). However, groundwater velocity for that study was much higher under the forced gradient (greater than 25 ft/d) than the current demonstration, which suggests that the increased retardation occurs only at high groundwater velocities (at least greater than 2 ft/d).

6.3.1.2 Passive Cell Distribution

Table 6-2 shows details for tracer arrival and first detection of *DHC* for the passive treatment cell. Data are presented only for wells that were sampled monthly for *DHC* bacteria. While tracer samples were collected more frequently for the passive cell CMT wells, *DHC* data were collected monthly from the deepest CMT port (Zone 1), and approximately quarterly from all other CMT ports. Because of this, the analysis of tracer and *DHC* arrival was only performed for Zone 1 of the CMT wells. For the passive cell, tracer injection was performed on 8/7/08, and bioaugmentation was performed on 1/13/09.

The distances from injection wells presented in Table 6-2 are relative to the nearest injection well. For PMW-1 through PMW-5, the nearest injection well was PIW-1. The direction of groundwater flow in the passive cell during operation of the active cell is to the southwest; therefore well PMW-4 is located along the axis of the treatment cell, while PMW-3 and PMW-5 are slightly off axis. PMW-2 is a crossgradient well, and PMW-1 is an upgradient well. For wells PMW-6 and PMW-7, the distances in Table 6-2 are from PIW-2, and for PMW-8 and PMW-9, the distances are from PIW-3.

As for the active cell, the first tracer arrival was the first measured iodide concentration above 4 mg/L. The peak tracer arrival was the date of the maximum concentration of tracer at those locations where it was detected. The tracer injection was designed to achieve a radius of influence of 5 f eet from the injection well. Because of this, the velocity and travel time calculations in Table 6-2 assume that tracer particles had traveled 5 ft of the distance between PIW-1 and the monitoring wells at "time zero," when ambient groundwater flow was assumed to be the dominant transport mechanism.

The retardation of bacteria was calculated first based on the velocity derived from the peak tracer arrival, and the first a rrival of *DHC* bacteria. The pe ak tracer ar rival w as us ed because it represents the a verage linear groundwater velocity (i.e., Darcy velocity divided by effective porosity). Because tracer injection was not performed in wells P IW-2 and P IW-3, no quantitative analysis of tracer first and peak arrival could be performed. Also, of the monitoring wells installed near these injection wells, only PMW-9 is directly downgradient of an injection well. Therefore, while travel times, distances, and time of first arrival are presented in Table 6-2 for all passive cell monitoring wells, only wells PMW-3, PMW-4, and PMW-5 are included in the retardation calculations.

Table 6-2. Passive Cell Tracer Test Data

Well	Distances from Nearest Injection Well (ft)	Tracer First Arrival Date	Travel Time based on tracer first arrival (days)	Velocity based on tracer first arrival (ft/day)	Tracer Peak Arrival Date	Travel Time based on tracer peak arrival (days)	Velocity based on tracer peak arrival (ft/day)	Date of First Arrival of Bacteri	First Arrival of Bacteri a (days)	"Velocity" of Bacteria (ft/d)	Retardation of Bacteria - Based on tracer peak arrival	Retardation of Bacteria - Based on tracer first arrival
	Distances from PIW-1											
PMW-2	15.6	9/5/2008	29	0.36	9/23/200	47	0.22	3/30/20 09	76	0.20	1.10	1.78
PMW-3 Z1	11.6	8/21/200 8	14	0.47	11/3/200 8	88	0.07	1/27/20 09	14	0.83	0.09	0.57
PMW-4 Z1	19.0	9/2/2008	26	0.54	10/17/20 08	71	0.20	1/27/20 09	14	1.36	0.15	0.40
PMW-5 Z1	24.9	9/23/200 8	47	0.42	10/17/20 08	71	0.28	1/27/20 09	14	1.78	0.16	0.24
										Average	0.13	0.40
										StDev	0.04	0.16
					Dis	tances from	PIW-2					
PMW-6	8.1	NA	7	0.44	NA	14	0.22	2/23/20 09	41	0.20	1.11	2.22
PMW-7	29.3	NA	55	0.44	NA	110	0.22	1/27/20 09	14	2.09	0.11	0.21
	Distances from PIW-3											
PMW-8	8.7	NA	8	0.44	NA	17	0.22	2/23/20 09	41	0.21	1.03	2.07
PMW-9	12.5	NA	17	0.44	NA	34	0.22	2/23/20 09	41	0.31	0.72	1.44

0.87

Tracer injection performed on 8/7/08

The injection was designed with a 5 ft ROI, so it is assumed that the tracer traveled 5 ft at time 0.

Red = not included in the average or standard deviation calculations

From Table 6-2, the retardation of *DHC* compared to peak tracer arrival was significantly less than 1, as the average was 0.13 with a standard deviation of 0.04. Even when compared to tracer first arrival, the retardation of *DHC* was 0.40 with a standard deviation of 0.16. This implies that the first arrival of bacteria was faster than the first arrival of tracer at all three CMT wells.

For the active cell, the same factors that could have contributed to similar observations identified for t he act ive cell a bove (sampling m ethods, *DHC* growth, a nd s ampling f requency) w ere considered for t he passive cell. For the passive cell, all three wells are CMT wells, so no difference existed in sampling methods. Also, *DHC* was detected 2 weeks following inoculation, during which time significant growth of *DHC* is unlikely. In terms of sampling f requency, sampling was performed 2 weeks following both tracer injection and bioaugmentation, and first arrival both of tracer and *DHC* had already occurred at PMW-3. It is possible that tracer arrived several days sooner at this well, and that the actual retardation factor for *DHC* was greater than 1. H owever, P MW-3 and P MW-5 were sampled 2 weeks following t racer injection and bioaugmentation, and *DHC* was detected at significant concentrations (10⁵ to 10⁶ cells/L), while tracer was not detected above background levels.

While the other monitoring wells were not included in the retardation analysis, it is interesting to note that arrival of *DHC* bacteria occurred at all wells within 41 days of inoculation except for PMW-2, which had *DHC* at 76 days. This represents an average "velocity" of 0.87 ft/d, which includes *DHC* transport between injection and monitoring wells off the axis of the treatment cell, and even crossgradient in some cases. For purposes of comparison, the average groundwater velocity that was calculated for the passive cell based on applying hydraulic parameters from the active cell tracer test to the passive cell was 0.25 ft/d. Based on the passive cell tracer test, the average first arrival of tracer correlates to a velocity of 0.44 ft/d, while peak arrival yields a velocity of 0.22 ft/d. This implies that the "velocity" of *DHC* is 2 to 4 times faster than that of conservative tracer. Perhaps the most important result is that bacterial transport in the passive cell was extremely rapid, with *DHC* colonization occurring at distances of up to 30 ft from injection points within two to five weeks from inoculation.

6.3.1.3 Bacterial Transport Summary

The tracer and *DHC* data indicate that bacterial transport was not significantly retarded compared to groundwater flow in either the active or plassive cells. In fact, many of the calculated retardation factors were less than one, especially in the passive cell. The average retardation under passive conditions was 0.13 to 0.40 depending on whether peak or first arrival tracer data are used, and for the active cell the averages were 0.86 to 1.54. These results suggest that *DHC* were transported more rapidly relative to groundwater flow under passive conditions compared to active recirculation. The groundwater velocity in the active cell was 1 to 1.8 ft/day, and for the passive cell it was 0.22 to 0.44 ft/d. This is a contrast of approximately a factor of 5, which represents a typical enhancement in flow that might be expected due to recirculation.

Another interesting observation was the fact that bacterial transport rate and extent was relatively independent of groundwater flow direction, especially in the passive cell. The off-axis CMT wells in the active cell had *DHC* velocities that were approximately half of what was observed at wells on the axis of the treatment cell. In the passive cell, one port in PMW-3 had a *DHC* velocity that was almost the same as the average *DHC* velocity for the passive cell, and PMW-5

had a ve locity that w as nearly twice the average. In a ddition, crossgradient wells such as PMW-2, PMW-7 and PMW-8 all showed DHC velocities similar to that of groundwater (0.2 to 0.3 ft/d). Therefore, DHC transport was not only less retarded in the direction of groundwater flow at slower groundwater velocities, it also occurred more rapidly in cross-gradient directions relative to the groundwater velocity.

Overall, the *DHC* results f rom bot h t reatment cel ls are consistent with the comparison of NAVWPNSTA Site 40 and Kelly Air Force Base in Section 2, and support the hypothesis that *DHC* bacterial transport is affected by groundwater velocity. Specifically, data from the passive cell suggest that bacterial transport was potentially faster than ambient groundwater velocity, while data from the active cell's howed DHC transport was approximately the same as groundwater velocity. Work published previous to this demonstration suggested that retardation factors of DHC under forced advection could be as high as 60-200 (Major et al, 2002). However, groundwater velocity for that study was much higher under the forced gradient (greater than 25 ft/d) than the current demonstration, which suggests that the increased retardation occurs only at high groundwater velocities (at least greater than 2 ft/d). Therefore, consideration of previously published work along with results from the current demonstration suggests that retardation of bacteria decreases as groundwater velocity decreases.

6.3.1.4 Bacterial Growth

Two methods were used to assess the extent of bacterial growth. The first one was to quantify the number of *DHC* cells that were present at the end of the demonstration, and compare that to the number of cells a dded during bi oaugmentation. F igure 6-1 s hows the *DHC* counts nine months after bioaugmentation, as represented by the 16S rRNA results. In order to determine the total number of *DHC* cells in each treatment cell, the area encompassed by each *DHC* contour was calculated, and was converted to a volume by multiplying by the treatment thickness of 15 ft and the porosity of 0.2. Then, the groundwater volume contained within a given *DHC* contour was multiplied by the average *DHC* concentration for that contour to determine the total number of *DHC* cells present in each specific area. Finally, the cell counts were then summed across each treatment cell. Table 6-3 shows the results of this calculation for the active cell, where 7.0 x 10¹⁴ total *DHC* cells were present at the end of the demonstration. Table 6-4 shows the results of this calculation for the passive cell, where 3.1 x 10¹⁴ total *DHC* cells were present at the end of the demonstration.

During bi oaugmentation, 100 L of bi oaugmentation culture was added to each treatment cell. This culture contained 5×10^{-10} *DHC* cells/L, which means that 5×10^{-12} total *DHC* cells were added to each treatment cell. Since both the active and passive treatment cells had *DHC* cells on the order of 10^{-14} total *DHC* cells, this implies that significant growth of a pproximately two orders of magnitude of *DHC* was stimulated during the demonstration.

The second method to assess the extent of bacterial growth was to determine whether *DHC* levels increased after first arrival at a given monitoring well. These trends are illustrated by Figures 5-8 and 5-14 for the active and passive cells respectively. Figure 5-8 shows that *DHC* concentrations increased by 5 to 6 orders of magnitude after it was first detected. While some of the increase is likely a "breakthrough curve" as the injected culture reaches the well, this increase is also believed to imply significant growth at this monitoring location because concentrations at

Table 6-3. Active Cell *DHC* Population Data

	a wa	Adjusted	Area	Depth	Volume		Adjusted Volume (20%	
Id	Cell/Liter	(Cell/Liter)	(m2)	(m)	(m3)	Volume (Liter)	Porosity) (Liter)	Total Cell Count
0	20,000,000.00	10,000,000.00	50.2	4.6	229.3	229,303.16	45,860.63	4.59E+11
1	20,000,000.00	30,000,000.00	186.0	4.6	850.5	850,491.27	170,098.25	5.10E+12
2	40,000,000.00	50,000,000.00	217.4	4.6	993.7	993,730.39	198,746.08	9.94E+12
3	60,000,000.00	70,000,000.00	267.3	4.6	1222.1	1,222,101.15	244,420.23	1.71E+13
4	140,000,000.00	150,000,000.00	14.7	4.6	67.0	67,025.44	13,405.09	2.01E+12
5	180,000,000.00	190,000,000.00	4.0	4.6	18.5	18,460.64	3,692.13	7.02E+11
6	180,000,000.00	190,000,000.00	0.4	4.6	2.0	2,011.85	402.37	7.65E+10
7	20,000,000.00	10,000,000.00	5.1	4.6	23.4	23,351.40	4,670.28	4.67E+10
8	20,000,000.00	30,000,000.00	32.8	4.6	150.0	149,996.76	29,999.35	9.00E+11
9	40,000,000.00	50,000,000.00	68.8	4.6	314.4	314,416.35	62,883.27	3.14E+12
10	60,000,000.00	70,000,000.00	121.3	4.6	554.6	554,638.56	110,927.71	7.76E+12
11	80,000,000.00	90,000,000.00	615.6	4.6	2814.3	2,814,342.99	562,868.60	5.07E+13
12	100,000,000.00	110,000,000.00	421.4	4.6	1926.6	1,926,628.61	385,325.72	4.24E+13
13	120,000,000.00	130,000,000.00	395.2	4.6	1806.8	1,806,832.44	361,366.49	4.70E+13
14	140,000,000.00	150,000,000.00	274.4	4.6	1254.5	1,254,512.46	250,902.49	3.76E+13
15	160,000,000.00	170,000,000.00	213.0	4.6	973.6	973,629.31	194,725.86	3.31E+13
16	180,000,000.00	190,000,000.00	114.9	4.6	525.5	525,545.07	105,109.01	2.00E+13
17	200,000,000.00	210,000,000.00	101.6	4.6	464.5	464,536.35	92,907.27	1.95E+13
18	220,000,000.00	230,000,000.00	96.3	4.6	440.4	440,437.75	88,087.55	2.03E+13
19	240,000,000.00	250,000,000.00	93.4	4.6	427.2	427,225.82	85,445.16	2.14E+13
20	260,000,000.00	270,000,000.00	91.8	4.6	419.8	419,792.37	83,958.47	2.27E+13
21	280,000,000.00	290,000,000.00	91.0	4.6	416.1	416,128.47	83,225.69	2.41E+13
22	300,000,000.00	310,000,000.00	90.9	4.6	415.5	415,474.69	83,094.94	2.58E+13
23	320,000,000.00	330,000,000.00	91.3	4.6	417.6	417,605.05	83,521.01	2.76E+13
24	340,000,000.00	350,000,000.00	92.5	4.6	422.7	422,729.42	84,545.88	2.96E+13
25	360,000,000.00	370,000,000.00	94.4	4.6	431.5	431,521.96	86,304.39	3.19E+13
26	380,000,000.00	390,000,000.00	97.4	4.6	445.4	445,444.27	89,088.85	3.47E+13
27	400,000,000.00	410,000,000.00	102.2	4.6	467.1	467,119.17	93,423.83	3.83E+13
28	420,000,000.00	430,000,000.00	99.6	4.6	455.2	455,163.42	91,032.68	3.91E+13
29	520,000,000.00	530,000,000.00	2.1	4.6	9.8	9,776.18	1,955.24	1.04E+12
30	500,000,000.00	510,000,000.00	17.4	4.6	79.7	79,731.35	15,946.27	8.13E+12
31	480,000,000.00	490,000,000.00	36.7	4.6	167.9	167,859.28	33,571.86	1.65E+13
32	460,000,000.00	470,000,000.00	60.9	4.6	278.7	278,656.44	55,731.29	2.62E+13
33	440,000,000.00	450,000,000.00	88.4	4.6	404.2	404,215.16	80,843.03	3.64E+13

Total DHC in Active Cell Area

7.0E+14

Table 6-4. Passive Cell DHC Population Data

		Adjusted	Area	Depth	Volume		Adjusted Volume (20% Porosity)	
Id	Cell/Liter	(Cell/Liter)	(m2)	(m)	(m3)	Volume (Liter)	(Liter)	Total Cell Count
0	20,000,000.00	10,000,000.00	22.4	4.6	102.6	102,614.38	20,522.88	2.05E+11
1	20,000,000.00	10,000,000.00	1.0	4.6	4.4	4,394.28	878.86	8.79E+09
2	20,000,000.00	30,000,000.00	106.1	4.6	484.9	484,919.10	96,983.82	2.91E+12
3	40,000,000.00	50,000,000.00	61.2	4.6	279.8	279,846.42	55,969.28	2.80E+12
4	60,000,000.00	70,000,000.00	62.1	4.6	283.9	283,888.53	56,777.71	3.97E+12
5	80,000,000.00	90,000,000.00	60.2	4.6	275.1	275,116.72	55,023.34	4.95E+12
6	380,000,000.00	390,000,000.00	2.4	4.6	10.7	10,744.32	2,148.86	8.38E+11
7	360,000,000.00	370,000,000.00	11.6	4.6	53.1	53,087.72	10,617.54	3.93E+12
8	340,000,000.00	350,000,000.00	23.9	4.6	109.2	109,243.85	21,848.77	7.65E+12
9	320,000,000.00	330,000,000.00	37.9	4.6	173.4	173,443.09	34,688.62	1.14E+13
10	300,000,000.00	310,000,000.00	53.7	4.6	245.7	245,659.06	49,131.81	1.52E+13
11	100,000,000.00	110,000,000.00	65.3	4.6	298.5	298,529.52	59,705.90	6.57E+12
12	280,000,000.00	290,000,000.00	73.5	4.6	335.8	335,818.78	67,163.76	1.95E+13
13	260,000,000.00	270,000,000.00	115.0	4.6	525.6	525,583.87	105,116.77	2.84E+13
14	240,000,000.00	250,000,000.00	116.2	4.6	531.1	531,144.38	106,228.88	2.66E+13
15	220,000,000.00	230,000,000.00	109.8	4.6	502.1	502,093.58	100,418.72	2.31E+13
16	200,000,000.00	210,000,000.00	118.0	4.6	539.4	539,378.33	107,875.67	2.27E+13
17	120,000,000.00	130,000,000.00	77.0	4.6	351.8	351,835.82	70,367.16	9.15E+12
18	180,000,000.00	190,000,000.00	134.2	4.6	613.4	613,427.30	122,685.46	2.33E+13
19	140,000,000.00	150,000,000.00	0.1	4.6	0.3	323.54	64.71	9.71E+09
20	160,000,000.00	170,000,000.00	156.5	4.6	715.7	715,733.89	143,146.78	2.43E+13
21	140,000,000.00	150,000,000.00	177.0	4.6	809.1	809,068.41	161,813.68	2.43E+13
22	120,000,000.00	130,000,000.00	83.8	4.6	383.1	383,056.26	76,611.25	9.96E+12
23	100,000,000.00	110,000,000.00	84.4	4.6	386.0	385,981.62	77,196.32	8.49E+12
24	80,000,000.00	90,000,000.00	96.3	4.6	440.3	440,287.12	88,057.42	7.93E+12
25	60,000,000.00	70,000,000.00	115.9	4.6	529.9	529,877.98	105,975.60	7.42E+12
26	40,000,000.00	50,000,000.00	132.9	4.6	607.6	607,634.06	121,526.81	6.08E+12
27	20,000,000.00	30,000,000.00	264.3	4.6	1208.5	1,208,505.44	241,701.09	7.25E+12
28	-	10,000,000.00	648.0	4.6	2962.7	2,962,749.05	592,549.81	5.93E+12
29	-	-	421.5	4.6	1927.3	1,927,251.35	385,450.27	0.00E+00

Total *DHC* in Passive Cell Area

3.1E+14

the inoculation points remained high the entire time, and were of a similar order of magnitude to the monitoring wells. Figure 5-14 shows a similar trend in that *DHC* concentrations increased by approximately 4 or ders of magnitude at PMW-8, although concentrations did decline between the June and October 2009 sampling events.

The DNA results shown for the rest of the active and passive cell wells are provided in Appendix H and Appendix I, respectively. For both treatment cells, increases of 2 to 5 orders of magnitude of *DHC* concentrations following first arrival were observed at all locations that were monitored monthly. The CMT ports that were only monitoring quarterly had *DHC* concentrations near their maximum levels during the first sampling event following bioaugmentation. Based on data from other wells, however, most growth occurred during this initial three month period, so these data are consistent with other wells.

6.3.2 Extent of Dechlorination

The second Phase 3 objective was to assess and quantify the extent of dechlorination using both the active and passive bioaugmentation approaches. To recap the results presented in Section 5, complete dechlorination of TCE to ethene was achieved in the downgradient two-thirds of the passive treatment cell, with ethene remaining as the predominant product in PMW-7, -8, and -9 in October 2009. In PMW-6, VC and ethene combined accounted for greater than 50% of the remaining compounds. In the upper third of the cell, little dechlorination was observed in spite of having electron donor distributed to all the CMT wells; iron reduction, sulfate reduction and methanogenesis in several locations; and low to moderate numbers of DHC. While determining the cause of this phenomenon was beyond the scope of this demonstration, it is very possible that inhibition f rom c o-contaminants s uch a sc hloroform c ould have limited DHC activity. Chloroform was present at concentrations as high as 1,500 µg/L and carbon tetrachloride as high as 15,000 µg/L in the passive cell near PIW-1. This is the only part of the demonstration area where these high concentrations were observed, and also the only a real where complete dechlorination was not achieved.

In the active cell, complete dechlorination (as indicated by eithene production) occurred to a distance of at least 30 ft from the injection wells. By October 2009, VC and ethene were by far the predominant compounds at all locations within 30 ft of the injection wells. At 75 ft downgradient (AMW-6), degradation products were increasing at the end of the demonstration, but with no electron do nor present and limited evidence of reducing conditions. This suggests that the presence of degradation products at this distance is simply due to migration from upgradient. Thus, complete dechlorination was stimulated to a distance between 30 and 75 ft.

In the ER-0513 work plan, decision rules were defined for this performance objective, based on trends observed in monitoring data as shown in Table 6-5. These decision rules are intended to provide a defined performance metric for the extent of dechlorination achieved.

Table 6-5. Decision Rules for Dechlorination Performance Objective

	Redox Conditions	Chloroethenes	Ethene	qPCR
Favorable trends	Sulfate decreasing or absent; Methane detected	Decreasing or not detected	Increasing or molar equivalent to initial TCE	DHC bacteria detected
Unfavorable trends	Sulfate present and not decreasing; no methane detected	Stable or increasing	Not detected	No <i>DHC</i> bacteria detected

Decision Rule 1: If the passive treatment cell shows all of the favorable trends in Table 6-5 at $\geq 2/3$ of all monitoring wells, then it will be determined that full-scale bi oaugmentation was successfully implemented using the passive approach. If less than $\frac{1}{2}$ of all monitoring wells in the passive cell show all favorable trends in Table 6-5, then it will be determined that full-scale bioaugmentation was not successfully implemented using the passive approach. If more than $\frac{1}{2}$ but less than $\frac{2}{3}$ of all monitoring wells show favorable trends, then further evaluation will be required.

Decision Rule 2: If the active recirculation treatment cell shows all of the favorable trends in Table 6-5 over a distance of greater than or equal to 75 ft from the reinjection wells, then it will be determined that full-scale bioaugmentation was successfully implemented using the active recirculation a pproach. If the active recirculation treatment cell does not show all of the favorable trends in Table 6-5 over a distance of at least 50 ft from the reinjection wells, then it will be determined that full-scale bioaugmentation was not successfully implemented using the active recirculation approach. All other combinations of potential outcomes will require further evaluation.

Each monitoring location in both treatment cells was assessed in order to determine whether favorable trends were achieved. Table 6-6 shows the results of this analysis for the passive cell. A "Y' in Table 6-6 indicates that a favorable trend was observed at a given monitoring location for a given parameter, and an "N" means that an unfavorable trend was observed. From Table 6-6, favorable trends were observed for more than 2/3 of all passive cell monitoring locations for redox conditions, chloroethene concentrations, and qPCR results. However, ethene production was only measured at half of the monitoring locations, and at some of these concentrations were between 5 and $10 \mu g/L$. In addition, concentrations of TCE decreased at some locations without a corresponding increase in daughter products.

Based on D ecision R ule 1, be tween ½ a nd 2/3 of a ll m onitoring wells in the passive cell exhibited favorable trends for all four parameters. According to Decision Rule 1, this condition requires f urther ev aluation. As discussed above, it is possible that high concentrations of chloroform limited *DHC* activity. Regardless of the cause of the inhibition, even though only half of the monitoring wells showed ethene production, this number is biased because all of these wells are located near PIW-1. In terms of treatment cell area, ethene production was observed in the passive cell near two of the three injection wells. This implies that dechlorination to ethene was observed t hroughout a pproximately two-thirds of the passive cell. Therefore, this performance objective was met in terms of area for the passive cell, though not in terms of monitoring wells.

Table 6-6. Passive Cell Results for Dechlorination Performance Objective

Well	PIW -1	PIW -2	PIW -3	PMW -1	PMW -2	PMW -3 (Z1)	PMW -3 (Z2)	PMW -3 (Z3)	PMW -4 (Z1)	PMW -4 (Z3)	PMW -4 (Z4)	PMW -5 (Z1)	PMW -5 (Z2)	PMW -5 (Z3)	PMW -6	PMW -7	PMW -8	PMW -9	Total Y
Redox Conditions	N	Y	Y	Y	Y	N	Y	N	Y	Y	Y	Y	Y	N	Y	Y	Y	Y	0.78
Chloroethenes	Y	Y	Y	N	N	N	Y	Y	Y	Y	N	Y	Y	Y	Y	Y	Y	Y	0.78
Ethene	Y	Y	Y	N	N	N	N	N	Y	N	N	Y	N	N	Y	Y	Y	Y	0.50
qPCR	Y	Y	Y	Y	Y	Y	Y	N	Y	Y	N	Y	Y	N	Y	Y	Y	Y	0.83

Notes:

Y - Favorable trends observed

N - Favorable trends not

observed

Table 6-7 shows the results of the extent of dechlorination analysis for the active cell. From Table 6-7, favorable trends were observed everywhere except AMW-6 and the extraction wells. Based on data presented in Section 5, AMW-6 and the extraction wells were beyond the area that was impacted by lactate injections and bioaugmentation, so these results are expected.

While more than 2/3 of the monitoring wells showed favorable trends in the active cell, Decision Rule 2 was based on the distance from the injection wells that was impacted. All wells that exhibited favorable trends are within 36 ft of the injection wells. Therefore, the portion of the active cell with favorable trends extends somewhere beyond 36 ft, but is less than 75 ft, which is the distance to the next well (AMW-6). According to Decision Rule 2 in Table 6-5, these results require further evaluation. Since too many utilities were present at the site in order to install any monitoring wells between 36 ft and 75 ft from the injection points, the precise location of the area that was impacted by the demonstration is unknown. Because of this, it was determined that this performance objective was partially met for the active cell.

Overall then, this performance objective was partially met. What is more important, however, is that the data are more than sufficient to make a comparison of the relative pros and cons of the two bioaugmentation strategies, which is discussed in the next section.

6.3.3 Comparison of Performance of Active and Passive Approaches

A third de cision was identified in the Demonstration Plan: to determine whether, and to the extent possible, under what conditions the passive approach is more technically effective and cost effective than the active recirculation approach. Decision #3 is based on the outcomes of Decisions 1 and 2, as well as on cost. Because of the multiple combinations of outcomes, and because of the fact that Decision Rules 1 and 2 are qualitative and are based on trends rather than explicit a ction levels, no decision rule was presented for Decision #3. However, an overall evaluation is made considering all available data in order to determine whether the passive approach was more technically effective and more cost effective than the active approach. Costs are discussed in Section 7, and technical performance is summarized below.

Table 6-7. Active Cell Results for Dechlorination Performance Objective

Well	AMW-1	AMW- 2	AMW-3 (Z1)	AMW-3 (Z2)	AMW-3 (Z3)	AMW-4 (Z1)	AMW-4 (Z2)	AMW-4 (Z3)	AMW-5 (Z1)	AMW-5 (Z2)	AMW-6	AEW	Total Y
Redox Conditions	Y	Y	Y	Y	Y	Y	Y	Y	N	Y	N	N	0.75
Chloroethenes	Y	Y	Y	Y	Y	Y	Y	Y	Y	Y	N	N	0.83
Ethene	Y	Y	Y	Y	Y	Y	Y	Y	Y	Y	N	N	0.83
qPCR	Y	Y	Y	Y	Y	Y	Y	Y	Y	Y	N	N	0.83

Notes:

Y - Favorable trends

observed

N - Favorable trends not observed

Based on all data for both the active and passive treatment cells, the following conclusions can be made regarding technical performance of the demonstration:

- Electron donor distribution from an individual injection point was similar using both the passive and active approaches (greater than 25 ft in both cases)
- Electron do nor and *DHC* distribution varied vertically for both strategies based on data from the CMT wells; this did not have a negative impact on dechlorination in the active cell, but dechlorination was minimal in all the CMT wells in the passive cell (likely due to inhibition caused by co-contaminants)
- Higher electron donor concentrations were achieved in the passive cell, which required significantly less donor compared to the active approach
- Strongly r educing r edox c onditions were e stablished within s imilar timeframes using both approaches
- Dechlorination pe rformance w as s imilar f or bot h a pproaches, w ith t he e xception of possible inhibition in part of the passive cell
- Bacterial distribution was similar from a given injection location both in terms of time to first arrival, and in terms of area influenced
- In terms of area impacted, the passive approach stimulated dechlorination and bacterial distribution over a larger percentage of the treatment cell compared to the active approach, which was limited to the area near the injection wells

It is likely that the hydrogeology of this site played an important role in the similar technical performance of the passive and active bioaugmentation strategies. In particular, it was observed that *DHC* was transported rapidly in both scenarios, with first arrival of *DHC* showing little or no retardation compared to first arrival of a cons ervative tracer. In addition, these first arrivals revealed the presence of some relatively high-flow solute transport pathways in the subsurface. It is possible that having some such higher-velocity flow paths is an important ingredient for the success of a passive bioaugmentation strategy. Without such paths, DHC transport might be slowed significantly. It is possible that a nactive strategy might a chieve more rapid DHC transport in a hydrogeologic setting with uniformly slow solute transport, though that could not be evaluated in this demonstration. A tracer test is a useful characterization technique for any full-scale bioaugmentation application to assist not only in the selection of a passive vs. an active approach, but also for design of injection well spacing, placement of well screens, monitoring well locations, and so forth. Tracer testing with three-dimensional monitoring, as with CMT wells, is particularly useful for this purpose, as was also illustrated and documented in the final reports for ESTCP project ER-0218.

Overall, technical performance of both approaches was similar in all regards. However, as discussed in Section 6.4, operations and maintenance (O&M) requirements were higher for the active approach, and the system was not as u ser-friendly compared to the passive approach. Also, as presented in Section 7, c osts for the active approach were higher than for the passive approach.

6.4 QUALITATIVE PERFORMANCE OBJECTIVES

One qualitative performance objective was established for the ER-0513 project. This objective was to assess the ease of use for both passive and active approaches. This includes operational time re quired in the field, time s pent c onducting maintenance and r epair a ctivities, and the amount of training required to operate each system. Data collected in support of this objective include feedback from field personnel; injection and operational logs, and the field team leader logbook.

During the course of the demonstration, the active recirculation system required more time for troubleshooting and maintenance than the passive system did. One major shutdown occurred in late 2008 due to malfunction of overflow shutoff switches and the autodialer (refer to Section 5.6.3). This required modification of the recirculation system to include an additional overflow tank, and additional instrumentation. In addition, several minor equipment malfunctions occurred during the course of the demonstration, such as flowmeter clogging, temporary extraction pump shutdowns, and PLC errors. The active recirculation system also required more training for field personnel to understand the PLC programming, how to properly dose the electron do nor, and how to troubleshoot the system. Although it did not occur during this demonstration, it is our experience from working at other sites that biofouling is also more common in recirculation systems than passive injection systems.

In contrast, the passive system required no electronics, and only had one minor repair to replace flowmeters. Less training was required for the passive system, because it consisted of a simple manifold to inject three wells at a time. The passive system did require a source of potable water to use for the injections, but one was available nearby.

The success criterion for this performance objective was to quantify the operational requirements for each approach. Data collected during the course of the ER-0513 demonstration did allow for an assessment of the ease of use of both approaches, and it was determined that the passive system was easier to use and required less maintenance. Therefore, this performance goal was met.

7.0 COST ASSESSMENT

A critical evaluation criterion for any cleanup technology is cost. In this section, implementation costs for bioremediation of chlorinated solvent source areas are estimated based on the costs of the demonstration. Section 7.1 includes a review of the approximate costs associated with the demonstration project. Section 7.2 provides a discussion of the primary cost drivers that influence effective implementation of EAB at sites, and includes a discussion of the positive and negative characteristics of active and passive treatment methods demonstrated for this project. Finally, Section 7.3 provides cost information for successful implementation of the remedy at a theoretical site.

7.1 COST REPORTING

Table 7-1 pr ovides the estimated implementation costs of the technology for the Site 70 demonstration project at NAVWPNSTA Site 70. These costs are the approximate costs for performing a detailed demonstration of the technology, including more intensive sampling and analysis than would typically be needed for a more "standard" application of the technology. Projected costs for a more typical application of the technology at a model site are provided in Section 7.3.

Detailed discussions of each of the cost element tasks in the table have been provided in previous sections of this report. For clarity, a summary of each is provided below:

Start-up: consists of work plan de velopment and t reatability/DNA s equencing s tudies. Work plan de velopment included finalization of the demonstration design, negotiation of anticipated project activities and costs, and development of supporting documentation. The treatability study consisted of bench-scale testing for dechlorination, which was recommended due to the high sulfate and chloride concentrations present at the site. The DNA sequencing study was conducted to determine whether native species of *DHC* were present at the site prior to implementation of the demonstration. Presence of the bacteria could have impacted the ability to assess growth and distribution of the bioaugmentation culture during the demonstration project.

General C onstruction: c onsists of w ell ins tallation, tracer testing and hydraulic characterization, and groundwater modeling. Well installation included monitoring, extraction, and injection well installation that was necessary for completion of the demonstration. Tracer testing and hydraulic c haracterization was performed to gather data on f low c haracteristics within the active and passive treatment cells. Modeling work was performed to indicate potential groundwater extraction rates and to anticipate electron donor distribution in the subsurface.

Active Cell Construction: consists of injection system construction, lactate injections, bioaugmentation, system troubleshooting/maintenance, and sampling. The active cell was constructed to include groundwater extraction and reinjection components, and to facilitate injection of electron donor for bioremediation. Bioaugmentation was necessary to provide *DHC* with the *vcrA* gene, which is necessary to obtain complete dechlorination to ethene. General system troubleshooting and maintenance was necessary for upkeep of the treatment system. Sampling was included for evaluation of the system performance.

Table 7-1. Approximate Implementation Costs for EAB at NAVWPNSTA Site 70

Cost Element	Sub-Category	Detail	Costs
Start-Up Costs			\$100,000
	Treatability/DNA Sequencing	Procurement- 80 hr	\$6,000
	Study	Subcontractors (lab services)	\$20,000
	Work Plan	Project Manager- 220 hr	\$27,500
		Technical Reviewer- 40 hr	\$8,000
		Project Engineer- 340 hr	\$34,000
		Drafting/Clerical- 60 hr	\$4,500
General			\$214,200
Construction Costs	Well Installation/Development	Project Geologist- 500 hr	\$50,000
		Subcontractor	\$112,000
		Materials/ODCs	\$20,000
	Tracer Testing/Hydraulic Characterization	Project Manager - 40 hr	\$5,000
		Project Engineer - 40 hr	\$4,000
		Project Geologist - 160 hr	\$16,000
		Materials/ODCs	\$4,200
	Screening Level Groundwater Modeling	Project Hydrogeologist - 24 hr	\$3,000
Active Cell			\$341,300
Construction/ O&M	Oversight/Supervision	Project Manager - 200 hr	\$25,000
O&M	Lactate Injection System Purchase/ Construction	Subcontractor	\$40,000
	Lactate Injection (1x per week)	Project Engineer- 10 hr/event, 40 events	\$40,000
		Lactate- 50 gal per event, 40 events	\$24,000
	Bioaugmentation	Project Engineer- 20 hr	\$2,000
		Bacterial Culture	\$15,000
	System Troubleshooting/ Maintenance (1 major and 3 minor events during demo)	Project Engineer - 80 hr	\$8,000
		Technician - 80 hr	\$4,800
		Materials/ODCs	\$10,000
	Sampling (12 total events)	Project Engineer - 240 hr	\$24,000
		Project Geologist - 240 hr	\$24,000
		Analytical (all analytes, including CSIA and qPCR)	\$106,500
		Materials/ODCs (\$1,500 per event)	\$18,000

Cost Element	Sub-Category	Detail	Costs
Passive Cell			\$251,300
Construction/ O&M	Oversight/Supervision	Project Manager - 100 hr	\$12,500
OXIVI	Lactate Injection System Purchase/ Construction	Subcontractor	\$15,500
	Lactate Injection (1x per week)	Project Engineer- 20 hr/event, 12 events	\$24,000
		Lactate- 50 gal per event, 12 events	\$7,200
	Bioaugmentation	Project Engineer- 20 hr	\$2,000
		Bacterial Culture	\$15,000
	System Troubleshooting/ Maintenance (1 minor event during demo)	Project Engineer - 10 hr	\$1,000
		Technician - 10 hr	\$600
		Materials/ODCs	\$1,000
	Sampling (12 total events)	Project Engineer - 240 hr	\$24,000
		Project Geologist - 240 hr	\$24,000
		Analytical (all analytes, including CSIA and qPCR	\$106,500
		Materials/ODCs (\$1,500 per event)	\$18,000
Performance			\$210,000
Assessment, Reporting, and Project Management	Includes final project reports, tech transfer, and data management/interpretation	Project Manager- 600 hr	\$75,000
		Technical Reviewer- 200 hr	\$40,000
		Project Engineer- 600 hr	\$60,000
		Drafting/Clerical- 200 hr	\$15,000
		Travel/ODC's	\$20,000
Demobilization	Site Cleanup and Restoration		\$5,000
Waste Disposal			NA
Long-term Monitoring			NA

Passive C ell C onstruction: similar to the a ctive cell c onstruction, with the exception of groundwater extraction and reinjection. The passive cell did not include these components, and utilized natural groundwater flow to distribute electron donor and bacteria. The costs for passive cell construction and O&M were considerably less than the active cell.

Performance A ssessment, Reporting, an d P roject Management: includes ongoi ng management a nd r eview of a nalytical d ata, as well as p eriodic p roject reporting. This a lso includes preparation of the final project reports.

Demobilization: includes removing equipment and materials from the site, as well as site restoration.

Waste Disposal: Includes removal and disposal of all investigation derived waste. These costs are standard, fairly insignificant, and were not tracked during the demonstration.

Long-Term Monitoring: Includes monitoring conducted after the demonstration is completed. These costs are standard and were not tracked during the demonstration.

7.2 COST DRIVERS

As with most *in situ* remediation technologies, the most important a spect of implementing bioaugmentation in chlorinated solvent source a reas is delivery and distribution. That is, the electron donor and bacteria must be distributed throughout the target treatment zone to stimulate the desired degradation. Therefore, the major cost drivers are likely to be the infrastructure and materials required to achieve distribution of amendments. These are largely driven by the scale of a sitelaterally and vertically, as well as the hydraulic conductivity and the degree of heterogeneity. The "bulk" hydraulic conductivity of the treatment zone will determine the spacing of injection wells, and will have a strong influence on the required treatment duration. The heterogeneity will mostly impact the treatment duration because a high degree of heterogeneity will increase the potential for preferential flow. A high degree of preferential flow will result in a cleanup timeframe that is dependent upon diffusion more than advection, which will increase treatment duration, thereby increasing costs.

Similarly, the sheer mass of contamination can be a cost driver. As long as the source consists primarily of solvents at residual saturation or sorbed to the soil, mass removal can be fairly rapid (subject to the potential constraints of hydraulic conductivity and he terogeneity discussed above). However, if DNAPL is present in pools, the cleanup timeframe becomes limited by dissolution rates. While these rates can be accelerated during bioremediation (see the ER-0218 final report), cleanup timeframes will still be long for large pools of DNAPL.

Another potential cost driver is a need for hydraulic containment. If a sufficient downgradient buffer zone is not a vailable at a site and extraction of groundwater is required to prevent the temporary increase in mass flux caused by EAB from impacting some nearby downgradient receptor, costs would increase. This is especially true if for some reason the extracted water cannot simply be reinjected in the source area.

Vapor int rusion concerns can also be a potential c ost d river. Bioremediation of c hlorinated solvents via EAB generates VC and methane. For shallow, unconfined groundwater sites, this creates the potential for these gases to reach fairly high concentrations in the unsaturated zone above the water table. If potential receptors were present above the treatment zone and soil vapor extraction were required, this would increase technology costs.

7.3 COST ANALYSIS

This section provides an estimate for "typical" passive and active bioaugmentation approaches at an example site with similar characteristics to that of NAVWPNSTA Site 70. The estimate is based on the costs associated with the demonstration project, but does not include the level of rigor r equired f or t echnology va lidation. T able 7.2 prov ides the si te cha racteristics and assumptions for the example site.

Table 7-2. Parameters Used as the Basis for Calculating Technology Implementation Costs.

	Active Approach	Passive Approach
Site Area (acre)	0.5	0.5
Site Area (sq ft)	21,780	21,780
Contaminated Thickness Treated (ft)	20	20
Treatment Volume (cubic yards)	16,200	16,200
Number of Injection Wells (scaled up from demonstration)	10	19
Number of Multilevel Monitoring Wells	2	2
Number of Fully Penetrating Monitoring Wells	8	8
Number of Extraction Wells (active cell only)	10	0
Duration of Operations (years)	5	5
Frequency/Concentration of Electron Donor Injection	Weekly/(3%)	Monthly (1%)
Frequency of Monitoring Events	quarterly	quarterly
	Same as Demonstration, but no CSIA and DNA only	Same as Demonstration, but no CSIA and DNA only
Monitoring Analytes	for first year	for first year

An effort was made to be conservative in several of the parameters so as to a void be ing too optimistic in the estimate. For example, the number of monitoring wells (especially the multilevel wells) is higher than many cleanups at the assumed scale. In addition, the Site 70 costs included tracer testing, modeling, a treatability study, and DNA sequencing, as noted in Table 7-1. These activities a renot a lways performed in typical applications, but can significantly improve technology performance, and should be considered prior to implementation of a remedy. The tracer testing and modeling efforts could be especially beneficial to a similar project, as they may a id in determination of flow rates, donor distribution effectiveness, estimated cleanup timeframes, and whether a passive or active treatment method would be more appropriate.

In other cases, the demonstration costs were reduced to reflect, for example, the frequency of sampling that would be typical of implementation, as opposed to the frequent sampling required to quantify bacterial growth and distribution under different conditions. Also, this project included two separate drilling mobilizations in order to properly construct both treatment cells; this would not be required for a typical implementation.

The number of injection wells required for each approach was scaled up based on the ER-0513 project. For the active approach, this was based on the fact that approximately one-half to two-thirds of the treatment cell was impacted during the demonstration, using two extraction and two injection wells. For the theoretical site, this led to 10 injection and extraction wells for the active approach, and 19 injection wells for the passive approach. The same lactate injection frequency was assumed for each approach (weekly for active, and monthly for passive). Monitoring would be conducted quarterly, rather than monthly as was done during the demonstration. Also, CSIA would not be performed, and qPCR for *DHC* would only be performed during the first year of operations.

This cost analysis focuses on comparing and contrasting the passive and active approaches for bioaugmentation in the context of implementing bi oremediation for cleanup of a chlorinated solvent source area. For a comparison of bi oremediation to other remediation technologies for source area cleanup, see the Cost and Performance Report for ESTCP project ER-0218.

Life cycle costing provides the greatest utility when a project has a significant initial capital or short-term operating cost, followed by a much longer period of lower operating costs. This is not really the case either for the comparison of active and passive bioaugmentation approaches (in any case, they would be assumed to have the same long-term monitoring needs if that were included). For both cases, the costs were assumed to be incurred over 5-6 years (including preliminary characterization, well drilling, etc.). Thus, the total costs reported below essentially are the life cycle costs. In both cases, the capital cost is relatively small and the operational period is still not very long, so again the utility of a net present value calculation is minimal and was not performed.

Tables 7-3 and 7-4 present the projected implementation costs for bioaugmentation using the active and passive approach, respectively. Most of the costs are similar (e.g. start-up, general construction, monitoring, and performance assessment) because they are common to both active and passive approaches. However, for a theoretical site of this size, the construction and O&M costs for the active approach are approximately three times as high as for the passive approach. The result is an estimated cost for the active approach of \$2.5 M, compared to \$1.5M for the passive approach. The primary drivers for this cost increase are the significantly higher amount of lactate required, and the higher costs for maintenance and oversight of recirculation systems. The magnitude of the cost differences for O&M activities increases as the size of the area treated increases. As alluded to in Section 6, the benefits of implementing an active approach do not appear to be justified by the increased costs, at least for a site like NAVWPNSTA Seal Beach. Bacterial distribution was not significantly faster, and dechlorination performance was similar to the passive approach.

It should be noted that some sites might have conditions that would lead to more significant benefits for recirculation systems. For sites with very high groundwater flow velocities, recirculation might be needed to manage residence within the treatment zone avoid chlorinated degradation products migrating off-site. Such a site would also allow electron donor to be distributed over a much larger distance prior to being degraded than was possible at Seal Beach, which would increase the benefit.

Table 7-3. Projected Implementation Costs for Bioaugmentation using Active Recirculation Approach

Sub-Category	Detail	Costs
		\$100,000
Treatability/DNA Sequencing	Procurement- 80 hr	\$6,000
Study	Subcontractors (lab services)	\$20,000
Work Plan	Project Manager- 220 hr	\$27,500
	Technical Reviewer- 40 hr	\$8,000
	Project Engineer- 340 hr	\$34,000
	Drafting/Clerical- 60 hr	\$4,500
		\$201,700
Well Installation/Development	Project Geologist- 500 hr	\$50,000
•	Subcontractor	\$112,000
	Materials/ODCs	\$20,000
Tracer Testing/Hydraulic Characterization	Project Manager - 20 hr	\$2,500
	Project Engineer - 20 hr	\$2,000
	Project Geologist - 80 hr	\$8,000
	Materials/ODCs	\$4,200
Screening Level Groundwater Modeling	Project Hydrogeologist - 24 hr	\$3,000
		\$1,751,700
Oversight/Supervision	Project Manager - 800 hr	\$100,000
Lactate Injection System Purchase/ Construction	Subcontractor	\$160,000
Lactate Injection (1x every week)	Project Engineer- 10 hr/event, 260 events	\$260,000
	Lactate- 250 gal per event, 260 events	\$780,000
Bioaugmentation	Project Engineer- 80 hr	\$8,000
	Bacterial Culture	\$60,000
System Troubleshooting/ Maintenance (1 major and 3 minor events during demo)	Project Engineer - 320 hr	\$32,000
	Technician - 320 hr	\$19,200
	Materials/ODCs	\$40,000
Sampling (21 total events)	Project Engineer - 630 hr	\$63,000
	Project Geologist - 630 hr	\$63,000
	Analytical (all analytes, excluding CSIA and qPCR only for Year 1)	\$135,000
	Materials/ODCs (\$1,500 per event)	\$31,500
	Treatability/DNA Sequencing Study Work Plan Well Installation/Development Tracer Testing/Hydraulic Characterization Screening Level Groundwater Modeling Oversight/Supervision Lactate Injection System Purchase/ Construction Lactate Injection (1x every week) Bioaugmentation System Troubleshooting/ Maintenance (1 major and 3 minor events during demo)	Treatability/DNA Sequencing Study Work Plan Project Manager- 220 hr Technical Reviewer- 40 hr Project Engineer- 340 hr Drafting/Clerical- 60 hr Well Installation/Development Well Installation/Development Well Installation/Development Tracer Testing/Hydraulic Characterization Project Manager - 20 hr Project Geologist - 80 hr Materials/ODCs Screening Level Groundwater Modeling Oversight/Supervision Lactate Injection System Purchase/ Construction Lactate Injection (1x every week) Bioaugmentation Project Engineer - 80 hr Bacterial Culture System Troubleshooting/ Maintenance (1 major and 3 minor events during demo) Project Engineer - 320 hr Materials/ODCs Sampling (21 total events) Project Engineer - 630 hr Project Engineer - 630 hr Project Engineer - 630 hr Project Geologist - 630 hr Analytical (all analytes, excluding CSIA and qPCR only for Year 1)

Cost Element	Sub-Category	Detail	Costs
Performance Assessment,			\$420,000
Reporting, and Project Management	Includes final project reports, tech transfer, and data management/interpretation	Project Manager- 1200 hr	\$150,000
		Technical Reviewer- 400 hr	\$80,000
		Project Engineer- 1200 hr	\$120,000
		Drafting/Clerical- 400 hr	\$30,000
		Travel/ODC's	\$40,000
Demobilization	Site Cleanup and Restoration		\$20,000
Waste Disposal			NA
Long-term Monitoring			NA
Total			\$2,493,400

On the other hand, sites with very low groundwater velocities might make a passive system impractical because very little distribution can be achieved without enhancing the hydraulic gradient. What this demonstration indicates is that for sites that are closer to the "average" in terms of groundwater velocity, passive bioaugmentation systems are likely to be more cost-effective than active systems.

Table 7-4. Projected Implementation Costs for Bioaugmentation using Passive Approach

Cost Element	Sub-Category	Detail	Costs
Start-Up Costs			\$100,000
	Treatability/DNA Sequencing	Procurement- 80 hr	\$6,000
	Study	Subcontractors (lab services)	\$20,000
	Work Plan	Project Manager- 220 hr	\$27,500
		Technical Reviewer- 40 hr	\$8,000
		Project Engineer- 340 hr	\$34,000
		Drafting/Clerical- 60 hr	\$4,500
General			\$201,700
Construction Costs	Well Installation/Development	Project Geologist- 500 hr	\$50,000
		Subcontractor	\$112,000
		Materials/ODCs	\$20,000
	Tracer Testing/Hydraulic Characterization	Project Manager - 20 hr	\$2,500
		Project Engineer - 20 hr	\$2,000
		Project Geologist - 80 hr	\$8,000
		Materials/ODCs	\$4,200
	Screening Level Groundwater Modeling	Project Hydrogeologist - 24 hr	\$3,000
Passive			\$761,300
Approach Construction/	Oversight/Supervision	Project Manager - 400 hr	\$50,000
O&M	Lactate Injection System Purchase/ Construction	Subcontractor	\$62,000
	Lactate Injection (1x every week)	Project Engineer- 20 hr/event, 48 events	\$96,000
		Lactate- 317 gal per event, 48 events	\$182,400
	Bioaugmentation	Project Engineer- 80 hr	\$8,000
		Bacterial Culture	\$60,000
	System Troubleshooting/ Maintenance (1 major and 3 minor events during demo)	Project Engineer - 40 hr	\$4,000
		Technician - 40 hr	\$2,400
		Materials/ODCs	\$4,000
	Sampling (21 total events)	Project Engineer - 630 hr	\$63,000
		Project Geologist - 630 hr	\$63,000
		Analytical (all analytes, excluding CSIA and qPCR only for Year 1)	\$135,000
		Materials/ODCs (\$1,500 per event)	\$31,500

Cost Element	Sub-Category	Detail	Costs
Performance Assessment,			\$420,000
Reporting, and Project Management	Includes final project reports, tech transfer, and data management/interpretation	Project Manager- 1200 hr	\$150,000
		Technical Reviewer- 400 hr	\$80,000
		Project Engineer- 1200 hr	\$120,000
		Drafting/Clerical- 400 hr	\$30,000
		Travel/ODC's	\$40,000
Demobilization	Site Cleanup and Restoration		\$20,000
Waste Disposal			NA
Long-term Monitoring			NA
Total			\$1,503,000

8.0 IMPLEMENTATION ISSUES

This section discusses implementation issues for bioaugmentation. In general, the issues are similar when using either the passive or active approach. However, additional issues related to permitting may be encountered when a pplying the technology using the active recirculation approach.

8.1 REGULATIONS THAT APPLY TO BIOAUGMENTATION

The primary regulation or set of regulations that are applicable to bioaugmentation technology are related to underground injection control. Permits may be required for both electron donors and f or bi oaugmentation c ultures. Specifically in C alifornia, W aste Discharge Requirement (WDR) permits a re required. G eneral W DR permit N O. R 4-2007-0019 covers groundwater remediation at petroleum hydrocarbon fuel, VOC, and/or hexavalent chromium impacted sites. Any amendment listed in this permit can be used at a site without a separate permitting process. In cases where a general WDR permit does not cover the amendments or cultures required for a site, a site-specific WDR permit may be needed. It should be noted that permits are not required for remediation at C ERCLA si tes such as NAVWPNSTA Site 70; however the substantive requirements of the permits need to be met.

Bioaugmentation at sites that use recirculation also need to address the issue of how extracted water is handled. Some states may have regulations that state extracted water needs to be treated prior to reinjection. However, RCRA regulations [specifically 3020(b)] specifically allow for both injection of treatment agents, and reinjection of extracted water amended with bioremediation treatment agents if certain conditions are met: "Specifically, the groundwater must be treated prior to reinjection; the treatment must be intended to substantially reduce hazardous constituents in the ground water — either before or after reinjection; the cleanup must be protective of human health and the environment; and the injection must be part of a response action under CERCLA, Section 104 or 106, or a RCRA corrective action intended to clean up the contamination."

8.2 STAKEHOLDER/END-USER ISSUES

While bioaugmentation is an innovative technology that has not been extensively documented at full scale, *in situ* bioremediation has been implemented at many DoD sites across the country. In general, *in situ* bioremediation is well received by regulators and the public for many reasons, including:

- Low Risks Since most or all of the contaminant treatment occurs in the soil or groundwater, risks to human health and the environment during implementation are low compared to *ex situ* technologies.
- Low secondary waste generation Contaminant treatment occurs *in situ*, with little offsite disposal of residuals required.
- **Minimal impacts during operations** Compared to *ex situ* technologies, little infrastructure is required to implement and operate the bi oremediation systems, resulting in minimal disruption to businesses and residences.

• Overall risk reduction – *In situ* bioremediation has been shown to be reliable in significantly de creasing contaminant c oncentrations in relatively short timeframes, resulting in reductions of risk to human health and the environment.

While the merits of bioremediation have resulted in widespread acceptance of the technology, full-scale bioaugmentation does present issues that are not encountered for bioremediation alone. These issues can be categorized as either concerns about the technology itself, or decision-making factors related to implementation of the technology.

The pr imary c oncerns a bout f ull-scale bi oaugmentation are re lated to the introduction of exogenous bacteria to a site's groundwater. Stakeholders may object to the introduction of non-native bacteria to an aquifer. For the current demonstration project, this concern was addressed by c iting the pr ecedence f or pe rforming bi oaugmentation a t ot her sites, m ost not ably a t NAVWPNSTA Seal Beach Site 40, as well as the fact that bioaugmentation is the CERCLA selected remedy for Site 70. A nother concern related to the introduction of bacteria may be simply the ability to distribute them over a sufficient area to achieve full-scale treatment; this was the purpose of this demonstration project.

The primary end us er decision-making factors regarding bioaugmentation are when (or if) to perform t he a ctual i noculation e vents, a nd t he most e ffective a nd e fficient m ethod for distribution of t he ba cteria. The f irst f actor ha s be en so mewhat controversial within t he environmental community, and the "proper" decision will depend on the specifics of the site. While this factor is not the primary focus of the demonstration project, it is important in terms of bioaugmentation implementation. At a minimum, a site should not be bioaugmented until the appropriate redox conditions have been established (i.e., sulfate reduction or methanogenesis) through biostimulation alone. Once this has been achieved, opinions vary about the amount of continue bi ostimulation be fore bi oaugmenting. O n one e xtreme, s ome a dvocate bioaugmenting immediately after achieving the appropriate redox conditions without waiting to see if the appropriate dechlorinating bacteria are indigenous to the site. The reasoning for this approach is that bi oaugmentation will reduce lag times prior to the onset of complete dechlorination even if de chlorinating bacteria ar e present at the site. However, this approach could result in unnecessarily bioaugmenting a site, which could increase overall remediation costs. The other extreme for this factor is to perform biostimulation alone for months or even years in order to determine if D CE stall will eventually be overcome naturally at a site. The reasoning f or t his a pproach w ould be t o a void unne cessarily bi oaugmenting a s ite w hen dechlorinating bacteria will eventually proliferate. The potential disadvantage of this approach is that a site could remain in a state of DCE stall for a significant amount of time before complete dechlorination is achieved, thereby increasing life-cycle costs compared to bioaugmentation.

Given that the purpose of this demonstration is to compare full-scale bioaugmentation systems (as opposed to remediation of the site), the first approach was adopted for this project. Since the pre-conditioning a pproach was a dopted, it did provide the opportunity to sample for *DHC* bacteria and determine whether biostimulation alone would be sufficient. In this case, it was evident that bi oaugmentation would be required. A nother factor in flavor of performing bioaugmentation as soon as redox conditions are appropriate is that the cost of bioaugmentation is usually a small portion of overall project costs, and in many cases is cost effective compared to a longer b iostimulation phase. However, this does not imply that this is the approach

recommended f or a ll s ites. V ery large s ites, for w hich bi oaugmentation w ould r epresent a significant cost, may benefit from a longer biostimulation phase.

The second factor, the best method for large-scale distribution of bacteria, is the primary focus of this demonstration project. The results of the side-by-side comparison of the passive and active approaches were presented in Sections 5 and 6. This topic is also discussed in the forthcoming ESTCP monograph on bioaugmentation.

8.3 PROCUREMENT ISSUES

No s ignificant pr ocurement i ssues e xist f or b ioaugmentation. T his t echnology uses r eadily available techniques f or w ell installation, a nd s tandard c omponents f or pe rforming s ubstrate injections. Projects that use a recirculation approach require more equipment and above ground infrastructure, but i t i s all st andard and readily available from i ndustrial supp ly companies. Amendments are widely available from bioremediation vendors across the country, and several bioaugmentation cultures a re a vailable from multiple s uppliers. B ioaugmentation t echnology does r equire so mewhat spe cialized expertise t o prope rly interpret da ta and make ope rational changes in order to optimize performance.

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APPENDICES

Appendix A: Points of Contact

POINT OF CONTACT Name	ORGANIZATION Name Address	Phone Fax E-mail	Role in Project
Joey Trotsky	NAVFAC ESC	(805) 982-1258	PI
Kent Sorenson	CDM	(303)-383-2300 sorensonks@cdm.com	Co-PI
Ryan Wymore	CDM	(303)-383-2300 wymorera@cdm.com	Project Manager
Brenda Reese	NAVFAC SW	(619)-532-4209	Remedial Project Manager
Pei-Fen Tamashiro	NAVWPNSTA Seal Beach	(562) 626-7897	IR Program Coordinator

Appendix B Memoranda Submitted to ESTCP

June 6, 2008

Ms. Andrea Leeson, Ph.D. ESTCP Program Office 901 North Stuart Street, Suite 303 Arlington, VA 22203

Subject: Baseline sampling and tracer test results for ER-0513

Dear Andrea:

This White Paper presents results of baseline sampling and tracer testing for Environmental Security Technology Certification Program (ESTCP) project ER-0513, with the intent of documenting whether the selected site will be appropriate for meeting the demonstration objectives. This project is being conducted at Naval Weapons Station Seal Beach, Site 70. The purpose of this demonstration is to compare the low-cost, passive approach for bioaugmentation to the more common recirculation approaches for full-scale TCE source area application. Performance of the two approaches is being measured in terms of growth and distribution of Dehalococcoides bacteria, time required to achieve complete dechlorination in the test area, and cost. Specifically, the technical objectives of this project are to:

- Demonstrate cost-effective large-scale bacterial distribution
- Demonstrate induction of complete dechlorination
- Compare and contrast passive and active approaches
- Provide technology transfer

Project field work began in February 2008 with construction of the active recirculation treatment cell. This was followed by the initiation of the "pre-conditioning" phase during which electron donor is being added to both the active and passive treatment cells in order to establish appropriate reducing conditions in the aquifer prior to bioaugmentation.

The active recirculation cell extracts and reinjects groundwater continuously. Electron donor (1% to 3% sodium lactate) is being pulsed into the reinjection line approximately once per month. For the passive treatment cell, sodium lactate is being injected into each of three injection wells once per month, with the injection concentration and electron donor mass being

the same for both treatment cells. Once conditions are sufficiently reducing (as evidenced by ferrous iron concentrations greater than 0.5 mg/L, and a decrease in sulfate of at least 10% from baseline), the treatment cells will be bioaugmented using a commercially available bioaugmentation culture (Shaw's SDC-9).

Approximately four months of field activities have been conducted to date for the ER-0513 project. This includes installation of the active recirculation system, well installation for the passive cell, baseline groundwater sampling, tracer testing, and pre-conditioning lactate injection. This white paper describes these activities in detail, and presents results obtained to date.

Active Recirculation System

The wells for the active cell were installed in September and October 2007, along with two of the passive cell wells. The active cell recirculation system itself was constructed, installed, and tested in March and April 2008. The system operates by extracting groundwater from wells AEW-1 and AEW-2 into a 275 gallon surge tank; the surge tank water is reinjected into AIW-1 and AIW-2, which is a distance of 100 ft upgradient from the extraction wells (refer to Figure 1 for well locations). Once the system was functional, it was operated for several days, and water levels were measured in active cell monitoring wells, and in the two existing passive cell monitoring wells, in order to determine the groundwater flow direction in the area of the proposed passive cell wells. Water level data were collected in several wells using transducers, and in other wells by taking water levels using a synoptic water level meter.

This phased approach for treatment cell construction allowed for the opportunity to assess groundwater flow direction in the area of the planned passive cell wells before installing the remaining ten wells. This helped avoid a scenario in which the entire passive treatment cell was installed, only to find out that groundwater did not flow parallel to the treatment cell axis.

Figure 2 shows the measured water levels from both the active cell and the previously installed passive cell wells, under ambient conditions, and with the active cell recirculation system operating (pumped conditions). Note that this figure shows water levels in elevation in feet below mean sea level, implying that groundwater flows in the direction of increasing numbers on the figure. From this figure, the groundwater flow direction was southerly under ambient conditions in both treatment cell areas, and perhaps slightly southwest in the passive cell area under pumped conditions. This was in contrast to the southeastern direction that was assumed based on data available at the time the ESTCP Demonstration Plan was submitted.

Passive Cell Well Installation

In order to account for the more southerly flow direction, placement of some of the active cell wells was adjusted slightly from their original planned locations. These adjustments were made considering interpreted groundwater flow directions as well as accounting for the many underground utilities in the area. The planned and actual locations are presented in Figure 3. The most significant change was moving continuous multi-channel tubing (CMT) well PMW-3 from its planned location southeast of injection well PIW-1 to a location southwest of PIW-1. Also, well PMW-2 was moved from its planned location on the treatment cell axis to a location southwest of PIW-1. Finally, wells PIW-2 and PMW-6 were moved a few feet to the west of their planned locations in order to avoid utilities.

The actual drilling and development of the remaining ten passive cell wells (four monitoring wells, three injection wells, and three CMT wells) was performed from March 24, 2008 through April 11, 2008 After installation of the remaining passive cell wells, a new round of water level measurements was collected under pumped conditions. These are presented in Figure 4, which shows water levels in elevation in feet below mean sea level. From Figure 4, the groundwater flow direction in the area of the passive cell is south to southeast, as opposed to the more southwesterly direction observed when only two wells were installed. Therefore, the placement of injection and monitoring wells in the passive cell should allow for meaningful results to be observed in all monitoring locations.

Baseline Sampling

Baseline sampling for the active cell was completed the week of April 7, 2008. This included sampling the three standard monitoring wells, all ports in the three CMT wells, and the water being produced from the extraction wells (refer to Figure 1 for well locations). Baseline sampling for the passive cell was completed the week of April 21, 2008. This included sampling the six standard monitoring wells, all ports in the three CMT wells, and the three injection wells (refer to Figure 1 for well locations). Both baseline events were conducted with the active cell recirculation system operating. Analytes sampled included VOCs, ethene/ethane/methane, anions (sulfate, chloride, nitrate/nitrite), alkalinity, COD, DNA samples, compound-specific isotope analysis, and iodide tracer (for background measurements).

The ESTCP Demonstration plan called for three sample ports in each CMT well. During installation of both the active and passive cell CMT wells, four sample ports were completed in all CMT wells except PMW-4, which has five sample ports. This was done in order to account for the possibility that some ports would not produce enough water for sampling. During the baseline sampling events, it was determined that the uppermost port in each active cell CMT

well did not produce sufficient water to complete a full set of samples. However, because extra ports were installed in each well, data are available from multiple depths in each CMT well.

Results of baseline sampling are summarized here and are presented in Figures 5 through 9. The VOC contaminant distribution (TCE and c-DCE) is shown in Figures 5 and 6. For the active treatment cell (Figure 5), concentrations were generally around 1,000 to 3,000 μ g/L for TCE, with other contaminants present at low levels, but concentrations increased significantly at the southern end of the cell. The highest concentration measured anywhere in the ESTCP demonstration area was 140,000 μ g/L at well AMW-6. This is adjacent to a previous chemical oxidation pilot test and was known to be the highest concentration area within the source. The sample collected from the water being extracted from wells AEW-1 and AEW-2 had a TCE concentration of 10,000 μ g/L.

For the passive cell (Figure 6), TCE concentrations were around 1,000 μ g/L at each end of the treatment cell (wells PMW-1 and PMW-9). However, TCE concentrations were much higher in the center of the passive cell (15,000 μ g/L to 63,000 μ g/L). Concentrations of other VOC contaminants were low in all passive cell wells.

Vertical profiles of contaminants in CMT wells are shown in Figures 7 and 8. For the active cell (Figure 7), upper zones generally have low levels of contaminants and also produce very little water when purged. TCE concentrations were approximately 600 to 1,800 μ g/L in middle to lower zones. For the passive cell (Figure 8), TCE concentrations are generally an order of magnitude higher than the active cell; upper zones had TCE concentrations of 1,000 to 10,000 μ g/L, while middle and lower zones had TCE as high as 63,000 μ g/L.

Results for other parameters show that the aquifer is generally mildly reducing with low levels of available carbon. Dissolved oxygen is less than 1 mg/L and ferrous iron is generally less than 0.1 mg/L at all locations. Sulfate is very high at this site, with concentrations ranging from approximately 1,600 mg/L to as high as 8,700 mg/L near the area where the chemical oxidation pilot test was conducted. Methane was detected at some wells up to 230 μ g/L, while COD ranged from non-detect to 100 mg/L. Overall, the pH is near neutral, and ORP ranges from -150 to +300 mV. The only exception to these general trends is well PMW-9, which has relatively high concentrations of methane of 2.8 mg/L, and somewhat depressed sulfate of 1,100 mg/L. While TCE is lower at this location than others in the passive cell, very low concentrations of reductive daughter products are present, and COD is low as well (16 mg/L). This suggests that while redox conditions may be approaching methanogenesis at location, little dechlorination is occurring.

Baseline DNA sampling showed that indigenous *Dehalococcoides* were only detected at low levels at two monitoring locations – the active extraction wells had 448 + 75 cells/L, and the passive cell well PMW-3 had 110 + 28 cells/L. These cell counts are just above the minimum quantification level for the quantitative polymerase chain reaction (qPCR) analysis, and are four to six orders of magnitude lower than what is typically observed following bioaugmentation. Also, it is important to note that the vinyl chloride reductase (vcrA) gene was not detected in any samples. This is important because the vcrA gene was identified during the DNA studies as the proposed "biomarker" that will be used to distinguish the bioaugmentation culture from any indigenous *Dehalococcoides* that grow during the demonstration.

The DNA sampling will be continued throughout the pre-conditioning phase in order to monitor increases in *Dehalococcoides* in response to the lactate injections. Also, monitoring for vcrA will be continued to ensure that this functional gene is not detected even if *Dehalococcoides* increases. If these data indicate that the indigenous strain begins to exhibit the vcrA gene, then a more sophisticated analytical approach that involves sequencing the genes will be considered for future samples to distinguish the inoculated *Dehalococcoides* from the indigenous.

Finally, while the full report containing the baseline compound-specific isotope analyses results is not yet available, preliminary results show that the TCE present near the active extraction wells is "heavier" than in other places. This implies that a mechanism which results in fractionation of TCE (i.e. preferential transformation of the TCE molecules with the "lighter" carbon-12 isotope) is or was active in the past in this area. This is consistent with the fact that this area of the site is near the former chemical oxidation pilot test, because chemical oxidation is known to cause fractionation of TCE, similar to what biodegradation causes. Thus, it appears that the effects of the chemical oxidation are still evident in the isotope signatures at this monitoring location. This should not affect data interpretation for the ER-0513 demonstration because future biodegradation will cause further fractionation of TCE, and will also produce daughter products, whose isotope signatures can then be monitored over time.

Active Cell Tracer Test

A tracer test was performed in the active cell in order to determine hydraulic properties and to confirm travel times from the injection to monitoring wells. The ESTCP Demonstration Plan described that either bromide or iodide would be used as the tracer. Since it was determined that the high chloride concentrations at Site 70 (historically as high as 10,000 mg/L) would cause significant interference with a bromide ion specific electrode, iodide was selected as the tracer. Samples were collected for iodide during the baseline sampling to determine the background response to the iodide probe (all samples were approximately 2-4 mg/L).

Approximately 500 gallons of potassium iodide was injected into the active cell on April 10, 2008. The average concentration of iodide in the injected solution was approximately 13,100 mg/L. Samples for iodide tracer were collected once per day from well AMW-2 for approximately four weeks. Periodic CMT monitoring has been performed for the seven weeks since the tracer injection.

Tracer breakthrough curves are shown in Figure 9 for the active cell tracer test. Tracer breakthrough was observed in AMW-2 (18 ft from injection wells) within 2 weeks. Breakthrough was observed at AMW-4 Zone 2 (28 ft) within approximately 2.5 weeks, Zone 1 (33 ft) within 3 weeks, and Zone 3 (24 ft) within 4 weeks. In addition, tracer breakthrough has occurred in AMW-5 Zone 2 and AMW-3 Zone 3 in approximately five weeks, and initial tracer arrival has occurred in the other ports in these CMT wells. These results show that the lower zones are more transmissive, which is also where the higher contaminant concentrations are found in these wells. The long tail on AMW-2 is likely the result of different tracer arrivals in the various lithologic units.

A preliminary analysis of the tracer test data was performed in order to estimate aquifer properties for the purpose of calculating potential ranges of travel times within the passive cell. The model used was developed for an instantaneous point source (Baetlse, 1969). The analytical equation is found in Domenico and Schwartz (1990, p. 650). A hydraulic conductivity of 10 ft/d was assumed as a starting point based on a pumping test performed in the source area at the site several years ago. An effective porosity of 0.20 was assumed based on CDM's experience with this soil type. A longitudinal dispersivity value equivalent to approximately 10% of the scale of the cell was assumed, and the transverse dispersivity was assumed to be 10% of the longitudinal. The hydraulic gradient used was 0.04 based on water level measurements during pumping. The final variable in this model is distance from the axis (or centerline) of transport. Given the two injection wells in the active cell, this analytical model does not perfectly represent the real system, and the distance from the axis has a questionable meaning. Also, solutions using this model will be nonunique as multiple combinations of the conductivity, effective porosity, and distance from the centerline can produce very similar results. Nevertheless, it is believed that this approach is useful to estimate aquifer properties reasonably, especially given the fact that the hydraulic conductivity has previously been measured by a multiple well pumping test at the site.

Using this approach, inverse modeling was performed to estimate a range of hydraulic conductivities based on matching model predictions to measured iodide breakthrough at several of the monitoring locations. The results of this exercise are shown in Figure 10. For the three active cell monitoring locations shown, the hydraulic conductivity ranged from 5 to 10

ft/d. Thus, the tracer test data could be reasonably matched using hydraulic property values consistent with the soil type and previous hydraulic testing at the site. A somewhat more rigorous semi-analytical model is currently being developed to confirm the expected implications of the estimated aquifer properties for the passive cell.

Based on the estimated values of parameters determined by the tracer test as listed above, travel times from passive cell injection wells to passive cell monitoring wells can be estimated. The most significant factor affecting the travel time is the injection event itself. The target injection volume of 1,000 gallons per well is based on achieving a radius of influence of 5 ft. Therefore, it is assumed that the injected substrate will be distributed 5 ft from the injection point at time zero. Given the range of hydraulic conductivities that were estimated based on the tracer test, along with the measured groundwater elevations presented in Figure 4, groundwater velocity in the passive cell is approximately 4-8 ft/month, or 45-90 ft/yr. This is well within the range of ambient groundwater velocity at other sites where bioremediation and bioaugmentation have been successful, and is in fact two to four times higher than what was originally assumed in the ER-0513 ESTCP Demonstration Plan.

The transport during injection combined with advection under ambient conditions results in travel times from injection wells PIW-1 and PIW-3 to their corresponding monitoring wells ranging from one to three months, assuming conductivity is 10 ft/d. Even if the low estimate of 5 ft/d for conductivity is assumed, travel times from PIW-1 and PIW-3 range from two to five months. Well PIW-2 has a monitoring well located 8 ft away (PMW-6), and another monitoring well located 29 ft away (PMW-7). Depending on the local flow direction in this area, travel times to PMW-6 could be less than one month, while travel times to PMW-7 could be three to seven months.

Pre-conditioning lactate injections and sampling

The initial lactate injection in the active cell was performed on April 23, 2008. Approximately 3,000 gallons was injected at a weight concentration of 1% (i.e. 10,000 mg/L). The initial passive cell lactate injection has not yet been performed, pending resolution of the injection approach with the Remedial Project Manager, the onsite Seal Beach environmental coordinator, and the ESTCP project team.

A monthly sampling event (pre-conditioning monthly event #1) in the active cell wells was performed the week of May 12, 2008. This included the three standard monitoring wells, extraction wells, and one port only from each of the CMT wells. Preliminary results from this sampling round suggest that effects of recirculation are beginning to be observed in the nearest monitoring well AMW-2, in that contaminant profiles and geochemistry are becoming more like

that of the water extracted from AEW-1 and AEW-2. Thus far, effects of the first lactate injection are not evident in this well.

Recommendations

The data collected during field construction, baseline sampling, and tracer testing indicate that meaningful results will be obtained during the 12-month duration of the bioaugmentation portion of the ER-0513 project, allowing for the project objectives to be met Most importantly, the aquifer hydraulics as determined from the tracer test are such that effects of lactate injections and bioaugmentation will be observed at most monitoring wells within three to six months (if not earlier). In addition, VOC concentrations are sufficiently high to support growth of the injected bioaugmentation culture, and the mildly reducing redox conditions can be driven to methanogenesis through the pre-conditioning lactate additions. Finally, the DNA studies and DNA sampling conducted to date suggest that the vcrA functional gene can be used to track the added bioaugmentation culture as planned.

Based on all of these factors, it is recommended that the ER-0513 project be continued as outlined in the ESTCP Demonstration Plan. Pre-conditioning lactate injections will be performed for an additional two months, and a final pre-conditioning sampling event will be conducted to ensure that the vcrA gene has not proliferated prior to bioaugmentation. Also, iodide tracer will be injected into one of the passive cell injection wells in order to confirm the predicted travel times from injection to monitoring wells. The sampling frequency following bioaugmentation is currently planned for monthly, but a recommendation to modify that might be made depending on the sampling results for the pre-conditioning phase. Based on the current schedule of activities, it is anticipated that bioaugmentation will be performed in late July to early August 2008.

Very truly yours,

Joey Trotsky NAVFAC ESC

Kent S. Sorenson, Jr., Ph.D., P.E. Vice President CDM

cc: Ryan A. Wymore, P.E., CDM

Attachment

Figure 1 – Site map

Figure 2 – Active Recirculation Water Levels

Figure 3 – Passive Cell Well Installation

Figure 4 – Actual Water Levels

Figure 5 – Active Cell VOC Concentrations

Figure 6 - Passive Cell VOC Concentrations

Figure 7 – Active Cell Vertical Profiles

Figure 8 - Passive Cell Vertical Profiles

Figure 9 - Active Cell Tracer Breakthrough Curves

Figure 10 - Preliminary Tracer Test Data Analysis

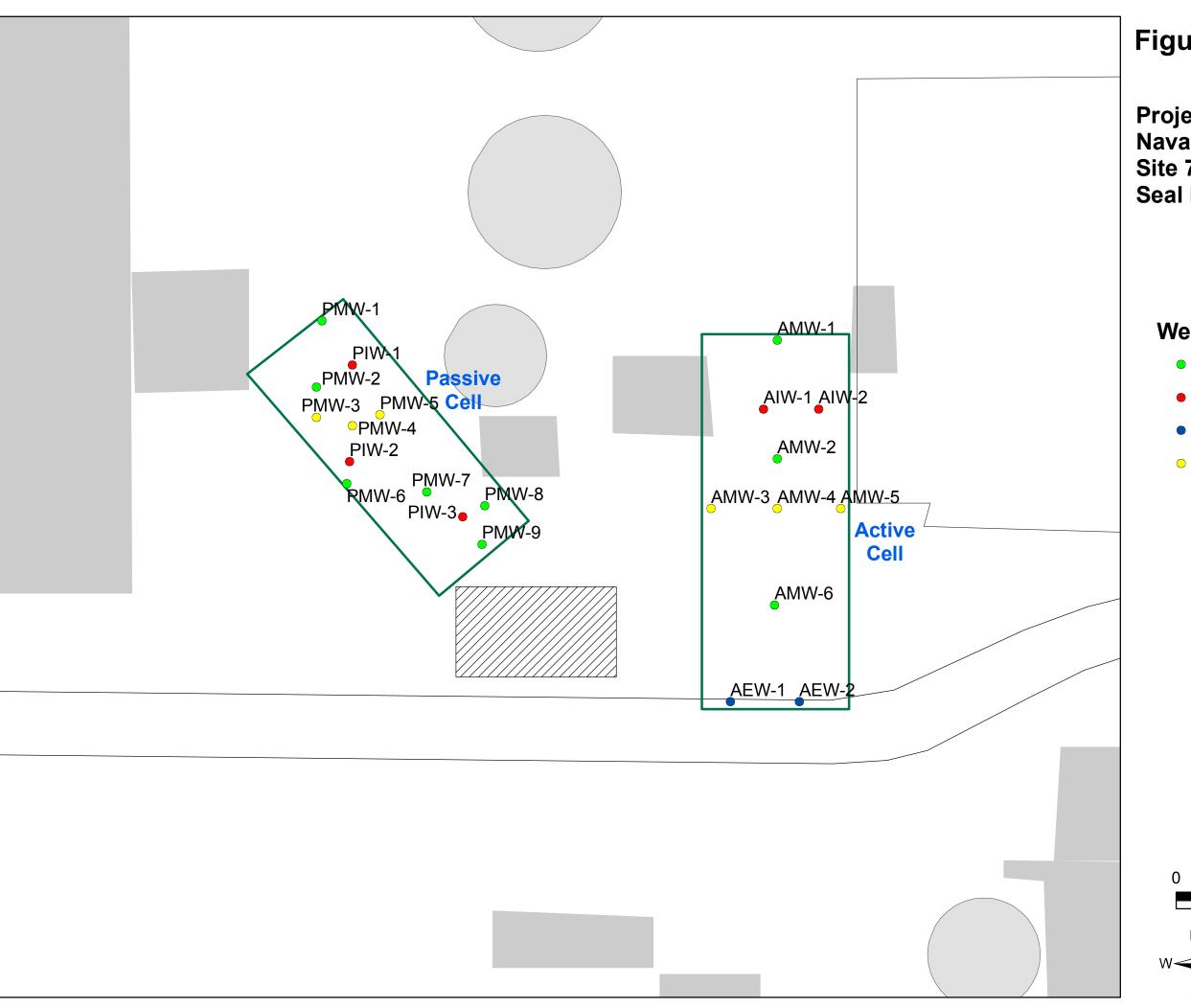
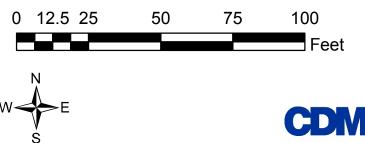


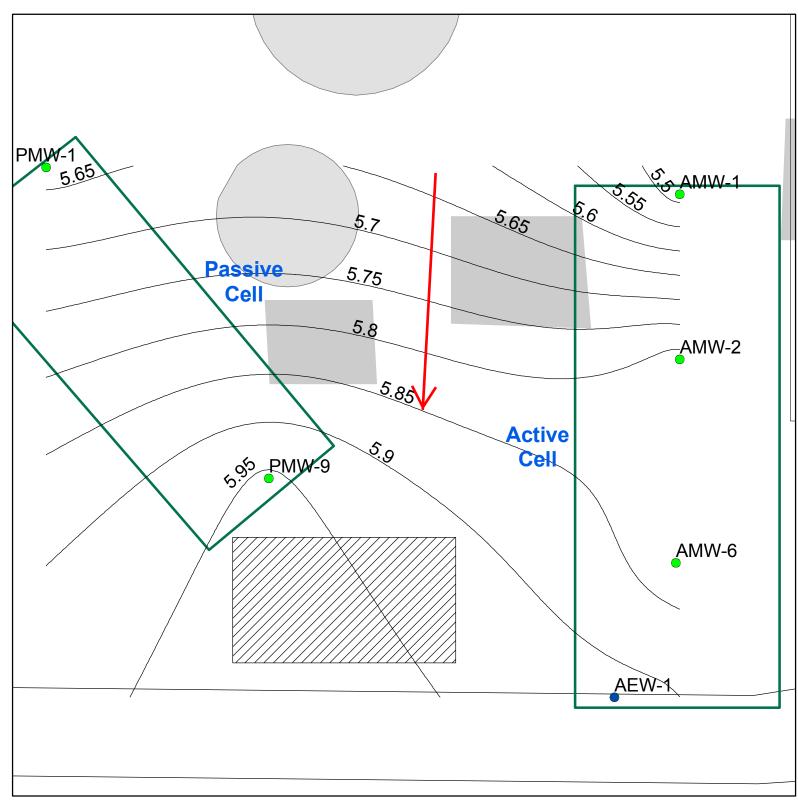
Figure 1 - Site Map

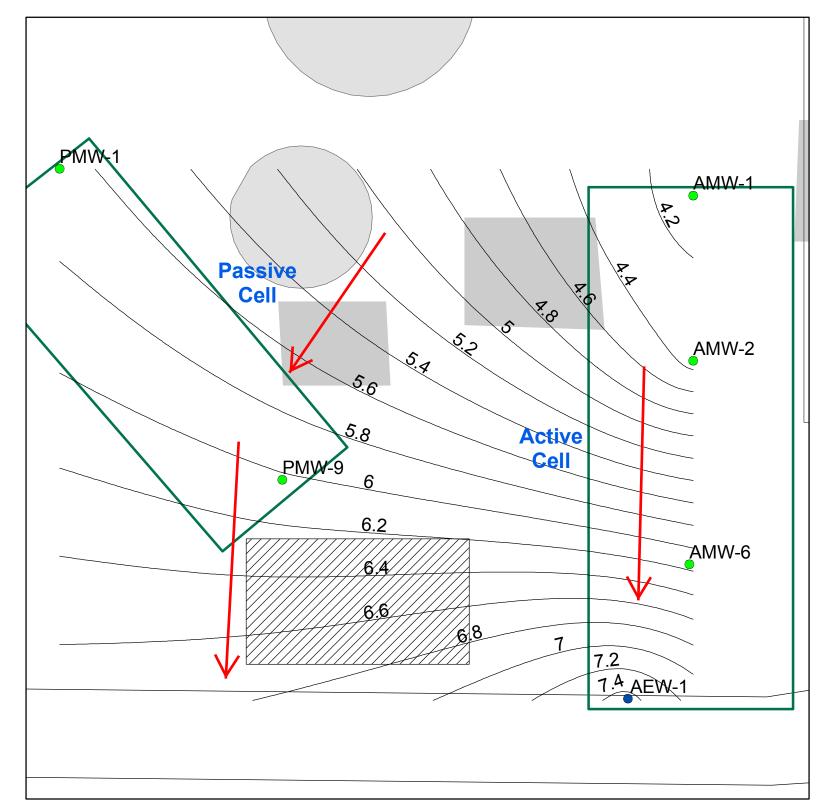
Project ER-0513 Naval Weapons Station Seal Beach Site 70 Seal Beach, California

Well Types

- Monitoring Well
- Injection Well
- Extraction Well
- CMT Well







Ambient Conditions

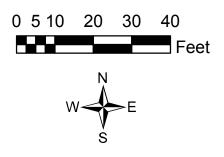
Pumping Conditions

Well Types

- Monitoring Well
- Extraction Well

Figure 2 - Active Recirculation Water Levels (feet below mean sea level)

Project ER-0513
Naval Weapons Station Seal Beach
Site 70
Seal Beach, California





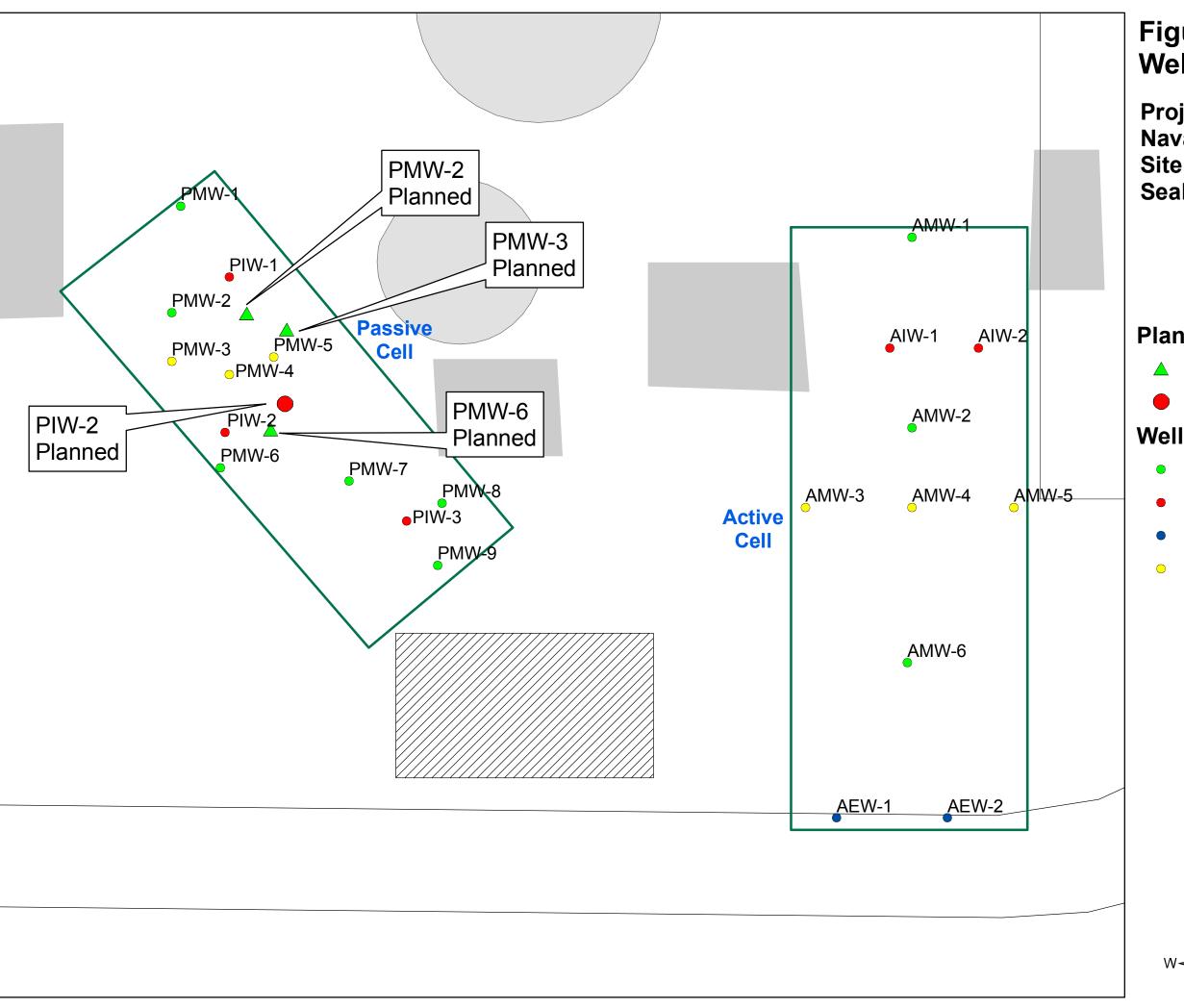


Figure 3 - Passive Cell Well Installation

Project ER-0513 Naval Weapons Station Seal Beach Site 70 Seal Beach, California

Planned Locations of Passive Wells

- Monitoring Well
- Injection Well

Well Types

- Monitoring Well
- Injection Well
- Extraction Well
- CMT Well







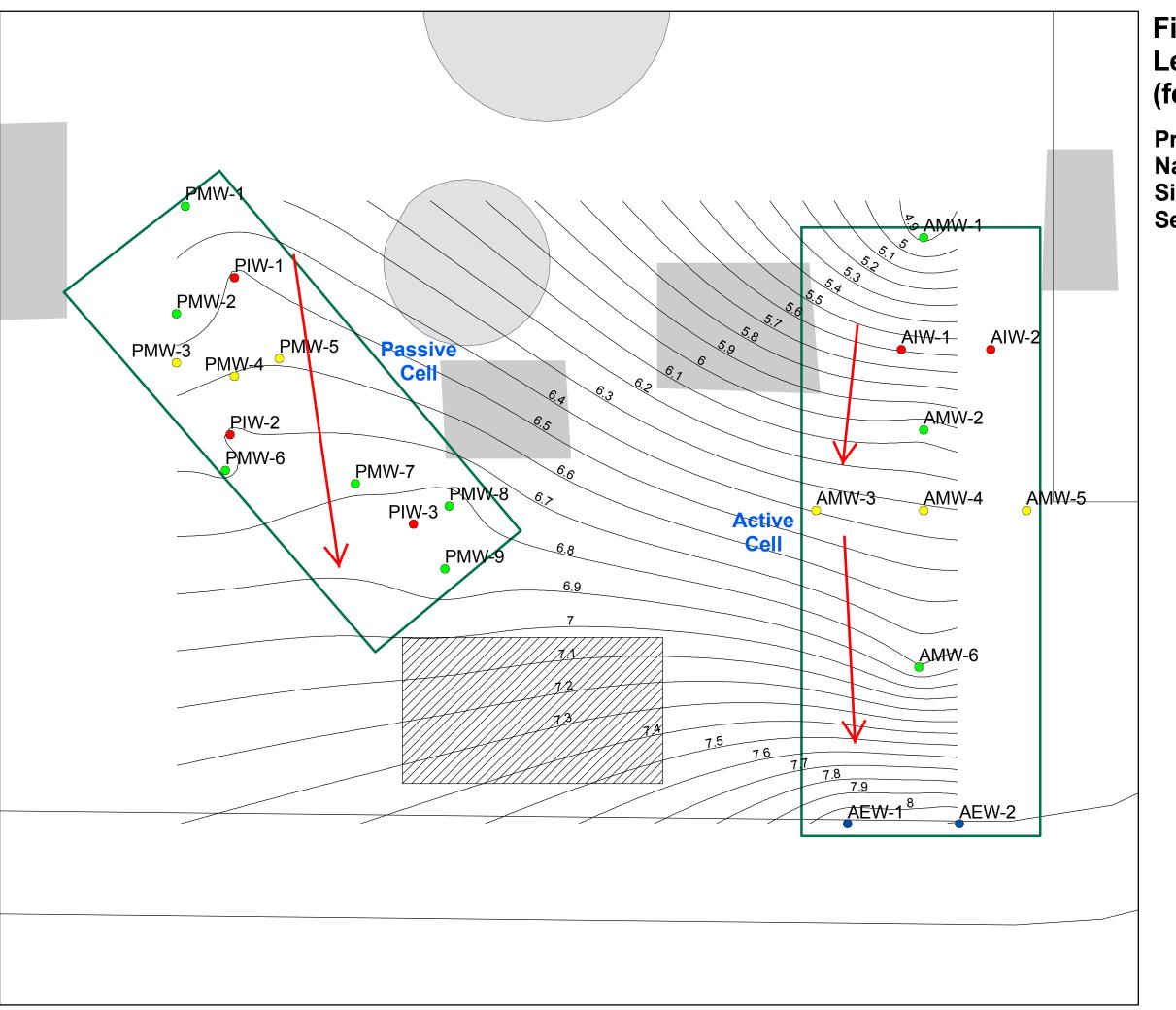
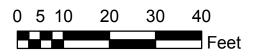


Figure 4 - Actual Water Levels (feet below mean sea level)

Project ER-0513 Naval Weapons Station Seal Beach Site 70 Seal Beach, California

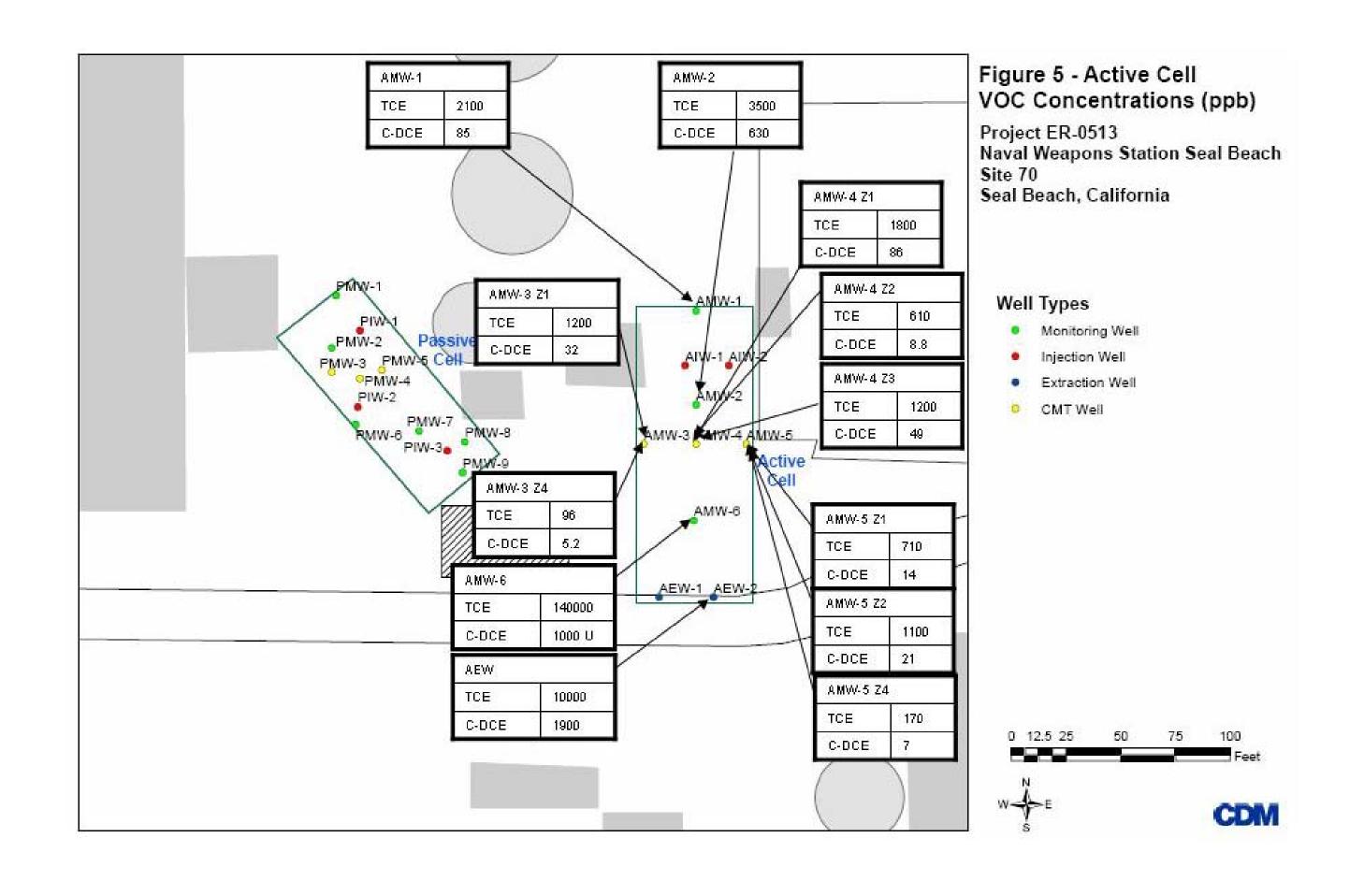
Well Types

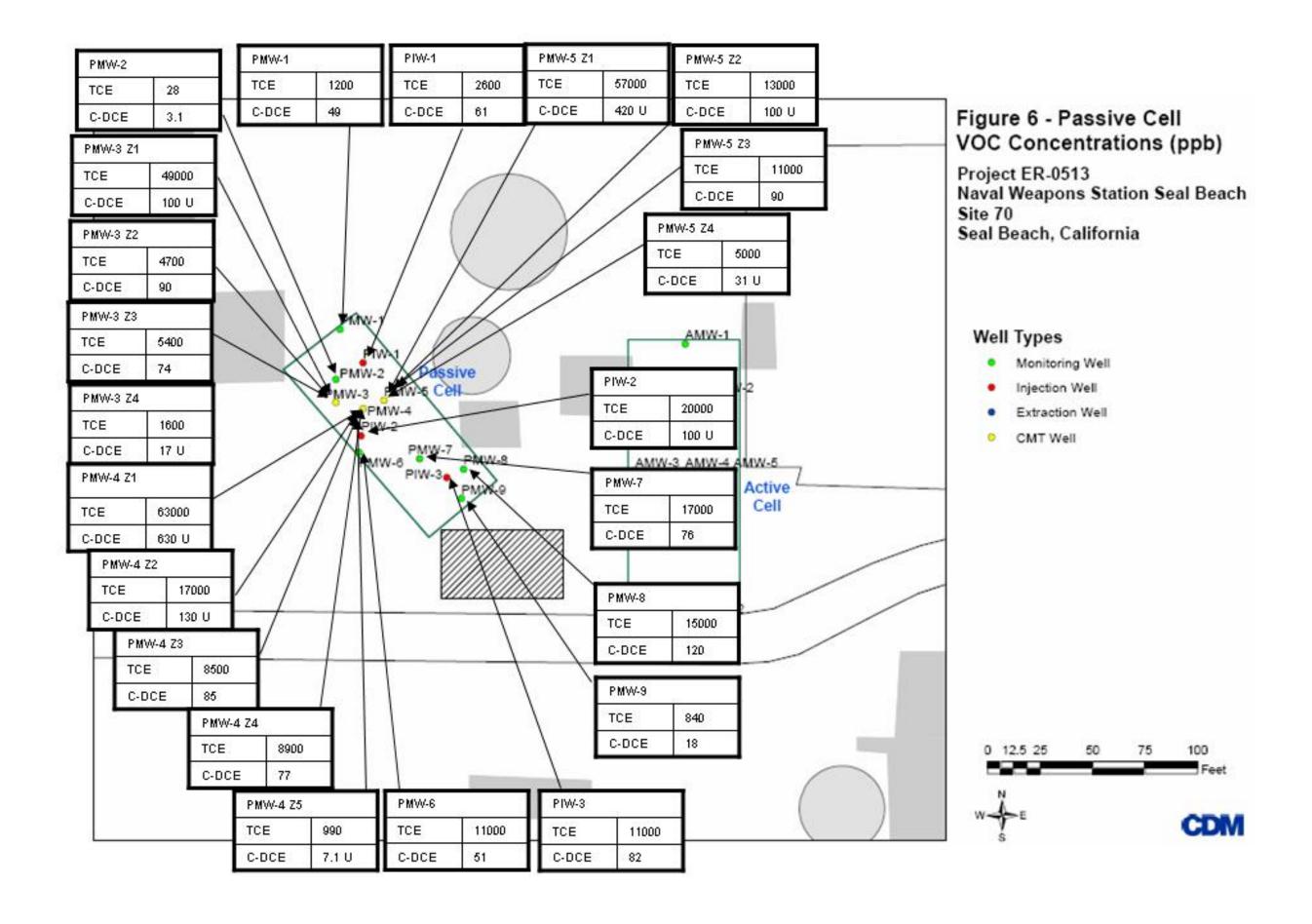
- Monitoring Well
- Injection Well
- Extraction Well
- CMT Well











Active Cell Concentration (ppb)

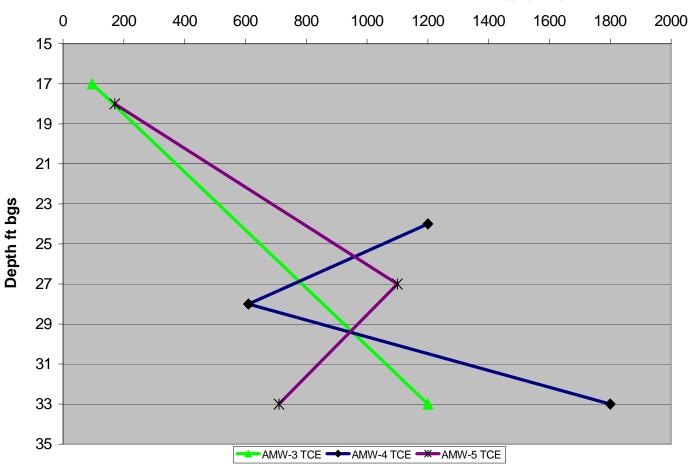


Figure 7 - Active Cell Vertical Profiles

Passive Cell Concentration (ppb)

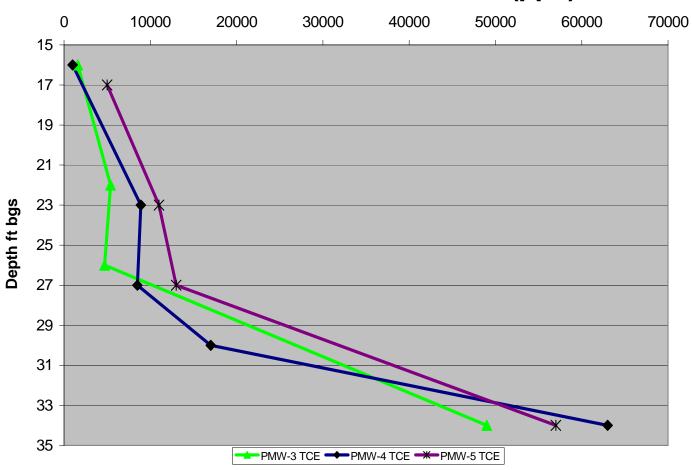


Figure 8 - Passive Cell Vertical Profiles

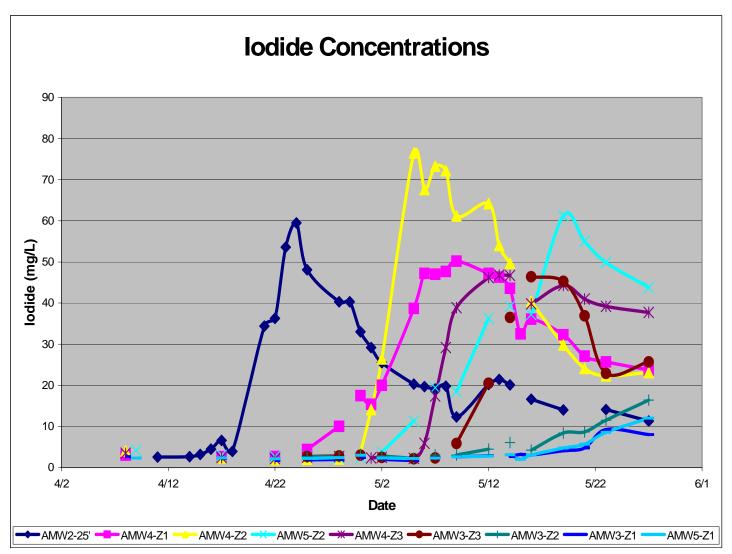


Figure 9 - Active Cell Tracer Breakthrough Curves

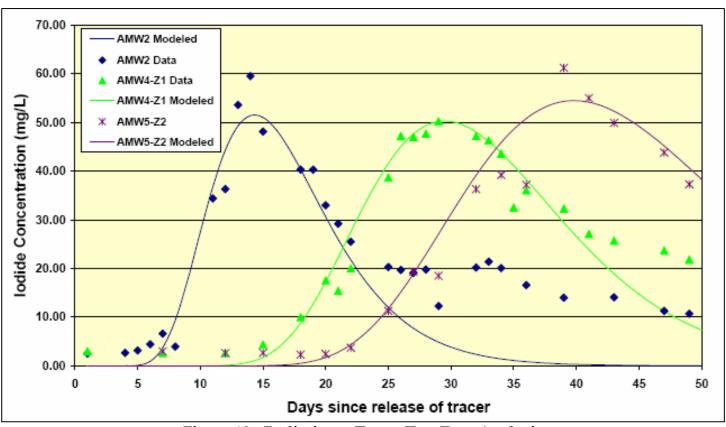


Figure 10 - Preliminary Tracer Test Data Analysis K = 7.5-10 ft/d (pumping test), n = 0.20, dh/dL = 0.04

December 29, 2008

Ms. Andrea Leeson, Ph.D. ESTCP Program Office 901 North Stuart Street, Suite 303 Arlington, VA 22203

Subject: Pre-Conditioning Results for ER-0513

This White Paper presents results of the "pre-conditioning" phase for Environmental Security Technology Certification Program (ESTCP) project ER-0513, with the intent of documenting that conditions are appropriate for bioaugmentation, as directed by the ESTCP program office in an email dated August 5, 2008. This project is being conducted at Naval Weapons Station Seal Beach, Site 70. The purpose of this demonstration is to compare the low-cost, passive approach for bioaugmentation to the more common recirculation approaches for full-scale TCE source area application.

Project field work began in February 2008 with construction of the active recirculation treatment cell. This was followed by the initiation of the "pre-conditioning" phase during which electron donor was added to both the active and passive treatment cells in order to establish appropriate reducing conditions in the aquifer prior to bioaugmentation. The active recirculation cell extracts and reinjects groundwater continuously, and electron donor (1% to 3% sodium lactate) is being pulsed into the reinjection line periodically. To date, three active cell injections have been performed from late April to mid-October 2008. For the passive treatment cell, sodium lactate was injected into each of three injection wells once per month between August and October 2008, with the injection concentration and electron donor mass being the same for both treatment cells. Groundwater conditions were monitored following each injection event during the pre-conditioning phase, in order to determine when sufficiently reducing conditions were achieved. In the June 9, 2008 white paper submitted to ESTCP, these conditions were defined as ferrous iron concentrations greater than 0.5 mg/L and a decrease in sulfate of at least 10% from baseline. Once conditions are shown to be sufficiently reducing, the treatment cells will be bioaugmented using a commercially available bioaugmentation culture (Shaw's SDC-9).

Pre-conditioning lactate injections and sampling

The initial lactate injection in the active cell was performed on April 23, 2008. Two additional injections were conducted on July 17, 2008 and October 17, 2008. Approximately 3,000 gallons

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were injected at a weight concentration of approximately 1% (i.e., 10,000 mg/L) as lactate. The initial passive injection was performed on August 6, 2008. Two additional injections were conducted on September 8, 2008 and October 21, 2008. Approximately 3,200 gallons (1,066 gallons per well) were injected at a weight concentration of 1% during each event.

Baseline sampling as completed for the active cell the week of April 7, 2008 and for the passive cell the week of April 21, 2008. Well details are shown in Table 1, and well locations are shown in Figure 1. The baseline sampling included sampling the three standard monitoring wells, all ports below the water table in the three CMT wells, and the water being produced from the extraction wells (refer to Figure 1 for well locations). Sampling was also conducted in September, October, and November 2008 to monitor groundwater conditions during pre-conditioning. The September and October events included sampling the same wells as the baseline event, except only the deepest zones (Zone 1) in the CMT wells were sampled. The November 2008 event was the final sampling event during pre-conditioning and included all the wells (and zones) included in the baseline event.

All active cell sampling events were conducted with the active cell recirculation system operating. Analytes sampled during all events included volatile organic compounds (VOCs), ethene/ethane/methane, anions (sulfate, chloride, and nitrate/nitrite), alkalinity, chemical oxygen demand (COD), and DNA samples. During the baseline and final sampling events, stable carbon isotope analysis was also performed.

Active Recirculation Cell Results

Electron Donor

Electron donor results as chemical oxygen demand (COD) are shown in Table 2. In general, COD concentrations did not increase significantly during pre-conditioning activities in the active cell. Given that donor injections were conducted approximately 6-8 weeks apart with continuous recirculation being conducted throughout this time, it is believed that the lactate may have been diluted and "washed out" from the monitoring wells. Because of this, smaller, more frequent injections will be performed during the bioaugmentation phase. Despite this observation in the monitoring wells, the redox data and VOC results clearly show that the lactate injections have had positive impacts in the active cell nearer the injection wells, in terms of driving conditions to be appropriate for bioaugmentation (see below).

Redox Parameters

Redox parameter results are also shown in Table 2. Ferrous iron was not detected at any wells during baseline sampling except for in the deepest zone (Z1) of AMW-4 and AMW-5.

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However, ferrous iron concentrations increased by November 2008 to greater than 3 mg/L at AMW-2, which is the closest downgradient well to the injection wells. Also, ferrous iron concentrations increased to above 0.5 mg/L at all three of the CMT wells further downgradient.

Sulfate concentrations decreased over 65% from 7,400 mg/L to 2,600 mg/L at AMW-2. Sulfate concentrations also decreased more than 20% from baseline conditions at AMW-3 Z1 (24%), AMW-4 Z2 (52%), AMW-4 Z3 (53%) and at AMW-5 Z2 (46%), and upgradient well AMW-1 (38%). Sulfate concentrations did increase at the deepest zone of AMW-5 Z1 from 3,600 mg/L to 4,900 mg/L. Sulfate concentrations remained relatively stable at AMW-4 Z1 and AMW-6.

Other electron acceptors nitrate and methane were also analyzed. Nitrate was not detected at any well during the final pre-conditioning sampling event. Methane concentrations were also below 50 μ g/L at all wells except the extraction points.

Overall, these results show that redox conditions in the active cell at wells near the injection points are iron- to sulfate-reducing, which is appropriate for bioaugmentation. While the entire active cell is not yet at the appropriate redox conditions, it is only a requirement for the portion of the aquifer where the culture will be injected to have the appropriate redox conditions. The remainder of the active cell will achieve the appropriate conditions as the bioaugmentation phase progresses.

Contaminants and Degradation Products

Results of baseline (April 2008) and final pre-conditioning (November 2008) sampling events are summarized and are presented in Table 3 for VOC compounds. Trichloroethene (TCE) concentrations were generally around 1,000 to 3,000 μ g/L during baseline sampling, with the exception of the extraction wells (10,000 μ g/L) and well AMW-6 (140,000 μ g/L); other contaminants were present at low levels. The November 2008 final pre-conditioning VOC contaminant distribution is shown in Figure 2 for tetrachloroethene (PCE), TCE, and cis-1,2-dichloroethene (c-DCE). In general, TCE concentrations were higher than baseline in all wells except AMW-6 (decrease from 140,000 μ g/L to 120,000 μ g/L) and AMW-2 (3,500 μ g/L to 1,300 μ g/L). This is due to the fact that the recirculation system is extracting groundwater with higher TCE concentrations and reinjecting it upgradient. The highest TCE concentration measured anywhere in the ESTCP demonstration area remained at well AMW-6 (120,000 μ g/L). Vinyl chloride (VC) was not detected at any wells except for AMW-2 (35 μ g/L). The sample collected from the water being extracted from wells AEW-1 and AEW-2 had a TCE

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concentration of 35,000 μ g/L, higher than the baseline concentration of 10,000 μ g/L. The c-DCE concentration remained stable at 1,700 μ g/L.

The VOCs present at AMW-2 during the November 2008 round are significant in that they show partial dechlorination is already occurring. The results of this event show that significant c-DCE is present at this location, as well as low levels of VC. This is consistent with the other data from this well, which show that conditions are sulfate-reducing.

Vertical profiles of primary contaminant TCE in active cell CMT wells are shown in Figure 3 for April 2008 and November 2008 sampling events. Under baseline conditions, the vertical profile for all 3 CMT wells showed lower overall TCE concentrations and an increase in TCE concentrations with depth. In November 2008, upper zones generally had similar TCE concentrations as deeper zones, probably due to the recirculation. TCE concentrations ranged from 520 to $10,000~\mu g/L$ in the three active zone CMT wells.

DNA Results

DNA analysis results using quantitative polymerase chain reaction (qPCR) are provided in Table 4. These results show that indigenous Dehalococcoides were only detected at low levels at two monitoring locations (AMW-2 and AMW-4 Z1) and the extraction wells – the monitoring wells had up to $3.4 \times 10^3 \pm 810$ cells/L and the extraction wells had $1.1 \times 10^4 \pm 5300$ cells/L. Although these cell counts are higher than the baseline concentrations, some increase in Dehalococcoides was expected after lactate injection. However, it is important to note that the vinyl chloride reductase (vcrA) gene was not detected in any samples. This is key because the vcrA gene was identified during the DNA studies as the proposed "biomarker" that will be used to distinguish the bioaugmentation culture from any indigenous Dehalococcoides that grow during the demonstration.

Active Cell Summary

Active cell results indicate that appropriate conditions have been achieved for successful bioaugmentation, particularly in wells near the reinjection locations. Ferrous iron increases were observed to above $0.5~\rm mg/L$ in all wells except AMW-6 and upgradient well AMW-1. Also, sulfate concentrations decreased more than 10% except in AMW-6 and the extraction wells. While COD concentrations did not increase above $60~\rm mg/L$ in any active cell well, the significantly increased c-DCE concentration at AMW-2 and other wells indicates that partial dechlorination is occurring near the injection wells.

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Passive Cell Results

Electron Donor

Electron donor (COD) results are shown from the baseline and final pre-conditioning events in Table 2. Baseline conditions showed that COD was at or below 100 mg/L throughout the passive cell. In November 2008, COD concentrations increased to above 1,000 mg/L in wells PMW-6, PMW-7, PIW-2, and PIW-3 and increased to near or above 100 mg/L in wells PMW-2, PMW-3 Z1, and PIW-1. COD only decreased slightly in wells PMW-1, PMW-9, and in the upper zones of all three CMT wells. These results indicate that donor has increased significantly in the areas surrounding the injection wells throughout the passive cell.

Redox Parameters

Electron acceptor results are also shown in Table 2. Ferrous iron was not detected at any wells during baseline sampling except for in the upper zone of PMW-5. The November 2008 results show that ferrous iron concentrations increased to above 0.5 mg/L at PMW-2, PMW-6, and PMW-8, which are the closest downgradient wells to injection wells PIW-1, PIW-2, and PIW-3, respectively. Also, ferrous iron concentrations increased to above 0.5 mg/L for at least one zone of all three CMT wells further downgradient.

Baseline sulfate concentrations were high in the passive cell, ranging from 1,100 mg/L in PMW-9 to 5,800 mg/L in PMW-5 Z3. Following the lactate injections, sulfate concentrations decreased from baseline conditions between 35% and 99% in the three injection wells. Sulfate concentrations also decreased more than 10% from baseline conditions in PMW-7 (13%) and PMW-8 (21%), while remaining relatively stable in the three CMT wells and PMW-6. Sulfate concentrations did increase over 100% in wells PMW-2 and PMW-9. Also, upgradient well PMW-1 increased in sulfate concentration from baseline by 24%.

Other electron acceptors nitrate and methane were also analyzed. Nitrate was not detected in any well during the final pre-conditioning sampling event except upgradient well PMW-1 (0.72 mg/L). Methane concentrations were above 0.1 mg/L in all monitoring wells except AMW-1 and AMW-2.

Overall, these redox conditions show that most of the passive cell wells are iron- to sulfate-reducing, and possibly even methanogenic based on methane concentrations of greater than 200 $\mu g/L$ at some wells. These results indicate that conditions are appropriate for bioaugmentation in the passive cell.

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Contaminants and Degradation Products

Passive sampling VOC results are summarized and are presented in Table 3 for the baseline and final pre-conditioning sampling events. The VOC contaminant distribution is shown in Figure 2 for PCE, TCE, and c-DCE. During the baseline event, TCE concentrations were approximately 1,000 μ g/L at each end of the treatment cell (wells PMW-1 and PMW-9). However, TCE concentrations were much higher in the center of the passive cell (15,000 μ g/L to 63,000 μ g/L). Concentrations of other VOC contaminants were low in all passive cell wells.

The results indicate that TCE concentrations were similar to baseline in all wells except the injection wells, which all decreased two orders of magnitude, and PMW-2, which increased from 28 μ g/L to 1,600 μ g/L. The highest concentration of TCE was still in the center of the cell, with concentrations in the lowest zone of the three CMT wells, ranging from 37,000 μ g/L in PMW-5 to 60,000 μ g/L in PMW-3. As opposed to the active recirculation cell, concentrations of degradation product c-DCE did not increase significantly from baseline conditions. No vinyl chloride was detected in the passive cell.

Vertical profiles of TCE in passive cell CMT wells are shown in Figure 4 for April 2008 and November 2008 sampling events. For the passive cell, TCE concentrations are generally an order of magnitude higher in the lower zone (Z1) than the upper zone (Z3-Z4) in all wells; upper zones had TCE concentrations of 4,800 to 9,100 μ g/L, while lower zones had TCE as high as 63,000 μ g/L. This profile is similar to the profile observed during baseline conditions.

DNA Results

DNA results (Table 4) show that indigenous *Dehalococcoides* were not detected in any wells in the passive cell during the November 2008 sampling event, including functional gene *vcrA*. This is important because the *vcrA* gene was identified during the DNA studies as the proposed "biomarker" that will be used to distinguish the bioaugmentation culture from any indigenous *Dehalococcoides* that grow during the demonstration.

Passive Cell Summary

Passive cell results indicate that conditions are becoming more reducing, with the most positive results observed near the injection wells. In these wells, ferrous iron increased to above 0.5 mg/L and sulfate decreased more than 10% except in PMW-2 and PMW-6. COD increased significantly at wells near the injection points also, and significant COD still remains at two of the three injection wells. This indicates that sufficient electron donor is being supplied for bioaugmentation.

Ms. Andrea Leeson, Ph.D. December 29, 2008 Page 7

Recommendations

The data collected during the pre-conditioning phase indicate conditions at and near the injection wells are appropriate for bioaugmentation . Electron acceptor results in both cells show that ferrous iron concentrations have generally increased to above 0.5 mg/L, with higher concentrations observed closer to the injection wells. Additionally, sulfate concentrations generally decreased over 10% near the injection wells in both cells from baseline conditions, indicating that the lactate additions are making the subsurface more reducing. The active recirculation cell results indicate that increased dechlorination is occurring following the lactate injections, but dechlorination beyond c-DCE has not generally been observed.

Most importantly, the DNA results indicate that low populations of Dehalococcoides are present in the treatment cells as expected, but that the vcrA gene has not been detected anywhere. This indicates that the vcrA functional gene can be used to track the added bioaugmentation culture as planned.

Based on all of these factors, it is recommended that bioaugmentation be performed in early January in both the active and passive treatment cells using the commercially available culture SDC-9. Please provide us with confirmation that we can move forward with bioaugmentation as planned.

Very truly yours,

Joey Trotsky Kent S. Sorenson, Jr., Ph.D., P.E.

NAVFAC ESC Vice President

Camp Dresser & McKee Inc.

cc: Ryan A. Wymore, P.E., CDM

Attachments

Figures

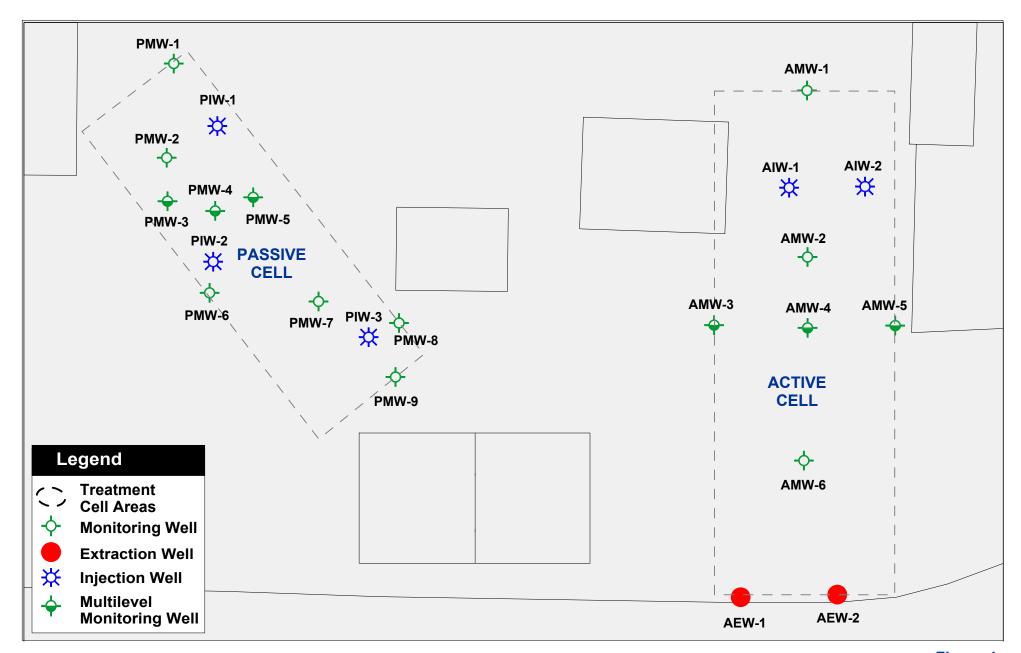
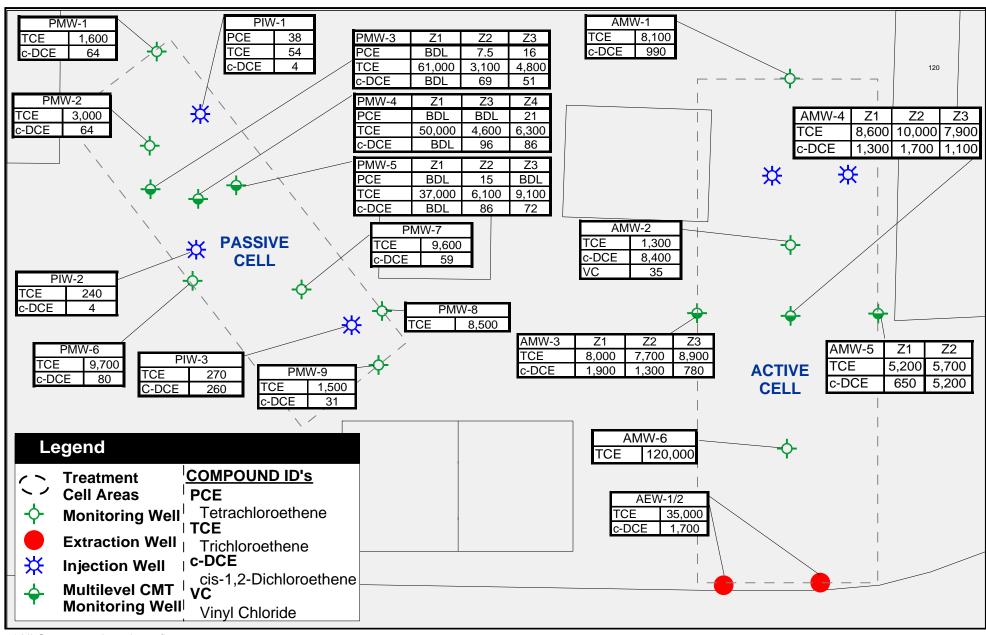






Figure 1 Site Map



*All Concentrations in ug/L
Only detected compounds are displayed.
BDL - Below detection limit

Scale in Feet

12.5 Feet 0 25 Feet



Figure 2 November 2008 VOC Concentrations

CMT Well TCE Concentrations (ug/L)

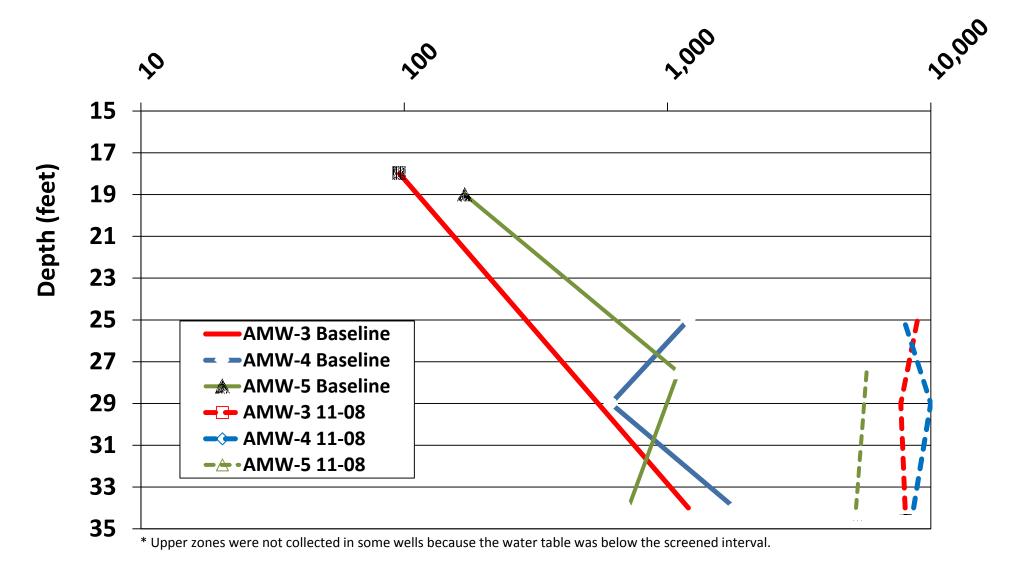


Figure 3
Active Cell Vertical TCE Profiles



CMT Well TCE Concentrations (ug/L)

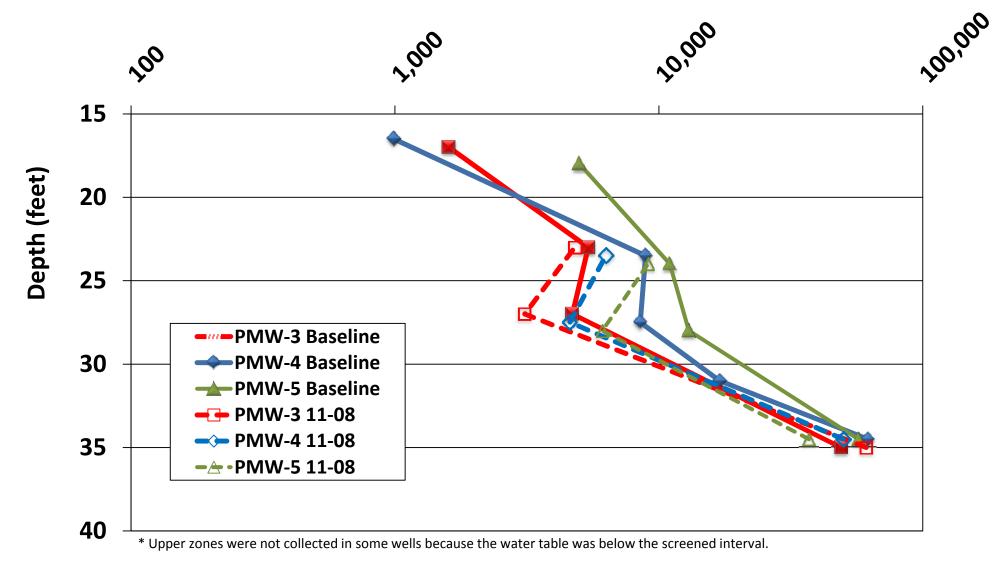


Figure 4
Passive Cell Vertical TCE Profiles



Tables

Well ID	Well Type	Screen Interval (ft bgs)
Ac	tive Recircula	tion Cell
AMW-1	Monitoring	15.1-35.1
AMW-2	Monitoring	15-35
AMW-3	CMT Z1	33-34
AMW-3	CMT Z2	28-29
AMW-3	CMT Z3	24-25
AMW-3	CMT Z4	17-18
AMW-4	CMT Z1	33-34
AMW-4	CMT Z2	28-29
AMW-4	CMT Z3	24-25
AMW-4	CMT Z4	18-19
AMW-5	CMT Z1	33-34
AMW-5	CMT Z2	26.5-27.5
AMW-5	CMT Z3	22-23
AMW-5	CMT Z4	18-19
AMW-6	Monitoring	15.5-35.5
AEW-1	Extraction	14.7-34.7
AEW-2	Extraction	15.3-35.3
AIW-1	Injection	15.6-35.6
AIW-2	Injection	15.5-35.5

Well ID	Well Type	Screen Interval (ft bgs)
	Passive C	
PMW-1	Monitoring	15.3-35.3
PMW-2	Monitoring	15-35
PMW-3	CMT Z1	34-35
PMW-3	CMT Z2	26-27
PMW-3	CMT Z3	22-23
PMW-3	CMT Z4	16-17
PMW-4	CMT Z1	33.5-34.5
PMW-4	CMT Z2	30-31
PMW-4	CMT Z3	26.5-27.5
PMW-4	CMT Z4	22.5-23.5
PMW-4	CMT Z5	15.5-16.5
PMW-5	CMT Z1	33.5-34.5
PMW-5	CMT Z2	27-28
PMW-5	CMT Z3	23-24
PMW-5	CMT Z4	17-18
PMW-6	Monitoring	15-35
PMW-7	Monitoring	15-35
PMW-8	Monitoring	15-35
PMW-9	Monitoring	15-35
PIW-1	Injection	15-35
PIW-2	Injection	15-35
PIW-3	Injection	15-35

CMT- Continuous Multichannel Tubing bgs - below ground surface

Sample	Well Type	Nitrate (r		ate (mg/L)		Ferrous Iron (mg/L)			Sulfate (mg/L)			ethane (ug	ı/L)	COD (mg/L)		
Location	Well Type	4/2008	11/2008	% Diff	4/2008	11/2008	% Diff	4/2008	11/2008	% Diff	4/2008	11/2008	% Diff	4/2008	11/2008	% Diff
Active Recirculation Cell																
AMW-1	Monitoring	0.89	0	-100%	0	0	NA	8,700	5,400	-38%	<5	<5	NA	34	32	-6%
AMW-2	Monitoring	0	0	NA	0	3.3	NA	7,400	2,600	-65%	<5	6	NA	40	47	18%
AMW-3	CMT Z1	0.21	0	-100%	0	0	NA	7,900	6,000	-24%	20	9	-55%	60	57	-5%
AMW-3	CMT Z2	NS	0	NA	NS	0	NA	NS	4,900	NA	NS	8	NA	NS	40	NA
AMW-3	CMT Z3	NS	0	NA	NS	0.70	NA	NS	3,700	NA	NS	13	NA	NS	38	NA
AMW-4	CMT Z1	0.14	0	-100%	0.17	3.10	1724%	6,300	5,900	-6%	21	9	-57%	48	47	-2%
AMW-4	CMT Z2	0.13	0	-100%	0	1.42	NA	6,900	3,300	-52%	41	10	-76%	44	36	-18%
AMW-4	CMT Z3	0	0	NA	0	0.37	NA	7,000	3,300	-53%	19	<5	NA	38	38	0%
AMW-5	CMT Z1	0.16	0	-100%	0.24	0	-100%	3,600	4,900	36%	28	14	-50%	42	47	12%
AMW-5	CMT Z2	0.18	0	-100%	0	3.23	NA	7,100	3,800	-46%	48	13	-73%	40	42	5%
AMW-6	Monitoring	0.35	0	-100%	0	0	NA	3,300	3,300	0%	40	33	-18%	58	47	-19%
AEW	Extraction	0.14	0	-14%	0	0	NA	1,600	1,500	-6%	140	100	-29%	28	34	21%
Passive Cell																
PMW-1	Monitoring	0.53	0.72	37%	0	0	NA	3,800	4,700	24%	360	14	-96%	28	25	-11%
PMW-2	Monitoring	0.04	0	-100%	0	2.19	NA	1,600	5,100	219%	2,300	71	-97%	18	120	567%
PMW-3	CMT Z1	0.03	0	-100%	0	1.92	NA	2,000	2,100	5%	220	220	0%	64	170	166%
PMW-3	CMT Z2	0.04	0	-100%	0	0.18	NA	4,200	3,800	-10%	80	86	8%	67	30	-55%
PMW-3	CMT Z3	0	0	NA	0	1.18	NA	3,900	4,400	13%	160	98	-39%	100	68	-32%
PMW-4	CMT Z1	0.09	0	-100%	0	0.62	NA	2,000	2,000	0%	180	290	61%	58	74	28%
PMW-4	CMT Z3	0	0	NA	0	0.10	NA	5,600	5,100	-9%	90	75	-17%	79	53	-33%
PMW-4	CMT Z4	0	0	NA	0	0.12	NA	5,000	4,400	-12%	190	130	-32%	68	57	-16%
PMW-5	CMT Z1	0.57	0	-100%	0	0	NA	2,100	2,200	5%	130	270	108%	38	44	16%
PMW-5	CMT Z2	0	0	NA	0	0.09	NA	5,700	6,000	5%	60	57	-5%	100	95	-5%
PMW-5	CMT Z3	0	0	NA	0.02	0.70	4567%	5,800	5,700	-2%	70	83	19%	87	83	-5%
PMW-6	Monitoring	0.10	0	-100%	0	0.99	NA	3,000	3,300	10%	170	130	-24%	56	78	39%
PMW-7	Monitoring	0.03	0	-100%	0	1.94	NA	3,000	2,600	-13%	210	140	-33%	50	1,100	2100%
PMW-8	Monitoring	0	0	NA	0	3.07	NA	2,400	1,900	-21%	430	150	-65%	46	1,400	2943%
PMW-9	Monitoring	0.01	0	-100%	0	0	NA	1,100	3,000	173%	2,800	370	-87%	16	13	-19%
PIW-1	Injection	0.11	0	-100%	0	0.02	NA	3,400	2,200	-35%	15	94	527%	28	99	254%
PIW-2	Injection	0.13	0	-100%	0	2.92	NA	3,900	600	-85%	230	6	-97%	71	4,900	6801%
PIW-3	Injection	0	0	NA	0	3.30	NA	3,100	15	-100%	150	14	-91%	30	5,700	18900%

NS - Not sampled during this event

NA - Percent difference not calculated

CMT - Continuous multichannel tubing

'<' - Below the reporting limit

Z1 is the deepest channel of each CMT well, Z3 is the shallowest. Not all channels were able to be sampled because thewater level was below the bottom of the channel.

Table 2 Electron Acceptor and Donor Results

Sample			CE	TO			,2-DCE	Vinyl Chloride		
Location	Well Type	(ug/L)			/L)		ug/L)		g/L)	
		4/2008	11/2008	4/2008	11/2008	4/2008	11/2008	4/2008	11/2008	
	T			/e Recircu				T		
AMW-1	Monitoring	BDL	BDL	2,100	8,100	83	990	BDL	BDL	
AMW-2	Monitoring	BDL	BDL	3,450	1,300	630	8,400	BDL	35	
AMW-3	CMT Z1	BDL	BDL	1,200	8,000	32	1,900	BDL	BDL	
AMW-3	CMT Z2	NS	BDL	NS	7,700	NS	1,300	NS	BDL	
AMW-3	CMT Z3	NS	BDL	NS	8,900	NS	780	NS	BDL	
AMW-4	CMT Z1	BDL	BDL	1,800	8,600	86	1,300	BDL	BDL	
AMW-4	CMT Z2	BDL	BDL	610	10,000	9	1,700	BDL	BDL	
AMW-4	CMT Z3	BDL	BDL	1,200	7,900	49	1,100	BDL	BDL	
AMW-5	CMT Z1	BDL	BDL	710	5,200	14	650	BDL	BDL	
AMW-5	CMT Z2	BDL	BDL	1,100	5,700	21	5,200	BDL	BDL	
AMW-6	Monitoring	BDL	BDL	140,000	120,000	BDL	BDL	BDL	BDL	
AEW	Extraction	BDL	BDL	10,000	35,000	1,900	1,700	BDL	BDL	
				Passive	Cell					
PMW-1	Monitoring	19	BDL	1,150	1,600	49	64	BDL	BDL	
PMW-2	Monitoring	33	BDL	28	3,000	3	65	BDL	BDL	
PMW-3	CMT Z1	BDL	BDL	49,000	61,000	BDL	BDL	BDL	BDL	
PMW-3	CMT Z2	BDL	7.5	4,700	3,100	90	69	BDL	BDL	
PMW-3	CMT Z3	20	16	5,400	4,800	98	51	BDL	BDL	
PMW-4	CMT Z1	BDL	BDL	62,000	50,000	BDL	BDL	BDL	BDL	
PMW-4	CMT Z3	BDL	BDL	8,500	4,600	85	96	BDL	BDL	
PMW-4	CMT Z4	BDL	21	8,900	6,300	77	86	BDL	BDL	
PMW-5	CMT Z1	BDL	BDL	57,000	37,000	BDL	BDL	BDL	BDL	
PMW-5	CMT Z2	BDL	15	13,000	6,100	BDL	86	BDL	BDL	
PMW-5	CMT Z3	BDL	BDL	11,000	9,100	90	72	BDL	BDL	
PMW-6	Monitoring	BDL	BDL	11,000	9,700	51	80	BDL	BDL	
PMW-7	Monitoring	BDL	BDL	17,000	9,600	76	59	BDL	BDL	
PMW-8	Monitoring	BDL	BDL	15,000	8,500	120	BDL	BDL	BDL	
PMW-9	Monitoring	BDL	BDL	840	1,500	18	31	BDL	BDL	
PIW-1	Injection	BDL	38	2,600	54	61.0	3.6	BDL	BDL	
PIW-2	Injection	BDL	BDL	20,000	240	BDL	3.9	BDL	BDL	
PIW-3	Injection	BDL	BDL	11,000	270	82	260	BDL	BDL	

PCE - tetrachloroethene

TCE - trichloroethene

c-DCE - cis-1,2-dichloroethene

CMT - Continuous Multichannel Tubing

ug/L - micrograms per liter

BDL - below detection limits

NS - Not Sampled

	DNA	Universal	Dehalococcoides			Dehalococcoides			Dehalo	сос	coides	Dehalococcoides		
	DIVA	PCR#	16S rRNA				tceA			ovcA	1	vcrA		
Sample ID	ng/L groundwater		copy/L groundwater*			copy/L g	jroui	ndwater*	copy/L groundwater*			copy/L		
ACTIVE RECIRCULATION CELL														
AMW1	1028	+		ND			ND			ND		ND		
AMW2	4715	+	3.36E+03	±	8.10E+02	2.36E+03	±	4.70E+02	4.60E+02	±	4.70E+02	ND		
AMW3-Z1	417	+		ND			ND			ND		ND		
AMW3-Z2	1073	+		ND			ND			ND		ND		
AMW3-Z3	1940	+		ND			ND			ND		ND		
AMW4-Z1	2258	+	2.07E+03	±	3.99E+02	2.15E+03	±	5.59E+02	4.00E+02	±	2.06E+02	ND		
AMW4-Z2	2463	+		ND			ND			ND		ND		
AMW4-Z3	2293	+		ND			ND			ND		ND		
AMW5-Z1	989	+		ND			ND			ND		ND		
AMW5-Z2	5718	-		ND			ND			ND		ND		
AMW6	375	+		ND			ND			ND		ND		
AEW	293	+	1.60E+04	±	3.53E+02		ND			ND		ND		
					PASSI	/E CELL	•					-		
PMW1	350	+		ND			ND			ND		ND		
PMW2	6877	+		ND			ND			ND		ND		
PMW3-Z1	6807	+		ND			ND			ND		ND		
PMW3-Z2	2319	+		ND			ND			ND		ND		
PMW3-Z3	887	_		ND			ND			ND		ND		
PMW4-Z1	5816	+		ND			ND			ND		ND		
PMW4-Z3	3435	+		ND			ND			ND		ND		
PMW4-Z4	4258	+		ND			ND			ND		ND		
PMW5-Z1	1813	+		ND			ND			ND		ND		
PMW5-Z2	7000	+		ND			ND			ND		ND		
PMW5-Z3	12813	+		ND			ND			ND		ND		
PMW6	1976	+		ND			ND			ND		ND		
PMW7	10500	+		ND			ND			ND		ND		
PMW8	8711	+		ND			ND			ND		ND		
PMW9	478	-		ND			ND			ND		ND		
PIW1	2414	-		ND			ND			ND		ND		
PIW2	19167	+		ND			ND			ND		ND		
PIW3	30973	+		ND			ND			ND		ND		

^{*:} Cells highlighted in yellow and in italics indicate that the value presented is below the reporting limit.

Table 4 qPCR DNA Results



^{#:} a '+' sign indicates that amplification of Bacteria was successful, and a '-' sign indicates that amplification was not successful. ND: indicates sample was non-detect for the target.

Appendix C Site Photographs



Picture 1. AIW-2 well.



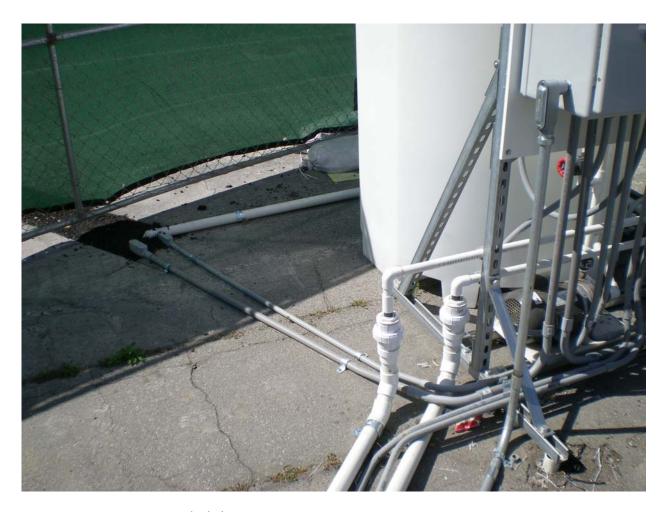
Picture 2. CMT Well.



Picture 3. Injection system control panel.



Picture 4. Dosatron setup.



Picture 5. Extraction piping daylighting.



Picture 6. Extraction piping daylighting.



Picture 7. Extraction well trench.



Picture 8. Normal monitoring well completion.



Picture 9. Peristaltic pump for groundwater purging.



Picture 10. Peristaltic pump sampling setup.



Picture 11. Piping to AIW-1.



Picture 12. Piping to AIW-2.



Picture 13. Piping between injection and extraction wells.



Picture 14. Piping between injection and extraction wells.



Picture 15. Groundwater purge setup with YSI.



Picture 16. Purging groundwater into bucket.



Picture 17. Sample collection.



Picture 18. Surge tank and control panel front.



Picture 19. Surge tank and control panel front.



Picture 20. Surge tank and control panel side.



Picture 21. Treatment compound area.



Picture 22. Treatment compound area.



Picture 23. Treatment compound area.



Picture 24. YSI with flow-thru cell for groundwater purging.

Appendix D Laboratory Study Reports

Final Report

Microcosm Study

With Groundwater from

Naval Weapons Station Seal Beach Site, Irvine CA

Wells EW-70-01 & MW-70-27

received 2/9/06

Subcontract No.: 6225-001-002-AL

July 7, 2006

Prepared for:

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Final Report of Microcosm Study with Groundwater from Naval Weapons Station Seal Beach Site Irvine CA, Wells EW-70-01 & MW-70-27 received 2/9/06 Subcontract No.: 6225-001-002-AL

July 7, 2006

Contents

Summary: Purpose, Approach, Results, Conclusions	3
Groundwater Characteristics	4
Methods	4
Results	5
Data Tables and Figures	7

Abbreviations

TCE, trichloroethene	DCE, cis-dichloroethene	VC, chloroethene
Ethe, ethene	Meth, methane	SO ₄ , sulfate
Ac, acetate	Lac, lactate	Pro, propionate
Bu, butyrate	NH ₄ -N ammonia nitrogen	PO ₄ phosphate
YE, yeast extract	B_{12} , vitamin B_{12}	

Microcosm Study with Groundwater from Naval Weapons Station Seal Beach Site Wells EW-70-01 & MW-70-27

Summary

Purpose and Approach.

The purpose of this microcosm test was to determine whether two of BCI's bioaugmentation cultures could achieve dechlorination in well samples from the Seal Beach Site.

CDM selected two wells for testing (1) EW-70-01, which had a high chloride content of 2,200 mg/L and high sulfate content of 1,650 mg/L, and (2) MW-70-27, which had high chloride of 4,400 mg/L and extremely high sulfate of 9,300 mg/L. Both wells contained total chlorinated ethene concentrations of less than 30 mg/L.

BCI selected two of its *D.ethenogenes* cultures for testing; Culture "S" (a TCE-degrader) and Culture "B" (a mixed chloroethene-degrader), both of which had capabilities with high chloride concentrations. Both cultures were augmented with a sulfate-reducing culture active at high sulfate concentrations.

Anaerobic microcosms were constructed to test each culture with each groundwater sample, using whey as donor (food source), and adding small amounts of minerals needed by bacteria (ammonia and phosphate) as well as yeast extract and vitamin B_{12} . Killed control microcosms were also constructed for each well sample. Microcosms were monitored by removing small samples and analyzing for chlorinated organics and ethene by gas chromatography, and organic acids and sulfate by capillary ion electrophoresis.

Results and Conclusions

For EW-70-01, which contained 'only' 1,650 mg/L sulfate and 2,200 mg/L chloride, BCI Cultures "S" and "B" were equally successful in dechlorinating 16 mg/L TCE and 6 mg/L cDCE completely to ethene in 112 days.

For MW-70-27, which contained very high sulfate of 9,270 mg/L and very high chloride of 4,350 mg/L, Culture "S" succeeded in dechlorinating 73% of the 26 mg/L TCE in 112 days, whereas Culture "B" dechlorinated less than 1 % of the initial TCE to ethene. Therefore, Culture "S" appears to be the better choice for MW-70-27.

The utilization of whey was highly efficient in both ground waters, resulting in very little accumulation of organic acids, mainly due to the utilization of both lactate and acetate by the sulfate-reducing bacteria, and both propionate and butyrate by the organic-acid-oxidizers in the dechlorinating cultures.

Sample Receipt and Groundwater Characteristics

Samples. Groundwater used in this microcosm study was collected on 2/7/06 at Naval Weapons Station Seal Beach Site from MW-70-27 and EW-70-01 into 1 L serum bottles which had been filled with Argon and contained FeS reducing agent to give 0.25 mM. The samples were received 2/9/06. The EW-70-01 sample contained some black solids, indicating that anaerobic conditions had been maintained during sampling and shipping, and received 0.05 mM additional reducing agent. Samples from MW-70-27 arrived having an orange precipitate, indicating that the groundwater was aerobic. These samples received 0.44 mM additional reducing agent to create anaerobic conditions.

Groundwater Characteristics. Results of Groundwater analysis on are given in Table A. The absence of organic acids indicates that both well areas may be donor-limited, while the presence of ammonia and phosphate indicate that these areas are not mineral-limited. The presence of VC in EW-70-01 indicates that there may have been DCE-dechlorinating bacteria (*D. ethenogenes*) in this well area in the past, or that these organisms may currently be present up-gradient. The presence in MW-70-27 of ammonia, rather than nitrate, indicates that the area is at least slightly anaerobic.

Table A. Seal Beach Site Groundwater Characteristics, 2/9/06, mg/L												
MG/L	meth	ethe	VC	cDCE	TCE	Cl	SO_4	organic acids	NO ₃	NH ₄ -N	PO_4	pН
EW-70-01	.11	.01	.180	6.2	16	2,200	1,650	0	0	.1	.15	6.7
MW-70-27	.02	0	.005	0.2	29	4,350	9,270	0	0	.3	.88	7.0

Methods

Microcosm Construction and Maintenance. Microcosms were constructed by transferring 100 ml of groundwater to 160 ml serum bottles using anaerobic technique, were sealed with Teflon-coated rubber septa affixed with crimped caps, and were overpressurized with 5 cc anoxic gas. Controls were killed by lowering the pH to 3. Microcosms were maintained in darkness, with aqueous portion in contact with the septa, at 22 ± 1 °C, and were shaken 3 times per week.

Amendments. Live microcosms received amendment stock solutions which were prepared using anaerobic procedures, and added by syringe to microcosms, giving 40 mg/L ammonia-nitrogen, 60 mg/L phosphate, 50 mg/L yeast extract, and 50 ppb vitamin B₁₂. Whey (aqueous) was received from a dairy, titrated to pH 8.8, made anaerobic, and stored frozen. Whey was added to microcosms in small increments as needed.

Day 1, Bioaugmenting with Sulfate-Reducing Bacteria: In order to lower the ORP to the level required by dechlorinating bacteria, the microcosms were bioaugmented with a BCI culture of salt-tolerant sulfate-reducers.

Days 8 and 23, Bioaugmenting with D.Ethenogenes. On day 8, one microcosm for each well received 0.3 ml of BCI Culture "S", and the second microcosm for each well received 0.3 ml of BCI Culture "B". This bioaugmentation was repeated on Day 23.

Maintenance: During the test, organic acids were monitored, and additional donor was added as needed to maintain detectable propionate, lactate, and/or butyrate. EW-70-01 microcosms received 0.4 ml whey on days 0, 8, 23, 50, 67, and 0.1 ml on day 81. MW-70-27 microcosms received 0.4 ml whey on days 0, 8, 23, 50, 67, and 0.5 ml on days 81, 88, and 96.

Removal of H_2S . Because H_2S resulting from sulfate reduction was removed by adding FeCl to precipitate FeS, subsequently requiring the addition of OH to re-adjust pH. This procedure is not necessary in situ, as metals in the soil will react with the sulfide. Starting on day 53, 16.8 mM FeCl were added to EW-70-01, and 23.5 mM FeCl were added to MW-70-27.

Microcosm Monitoring.

Methane, ethene, and chlorinated compounds were monitored by removing $100~\mu L$ samples of microcosm headspace and injecting into a HP 5890 gas chromatograph according to EPA Method 5021A. Standards were prepared similarly, and analyzed in the same manner as samples. ChemStation software was used to calculate response factors and quantitate results. Concentrations reported are those that would be present if each compound were completely in the aqueous phase (not partially in the headspace).

Nitrate, Sulfate and organic acids were determined by removing $100~\mu L$ aqueous samples and analyzing by capillary ion electrophoresis according to EPA Method 6500 (which does not separate lactate and propionate). Compounds were identified by retention time ratio in comparison with standards analyzed with each batch. Response factors were calculated and results quantified by Millennium software. pH was determined by removing $150~\mu L$ aqueous samples by syringe and measured with a ThermoOrion model 290A pH meter and a Sure-flow Ross semi-micro electrode.

Results and Discussion:

EW-70-01 and MW-70-27 results are presented in Table 1 & Figure 1, and Table 2 & Figure 2 respectively.

<u>Controls</u>: The concentrations of contaminants and daughter products in the killed controls for either EW-70-0 or MW-70-27 did not change during the 112 day test period.

<u>Utilization of Whey in Ground Water</u>. Whey is initially broken down to a mixture of organic acids, formate, acetate, propionate, lactate, and butyrate. Acetate and lactate can be utilized by sulfate-reducing bacteria. Propionate and butyrate are further broken down to acetate, CO₂, and H₂, which is the donor used by dechlorinating bacteria.

EW-70-01 Results (Initial 1,600 mg/L SO₄, 16 mg/L TCE, 6 mg/L cDCE)

EW-70-01 with Culture "S"

During the first three weeks, 300 mg/L of sulfate were reduced, and 20% of the TCE was dechlorinated to cDCE (a step which does not require *D. ethenogenes*). By day 112, the remaining sulfate was reduced and all remaining TCE had been dechlorinated to cDCE, then to VC and finally to Ethene.

EW-70-01 with Culture "B"

During the first three weeks, 270 mg/L Sulfate were reduced and no significant dechlorination occurred. Subsequently, by day 112, all of the remaining sulfate was reduced, and all of the TCE, DCE, and VC had been dechlorinated to ethene.

Utilization of Whey in EW-70-01

During the initial stage of sulfate reduction, acetate accumulated, indicating that sulfate-reducing bacteria were converting Lactate to acetate. Subsequently, acetate was utilized by the sulfate-reducers. Propionate and butyrate were apparently utilized to produce H₂ so quickly, that detectable concentrations were seen only on days 7 and 21 with culture S, and on day 109 with culture B. After dechlorination was complete, methane generation increased.

Culture Selection for EW-70-01

The two BCI Cultures, "S" and "B", dechlorinated TCE and cDCE in EW-70-01 with equal success.

MW-70-27 Results (Initial 9,270 mg/L SO₄, 29 mg/L TCE, 0.2 mg/L cDCE)

MW-70-27 with Culture "S"

By day 21, about 400 mg/L sulfate had been reduced, and all TCE had been dechlorinated to cDCE. Subsequently, by day 112, additional 3,000 mg/L sulfate had been reduced and all cDCE had been dechlorinated to 73 % ethene and 27 % VC, with dechlorination continuing.

MW-70-27 with Culture "B"

By day 21, about 260 mg/L sulfate had been reduced, and all of the TCE had been dechlorinated to cDCE. Subsequently, by day 112, additional 3,000 mg/L sulfate had been reduced, but only 21% of the cDCE had been dechlorinated (to VC).

Utilization of Whey in MW-70-27

With both Culture "S" and Culture "B", acetate accumulated initially, but was subsequently utilized. Organic acids from whey were utilized too quickly to accumulate.

Culture Selection for MW-70-27

In the high-chloride, high-sulfate groundwater, Culture "S" succeeded in dechlorinating 73% of the 26 mg/L TCE in 112 days, whereas Culture "B" dechlorinated less than 1 % of the initial TCE to ethene. Therefore, Culture "S" appears to be the better choice for MW-70-27.



Seal Beach Site 70 Project Quantitative PCR Analytical Summary

31 January, 2007

Overview:

The objective of this project was to detect the number of *Dehalococcoides sp.* (DHC) 16S rRNA gene copies and reductase functional genes (*tceA*, *vcrA*, and *bvcA* copies) contained in groundwater collected from the Seal Beach Site 70 site, Seal Beach, CA, using quantitative polymerase chain reaction (QPCR). The client is CDM. Table 1 describes the sample matrix and the condition of the samples upon arrival to the analytical laboratory.

Table 1. Description of Seal Beach Site 70 samples and volume filtered for DNA extraction.

Well Location	Matrix/Date Sampled	Condition Received/ Observations	Volume
MW70-27	Groundwater	Dry ice preserved filter	18
EW70-01	Groundwater	Dry ice preserved filter	27

The two samples arrived in good condition within the specified holding time. Upon arrival, the samples were frozen for storage at -80°C until the DNA extraction was performed. Following DNA extraction, the samples were first subjected to polymerase chain reaction (PCR) using universal bacterial probes in order to verify that amplifiable DNA was present in the samples. In addition, for the 16S rRNA gene, a "nested" QPCR approach can be applied in which the universal bacterial PCR-amplified DNA is used as the template in a QPCR reaction. Although the results from the nested QPCR cannot be quantified per se, they can be used to lower the detect limit for the QPCR in order to determine if the *Dehalococcoides* 16S rRNA gene is present at concentrations lower than the method detect limit (MDL) using the groundwater DNA extractions. The results of these studies are described here.



Methods:

DNA Extraction: 250 to 500 mL of groundwater was filtered in the field using sterile 0.2-μm acetate filters and filter apparatus (Table 1). The filters were frozen at -80°C and then shattered. Next, each sample tube was amended with 2 mL of DNA-free water, vortexed vigorously for 5 minutes, and the liquid volume was partitioned into DNA extraction tubes. DNA extractions were performed using the Bio101 DNA Extraction Kit according to the manufacturer's instructions. Community DNA was eluted in nuclease-free water (50 μL) and stored at -20°C.

Amplification of Bacteria: The PCR was used to amplify nearly full-length 16S rRNA genes from Bacteria. Each 25- μ L PCR reaction included 0.4 mg mL-1 molecular-grade BSA (Sigma Chemicals), 1X PCR buffer (Promega), 1.5 mM MgCl₂, 0.5 μ M each forward and reverse primer (Invitrogen), 1 U Taq DNA polymerase (Promega), 0.2 mM each dNTP (Invitrogen), 1 μ L template DNA, and molecular-grade water (Promega). Amplification was performed on a PerkinElmer Model 9600 thermocycler using the following regime: 94°C (5 min) followed by 25 cycles of 94°C (1 min), 53.5°C (1 min), and 72°C (1 min). The reaction was finished with an additional 7 minutes at 72°C. PCR products were examined in a 1.2% agarose gel stained with ethicium bromide to confirm specificity of the amplification reactions.

Detection of Dehalococcoides: The QPCR methods for assessing the 16S rRNA gene, and the reductase genes tceA, bvcA, and vcrA, are very sensitive in detecting specific DNA fragments. The detection limit for the methods used is approximately 2 gene copies per μL of the DNA extraction. The reporting limit is 50 gene copies per μL of the DNA extraction.

A mixed laboratory culture containing *Dehalococcoides* was used to obtain the quantitative standard used in these analyses. Plasmid DNA containing DNA inserts of targets 16S rRNA gene, *tceA*, *bvcA*, and *vcrA* from *Dehalococcoides* were purified and quantified fluorometrically. Based on the known size of the plasmid and insert, DNA concentrations were converted to insert copy numbers. A dilution series spanning seven orders of magnitude was generated using known concentrations of each plasmid. Amplification and detection of the DNA was performed using the Cepheid System. The acceptance criterion for the standard curve is a linear R² value of greater than 0.995.

TaqMan Protocol. The 16S rRNA gene QPCR reaction was performed using TaqMan chemistry (Applied Biosystems). All reagents and materials used in the QPCR amplification are purchased from Applied Biosystems. Reaction volumes of 25 μ L contained forward and reverse primers at a concentration of 700 nM, a probe at a concentration of 200 nM, 1 x TaqMan Universal PCR Master Mix, and 5 μ L of sample DNA. The settings for cycle number and reaction conditions used for all runs were 95°C for 10 minutes, and 45 cycles of 95°C for 15 seconds and 58°C for 1 minute. Standards and unknowns were run in triplicate to ensure reproducibility. Cycle thresholds (C_t) were set to minimize the standard deviation of standard curve triplicate C_t values, and also to obtain a standard curve slope as close to negative 3.5 as possible.

SYBR Green Protocol. The functional reductase genes tceA, bvcA, and vcrA were assessed using SYBR green chemistry (Applied Biosystems). Reaction volumes of 25 μ L contained forward and reverse primers at a concentration of 700 nM, a probe at a concentration of 200 nM, 1 x SYBR green Universal PCR Master Mix, and 5 μ L of sample DNA. The settings for cycle number and reaction conditions used for all runs were 95°C for 10 minutes, and 45 cycles of 95°C for 15 seconds and 58°C for 1 minute. Standards and unknowns were run in triplicate to ensure reproducibility. Cycle thresholds (C_t) were set to minimize the standard deviation of standard curve triplicate C_t values, and also to obtain a standard curve slope as close to negative 3.5 as possible.



Results:

The two samples arrived at the lab in good condition frozen with dry ice still in the cooler. The filters were immediately placed in a -80°C freezer and stored until the DNA extraction was performed. Table 2 summarizes the results of the project samples. The DNA extraction negative control and all PCR negative controls did not amplify any product. In addition, all calibration control checks were within acceptable values.

Table 2. Results of molecular analyses for Seal Beach site samples.

Well Location	DNA (ng/L groundwater)	PCR Bacteria [#]	Dehalococcoides 16S rDNA (copy/L groundwater)*	Dehalococcoides tceA (copy/L groundwater)*	Dehalococcoides bvcA (copy/L groundwater)*	Dehalococcoides vcrA (copy/L groundwater)*
MW70-27	10	+	0.00 (+)#	0.00	0.00	0.00
EW70-01	26	+	$4.59 \times 10^2 \pm 2.91 \times 10^2$	$7.50 \times 10^{3^{\circ}}$	8.95 × 10 ³	0.00

^{* :} a * indicates that the value presented is below the reporting limit.

The DNA concentration of the DNA extraction in ng/L of groundwater is reported as an indicator of relative biomass levels for the samples so that relative comparisons can be made. The DNA concentrations ranged from 10 ng/L groundwater for sample MW70-27 to 26 ng/L for sample EW70-01 (Table 2). This indicates very low biomass in the samples, especially considering the large volumes of groundwater that were filtered. All DNA extractions yielded sufficient DNA to amplify *Bacteria*, confirming that despite the very low biomass, amplifiable DNA was obtained from each sample.

DHC was detected in the DNA extraction for sampleEW70-01 at low concentrations (459 16S rRNA gene copies/L groundwater) and was not detected in sample MW70-27 (Table 2). However, DHC was detected in sample MW70-27 in the nested QPCR, which indicates that this microbe is present but below the MDL using the DNA extraction alone. In addition, the reductase genes *tceA* and *bvcA* were detected in the samples, but *vcrA* was not.

^{#:} a '+' sign indicates that amplification of Bacteria and Dehalococcoides (in the nested QPCR) was successful, and a '-' sign indicates that amplification was not successful.

[^] these samples were not run in triplicate due to limited volumes of DNA.

Evaluation of 16S rRNA gene and vcrA gene Sequences Obtained from Seal Beach Site 70 and Three Bioaugmentation Cultures

Introduction:

Molecular analyses was conducted to evaluate *Dehalococcoides* spp. found in the Seal Beach Site 70 site with those found in various bioaugmenation cultures including BCI, Shaw and KB-1. These analyses were conducted in order to determine if indigenous *Dehalococcoides* spp. could be distinguished from those present in several bioaugmentation cultures for the purpose of tracking the growth and transport of the bioaugmented *Dehalococcoides* spp. following inoculation into groundwater at the Seal Beach Site 70 site.

Several methods were used to evaluate *Dehalococcoides* including quantitative PCR analysis and clone library analysis to evaluate various *Dehalococcoides* genes including the 16S rRNA gene, and functional reductase genes *vcrA*, *bvcA* and *tceA*. The following describes clone library analysis used to evaluate the 16S rRNA gene of the Seal Beach Site 70 *Dehalococcoides* and the three bioagumentation cultures and evaluation of *vcrA* sequence analysis of the Shaw and KB-1 bioaugmentation cultures.

Methods:

Clone libraries were constructed for samples EW70-01, BCI and the Shaw bioaugmentation culture to determine the 16S rRNA gene sequence composition of *Dehalococcoides* spp. amplified using primers Fp DHC 1/Rp DHC 1377 shown in Table 1 (Hendrickson et al 2002). (Table 1). In addition, a clone library was constructed using the Shaw bioaugmentation culture using vcrA reductase-gene specific primers. The TOPO® TA kit with TOP10 chemically competent E. coli was used for clone library construction (InvitrogenTM) and the clone libraries were constructed according to the manufacturers instructions.

The clones were selected by blue-white screening, and only those colonies containing plasmids with inserts (white colonies) were selected and plated on LB/SGAL/Kan media plates (Sigma-Aldrich). Plasmids were purified from 5 transformants from each of the 16S rRNA libraries and the vcrA reductase library. Plasmid DNA was extracted and purified from cultures of each clone grown in 1 mL of TPYNG medium containing kanamycin using the QIAprep Spin Miniprep Kit (Qiagen). The purified plasmids were sequenced using primers identified in Table X. to obtain greater than 2X average coverage of the entire insert. Sequencing reactions employed the ABI PRISM BigDye Terminator Cycle Sequencing Ready Reaction Kit (Applied Biosystems) and Model 3100 Automated DNA Sequencer (Applied Biosystems).

The sequences were assembled and aligned using BioEdit software. Sequences were initially aligned against known sequences (GenBank database) using the BLAST tool provided by the National Center for Biotechnology Information. For the 16S rRNA clone libaries a multiple sequence alignment of clones from the Seal Beach site, and the three bioaugmentation cultures was performed with the European Molecular Biology

Laboratory-European Bioinformatics Institute (EMBL-EBI) Clustal W alignment tool. For the vcrA library, the vcrA sequence for *Dehalococcoides* strain VS, an uncultured vcrA sequence and the KB-1 published vcrA sequence were downloaded and included in an alignment with the clones obtained from the Shaw clone library constructed here. The 16S rRNA gene and vcrA gene sequence similarity was then assessed for the sequences within the two alignments. In addition, base pair mismatches were identified and evaluated.

Table 1. Primer targets used for generation of clone libraries.

Primer	Target	Sequence	Use	Reference
Fp DHC	16S rDNA DHC	5'GATGAACGCTAGCGGCG3'	Cloning/ Sequencing	Hendrickson et al 2002
Rp DHC 1377	16S rDNA DHC	5'GGTTGGCACATCGACTTCAA3'	Cloning/ Sequencing	Hendrickson et al 2002
Rp DHC 692	16S rDNA DHC	5'TCAGTGACAACCTAGAAAAC3'	Sequencing	Hendrickson et al 2002
515F	16S rDNA Universal Bacteria	5'GTGCCAGCMGCCGCGGTAA3'	Sequencing	
vcrABF	vcrA reductase	5'CTATGAAGGCCCTCCAGATGC3'	Cloning/ Sequencing	Muller et al 2004
vcrABR	vcrA reductase	5'GTAACAGCCCCAATATGCAAGTA3'	Cloning/ Sequencing	Muller et al 2004
vcrAF	vcrA reductase	5'CTCGGCTACCGAACGGATT3'	Sequencing/QPCR	Lee et al 2006
vcrAR	vcrA reductase	5'GGGCAGGAGGATTGACACAT3'	Sequencing/QPCR	Lee et al 2006

Results:

16S rRNA gene analysis. In order to evaluate the utility of 16S rDNA methods for tracking *Dehalococcoides* populations indigenous to the Seal Beach site and those found in the bioaugmentation cultures, clone libraries were constructed from the Seal Beach Site 70 groundwater sample collected from well EW70-01, and from the bioaugmentation cultures obtained from BCI and Shaw. 16S rDNA sequences were obtained from five clones from the EW70-01 and BCI libraries and four clones from the Shaw library. The approximately 1300 bp DNA sequence obtained from each clone was initially aligned against known sequences using the BLAST tool (Table 2) in order to determine the closest match with sequences in the GenBank database. In addition to the sequences obtained from the libraries, an alignment was generated using a ClustalW algorithm (http://www.ebi.ac.uk/clustalw/) with published sequences from bioaugmentation culture KB-1 in order to determine the sequence similarity between the environmental clone sequences and those observed in the various bioaugmentation cultures (Table 3 and Figure 1).

Results from the 16S rDNA clone library GenBank analysis suggests that most of the *Dehalococcoides* spp. identified in the Seal Beach and bioaugmentation clone libraries were most closely related to *Dehalococcoides ethenogenes* strain 195, or *Dehalococcoides* sp. TM-EtOH (Table 2) with greater than 98-99% sequence similarity. In addition, the ClustalW alignment conducted with the generated clone sequences and two sequences published for the KB-1 culture (KB1-PCE and KB1-VC) suggests that all of the 16S rDNA sequences evaluated were 97% or greater to eachother (Table 2). The alignment shown in Figure 1 illustrates the DNA base pair differences between the sequences by highlighting them in yellow. These data illustrate that the *Dehalococcoides* spp. 16S rDNA sequences are highly similar, and while there are some regions between different sequences that are significantly different, it would be difficult to distinguish between the observed sequences found within the different bioaugmentation cultures and those indigenous to the Seal Beach site by 16S rDNA molecular analysis alone.

Table 2. Genebank results for the bacterial 16S rDNA clone library results for the Seal Beach site (EW70-01) sample and the bioaugmentation cultures BCI, and Shaw.

Name	Accession	Closest GenBank match	Base pair % similarity
EW07-01#8	CP000027.1	Dehalococcoides ethenogenes 195, complete genome	1265/1281 (98%),
EW-70-01#6	AY882433.1	Dehalococcoides sp. TM-EtOH 16S ribosomal RNA gene, partial	1270/1279 (99%),
EW-70-01#7	AY882433.1	Dehalococcoides sp. TM-EtOH 16S ribosomal RNA gene, partial	1276/1278 (99%),
EW-70-01#2	CP000027.1	Dehalococcoides ethenogenes 195, complete genome	1286/1292 (99%),
EW-70-01#3	CP000027.1	Dehalococcoides ethenogenes 195, complete genome	1272/1275 (99%),
BCI #3	AF388530.1	Uncultured Dehalococcoides sp. clone DHC-asd 16S ribosomal RNA	1266/1276(99%),
BCI #17	AY882433.1	Dehalococcoides sp. TM-EtOH 16S ribosomal RNA gene, partial	1272/1277(99%),
BCI#15	CP000027.1	Dehalococcoides ethenogenes 195, complete genome	1271/1278 (99%),
BCI#1	CP000027.1	Dehalococcoides ethenogenes 195, complete genome	1256/1263 (99%),
BCI#16	CP000027.1	Dehalococcoides ethenogenes 195, complete genome	1273/1276 (99%
Shaw16s#1	CP000027.1	Dehalococcoides ethenogenes 195, complete genome	1006/1011 (99%),
Shaw16s#2	AY882433.1	Dehalococcoides sp. TM-EtOH 16S ribosomal RNA gene, partial	1277/1278 (99%),
Shaw16s#3	CP000027.1	Dehalococcoides ethenogenes 195, complete genome	278/1279 (99%),
Shaw16s#4	CP000027.1	Dehalococcoides ethenogenes 195, complete genome	1327/1331 (99%),

Table 3. Sequence Similarity of 16S rRNA gene sequences from BCI, Shaw, KB-1 and EW70-01.

SeqA	Name	Length(bp)	SeqB	Name	Length(bp)	Score (%)
1	BCI#1	1388	===== 2	======= BCI#3	1386	97
1	BCI#1 BCI#1	1388	3	BCI#3	1336	98
1	BCI#1 BCI#1	1388	4	BCI#15	1388	98
1	BCI#1	1388	5	BCI#10 BCI#17	1386	98
1	BCI#1	1388	6	EW70-01#2		98
1	BCI#1	1388	7	EW70-01#3		98
1	BCI#1	1388	8	EW70-01#6		98
1	BCI#1	1388	9	EW70-01#7		98
1	BCI#1	1388	10	EW70-01#8		98
1	BCI#1	1388	11	Shaw16s#1		98
1	BCI#1	1388	12	Shaw16s#2		98
1	BCI#1	1388	13	Shaw16s#3		98
1	BCI#1	1388	14	Shaw16s#4		98
1	BCI#1	1388	15	KB1-VC	1386	97
1	BCI#1	1388	16	KB1-PCE	1385	97
2	BCI#3	1386	3	BCI#15	1336	98
2	BCI#3	1386	4	BCI#16	1388	98
2	BCI#3	1386	5	BCI#17	1386	98
2	BCI#3	1386	6	EW70-01#2		98
2	BCI#3	1386	7	EW70-01#3	3 1388	98
2	BCI#3	1386	8	EW70-01#6	1386	98
2	BCI#3	1386	9	EW70-01#7	7 1387	98
2	BCI#3	1386	10	EW70-01#8	3 1373	97
2	BCI#3	1386	11	Shaw16s#1	1387	98
2	BCI#3	1386	12	Shaw16s#2	1388	98
2	BCI#3	1386	13	Shaw16s#3		98
2	BCI#3	1386	14	Shaw16s#4		98
2	BCI#3	1386	15	KB1-VC	1386	98
2	BCI#3	1386	16	KB1-PCE	1385	98
3	BCI#15	1336	4	BCI#16	1388	99
3	BCI#15	1336	5	BCI#17	1386	98
3	BCI#15	1336	6	EW70-01#2		98
3	BCI#15	1336	7	EW70-01#3		99
3	BCI#15	1336	8	EW70-01#6		98
3	BCI#15	1336	9	EW70-01#7		99
3	BCI#15	1336	10	EW70-01#8		98
3	BCI#15	1336	11	Shaw16s#1		98
3	BCI#15	1336	12	Shaw16s#2		99
3	BCI#15	1336	13	Shaw16s#3		99
3	BCI#15	1336	14	Shaw16s#4		99
3	BCI#15	1336	15	KB1-VC	1386	97
3	BCI#15	1336	16	KB1-PCE	1385	97
4	BCI#15 BCI#16	1388	5	BCI#17	1386	99
4			6			99
	BCI#16	1388		EW70-01#2		
4	BCI#16	1388	7	EW70-01#3		99
4	BCI#16	1388	8	EW70-01#6		99
4	BCI#16	1388	9	EW70-01#7		99
4	BCI#16	1388	10	EW70-01#8		98
4	BCI#16	1388	11	Shaw16s#1		99
4	BCI#16	1388	12	Shaw16s#2		99
4	BCI#16	1388	13	Shaw16s#3		99
4	BCI#16	1388	14	Shaw16s#4	1279	99

4	BCI#16	1388	15	KB1-VC	1386	98
4	BCI#16	1388	16	KB1-PCE	1385	97
5	BCI#17	1386	6	EW70-01#2	1388	99
5	BCI#17	1386	7	EW70-01#3	1388	99
5	BCI#17	1386	8	EW70-01#6	1386	99
5	BCI#17	1386	9	EW70-01#7	1387	99
5	BCI#17	1386	10	EW70-01#8	1373	98
5	BCI#17	1386	11	Shaw16s#1	1387	99
				**		
5	BCI#17	1386	12	Shaw16s#2	1388	99
5	BCI#17	1386	13	Shaw16s#3	1388	99
5	BCI#17	1386	14	Shaw16s#4	1279	99
5	BCI#17	1386	15	KB1-VC	1386	97
5	BCI#17	1386	16	KB1-PCE	1385	97
6	EW70-01#2	1388	7	EW70-01#3	1388	99
6	EW70-01#2	1388	8	EW70-01#6	1386	98
6	EW70-01#2	1388	9	EW70-01#7	1387	99
6	EW70-01#2	1388	10	EW70-01#8	1373	98
6	• • • • • • • • • • • • • • • • • • • •		11	Shaw16s#1	1387	99
	EW70-01#2	1388				
6	EW70-01#2	1388	12	Shaw16s#2	1388	99
6	EW70-01#2	1388	13	Shaw16s#3	1388	99
6	EW70-01#2	1388	14	Shaw16s#4	1279	99
				**		
6	EW70-01#2	1388	15	KB1-VC	1386	97
6	EW70-01#2	1388	16	KB1-PCE	1385	97
7	EW70-01#3	1388	8	EW70-01#6	1386	99
	• • • • • • • • • • • • • • • • • • • •			**		
7	EW70-01#3	1388	9	EW70-01#7	1387	99
7	EW70-01#3	1388	10	EW70-01#8	1373	98
7	EW70-01#3	1388	11	Shaw16s#1	1387	99
7	EW70-01#3	1388	12	Shaw16s#2	1388	99
7	EW70-01#3	1388	13	Shaw16s#3	1388	99
7	EW70-01#3	1388	14	Shaw16s#4	1279	99
7	EW70-01#3	1388	15	KB1-VC	1386	97
	• • • • • • • • • • • • • • • • • • • •					
7	EW70-01#3	1388	16	KB1-PCE	1385	97
8	EW70-01#6	1386	9	EW70-01#7	1387	99
8	EW70-01#6	1386	10	EW70-01#8	1373	98
	• •			**		
8	EW70-01#6	1386	11	Shaw16s#1	1387	98
8	EW70-01#6	1386	12	Shaw16s#2	1388	99
8	EW70-01#6	1386	13	Shaw16s#3	1388	99
		1386				
8	EW70-01#6		14	Shaw16s#4	1279	98
8	EW70-01#6	1386	15	KB1-VC	1386	97
8	EW70-01#6	1386	16	KB1-PCE	1385	97
9	EW70-01#7	1387	10	EW70-01#8	1373	98
9	EW70-01#7	1387	11	Shaw16s#1	1387	99
9	EW70-01#7	1387	12	Shaw16s#2	1388	99
9	EW70-01#7	1387	13	Shaw16s#3	1388	99
9	EW70-01#7	1387	14	Shaw16s#4	1279	99
9	EW70-01#7	1387	15	KB1-VC	1386	97
9	EW70-01#7	1387	16	KB1-PCE	1385	97
10	EW70-01#8	1373	11	Shaw16s#1	1387	98
10	EW70-01#8	1373	12	Shaw16s#2	1388	98
10	EW70-01#8	1373	13	Shaw16s#3	1388	99
10	EW70-01#8	1373	14	Shaw16s#4	1279	98
10	EW70-01#8	1373	15	KB1-VC	1386	97
10	EW70-01#8	1373	16	KB1-PCE	1385	97
11	Shaw16s#1	1387	12	Shaw16s#2	1388	99
11	Shaw16s#1	1387	13	Shaw16s#3	1388	99
11	Shaw16s#1	1387	14	Shaw16s#4	1279	99
11	Shaw16s#1	1387	15	KB1-VC	1386	97

11	Shaw16s#1	1387	16	KB1-PCE	1385	97		
12	Shaw16s#2	1388	13	Shaw16s#3	1388	99		
12	Shaw16s#2	1388	14	Shaw16s#4	1279	99		
12	Shaw16s#2	1388	15	KB1-VC	1386	98		
12	Shaw16s#2	1388	16	KB1-PCE	1385	98		
13	Shaw16s#3	1388	14	Shaw16s#4	1279	99		
13	Shaw16s#3	1388	15	KB1-VC	1386	98		
13	Shaw16s#3	1388	16	KB1-PCE	1385	97		
14	Shaw16s#4	1279	15	KB1-VC	1386	98		
14	Shaw16s#4	1279	16	KB1-PCE	1385	98		
15	KB1-VC	1386	16	KB1-PCE	1385	99		
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Figure 1. Sequer	nce Alignment 16S rDNA for the Shaw, BCI, KB1 and EW70-01.
FW70-01#8	$-$ CCCTTG Δ TG Δ CCCCT Δ CCCCCCCTT Δ TGC Δ TCC Δ Δ CTCG Δ Δ CCG $-$ TCTT Δ Δ CC Δ Δ T $-$ 58

EW70-01#8	-CCCTTGATGAACGCTAGCGGCGTGCCTTATGCATGCAAGTCGAACGG-TCTTAAGCAAT	58
Shaw16s#1	$\tt GCCCTTGATGAACGCTAGCGGCGTGCCTTATGCATGCAAGTCGAACGG-TCTTAAGCAAT$	59
BCI#1	-CCCTTGATGAACGCTAGCGGCGTGCCTTATGCATGCAAGTCGAACGG-TCTTAAGCAAT	58
BCI#16	CTTGATGAACGCTAGCGGGGGGGCCTTATGCATGCAAGTCGAACGG-TCTTAAGCAAT	56
EW70-01#6	CTTGATGAACGCTAGCGGGCGTGCCTTATGCATGCAAGTCGAACGG-TCTTAAGCAAT	56
EW70-01#3	CCTTGATGAACGCTAGCGGCGTGCCTTATGCATGCAAGTCGAACGG-TCTTAAGCAAT	57
BCI#17	$-\texttt{CCCTTGATGAACGCTAGCGGCGTGCCTTATGCATGCAAGTCGAACGG-TCTT}{\textbf{AA}\textbf{GCAAT}}$	58
Shaw16s#2	$-\texttt{CCCTTGATGAACGCTAGCGGCGTGCCTTATGCATGCAAGTCGAACGG-TCTT}{\textbf{AAGCAAT}}$	58
EW70-01#7	$-\texttt{CCCTTGATGAACGCTAGCGGCGTGCCTTATGCATGCAAGTCGAACGG-TCTT}{\textbf{AAGCAAT}}$	58
Shaw16s#3	$-\texttt{CCCTTGATGAA} \\ \texttt{CGCTAGCGGCGTGCCTTATGCATGCAAGTCGAACGG-TCTTAAGCAAT}$	58
BCI#15	$-\texttt{CCCTTGATGAACGCTAGCGGCGTGCCTTATGCATGCAAGTCGAACGG-TCTT}{\textbf{AA}\textbf{GCAAT}}$	58
EW70-01#2	-CCCTTGATGAACGCTAGCGGCGTGCCTTATGCATGCAAGTCGAACGG <mark>G</mark> TCTTAAGCA <mark>-</mark> T	58
Shaw16s#4		2
BCI#3	CTTGATGAACGCTAGCGGCGTGCCTTATGCATGCGAGTCGAACGG-TCTTAAGCAAT	56
KB1-VC	GATGAACGCTAGCGGCGTGCCTTATGCATGCAAGTCGAACGG-TCTTAAGCAAT	53
KB1-PCE	GATGAACGCTAGCGGCGTGCCTTATGCATGCAAGTCGAACGG-TCTTAAGCAAT	53
	*	
EW70-01#8	TAAGA-TAGTGGCAAACGGGTGAGTAACGCGTAAGTAACCTACCTCTAAGTGGGGGGATAG	117
Shaw16s#1	TAAGA-TAGTGGCAAACGGGTGAGTAACGCGTAAGTA <mark>G</mark> CCTACCTCTAAGTGGGGGGATAG	118
BCI#1	TAAGA-TAGTGGCAAACGGGTGAGTAACGCGTAAGTAACCTACCTCTAAGTGGGGGGATAG	117
BCI#16	TAAGA-TAGTGGCAAACGGGTGAGTAACGCGTAAGTAACCTACCTCTAAGTGGGGGGATAG	115
EW70-01#6	TAAGA-TAGTGGCAAACGGGTGAGTAACGCGTAAGTAACCTACCTCTAAGTGGGGGGATAG	115
EW70-01#3	${\tt TAAGA} {\tt A}{\tt TAGTGGCAAACGGGTGAGTAACGCGTAAGTAACCTACCTCTAAGTGGGGGGATAG}$	117
BCI#17	TAAGA-TAGTGGCAAACGGGTGAGTAACGCGTAAGTAACCTACCTCTAAGTGGGGGGATAG	117
Shaw16s#2	TAAGA-TAGTGGCAAACGGGTGAGTAACGCGTAAGTA <mark>G</mark> CCTACCTCTAAGTGGGGGGATAG	117
EW70-01#7	TAAGA-TAGTGGCAAACGGGTGAGTAACGCGTAAGTAACCTACCTCTAAGTGGGGGGATAG	117
Shaw16s#3	TAAGA-TAGTGGCAAACGGGTGAGTAACGCGTAAGTAACCTACCTCTAAGTGGGGGGATAG	117
BCI#15	TAAGA-TAGTGGCAAACGGGTGAGTAACGCGTAAGTAACCTACCTCTAAGTGGGGGGATAG	117
EW70-01#2	TAAGA-TAGTGGC <mark>T</mark> AACGGGTGAGTAACGCGTAAGTAACCTACCTCTAAGTGGGGGGATAG	117
Shaw16s#4	TAAGA-TAGTGGCAAACGGGTGAGTAACGCGTAAGTAACCTACCTCTAAGTGGGGGGATAG	61
BCI#3	TAAGA-TAGTGGCAAACGGGTGAGTAACGCGTAAGTAACCTACCTCTAAGTGGGGGGATAG	115
KB1-VC	TAAGA-TAGTGGC <mark>G</mark> AACGGGTGAGTAACGCGTAAGTAACCTACCTCTAAGTGGGGGGATAG	112
KB1-PCE	TAAGA-TAGTGGC <mark>G</mark> AACGGGTGAGTAACGCGTAAGTAACCTACCTCTAAGTGGGGGGATAG	112

EW70-01#8 Shaw16s#1 BCI#1 BCI#16 EW70-01#6 EW70-01#3 BCI#17 Shaw16s#2 EW70-01#7 Shaw16s#3 BCI#15 EW70-01#2 Shaw16s#4 BCI#3 KB1-VC KB1-PCE	CTT <mark>T</mark> GGGAAACTGAAGGTAATACCGCATGTGATGGGCTGACATAAGTCGGTTCATTAAAG CTTCGGGAAACTGAAGGTAATACCGCATGTGATGGGCTGACATAAGTCGGTTCATTAAAG	178 177 175 175 177 177 177 177 177 177 177
EW70-01#8 Shaw16s#1 BCI#1 BCI#16 EW70-01#6 EW70-01#3 BCI#17 Shaw16s#2 EW70-01#7 Shaw16s#3 BCI#15 EW70-01#2 Shaw16s#4 BCI#3 KB1-VC KB1-PCE	CCGCAAGGTGCTTGGTGAGGGGCTTGCGTCCG-ATTAGCTAGTTGGTGGGGTAATGGCCT CCGCAAGGTGCTTGGTGAGGGGCTTGCGTCCG-ATTAGCTAGTTGGTGGGGTAATGGCCT CCGCAAGGTGCTTGGTGAGGGGCTTGCGTCCG-ATTAGCTAGTTGGTGGGGGTAATGGCCT CCGCAAGGTGCTTGGTGAGGGGCTTGCGTCCG-ATTAGCTAGTTGGTGGGGGTAATGGCCT CCGCAAGGTGCTTGGTGAGGGGCTTGCGTGAGGAAAAAAAA	235 236 236 236 236 236 236
EW70-01#8 Shaw16s#1	ACCAAGGCTTCGATCGGTAGCTGGTCTGAGAGGATGATCAGCCACACTGGGACTGAGACA ACCAAGGCTTCGATCGGTAGCTGGTCTGAGAGGATGATCAGCCACACTGGGACTGAGACA	

BCI#1 BCI#16 EW70-01#6 EW70-01#3 BCI#17 Shaw16s#2 EW70-01#7 Shaw16s#3 BCI#15 EW70-01#2	ACCAAGGCTTCGATCGGTAGCTGATCTGAGAGGATGATCAGCCACACTGGGACTGAGACA ACCAAGGCTTCGATCGGTAGCTGGTCTGAGAGGATGATCAGCCACACTGGGACTGAGACA	294 295 296 296 296 296 296 296
Shaw16s#4	ACCAAGGCTTCGATCGGTAGCTGGTCTGAGAGGATGATCAGCCACACTGGGACTGAGACA	
BCI#3	ACCAAGGCTTCGATCGGTAGCTGGTCTGAGAGGATGATCAGCCACACTGGGACTGAGACA	
KB1-VC	ACCAAGGCTTCGATCGGTAGCTGGTCTGAGAGGATGATCAGCCACACTGGGACTGAGACA	
KB1-PCE	ACCAAGGCTTCGATCGGTAGCTGGTCTGAGAGGATGATCAGCCACACTGGGACTGAGACA *********************************	29I
EW70-01#8	CGGCCCAGACTCCTACGGGAGGCAGCAGCAAGGAATCTTGGGCAACGAAAGCCTGA	352
Shaw16s#1	CGGCCCAGACTCCTACGGGAGGCAGCAGCAAGGAATCTTGGGCAATGGGCGAAAGCCTGA	
BCI#1	CGGCCCAGACTCCTACGGGAGGCAGCAACGAATCTTGGGCAATGGGCGAAAGCCTGA	
BCI#16	CGGCCCAGACTCCTACGGGAGGCAGCAGCAAGGAATCTTGGGCAATGGGCGAAAGCCTGA	
EW70-01#6	CGGCCCAGACTCCTACGGGAGGCAGCAGCAAGGAATCTTGGGCAATGGGCGAAAGCCTGA	355
EW70-01#3	CGGCCCAGACTCCTACGGGAGGCAGCAGCAAGGAATCTTGGGCAATGGGCGAAAGCCTGA	356
BCI#17	CGGCCCAGACTCCTACGGGAGGCAGCAGCAAGGAATCTTGGGCAATGGGCGAAAG <mark>T</mark> CTGA	356
Shaw16s#2	CGGCCCAGACTCCTACGGGAGGCAGCAGCAAGGAATCTTGGGCAATGGGCGAAAGCCTGA	356
EW70-01#7	CGGCCCAGACTCCTACGGGAGGCAGCAGCAAGGAATCTTGGGCAATGGGCGAAAGCCTGA	356
Shaw16s#3	$\tt CGGCCCAGACTCCTACGGGAGGCAGCAGCAAGGAATCTTGGGCAATGGGCGAAAGCCTGA$	356
BCI#15	$\tt CGGCCCAGACTCCTACGGGAGGCAGCAGCAAGGAATCTTGGGCAATGGGCGAAAGCCTGA$	356
EW70-01#2	CGGCCCAGACTCCTACGGGAGGCAGCAGCAAGGAATCTTGGGCAATGGGCGAAAGCCTGA	356
Shaw16s#4	$\tt CGGCCCAGACTCCTACGGGAGGCAGCAGCAAGGAATCTTGGGCAATGGGCGAAAGCCTGA$	300
BCI#3	CGGCCCAGACTCCTACGGGAGGCAGCAGCAAGGAATCTTGGGCAATGGGCG <mark>G</mark> AAGCCTGA	354
KB1-VC	CGGCCCAGACTCCTACGGGAGGCAGCAGCAAGGAATCTTGGGCCAATGGGCGAAAGCCTGA	351
KB1-PCE	CGGCCCAGACTCCTACGGGAGGCAGCAGCAAGGAATCTTGGGCCAATGGGCGAAAGCCTGA	351

EW70-01#8 Shaw16s#1 BCI#1 BCI#16	CCCAGCAACGCCGCGTGAGGGATGAAAGGCTTTCGGGTGTAAAACCTCTTTTCACAGGGACCCAGCACGCCCGCGTGAGGGATGAA-GGCTTTCGGGTTGTAAACCTCTTTTCACAGGGACCCAGCAACGCCGCGTGAGGGATGAA-GGCTTTCGGGTTGTAAACCTCTTTTCACAGGGACCCAGCAACGCCGCGTGAGGGATGAA-GGCTTTCGGGTTGTAAACCTCTTTTCACAGGGACCCAGCAACGCCGCGTGAGGGATGAA-GGCTTTCGGGTTGTAAACCTCTTTTCACAGGGA	416 415

EW70-01#6	CCCAGCAAC <mark>A</mark> CCGCGTGAGGGATGAA-GGCTTTCGGGTTGTAAACCTCTTTTCACAGGGA	414
EW70-01#3	CCCAGCAACGCCGCGTGAGGGATGAA-GGCTTTCGGGTTGTAAACCTCTTTTCACAGGGA	415
BCI#17	CCCAGCAACGCCGCGTGAGGGATGAA-GGCTTTCGGGTTGTAAACCT <mark>-</mark> TTTTCACAGGGA	414
Shaw16s#2	CCCAGCAACGCCGCGTGAGGGATGAA-GGCTTTCGGGTTGTAAACCTCTTTTCACAGGGA	415
EW70-01#7	CCCAGCAACGCCGCGTGAGGGATGAA-GGCTTTCGGGTTGTAAACCTCTTTTCACAGGGA	
Shaw16s#3	CCCAGCAACGCCGCGTGAGGGATGAA-GGCTTTCGGGTTGTAAACCTCTTTTCACAGGGA	
BCI#15	CCCAGCAACGCCGCGTGAGGGATGAA-GGCTTTCGGGTTGTAAACCTCTTTTC <mark>G</mark> CAGGGA	
EW70-01#2	CCCAGCAACGCCGCGTGAGGGATGAA-GGCTTTCGGGTTGTAAACCTCTTTTCACAGGGA	
Shaw16s#4	CCCAGCAACGCCGCGTGAGGGATGAA-GGCTTTCGGGTTGTAAACCTCTTTTCACAGGGA	
BCI#3	CCCAGCAACGCCGCGTGAGGGATGAA-GGCTTTCGGGTTGTAAACCTCTTTTCA <mark>T</mark> AGGGA	413
KB1-VC	CCCAGCAACGCCGCGTGAGGGATGAA-GGCTTTCGGGTTGTAAACCTCTTTTCA <mark>T</mark> AGGGA	410
KB1-PCE	CCCAGCAACGCCGCGTGAGGGATGAA-GGCTTTCGGGTTGTAAACCTCTTTTCA <mark>T</mark> AGGGA	410

EW70-01#8	AGAATAATGACGGTACCTGTGGAATAAGC <mark>C</mark> TCGGCTAACTACGTGCCAGCAGCCGCGGTA	472
Shaw16s#1	AGAATAATGACGGTACCTGTGGAATAAGCTTCGGCTAACTACGTGCCAGCAGCCGCGGTA	476
BCI#1	AGAATAATG <mark>T</mark> CGGTACCTGTGGAATAAGCTTCGGCTAACTACGTGCCAGCAGCCGCGGTA	475
BCI#16	AGAATAATGACGGTACCTGTGGAATAAGCTTCGGCTAACTACGTGCCAGCAGCCGCGGTA	473
EW70-01#6	A-AATAATGACGGTACCTGTGGAATAAGCTTCGGCTAACTACGTGCCAGCAGCCGCGGTA	473
EW70-01#3	AGAATAATGACGGTACCTGTGGAATAAGCTTCGGCTAACTACGTGCCAGCAGCCGCGGTA	475
BCI#17	AGAATAATGACGGTACCTGTGGAATAAGCTTCGGCTAACTACGTGCCAGCAGCCGCGGTA	474
Shaw16s#2	AGAATAATGACGGTACCTGTGGAATAAGCTTCGGCTAACTACGTGCCAGCAGCCGCGGTA	475
EW70-01#7	AGAATAATGACGGTACCTGTGGAATAAGCTTCGGCTAACTACGTGCCAGCAGCCGCGGTA	475
Shaw16s#3	AGAATAATGACGGTACCTGTGGAATAAGCTTCGGCTAACTACGTGCCAGCAGCCGCGGTA	475
BCI#15	AGAATAATGACGGTACCTGTGGAATAAGCTTCGGCTAACTACGTGCCAGCAGCCGCGGTA	475
EW70-01#2	AGAATAATGACGGTACCTGTGGAATAAGCTTCGGCTAACTACGTGCCAGCAGCCGCGGTA	475
Shaw16s#4	AGAATAATGACGGTACCTGTGGAATAAGCTTCGGCTAACTACGTGCCAGCAGCCGCGGTA	419
BCI#3	AGAATAATGACGGTACCTGTGGAATAAGCTTCGGCTAACTACGTGCCAGCAGCCGCGGTA	473
KB1-VC	AGAATAATGACGGTACCTGTGGAATAAGCTTCGGCTAACTACGTGCCAGCAGCCGCGGTA	470
KB1-PCE	AGAATAATGACGGTACCTGTGGAATAAGCTTCGGCTAACTACGTGCCAGCAGCCGCGGTA	470
	* ***** *********** ************	
EW70-01#8	ATACGTAGGAAGCAAGCGTTATCCGGATTTATTGGGCGTAAAGTGAGCGTAGGTGGTCTT	532
Shaw16s#1	ATACGTAGGAAGCAAGCGTTATCCGGATTTATTGGGCGTAAAGTGAGCGTAGGTGGTCTT	536
BCI#1	ATACGTAGGAAGCAAGCGTTATCCGGATTTATTGGGCGTAAAGTGAGCGTAGGTGGTCTT	535
BCI#16	ATACGTAGGAAGCAAGCGTTATCCGGATTTATTGGGCGTAAAGTGAGCGTAGGTGGTCTT	533
EW70-01#6	ATACGTAGGAAGCAAGCGTTATCCGGATTTATTGGGCGTAAAGTGAGCGTAGGTGGTCTT	533
EW70-01#3	ATACGTAGGAAGCAAGCGTTATCCGGATTTATTGGGCGTAAAGTGAGCGTAGGTGGTCTT	535

BCI#17	ATACGTAGGAAGCAAGCGTTATCCGGATTTATTGGGCGTAAAGTGAGCGTAGGTGGTCTT	534
Shaw16s#2	ATACGTAGGAAGCAAGCGTTATCCGGATTTATTGGGCGTAAAGTGAGCGTAGGTGGTCTT	535
EW70-01#7	ATACGTAGGAAGCAAGCGTTATCCGGATTTATTGGGCGTAAAGTGAGCGTAGGTGGTCTT	535
Shaw16s#3	ATACGTAGGAAGCAAGCGTTATCCGGATTTATTGGGCGTAAAGTGAGCGTAGGTGGTCTT	535
BCI#15	ATACGTAGGAAGCAAGCGTTATCCGGATTTATTGGGCGTAAAGTGGGCGTAGGTGGTCTT	535
EW70-01#2	ATACGTAGGAAGCAAGCGTTATCCGGATTTATTGGGCGTAAAGTGAGCGTAGGTGGTCTT	535
Shaw16s#4	ATACGTAGGAAGCAAGCGTTATCCGGATTTATTGGGCGTAAAGTGAGCGTAGGTGGTCTT	479
BCI#3	ATACGTAGGAAGCAAGCGTTATCCGGATTTATTGGGCGTAAAGTGAGCGTAGGTGGTCTT	533
KB1-VC	ATACGTAGGAAGCAAGCGTTATCCGGATTTATTGGGCGTAAAGTGAGCGTAGGTGGTCTT	530
KB1-PCE	ATACGTAGGAAGCAAGCGTTATCCGGATTTATTGGGCGTAAAGTGAGCGTAGGTGGTCTT	530

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EW70-01#8	TCAAGTTGGATGTGAAATTTCCCGGCTTAACCGGGACGTGTCATTCAATACTGTTGGACT	592
Shaw16s#1	TCAAGTTGGATGTGAAATTTCCCGGCTTAACCGGGACGTGTCATTCAATACTGTTGGACT	596
BCI#1	CCAAGTTGGATGTGAAATTTCCCGGCTTA <mark>G</mark> CCGGGACGTGTCATTCAATACTGTTGGACT	595
BCI#16	TCAAGTTGGATGTGAAATTTCCCGGCTTAACCGGGACGTGTCATTCAATACTGTTGGACT	593
EW70-01#6	TCAAGTTGGATGTGAAATTTCCCGGCTTAACCGGGACGTGTCATTCAATACTGTTGGACT	593
EW70-01#3	TCAAGTTG <mark>A</mark> ATGTGAAATTTCCCGGCTTAACCGGGACGTGTCATTCAATACTGTTGGACT	595
BCI#17	TCAAGTTGGATGTGAAATTTCCCGGCTTAACCGGGACGTGTCATTCAATACTGTTGGACT	594
Shaw16s#2	TCAAGTTGGATGTGAAATTTCCCCGGCTTAACCGGGACGTGTCATTCAATACTGTTGGACT	595
EW70-01#7	TCAAGTTGGATGTGAAATTTCCCGGCTTAACCGGGACGTGTCATTCAATACTGTTGGACT	595
Shaw16s#3	TCAAGTTGGATGTGAAATTTCCCGGCTTAACCGGGACGTGTCATTCAATACTGTTGGACT	595
BCI#15	TCAAGTTGGATGTGAAATTTCCCGGGCTTAACCGGGACG <mark>A</mark> GTCATTCAATACTGTTGGACT	595
EW70-01#2	TCAAGTTGGATGTGAAATTTCCCGGCTTAACCGGGACGTGTCATCCAATACTGTTGGACT	595
Shaw16s#4	TCAAGTTGGATGTGAAATTTCCCGGCTTAACCGGGACGTGTCATTCAATACTGTTGGACT	539
BCI#3	TCAAGTTGGATGTGAAATTTCCCGGCTTAACCGGGACGTGTCATTCAATACTGTTGGACT	593
KB1-VC	TCAAGTTGGA <mark>TGTGAAATTTCCCGGGCTTAACCGGGACG</mark> GTCATTCAATACTGTTGGACT	590
KB1-PCE	TCAAGTTGGA <mark>-</mark> GTGAAATTTCCCGGCTTAACCGGGACG <mark>A</mark> GTCATTCAATACTGTTGGACT	589
	****** * ********** ****** ***** ****	
EW70-01#8	AGAGTACAGCAGGAGAAAACGGAATTCCCGGTGTAGTGGTAAAATGCGTAGATATCGGGA	652
Shaw16s#1	AGAGTACAGCAGGAGAAAACGGAATTCCCGGTGTAGTGGTAAAATGCGTAGATATCGGGA	
BCI#1	AGAGTACAGCAGGAGAAAACGGAATTCCCGGTGTAGTGGTAAAATGCGTAGATATCGGGA	
BCI#16	AGAGTACAGCAGGAGAAAACGGAATTCCCGGTGTAGTGGTAAAATGCGTAGATATCGGGA	
EW70-01#6	AGAGTACAGCAGGAGAAAACGGAATTCCCGGTGTAGTGGTAAAATGCGTAGATATCGGGA AGAGTACAGCAGGAGAAAACGGAATTCCCGGTGTAGTGGTAAAATGCGTAGATATCGGGA	
EW70-01#3	AGAGTACAGCAGGAGAAAACGGAATTCCCGGTGTAGTGGTAAAATGCGTAGATATCGGGA AGAGTACAGCAGGAGAAAACGGAATTCCCGGTGTAGTGGTAAAATGCGTAGATATCGGGA	
BCI#17	AGAGTACAGCAGGAGAAAACGGAATTCCCGGTGTAGTGGTAAAATGCGTAGATATCGGGA AGAGTACAGCAGGAGAAAACGGAATTCCCGGTGTAGTGGTAAAATGCGTAGATATCGGGA	
Shaw16s#2	AGAGTACAGCAGGAGAAAACGGAATTCCCCGGTGTAGTGGTAAAATGCGTAGATATCGGGA	
SIIGMT02#7	ADDDIAIADA1DJDIAAA1DDIDA1D1DG1DADJAAAADADDAJDAJDADA	000

EW70-01#7	AGAGTACAGCAGGAGAAAACGGAATTCCCGGTGTAGTGGTAAAATGCGTAGATATCGGGA	
Shaw16s#3	AGAGTACAGCAGGAGAAAACGGAATTCCCGGTGTAGTGGTAAAATGCGTAGATATCGGGA	
BCI#15	AGAGTACAGCAGGAG <mark>T</mark> AAACGGAATTCCCGGTGTAGTGGTAAAATGCG <mark>K</mark> AGATATCGGGA	655
EW70-01#2	AGAGTACAGCAGGAGAAAACGGAATTCCCGGTGTAGTGGTAAAATGCGTAGATATCGGGA	655
Shaw16s#4	AGAGTACAGCAGGAGAAAACGGAATTCCCGGTGTAGTGGTAAAATGCGTAGATATCGGGA	599
BCI#3	AGAGTACAGCAGGAGAAAACGGAATTCCCGGTGTAGTGGTAAAATGCGTAGATATCGGGA	653
KB1-VC	AGAGTACAGCAGGAGAAAACGGAATTCCCGGTGTAGTGGTAAAATGCGTAGATATCGGGA	650
KB1-PCE	AGAGTACAGCAGGAGAAAACGGAATTCCCGGTGTAGTGGTAAAATGCGTAGATATCGGGA	649

EW70-01#8	GGAACACCAGAGGCGAAGGCGGTTTTCTAGGTTGTCACTGACACTGAGGCTCGAAAGCGT	712
Shaw16s#1	GGAACACCAGAGGCGAAGGCGGTTTTCTAGGTTGTCACTGACACTGAGGCTCGAAAGCGT	716
BCI#1	GGAACACCAGAGGCGAAGGCGGTTTTCTAGGTTGTCACTGACACTGAGGCTCGAAAGCGT	715
BCI#16	GGAACACCAGAGGCGAAGGCGGTTTTCTAGGTTGTCACTGACACTGAGGCTCGAAAGCGT	713
EW70-01#6	GGAACACCAGAGGCGAAGGCGGTTTTCTAGGTTGTCACTGACACTGAGGCTCGAAAGCGT	713
		715
EW70-01#3	GGAACACCAGAGGCGAAGGCGGTTTTCTAGGTTGTCACTGACACTGAGGCTCGAAAGCGT	714
BCI#17	GGAACACCAGAGGCGAAGGCGGTTTTCTAGGTTGTCACTGACGCTGAGGCTCGAAAGCGT	71 4 715
Shaw16s#2	GGAACACCAGAGGCGAAGGCGGTTTTCTAGGTTGTCACTGACACTGAGGCTCGAAAGCGT	715
EW70-01#7	GGAACACCAGAGGCGAAGGCGGTTTTCTAGGTTGTCACTGACACTGAGGCTCGAAAGCGT	
Shaw16s#3	GGAACACCAGAGGCGAAGGCGGTTTTCTAGGTTGTCACTGACACTGAGGCTCGAAAGCGT	715
BCI#15	GGAACACCAGAGGCGAAGGCGGTTTTCTAGGTTGTCACTGACACTGAGGCTCGAAAGCGT	715
EW70-01#2	GGAACACCAGAGGCGAAGGCGGTTTTCTAGGTTGTCACTGACACTGAGGCTCGAAAGCGT	715
Shaw16s#4	GGAACACCAGAGGCGAAGGCGGTTTTCTAGGTTGTCACTGACACTGAGGCTCGAAAGCGT	659
BCI#3	GGAACACCAGAGGCGAAGGCGGTTTTCTAGGTTGTCACTGACACTGAGGCTCGAAAA	713
KB1-VC	GGAACACCAGAGGCGAAGGCGGTTTTCTAGGTTGTCACTGACACTGAGGCTCGAAAGCGT	710
KB1-PCE	GGAACACCAGAGGCGAAGGCGGTTTTCTAGGTTGTCACTGACACTGAGGCTCGAAAGCGT	709

EW70-01#8	GGGGAGCGAACAGAATTAGATACTCTGGTAGTCCACGCCTTAAACTATGGACACTAGGTA	772
Shaw16s#1	GGGGAGCGAACAGAATTAGATACTCTGGTAGTCCACGCCTTAAACTATGGACACTAGGTA	776
BCI#1	GGGGAGCGAACAGAATTAGATACTCTGGTAGTCCACGCCTTAAACTATGGACACTAGGTA	775
BCI#16	GGGGAGCGAACAGAATTAGATACTCTGGTAGTCCACGCCTTAAACTATGGACACTAGGTA	773
EW70-01#6	GGGGAGCGAACAGAATTAGATACTCTGGTAGTCCACGCCTTAAACTATGGACACTAGGTA	773
EW70-01#3	GGGGAGCGAACAGAATTAGATACTCTGGTAGTCCACGCCTTAAACTATGGACACTAGGTA	775
BCI#17	GGGGAGCGAACAGAATTAGATACTCTGGTAGTCCACGCCTTAAACTATGGACACTAGGTA	774
Shaw16s#2	GGGGAGCGAACAGAATTAGATACTCTGGTAGTCCACGCCTTAAACTATGGACACTAGGTA	775
EW70-01#7	GGGGAGCGAACAGAATTAGATACTCTGG <mark>C</mark> AGTCCACGCCTTAAACTATGGACACTAGGTA	775
Shaw16s#3	GGGGAGCGAACAGAATTAGATACTCTGGTAGTCCACGCCTTAAACTATGGACACTAGGTA	775

BCI#15 EW70-01#2 Shaw16s#4 BCI#3	GGGGAGCGAACAGAATTAGATACTCTGGTAGTCCACGCCTTAAACTATGGACACTAGGTA GGGGAGCGAACAGAATTAGATACTCTGGTAGTCCACGCCTTAAACTATGGACACTAGGTA GGGGAGCGAACAGAATTAGATACTCTGGTAGTCCACGCCTTAAACTATGGACACTAGGTA GGGGAGCGAACAGAATTAGATACTCTGGTAGTCCACGCA	775 719 773
KB1-VC KB1-PCE	GGGGAGCGAACAGAATTAGATACTCTGGTAGTCCACGCCTTAAACTATGGACACTAGGTA GGGGAGCGAACAGAATTAGATACTCTGGTAGTCCACGCCTTAAACTATGGACACTAGGTA *********************************	770 769
EW70-01#8	${\tt TAGGGAGTATCGACCCTCTCTGTGCCGAAGCTAACGCTTTAAGTGTCCCGCCTGGGGAGT}$	832
Shaw16s#1	TAGGGAGTATCGACCCTCTCTGTGCCGAAGCTAACGCTTTAAGTGTCCCGCCTGGGGAGT	836
BCI#1	TAGGGAGTATCGACCCTCTCTGTGCCGAAGCTAACGCTTTAAGTGTCCCGCCTGGGGAGT	835
BCI#16	TAGGGAGTATCGACCCTCTCTGTGCCGAAGCTAACGCTTTAAGTGTCCCGCCTGGGGAGT	833
EW70-01#6	TAGGGAGTATCGACCCTCTCTGTGCCGAAGCTAACGCTTTAAGTGTCCCGCCTGGGGAGT	833
EW70-01#3	TAGGGAGTATCGACCCTCTCTGTGCCGAAGCTAACGCTTTAAGTGTCCCGCCTGGGGAGT	835
BCI#17	TAGGGAGTATCGACCCTCTCTGTGCCGAAGCTAACGCTTTAAGTGTCCCGCCTGGGGAGT	834
Shaw16s#2	TAGGGAGTATCGACCCTCTCTGTGCCGAAGCTAACGCTTTAAGTGTCCCGCCTGGGGAGT	835
EW70-01#7	TAGGGAGTATCGACCCTCTCTGTGCCGAAGCTAACGCTTTAAGTGTCCCGCCTGGGGAGT	835
Shaw16s#3	TAGGGAGTATCGACCCTCTCTGTGCCGAAGCTAACGCTTTAAGTGTCCCGCCTGGGGAGT	835
BCI#15	TAGGGAGTATCGACCCTCTCTGTGCCGAAGCTAACGCTTTAAGTGTCCCGCCTGGGGAGT	835
EW70-01#2	TAGGGAGTATCGACCCTCTCT <mark>A</mark> TGCCGAAGCTAACGCTTTAAGTGTCCCGCCTGGGGAGT	835
Shaw16s#4	TAGGGAGTATCGACCCTCTCTGTGCCGAAGCTAACGCTTTAAGTGTCCCGCCTGGGGAGT	779
BCI#3	TAGGGAGTATCGACCCTCTCTGTGCCGAAGCTAACGCTTTAAGTGTCCCGCCTGGGGAGT	833
KB1-VC	TAGGGAGTATCGACCCTCTCTGTGCCGAAGCTAACGCTTTAAGTGTCCCGCCTGGGGAGT	
KB1-PCE	TAGGGAGTATCGACCCTCTCTGTGCCGAAGCTAACGCTTTAAGTGTCCCGCCTGGGGAGT	829

EW70-01#8	ACGGTCGCAAGGCTAAAACTCAAAGGAATTGACGGGGGCCCGCACAAGCAGCGGAGCGTG	892
Shaw16s#1	ACGGTCGCAAGGCTAAA <mark>G</mark> CTCAAAGGAATTGACGGGGGCCCGCACAAGCAGCGGAGCGTG	896
BCI#1	ACGGTCGCA <mark>G</mark> GGCTAAAACTCAAAGGAATTGACGGGGGCCCGCACAAGCAGCGGAGCGTG	895
BCI#16	ACGGTCGCAAGGCTAAAACTCAAAGGAATTGACGGGGGCCCGCACAAGCAGCGGAGCGTG	893
EW70-01#6	ACGGTCGCAAGGCTAAAACTCAAAGGAATTGACGGGGGCCCGCACAAGCAGCGGAGCGTG	893
EW70-01#3	ACGGCCGCAAGGCTAAAACTCAAAGGAATTGACGGGGGCCCGCACAAGCAGCGGAGCGTG	895
BCI#17	ACGGTCGCAAGGCTAAAACTCAAAGGAATTGACGGGGGCCCGCACAAGCAGCGGAGCGTG	894
Shaw16s#2	ACGGTCGCAAGGCTAAAACTCAAAGGAATTGACGGGGGCCCGCACAAGCAGCGGAGCGTG	895
EW70-01#7	ACGGTCGCAAGGCTAAAACTCAAAGGAATTGACGGGGGCCCGCACAAGCAGCGGAGCGTG	895
Shaw16s#3	ACGGTCGCAAGGCTAAAACTCAAAGGAATTGACGGGGGCCCGCACAAGCAGCGGAGCGTG	895
BCI#15	ACGGTCGCAAGGCTAAAACTCAAAGGAAT <mark>C</mark> GACGGGGGCCCGCACAAGCAGCGGAGCGTG	895
EW70-01#2	ACGGTCGCAAGGCTAAAACTCAAAGGAATTGACGGGGGCCCGCACAAGCAGCGGAGCGTG	895

Shaw16s#4 BCI#3 KB1-VC	ACGGTCGCAAGGCTAAAACTCAAAGGAATTGACGGGGGCCCGCACAAGCAGCGGAGCGTG ACGGTCGCAAGGCTAAAACTCAAAGGAATTGACGGGGGCCCGCACAAGCAGCGGAGCGTG ACGGTCGCAAGGCTAAAACTCAAAGGAATTGACGGGGGCCCGCACAAGCAGCGGAGCGTG	893
KB1-PCE	ACGGTCGCAAGGCTAAAACTCAAAGGAATTGACGGGGGCCCGCACAAGCAGCGGAGCGTG **** *** **** ***** ******* **********	889
EW70-01#8	TGGTTTAATTCGATGCTACACGAAGAACC <mark>C</mark> TACCAAGATTTGACATGCATG <mark>G</mark> AGTAGTGA	952
Shaw16s#1	TGGTTTAATTCGATGCTACACGAAGAACCTTACCAAGATTTGACATGCATG	956
BCI#1	TGGTTTAATTCGATGCTACACGAAGAACCTTACCAAGATTTGACATGCATG	955
BCI#16	TGGTTTAATTCGATGCTACACGAAGAACCTTACCAAGATTTGACATGCATG	953
EW70-01#6	TGGTTTAATTCGATGCTACACGAAGAACCTTACTAAGATTTGACATGCATG	953
EW70-01#3	TGGTTTAATTCGATGCTACACGAAGA <mark>G</mark> CCTTACCAAGATT <u>T</u> GACATGCATGAAGTAGTGA	
BCI#17	TGGTTTAATTCGATGCTACACGAAGAACCTTACCAAGATT <mark>C</mark> GACATGCATGAAGTAGTGA	954
Shaw16s#2	TGGTTTAATTCGATGCTACACGAAGAACCTTACCAAGATTTGACATGCATG	
EW70-01#7	TGGTTTAATTCGATGCTACAC <mark>A</mark> AAGAACCTTACCAAGATTTGACATGCATGAAGTAGTGA	
Shaw16s#3	TGGTTTAATTCGATGCTACACGAAGAACCCTTACCAAGATTTGACATGCATG	
BCI#15	TGGTTTAATTCGATGCTACACGAAGAA <mark>M</mark> CTTACCAAGATTTGACATGCATGAAGTAGTGA	
EW70-01#2	TGGTTTAATTCGATGCTACACGAAGAACCTTACCAAGATTTGACATGCATG	
Shaw16s#4	TGGTTTAATTCGATGCTACACGAAGAACCTTACC <mark>T</mark> AGATTTGACATGCATGAAGTAGTGA	
BCI#3	TGGTTTAATTCGATGCTACACGAAGAACCTTACCAAGATTTGACATGCATG	
KB1-VC	TGGTTTAATTCGATGCTACACGAAGAACCTTACCAAGATTTGACATGCATG	950
KB1-PCE	TGGTTTAATTCGATGCTACACGAAGAACCTTACCAAGATTTGACATGCATG	949
EW70-01#8	ACCGAAAGGGAAACGACCTGTTAAGTCAGGAGTTTGCACAGGTGCTGCATGGCTGTCGTC	1012
Shaw16s#1	ACCGAAAGGGAAACGACCTGTTAAGTCAGGAGTTTGCACAGGTGCTGCATGGCTGTCGTC	1016
BCI#1	ACCGAAAGGGAAACGACCTGTTAAGTCAGGAGTTTGCACAGGTGCTGCATGGCTGTCGTC	1015
BCI#16	ACCGAAAGGGAAACGACCTGTTAAGTCAGGAGTTTGCACAGGTGCTGCATGGCTGTCGTC	1013
EW70-01#6	ACCGAAAGGGAAACGACCTGTTAAGTCAGGAGTTTGCACAGGTGCTGCATGGCTGTCGTC	1013
EW70-01#3	ACCGAAAGGGAAACGACCTGTTAAGTCAGGAGTTTGCACAGGTGCTGCATGGCTGTCGTC	1015
BCI#17	ACCGAAAGGGAAACGACCTGTTAAGTCAGGAGTTTGCACAGGTGCTGCATGGCTGTCGTC	1014
Shaw16s#2	ACCGAAAGGGAAACGACCTGTTAAGTCAGGAGTTTGCACAGGTGCTGCATGGCTGTCGTC	1015
EW70-01#7	ACCGAAAGGGAAACGACCTGTTAAGTCAGGAGTTTGCACAGGTGCTGCATGGCTGTCGTC	1015
Shaw16s#3	ACCGAAAGGGA <mark>G</mark> ACGACCTGTTAAGTCAGGAGTTTGCACAGGTGCTGCATGGCTGTCGTC	1015
BCI#15	ACCGAAAGGGAAACGA <mark>T</mark> CTGTTAAGTCAGGAGTTTGCACAGGTGCTGCATGGCTGTCGTC	1015
EW70-01#2	ACCGAAAGGGAAACGACCTGTTAAGTCAGGAGTTTGCACAGGTGCTGCATGGCTGTCGTC	1015
Shaw16s#4	ACCGAAAGGGAAACGACCTGTTAAGTCAGGAGTTTGCACAGGTGCTGCATGGCTGTCGTC	959
BCI#3	ACCGAAAGGGAAACGACCTGTTAAGTCAGGAGTTTGCACAGGTGCTGCATGGCTGTCGTC	1013

KB1-VC KB1-PCE	ACTGAAAGGGGAACGACCTGTTAAGTCAGGAACTTGCACAGGTGCTGCATGGCTGTCGTC ACTGAAAGGGGAACGACCTGTTAAGTCAGGAACTTGCACAGGTGCTGCATGGCTGTCGTC ** ****** *** *** *******************	1010 1009
EW70-01#8	AGCTCGTGCCGTGAGGTGTTTGG-TTAAGTCCTGCAACGAGCGCAACCC-TTGTTGCTAG	1070
Shaw16s#1	AGCTCGTGCCGTGAGGTGTTTGG-TTAAGTCCTGCAACGAGCGCAACCC-TTGTTGCTAG	1074
BCI#1	AGCTCGTGCCGTGAGGTGTTTGG-TTAAGTCCTGCAACGAGCGCAACCC-TTGTTGCTAG	1073
BCI#16	AGCTCGTGCCGTGAGGTGTTTGG <mark>G</mark> TTAAGTCCTGCAACGAGCGCAACCC <mark>C</mark> TTGTTGCTAG	1073
EW70-01#6	AGCTCGTGCCGTGAGGTGTTTGG-TTAAGTCCTGCAACGAGCGCAACCC-TTGTTGCTAG	1071
EW70-01#3	AGCTCGTGCCGTGAGGTGTTTGG-TTAAGTCCTGCAACGAGCGCAACCC-TTGTTGCTAG	1073
BCI#17	AGCTCGTGCCGTGAGGTGTTTGG-TTAAGTCCTGCAACGAGCGCAACCC-TTGTTGCTAG	1072
Shaw16s#2	AGCTCGTGCCGTGAGGTGTTTGG-TTAAGTCCTGCAACGAGCGCAACCC-TTGTTGCTAG	1073
EW70-01#7	AGCTCGTGCCGTGAGGTGTTTGG-TTAAGTCCTGCAACGAGCGCAACCC-TTGTTGCTAG	1073
Shaw16s#3	AGCTCGTGCCGTGAGGTGTTTGG-TTAAGTCCTGCAACGAGCGCAACCC-TTGTTGCTAG	1073
BCI#15	AGCTCGTGCCGTGAGGTGTTTGG-TTAAGTCCTGCAACGAGCGCAACCC-TTGTTGCTAG	1073
EW70-01#2	AGCTCGTGCCGTGAGGTGTTTGG-TTAAGTCCTGCAACGAGCGCAACCC-TTGTTGCTAG	1073
Shaw16s#4	AGCTCGTGCCGTGAGGTGTTTGG-TTAAGTCCTGCAACGAGCGCAACCC-TTGTTGCTAG	1017
BCI#3	AGCTCGTGCCGTGAGGTGTTTGG-TTAAGTCCTGCAACGAGCGCAACCC-TTGTTGCTAG	1071
KB1-VC	AGCTCGTGCCGTGAGGTGTTTGG-TTAAGTCCTGCAACGAGCGCAACCC-TTGTTGCTAG	1068
KB1-PCE	AGCTCGTGCCGTGAGGTGTTTGG-TTAAGTCCTGCAACGAGCGCAACCC-TTGTTGCTAG	1067

EW70-01#8	TTAAATTTTCTAGCGAGACTGCCCCGCGAAACGGGGAGGAAGGTGGGGATGAC <mark>A</mark> TCAAGT	1130
Shaw16s#1	TTAAATTTTCTAGCGAGACTGCCCCGCGAAACGGGGAGGAAGGTGGGGATGACGTCAAGT	1134
BCI#1	TTAAATTTTCTAGCGAGAC <mark>G</mark> CCCCGCGAAACGGGGAGGAAGGTGGGGATGACGTCAAGT	1133
BCI#16	TTAAATTTTCTAGCGAGACTGCCCCGCGAAACGGGGAGGAAGGTGGGGATGACGTCAAGT	1133
EW70-01#6	TTAAATTTTCTAGCGAGACTGCCCCGCGAAACGGGGAGGAAGGTGGGGATGACGTCAAGT	1131
EW70-01#3	TTAAATTTTCTAGCGAGACTGCCCCGCGAAACGGGGAGGAAGGTGGGGATGACGTCAAGT	1133
BCI#17	TTAAATTTTCTAGCGAGACTGCCCCGCGAAACGGGGAGGAAGGTGGGGATGACGTCAAGT	1132
Shaw16s#2	TTAAATTTTCTAGCGAGACTGCCCCGCGAAACGGGGAGGAAGGTGGGGATGACGTCAAGT	1133
EW70-01#7	TTAAATTTTCTAGCGAGACTGCCCCGCGAAACGGGGAGGAAGGTGGGGATGACGTCAAGT	1133
Shaw16s#3	TTAAATTTTCTAGCGAGACTGCCCCGCGAAACGGGGAGGAAGGTGGGGATGACGTCAAGT	1133
BCI#15	TTAAATTTTCTAGCGAGACTGCCCCGCGAAACGGGGAGGAAGGTGGGGATGACGTCAAGT	1133
EW70-01#2	TTAAATTTTCTAGCGAGACTGCCCCGCGAAACGGGGAGGAAGGTGGGGATGACGTCAAGT	1133
Shaw16s#4	TTAAATTTTCTAGCGAGACTGCCCCGCGAAACGGGGAGGAAGGTGGGGATGACGTCAAGT	
BCI#3		1131
KB1-VC	TTAAATTTTCTAGCGAGACTGCCCCGCGAAACGGGGAGGAAGGTGGGGATGACGTCAAGT	1128
KB1-PCE	TTAAATTTTCTAGCGAGACTGCCCCGCGAAACGGGGAGGAAGGTGGGGATGACGTCAAGT	

EW70-01#8 CAGCATGGCCTTTATATCTTGGGCTACACACACGCTACAATGGACAGAACAATAGGTTGC 1190 Shaw16s#1 CAGCATGGCCTTTATATCTTGGGCTACACACACCCTACAATGGACAGAACAATAGGTTGC 1194 BCT#1 CAGCATGGCCTTTATATCTTGGGCTACACACACGCTACAATGGACAGAACAATAGGTTGC 1193 BCI#16 CAGCATGGCCTTTATATCCCTGGGCTACACACACGCTACAATGGACAGAACAATAGGTTGC 1193 CAGCATGGCCTTTATATCTTGGGCTACACACACGCTACAATGGACAGAACAATAGGTTGC 1191 EW70-01#6 EW70-01#3 CAGCATGGCCTTTATATCTTGGGCTACACACACCGCTACAATGGACAGAACAATAGGTTGC 1193 BCI#17 CAGCATGGCCTTTATATCTTGGGCTACACACACGCTACAATGGACAGAACAATAGGTTGC 1192 Shaw16s#2 CAGCATGGCCTTTATATCTTGGGCTACACACACGCTACAATGGACAGAACAATAGGTTGC 1193 EW70-01#7 CAGCATGGCCTTTATATCTTGGGCTACACACACCCTACAATGGACAGAACAATAGGTTGC 1193 Shaw16s#3 CAGCATGGCCTTTATATCTTGGGCTACACACACGCTACAATGGACAGAACAATAGGTTGC 1193 BCI#15 CAGCATGGCCTTTATATCTTGGGCTACACACACCCTACAATGGACAGAACAATAGGTTGC 1193 EW70-01#2 CAGCATGGCCTTTATATCTTGGGCTACACACACGCTACAATGGACAGAGCCAATAGGTTGC 1193 Shaw16s#4 CAGCATGGCCTTTATATCTTGGGCCCACACACACGCTACAATGGACAGAACAATAGGTTGC 1137 BCI#3 CAGCATGGCCTTTATATCTTGGGCTACACACACGCTACAATGGACAGAACAGTAGGTTGC 1191 KB1-VC CAGCATGGCCTTTATATCTTGGGCTACACACACCGCTACAATGGACAGAACAATAGGTTGC 1188 KB1-PCE CAGCATGGCCTTTATATCTTGGGCTACACACACGCTACAATGGACAGAACAATAGGTTGC 1187 EW70-01#8 AACAGTGTGAACTGGAGCTAATCCCCAAAGCTGTCCTCAGTTCGGATTGCAGGCTGAAAC 1250 Shaw16s#1 AACAGTGTGAACTGGAGCTAATCCCCAAAGCTGTCCTCAGTTCGGATTGCAGGCTGAAAC 1254 BCI#1 AACAGTGTGAACTGGAGCTAATCCCCAAAGCTGTCCTCAGTTCGGATTGCAGGCTGAAAC 1253 BCI#16 AACAGTGTGAACTGGAGCTAATCCCCAAAGCTGTCCTCAGTTCGGATTGCAGGCTGAAAC 1253 AACAGTGTGAACTGGAGCTAATCCCCAAAGCTGTCCTCAGTTCGGATTGCAGGCTGAAAC 1251 EW70-01#6 EW70-01#3 AACAGTGTGAACTGGAGCTAATCCCCAAAGCTGTCCTCAGTTCGGATTGCAGGCTGAAAC 1253 BCI#17 AACAGTGTGAACTGGAGCTAATCCCCAAAGCTGTCCTCAGTTCGGATTGCAGGCTGAAAC 1252 Shaw16s#2 AACAGTGTGAACTGGAGCTAATCCCCAAAGCTGTCCTCAGTTCGGATTGCAGGCTGAAAC 1253 EW70-01#7 AACAGTGTGAACTGGAGCTAATCCCCAAAGCTGTCCTCAGTTCGGATTGCAGGCTGAAAC 1253 Shaw16s#3 AACAGTGTGAACTGGAGCTAATCCCCAAAGCTGTCCTCAGTTCGGATTGCAGGCTGAAAC 1253 BCI#15 AACAGTGTGAACTGGAGCTAATCCCCAAAGCTGTCCTCAGTTCGGATTGCAGGCTGAAAC 1253 EW70-01#2 AACAGTGTGAACTGGAGCTAATCCCCAAAGCTGTCCTCAGTTCGGATTGCAGGCTGAAAC 1253 Shaw16s#4 AACAGTGTGAACTGGAGCTAATCCCCAAAGCGGTCCTCAGTTCGGATTGCAGGCTGAAAC 1197 BCI#3 AACAGTGTGAACTGGAGCTAATCCCCAAAGCTGTCCTCAGTTCGGATTGCAGGCTGAAAC 1251 KB1-VC AACAGTGCGAACTGGAGCTAATCCCCAAAGCTGTCCTCAGTTCGGATTGCAGGCTGAAAC 1248

KB1-PCE

AACAGTGCGAACTGGAGCTAATCCCCAAAGCTGTCCTCAGTTCGGATTGCAGGCTGAAAC 1247

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CCGCCTGCATGAAGTTGGAGTTGCTAGTA<mark>T</mark>C<mark>AGG</mark>ATATCAGCAAGGTGCGGTGAATACGT 1310
EW70-01#8
Shaw16s#1
                CCGCCTGCATGAAGTTGGAGTTGCTAGTAACCGCATATCAGCAAGGTGCGGTGAATACGT 1314
BCI#1
                CCGCCTGCATGAAGTTGGAGTTGCTAGTAACCGCATATCAGCAAGGTGCGGTGAATACGT 1313
BCI#16
                CCGCCTGCATGAAGTTGGAGTTGCTAGTAACCGCATATCAGCAAGGTGCGGTGAATACGT 1313
                CCGCCTGCATGAAGTTGGAGTTGCTAGTAACCGCATATCAGCAAGGTGCGGTGAATACGT 1311
EW70-01#6
EW70-01#3
                CCGCCTGCATGAAGTTGGAGTTGCTAGTAACCGCATATCAGCAAGGTGCGGTGAATACGT 1313
BCI#17
                CCGCCTGCATGAAGTTGGAGTTGCTAGTAACCGCATATCAGCAAGGTGCGGTGAATACGT 1312
                CCGCCTGCATGAAGTTGGAGTTGCTAGTAACCGCATATCAGCAAGGTGCGGTGAATACGT 1313
Shaw16s#2
EW70-01#7
                CCGCCTGCATGAAGTTGGAGTTGCTAGTAACCGCATATCAGCAAGGTGCGGTGAATACGT 1313
Shaw16s#3
                CCGCCTGCATGAAGTTGGAGTTGCTAGTAACCGCATATCAGCAAGGTGCGGTGAATACGT 1313
BCI#15
                CCGCCTGCATGAAGTTGGAGTTGCTAGTAACCGCATATCAGCAAGGTGCGGTGAATACGT 1313
EW70-01#2
                CCGCCTGCATGAAGTTGGAGTTGCTAGTAACCGCATATCAGCAAGGTGCGGTGAATACGT 1313
Shaw16s#4
                CCGCCTGCATGAAGTTGGAGTTGCTAGTAACCGCATATCAGCAAGGTGCGGTGAATACGT 1257
BCI#3
                CCGCCTGCATGAAGTTGGAGTTGCTAGTAACCGCATATCAGCAAGGTGCGGTGAATACGT 1311
KB1-VC
                CCGCCTGCATGAAGTTGGAGTTGCTAGTAACCGCATATCAGCA<mark>T</mark>GGTGCGGTGAATACGT 1308
KB1-PCE
                CCGCCTGCATGAAGTTGGAGTTGCTAGTAACCGCATATCAGCA<mark>T</mark>GGTGCGGTGAATACGT 1307
EW70-01#8
                TCTCGGGCCTTG-ACACCCCCCCGTCACGTCATGAAAGCCGGTAACACTTGAAGTCGAT 1369
Shaw16s#1
                TCTCGGGCCTTG-ACACACCGCCCGTCACGTCATGAAAGCCGGTAACACTTGAAGTCGAT 1373
BCI#1
                TCTCGGGCCTTGTACACACCGCCCGTCACGTCATGAAAGCCGGTAACACTTGAAGTCGAT 1373
BCI#16
                TCTCGGGCCTTGTACACACCGCCCGTCACGTCATGAAAGCCGGTAACACTTGAAGTCGAT 1373
EW70-01#6
                TCTCGGGCCT-GTACACACCGCCCGTCACGTCATGAAAGCCGGTAACACTTGAAGTCGAT 1370
EW70-01#3
               TCTCGGGCCTTGTACACACCGCCCGTCACGTCATGAAAGCCGGTAACACTTGAAGTCGAT 1373
                TCTCGGGCCT-GTACACACCGCCCGTCACGTCATGAAAGCCGGTAACACTTGAAGTCGAT 1371
BCI#17
                TCTCGGGCCT-GTACACACCGCCCGTCACGTCATGAAAGCCGGTAACACTTGAAGTCGAT 1372
Shaw16s#2
EW70-01#7
                TCTCGGGCCT-GTACACACCGCCCGTCACGTCATGAAAGCCGGTAACACTTGAAGTCGAT 1372
Shaw16s#3
                TCTCGGGCCTTGTACACACCGCCCGTCACGTCATGAAAGCCGGTAACACTTGAAGTCGAT 1373
BCI#15
                TCTCGGGCCTTGTACACACCGCC----- 1336
                TCTCGGGCCTTGTACACACCGCCCGTCACGTCATGAAAGCCGGTAACACTTGAAGTCGAT 1373
EW70-01#2
Shaw16s#4
                TCTCGGGCCTTG-ACACACCGCC-----
BCI#3
                TCTCGGGCCTTGTACACACCGCCCGTCACGTCATGAAAGCCGGTAACACTTGAAGTCGAT 1371
KB1-VC
                TCTCGGGCCTTGTACACACCGCCCGTCACGTCATGAAAGCCGGTAACACTTGAAGTCGAT 1368
KB1-PCE
                TCTCGGGCCTTGTACACACCGCCCGTCACGTCATGAAAGCCGGTAACACTTGAAGTCGAT 1367
EW70-01#8
                GTGC----- 1373
                GTGCCAACC-AAGGG--- 1387
Shaw16s#1
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BCI#1	GTGCCAACC-AAGGGC	1388
BCI#16	GTGCCAACC-AAGGGC	1388
EW70-01#6	GTGCCAACC-AAGGGC	1385
EW70-01#3	GTGCCAACC-AAGGGC	1388
BCI#17	GTGCCAACC-AAGGGC	1386
Shaw16s#2	GTGCCAACCCAAGGGC	1388
EW70-01#7	GTGCCAACC-AAGGGC	1387
Shaw16s#3	GTGCCAACC-AAGGGC	1388
BCI#15		
EW70-01#2	GTGCCAACC-AAGGGC	1388
Shaw16s#4		
BCI#3	GTGCCAACC-AAGGGC	1386
KB1-VC	GTGCCAACCGCAAGGAGG	1386
KB1-P		

vcrA Gene analysis. Quantitative PCR analysis suggested that the functional reductase gene vcrA was not detected within the Seal Beach site 70 environmental sample, but was present in high concentrations in all three bioaugmentation cultures. Therefore, this reductase gene was identified as the preliminary target for tracking the growth and transport of the bioaugmentation culture in the field. In order to determine if there are significant differences between the vcrA gene sequences present within the bioaugmentation cultures, clone libraries were constructed using vcrA-specific PCR primers. First, PCR was performed using vcrA primers identified in Table 1 to generate an approximately 1,400 bp PCR product of the vcrA gene in the Seal Beach Site 70 sample EW70-01, and bioaugmentation cultures Shaw and BCI. The Seal Beach Site 70 sample did not amplify, confirming that the vcrA gene was not detected using either the QPCR or PCR protocols described. The BCI bioaugmentation culture, however, did not amplify either. Therefore, while QPCR analysis identified high gene copy numbers of vcrA within this culture, the long primer set used for the clone library construction did not amplify, and therefore a clone library could not be constructed.

A clone library targeting *vcrA* was generated using the Shaw bioaugmentation culture, and four clones were sequenced. The approximately 1400 bp DNA sequence obtained from each clone was initially aligned against known sequences using the BLAST tool (Table 4) in order to determine the closest match with sequences in the GenBank database. In addition to the sequences obtained from the library, an alignment was generated using a ClustalW algorithm (http://www.ebi.ac.uk/clustalw/) with published sequence for vcrAKB1RdhAB14 *vcrA* from bioaugmentation culture KB-1, and from *Dehalococcoides* strain VS (Table 4 and Figure 2). The GenBank alignment suggested that all four Shaw vcrA sequences most closely matched the *vcrA* gene published for *Dehalococcoides* strain VS with greater than 99% sequence similarity (Table 4).

Figure 2 illustrates the DNA sequence alignment for the Shaw *vcrA* clone sequences, and the *vcrA* sequence from *Dehalococcoides* strain VS and the KB-1 *vcrA* published sequence. All of the sequences evaluated were highly similar, with little distinction between the different strains. These data will be archived and evaluated further should indigenous strains of *vcrA* be detected in the field at Seal Beach following biostimulation, but before bioaugmentation.

Table 4. Genebank results for the reductase gene *vcrA* clone library results for the Shaw bioaugmentation culture.

Clone	target	Closest GenBank match	% similarity	Citation
Shaw		Bacterium VS vinyl- chloride reductive	1 422 /1 442	Muller, et al 2004 AEM. 70 (8), 4880-
vcrA #2	vcrA	dehalogenase operon AY322364.1	1433/1442 (99%)	4888
		Bacterium VS vinyl- chloride reductive		Muller, et al 2004
Shaw vcrA #5	vcrA	dehalogenase operon <u>AY322364.1</u>	1384/1393 (99%),	AEM. 70 (8), 4880- 4888

Shaw vcrA #1	vcrA	Bacterium VS vinyl- chloride reductive dehalogenase operon AY322364.1	1381/1391 (99%)	Muller, et al 2004 AEM. 70 (8), 4880- 4888
Shaw vcrA #3	vcrA	Bacterium VS vinyl- chloride reductive dehalogenase operon AY322364.1	1375/1381 (99%)	Muller, et al 2004 AEM. 70 (8), 4880- 4888

Figure 2. Sequence alignment from Shaw and KB1 vcrA sequences and Strain VS.

ShawvcrA#3 ShawvcrA#5 ShawvcrA#1 ShawvcrA#2 OperonfromStrainVS vcrAKB1RdhAB14	CTTCAGATGAGAATGTCAGGTGAAGAGCAAAAGAAGCGAATTTTTCAGATGAGAATGTCAGGTGAAGAGCAAAAGAAGCGAATTTTCTTCAGATGAGAATGTCAGGTGAAGAGCAAAAGAAGCGAATTTTGGGCATAGGCTTCAGATGAGAATGTCAGGTGAAGAGCCAAAAGAAGCGAATTTT ATCATGGGGCAATAGGCTTCAGGTGAGAATGTCAGGTGAAGAGCAAAAGAAGCGAATTTT ATCATGGGGCAATAGGCTTCAGATGAGAATGTCAGGTGAAGAGCCAAAAGAAGCGAATTTT ********************************	44 42 44 53 960 292
ShawvcrA#3 ShawvcrA#5 ShawvcrA#1 ShawvcrA#2 OperonfromStrainVS vcrAKB1RdhAB14	GGCCGCTAAAAAAGAGAGKTTCCCTGGTTGGGACGGTGGGTTACACGGGAGAGGGGATCA GGCCGCCAAAAAAAGAGAGGTTCCCTGGTTGGGACGGTGGGTTACACGGGAGAGGGGATCA GGCCGCTAAAAAAGAGAGGTTCCCTGGTTGGGACGGTGGGTTACACGGGAGAGGGGATCA GGCCGCTAAAAAAAGAGAGGTTCCCTGGTTGGGACGGTGGGTTACACGGGAGAGGGGATCA GGCCGCTAAAAAAAGAGAGGTTCCCTGGTTGGGACGGTGGGTTACACGGGAGAGGGGATCA GGCCGCTAAAAAAAGAGAGGTTCCCTGGTTGGGACGGTGGGTTACACGGGAGAGGGGATCA ****** ******************************	102 104 113 1020
ShawvcrA#3 ShawvcrA#5 ShawvcrA#1 ShawvcrA#2 OperonfromStrainVS vcrAKB1RdhAB14	GCGGGCGGATGCACTATTTTACGCAGTAACTCAACCATTTCCTGGTAGTGATGAGGAAGG GCGGGCGGATGCACTATTTTACGCAGTAACTCAACCATTTCCTGGTAGTGATGAGGAAGG GCGGGCGGATGCACTATTTTACGCAGTAACTCAGCCACTTCCTGGTAGTGGTGAGGAAGG GCGGGCGGATGCACTATTTTACGCAGTAACTCAACCATTTCCTGGTAGTGGTGAGGAAGG GCGGGCGGATGCACTATTTTACGCAGTAACTCAACCATTTCCTGGTAGTGTGAGGAAGG GCGGGCGGATGCACTATTTTACGCAGTAACTCAACCATTTCCTGGTAGTGTGAGGAAGG **********************	162 164 173 1080
ShawvcrA#3 ShawvcrA#5 ShawvcrA#1 ShawvcrA#2 OperonfromStrainVS vcrAKB1RdhAB14	GCACGGACTATTCCAACCTTATCCTGATCAACCCGGTAAGTTTTACGCGAGATGGGGTTT GCACGGACTATTCCAACCTTATCCTGATCAACCCGGTAAGTTTTACGCGAGATGGGGTTT GCACGGACTATTCCAACCTTATCCTGATCAACCCGGTAAGTTTTACGCGAGATGGGGTTT GCGCGGACTATTCCAACCTTATCCTGATCAACCCGGTAAGTTTTACGCGAGATGGGGTTT GCACGGACTATTCCAACCTTATCCTGATCAACCCGGTAAGTTTTACGCGAGATGGGGTTT GCACGGACTATTCCAACCTTATCCTGATCAACCCGGTAAGTTTTACGCGAGATGGGGTTT ** ****************************	224 222 224 233 1140 472
ShawvcrA#3 ShawvcrA#5 ShawvcrA#1 ShawvcrA#2 OperonfromStrainVS vcrAKB1RdhAB14	GTATGGTCCGCCACATGATTCAGCGCCACCTGATGGGAGCGTACCAAAATGGGAGGGTAC GTATGGTCCGCCACGTGATTCAGCGCCACCTGATGGGAGCGTACCAAAATGGGAGGGTAC GTATGGTCCGCCACATGATTCAGCGCCACCTGATGGGAGCGTACCAAAATGGGAGGGTAC GTATGGTCCGCCACATGATTCAGCGCCACCTGATGGGAGCGTACCAAAATGGGAGGGTAC GTATGGTCCGCCACATGATTCAGCGCCACCTGATGGGAGCGTACCAAAATGGGAGGGTAC GTATGGTCCGCCACATGATTCAGCGCCACCTGATGGGAGCGTACCAAAATGGGAGGGTAC GTATGGTCCGCCACATGATTCAGCGCCACCTGATGGGAGCGTACCAAAATGGGAGGGTAC	

ShawvcrA#3 ShawvcrA#5 ShawvcrA#1 ShawvcrA#2 OperonfromStrainVS vcrAKB1RdhAB14	TCCAGAAGACAATTTTCTAATGCTGAGGGCAGCTGCAAAATATTTTGGTGCTGGTGGCGT TCCAGAAGACAATTTTCTAATGCTGAGGGCAGCTGCAAAATATTTTTGGTGCTGGTGGCGT TCCAGAAGACAATTTTCTAATGCTGAGGGCAGCTGCAAAATATTTTTGGTGCTGGTGGCGT TCCAGAAGACAATTTTCTAATGCTGAGGGCAGCTGCAAAATATTTTTGGTGCTGGTGGCGT TCCAGAAGACAATTTTCTAATGCTGAGGGCAGCTGCAAAATATTTTTGGTGCTGGTGGCGT TCCAGAAGACAATTTTCTAATGCTGAGGGCAGCTGCAAAATATTTTTGGTGCTGGTGGCGT ********************	342 344 353 1260
ShawvcrA#3	TGGTGCTCTTAACCTGGCAGATCCCAAATGCAAAAAACTAATATATAAGAAAGCTCAGCC	
ShawvcrA#5	TGGTGCTCTTAACCTGGCAGATCCCAAATGCAAAAAACTAATATATAAGAAAGCTCAGCC	
ShawvcrA#1	TGGTGCTCTTAACCTGGCAGATCCCAAATGCAAAAAACTAATATATAAGAAAGCTCAGCC	
ShawvcrA#2	TGGTGCTCTTAACCTGGCAGATCCCAAATGCAAAAAACTAATATATAAGAAAGCTCAGCC	
OperonfromStrainVS vcrAKB1RdhAB14	TGGTGCTCTTAACCTGGCAGATCCCAAATGCAAAAAACTAATATAAAGAAAG	
VCIARBIRGHAB14	1GG1GC1C11AACC1GGCAGA1CCCAAA1GCAAAAAAC1AA1A1A1AAAAAAGC1CAGCC **********************************	052
ShawvcrA#3	GATGACTCTAGGAAAAGGAACATACAGTGAAATAGGTGGACCAGGAATGATCGATGCAAA	464
ShawvcrA#5	GATGACTCTAGGAAAAGGAACATACAGTGAAATAGGTGGACCAGGAATGATCGATGCAAA	462
ShawvcrA#1	GATGACTCTAGGAAAAGGAACATACAGTGAAATAGGTGGACCAGGAATGATCGATACAAA	464
ShawvcrA#2	GATGACTCTAGGAAAAGGAACATACAGTGAAATAGGTGGACCAGGAATGATCGATGCAAA	473
OperonfromStrainVS	GATGACTCTAGGAAAAGGAACATACAGTGAAATAGGTGGACCAGGAATGATCGATGCAAA	1380
vcrAKB1RdhAB14	GATGACTCTAGGAAAAGGAACATACAGTGAAATAGGTGGACCAGGAATGATCGATGCAAA	712

ShawvcrA#3	ATTTTATCCCAGGGTTCCTGACCATGCCGTACCTATTAACTTTAAGGAAGCGGATTATAG	524
ShawvcrA#5	ATTTTATCCCAAGGTTCCTGACCATGCCGTACCTATTAACTTTAAGGAAGCGGATTATAG	
ShawvcrA#1	ATTTTATCCCAAGGTTCCTGACCATGCCGTACCTATTAACTTTAAGGAAGCGGATTATAG	
ShawvcrA#2		533
OperonfromStrainVS	AATTTATCCCAAGGTTCCTGACCATGCCGTACCTATTAACTTTAAGGAAGCGGATTATAG	
vcrAKB1RdhAB14	ATTTTATCCCAAGGTTCCTGACCATGCCGTACCTATTAACTTTAAGGAAGCGGATTATAG	772
	* ******** ****************************	
ShawvcrA#3	CTACTACAATGATGCAGAGTGGGTTATTCCAACAAAGTGTGAATCCATTTTCACTTTCAC	584
ShawvcrA#5	$\tt CTACTACAATGATGCAGAGTGGGTTATTCCAACAAAGTGTGAATCCATTTTCACTTTCAC$	582
ShawvcrA#1	$\tt CTACTACAATGATGCAGAGTGGGTTATTCCAACAAAGTGTGAATCCATTTTCACTTTCAC$	584
ShawvcrA#2	CTACTACAATGATGCAGAGTGGGTTATTCCAACAAA GTGTGAATCCATTTTCACTTTCAC	593
OperonfromStrainVS	$\tt CTACTACAATGATGCAGAGTGGGTTATTCCAACAAAGTGTGAATCCATTTTCACTTTCAC$	1500
vcrAKB1RdhAB14	$\tt CTACTACAATGATGCAGAGTGGGTTATTCCAACAAAGTGTGAATCCATTTTCACTTTCAC$	832

ShawvcrA#3	CCTACCTCAACCACAAGAACTCAATAAGAGGACGGGTGGTATAGCAGGTGCTGGATCATA	644
ShawvcrA#5	$\tt CCTACCTCAACCACAAGAACTCAATAAGAGGACGGGTGGTATAGCAGGTGCTGGATCATA$	642

ShawvcrA#1 ShawvcrA#2 OperonfromStrainVS vcrAKB1RdhAB14	CCTACCTCAACCACAAGAACTCAATAAGAGGACGGGTGGTATAGCAGGTGCTGGATCATA CCTACCTCAACCACAAGAACTCAATAAGAGGACGGGTGGTATAGCAGGTGCTGGATCATA CCTACCTCAACCACAAGAACTCAATAAGAGGACGGGTGGTATAGCAGGTGCTGGATCATA CCTACCTCAACCACAAGAACTCAATAAGAGGACGGGTGGTATAGCAGGTGCTGGATCATA *********************************	653 1560
ShawvcrA#3 ShawvcrA#5 ShawvcrA#1 ShawvcrA#2 OperonfromStrainVS vcrAKB1RdhAB14	TACTGTATACAAAGATTTCGCTAGGGTAGGCACTTTAGTCCAAATGTTTATTAAGAATCT TACTGTATACAAAGATTTCGCTAGGGTAGGCACTTTAGTCCAAATGTTTATTAAGAATCT TACTGTATACAAAGATTTCGCTAGGGTAGGCACTTTAGTCCAAATGTTTATTAAGAATCT TACTGTATACAAAGATTTCGCTAGGGTAGGCACTTTAGTCCAAATGTTTATTAAGAATCT TACTGTATACAAAGATTTCGCTAGGGTAGGCACTTTAGTCCAAATGTTTATTAAGTATCT TACTGTATACAAAGATTTCGCTAGGGTAGGCACTTTAGTCCAAATGTTTATTAAGTATCT **********************************	702 704 713 1620
ShawvcrA#3 ShawvcrA#5 ShawvcrA#1 ShawvcrA#2 OperonfromStrainVS vcrAKB1RdhAB14	AGGTTATCACGCTTTATATTGGCCAATTGGATGGGGACCGGGTGGTTGCTTTACCACTTT AGGTTATCACGCTTTATATTGGCCAATTGGATGGGGACCGGGTGGTTGCTTTACCACTTT AGGTTATCACGCTTTATATTGGCCAATTGGATGGGGACCGGGTGGTTGCTTTACCACTTT AGGTTATCACGCTTTATATTGGCCAATTGGACGGGACCGGGTGGTTGCTTTACCACTTT AGGTTATCACGCTTTATATTGGCCAATTGGATGGGGACCGGGTGGTTGCTTTACCACTTT AGGTTATCACGCTTTATATTGGCCAATTGGATGGGGACCGGGTGGTTGCTTTACCACTTT ***************************	762 764 773 1680
ShawvcrA#3 ShawvcrA#5 ShawvcrA#1 ShawvcrA#2 OperonfromStrainVS vcrAKB1RdhAB14	TGACGGGCAAGGTGAACAGGGTAGAACAGGTGCTGCTATCCATTGGAAGTTTGGTTCTTC TGACGGGCAAGGTGAACAGGTAGAACAGGTGCTGCTATCCATTGGAAGTTTGGTTCTTC TGACGGGCAAGGTGAACAGGTAGAACAGGTGCTGCTATCCATTGGAAGTTTGGTTCTTC TGACGGGCAAGGTGAACAGGGTAGAACAGGTGCTGCTATCCATTGGAAGTTTGGTTCTTC TGACGGGCAAGGTGAACAGGGTAGAACAGGTGCTGCTATCCATTGGAAGTTTGGTTCTTC TGACGGGCAAGGTGAACAGGGTAGAACAGGTGCTGCTATCCATTGGAAGTTTGGTTCTTC TGACGGGCAAGGTGAACAGGGTAGAACAGGTGCTGCTATCCATTGGAAGTTTGGTTCTTC ***************************	822 824 833 1740
ShawvcrA#3 ShawvcrA#5 ShawvcrA#1 ShawvcrA#2 OperonfromStrainVS vcrAKB1RdhAB14	ACAACGTGGTTCTGAAAGAGTAGTAA-CTGATTTACCGATAGCTCCTACCCCGCCAATTG ACAACGTGGTTCTGAAAGAGTAGTAA-CTGATTTACCGATAGCTCCTACCCCGCCAATTG ACAACGTGGTTCTGAAAGAGTAGTAA-CTGATTTACCGATAGCTCCTACCCCGCCAATTG ACAACGTGGTTCTGAAAGAGTAGTAAACTGATTTACCGATAGCTCCTACCCCGCCAATTG ACAACGTGGTTCTGAAAGAGTAATAA-CTGATTTACCGATAGCTCCTACCCCGCCAATTG ACAACGTGGTTCTGAAAGAGTAATAA-CTGATTTACCGATAGCTCCTACCCCGCCAATTG **********************************	881 883 893 1799
ShawvcrA#3 ShawvcrA#5 ShawvcrA#1 ShawvcrA#2	ATGCAGGTATGTTT-GAGTTTTGCAAAACCTGTCATATATGCCGTGACGTTTGCGTCTCT ATGCAGGTATGTTTTGAGTTTTGCMAAACCTGTCATATATGCCGTGACGTTTGCGTCTCT ATGCAGGTATGTTT-GAGCTTTGCAAAACCTGTCATATATGCCGTGACGTTTGCGTCTCT ATGCAGGTATGTTT-GAGTTTTGCAAAACCTGTCATATATGCCGTGACGTTTGCGTCTCT	941 942

OperonfromStrainVS vcrAKB1RdhAB14	ATGCAGGTATGTTT-GAGTTTTGCAAAACCTGTTATATATGCCGTGACGTTTGCGTCTCT ATGCAGGTATGTTT-GAGTTTTGCAAAACCTGTTATATATGCCGTGACGTTTGCGTCTCT **************************	
ShawvcrA#3 ShawvcrA#5 ShawvcrA#1 ShawvcrA#2 OperonfromStrainVS vcrAKB1RdhAB14	GGGGGTGTGCACCAAGAAGACGAACCAACTTGGGATTCAGGTAATTGGTGGAATGTACAA GGGGGTGTGCACCAAGAAGACGAACCAACTTGGGATTCAGGTAATTGGTGGAATGTACAA GGGGGTGTGCACCAAGAAGACGAACCAACTTGGGATTCAGGTAATTGGTGGAATGTACAA GGGGGTGTGCACCAAGAAGACGAACCAACTTGGGATTCAGGTAATTGGTGGAATGTACAA GGGGGTGTGCACCAAGAAGACGAACCAACTTGGGATTCAGGTAATTGGTGGAATGTACAA GGGGGTGTGCACCAAGAAGACGAACCAACTTGGGATTCAGGTAATTGGTGGAATGTACAA **********************************	1001 1002 1012 1918
ShawvcrA#3 ShawvcrA#5 ShawvcrA#1 ShawvcrA#2 OperonfromStrainVS vcrAKB1RdhAB14	GGATATCTCGGCTACCGAACGGATTGGAGTGGTTGCCATAACCAGTGCGGTATGTGTCA- GGATATCTCGGCTACCGAACGGATTGGAGTGGTTGCCATAACCAGTGCGGTATGTGTCA- GGATATCTCGGCTACCGAACGGATTGGAGTGGTTGCCATAACCAGTGCGGTATGTGTCA- GGATATCTCGGCTACCGAACGGATTGGAGTGGTTGCCATAACCAGTGCGGTATGTGTCAC GGATATCTCGGCTACCGAACGGATTGGAGTGGTTGCCATAACCAGTGCGGTATGTGTCA- GGATATCTCGGCTACCGAACGGATTGGAGTGGTTGCCATAACCAGTGCGGTATGTGTCA- ************************************	1060 1061 1072 1977
ShawvcrA#3 ShawvcrA#5 ShawvcrA#1 ShawvcrA#2 OperonfromStrainVS vcrAKB1RdhAB14	ATCCTCCTGCCCTTTTACTTATTTAGGTTTGGAAAATGCTTCATTAGTGCACAAAATAGT ATCCTCCTGCCCTTTTACTTATTTAGGTTTGGAAAATGCTTCATTAGTGCACAAAATAGT ATCCTCCTGCCCTTTTACTTATTTAGGTTTGGAAAATGCTTCATTAGTGCACAAAATAGT ATCCTCCTGCCCTTTTACTTATTTAGGTTTTGGAAAATGCTTCATTAGTGCACAAAATAGT ATCCTCCTGCCCTTTTACTTATTTAGGTTTTGGAAAATGCTTCATTAGTGCACAAAATAGT ATCCTCCTGCCCTTTTACTTATTTAGGTTTTGGAAAATGCTTCATTAGTGCACAAAATAGT *******************************	1120 1121 1132 2037
ShawvcrA#3 ShawvcrA#5 ShawvcrA#1 ShawvcrA#2 OperonfromStrainVS vcrAKB1RdhAB14	AAAAGGTGTTGTTGCTAACACGACTGTTTTTAATAGTTTTTTTACCAATATGGAGAAAGC AAAAGGTGTTGTTGCTAACACGACTGTTTTTAATAGTTTTTTTACCAATATGGAGAAAGC AAAAGGTGTTGTTGCTAACACGACTGTTTTTAATAGTTTTTTTACCAATATGGAGAAAGC AAAAGGTGTTGTTGCTAACACGACTGTTTTTAATAGTTTTTTTACCAATATGGAGAAAGC AAAAGGTGTTGTTGCTAACACGACTGTTTTTAATAGTTTTTTTACCAATATGGAGAAAGC AAAAGGTGTTGTTGCTAACACGACTGTTTTTAATAGTTTTTTTACCAATATGGAGAAAGC ******************************	1180 1181 1192 2097
ShawvcrA#3 ShawvcrA#5 ShawvcrA#1 ShawvcrA#2 OperonfromStrainVS vcrAKB1RdhAB14	ATTAGGATATGGTGATTTAACCATGGAAAATTCTAACTGGTGGAAAGAAGAAGACCGAT ATTAGGATATGGTGATTTAACCATGGAAAATTCTAACTGGTGGAAAGAAGAAGACCGAT ATTAGGATATGGTGATTTAACCATGGAAAATTCTAACTGGTGGAAAGAAGAAGACCGAT ATTAGGATATGGTGATTTAACCATGGAAAATTCTAACTGGTGGAAAGAAGAAGAAGACCGAT ATTAGGATATGGTGATTTAACCATGGAAAATTCTAACTGGTGGAAAGAAGAAGAAGACCGAT ATTAGGATATGGTGATTTAACCATGGAAAATTCTAACTGGTGGAAAGAAGAAGAAGACCGAT	1240 1241 1252 2157

ShawvcrA#3 ShawvcrA#5 ShawvcrA#1 ShawvcrA#2 OperonfromStrainVS vcrAKB1RdhAB14		1300 1301

ShawvcrA#3 ShawvcrA#5 ShawvcrA#1 ShawvcrA#2 OperonfromStrainVS vcrAKB1RdhAB14	TTGAAATGGATGCTATATATTTTTTCTTAACAATTGCATTAGCAGTTGGACTAACTA	1360 1361 1372
ShawvcrA#3 ShawvcrA#5 ShawvcrA#1 ShawvcrA#2 OperonfromStrainVS vcrAKB1RdhAB14	TATTTACCTGGTTTAAAAAGAATAATATCACTTTAAAGTGGAATGAGTGGGTACTTG-CA TATTTACCTGGTTTAAAAAGAATAATATCACTTTAAAGTGGAATGAGTGGGTACTTG-CA TATTTACCTGGTTTAAAAAGAATAATATCACTTTAAAGTGGAATGAGTGGGTACTTG-CA TATTTACCTGGTTTAAAAAGAATAATATCACTTTAAAGTGGAATGAGTGGGTACTTG-CA TATTTACCTGGTTTAAAAAGAATAATATCACTTTAAAGTGGAATGAGTGGGTACTTG-CA TATTTACCTGGTTTAAAAAGAATAATATCACTTTAAAGTGGAATGAGTGGGTACTTGGCA	1419 1420 1431
ShawvcrA#3 ShawvcrA#5 ShawvcrA#1 ShawvcrA#2 OperonfromStrainVS vcrAKB1RdhAB14	TATTGGGGCTGTTAAGGGGGGTAATCTTGGGCATATCTGTTTCCTGAG TATTGGGGCTGTTACAAGGGC TATTGGGGCTGTTACAAGGGC	1441

Appendix E Well Logs and Well Completion Information

Appendix E.1 Lithologic Logs

AEW	1		Site/Location: Naval Weapons Station, Seal Beach	
aval Facilitie	es Eng. S	Serv. Ctr.)	Project No.: 50999-56254-6225.001.TK4.EQUIP	Page of
RSI Drilling			Drill Method: Hollow Stem Auger	Elevation:
9/12/0	7		Sampling Method: CA-Modified Split Spoon — CC	Restrict to
9/12/6	7		Total Depth: Remarks:	
Kristeen Be	ennett	H.	Depth to Water:	
Recovery (feet)	PID (ppm)	USCS	Description	Well Constructio Detail
8	2 3	0.00	4" Asphalt cored (18" diameter) 4' Rood bose (sord gravel mixture) 4-5' Native clay (see below) 5-7.5: Clay: Black (7.542.511)	
56%0"	24.5	CL	f. to C, SA, poorly graded sand; 5% micaceous silt; dry. 7.5-10: similar to above w/ char in colo/ to very dark grayish brown (2.548/2) w/ organic material and	roze
	18.3	CYsc	10-12.5-5, milar to above 12.5-15: Clay w/ Jand; Darhgrayi brown (2.54412); 60% moderatel plastic clay; 30% tisand, Sp, we graded 1100% micace ous sitt; some laminar bedding; Fe-oxid mottling; organic debus	7
60"/60"	3.7	5c	19.5-20 - Clayey Sand: Dark Olivegra (54312): 55% f. tom, sand, 58, poor	my my
48"	7.4	SM	20(?) -25: 5ilty Sand; Olive grave 10500 f. to m. sand, SR, well grave 2500 micaceous Silty 1000 was	(5Y412)i
31"	0.2	Sm/m		net
	Recovery (feet) Solonia	2.3 So / 60 / 24.5 2.4 3.9 18.3 55 / 60 / 24.5 2.4 4.2 1.60 / 4.2 1.60 / 4.2 1.7 4.7 4.7 4.7 4.7 4.7 7.4 1.2 7.2	2.3 Sto (60" 1.2 50" 1.3 50" 1.	aval Facilities Eng. Serv. Ctr.) Project No.: 50999-56254-6225.001.TK4.EQUIP RSI Drilling 9 1/2 07 9 1/2 07 Total Depth: Recovery PID (feet) PID (feet)

Boring/ W	ell No.:	AEW	1		Site/Location: Naval Weapons Station, Seal Beach	
		val Facilitie		erv. Ctr.)	Project No.: 50999-56254-6225.001.TK4.EQUIP	Page 2 of 2
		SI Drilling	0		Drill Method: Hollow Stem Auger	Elevation:
Date Start		112107			Sampling Method: CA-Modified Split Spoon	
Date Ende	ed: 9	112/67			Total Depth: Remarks:	
On-site Geologist: Kristeen Bennett			Depth to Water:			
Depth (feet)	Blow Count	Recovery (feet)	PID (ppm)	USCS	Description	Well Construction Detail
3 5		37/60	1.3		35-37-No recovery - 7 Sitt? 37-39': Clay wil Gravel! Orice (544) 50% plastic clay 125% of tom, SR growel; 15% micarcous sitt; 15 fitoe, 5t to SR sand, poorly grad gravel; wet. 39-40'-Similar to 25-30' T.D. of botchole = 35 ft. bags T.D. of well	((3)) 5 A to 6%

	Vell No.:	AIW		· C · \	Site/Location: Naval Weapons Station, Seal Beach	ln 1
		val Facilitie	es Eng. S	erv. Ctr.)	Project No.: 50999-56254-6225.001.TK4.EQUIP	Page of Elevation:
		SI Drilling			Drill Method: Hollow Stem Auger	Elevation:
Date Star		111/07			Sampling Method: CA-Modified Split Spoon	
Date End	ed:	1/11/07			Total Depth: 35.1 Array Remarks:	
On-site Geologist: Kristeen Bennett			ennett		Depth to Water:	
Depth (feet)	Blow Count	Recovery (feet)	PID (ppm)	USCS	Description	Well Construction Detail
5	Not	20"/00"	0.7		3" Asphalt cored 18" diameter 5 of road base removed (Sand) gravel mixture) to 84 bags Hand augured to 84 bags 8.5-9.0(?) -Road base -> Eravelly Olive brown (2.545/4); 20% f. t	oc.,
0		52"/100"	6.0	CV/mi	poorly graded, SA gravel; 60% of to poorly graded, SA to SR sand 1150 micaceous silt i 50% plastic clay. 9.0-10.0 (?) Silty Clay i Black (2.5% 20% productly plastic clay i 25% o ceous silt i 50% d. to m. sand, well SA to SR imarst10-11.5 similar to about when color to Olive brown (2.5% 4)	(2.5/1)) wica- graded, erge in
2 0		(w)'/ (wo*	0.3	SC	-11.5-13 simular to above when color to Very dark grayish to (2543/2) 13-15 (?) Sandy Clay: Dark yellow (10423/4); 50% moderately place 30% of to c. sand, poorly grades 20% mica come Silt; moist ifer	vishbrown tic clay; d satosk;
2 5_		541/00"	2.3	CL	mothling w/ organic debris "balls, 15-17-5. miler to above w/ mon sund (40%) 17-18.5-5. miler to above w/ one color to boy dark grayish brows 18.5-20". Clay w/ sand: Light gra	rgein (loye3/2)
-	1	44"/60"	0.7	SM	75% plastic, franke (?) chy; 19 m. sand, SA to SP, well graded wicaceous silt; moist. 20-21.5 - similar to 15-17. 21.5-22.5 - similar 18.5-20"	50/0 t. to
3 O			4.1	me	703-420	

Boring/ Well No.:	AII			Site/Location: Naval Weapon		- 4
Client: NFESC (N		es Eng. S	Serv. Ctr.)	Project No.: 50999-56254-6225		Page 2 of 2
Drill Contractor: I				Drill Method: Hollow Stem A		Elevation:
Date Started:	9/11/07			Sampling Method: CA-Modi		
Date Ended:				Total Depth: 35.1	Remarks:	
On-site Geologist: Kristeen Bennett		Depth to Water:				
Depth (feet) Blow Count	Recovery (feet)	PID (ppm)	USCS	Descri	ption	Well Construction
3 5	47/60"	0.3 0.V 0.6 2.2	my ci	22.5-24.5: Silt: 0 Wicaceous Silt; 1 f. sure; SP, well -24-24.2 - Sand to 21.5-22.5 24.5 - 28/2/Silty Sa Loselo f. sand; well micaceous silt; 1 28(2)-30: Silt; Ofi micaceous silt; 30-33(2): Silt; (544/2); boelo micaceous silt; biodurbation (233-35(2)- Simi	510 plastic clay) Legraded; net Ly clay (exist sin Ly clay (exist sin Ly clay) (exist sin Solo plastic clay ve (54 4/3); 90% 10% plastic clay; (Clay: Olive gra plastic clay; 40% net; heavy ?) lex to 20-30	Just met

Boring/ Well No.:	11200	`		Site/Location: Naval Weapons Station, Seal Beach	
Client: NFESC (N	aval Facilitie	es Eng. S	Serv. Ctr.)	· · · · · · · · · · · · · · · · · · ·	
Drill Contractor: 1	RSI Drilling			Drill Method: Hollow Stem Auger	Elevation:
	1110107			Sampling Method: CA-Modified Split Spoon	
Date Ended:	1110107			Total Depth: 35 Remarks:	
On-site Geologist	Kristeen Bennett Depth to Water:				
Depth Blow (feet) Count	Recovery (feet)	PID (ppm)	USCS	Description	Well Construction Detail
3 5	60"/60"	6.5 3.2 0.9 2.1	nyon	27.8-24.5; Sitty clay: Que (515)3 50% plastic clay; took vicaceous sitt; 10% fi to me sand, SA, well 245-25; similar to 15.5 to 18 26-28 D(): Sitty Sand: Office gra 10% f. to m. sand, Sk; well graded micaceous silt; 5% plastic clay Fe-oxide mothling. 28.0(?)-36; SandySitt; Office gra 50% micaceous silt; 25% f. sand, 50% micaceous silt;	groded', (544/2) 25% met', me(544/2); we(19raded)

Boring/ Well No.:	AIW	1		Site/Location: Naval Weapons Station, Seal Beach			
Client: NFESC (Naval Facilities Eng. Serv. Ctr.)						Page 1 of 2	
Drill Contractor: I	RSI Drilling	1		Drill Method: Hollow Stem Auger		Elevation:	
Date Started:	9/10/0	7		Sampling Method: CA-Modified Split Spoon			
Date Ended:					Remarks:		
On-site Geologist	Kristeen Ber	nnett		Depth to Water:			
Depth Blow (feet) Count	Recovery (feet)	PID (ppm)	USCS	Descriptio	on	Well Construction Detail	
2 5	24"/w" (2) (2) (3) (4) (4) (6) " 7	13.7 13.7 13.7 13.1 13.1 13.2 17.8 155.7	SC CL SC CYSC CYML SC	5.5-7.2(10.2)? Clay: 5.5-7.2(10.2)? Clay: 80°10 moderately sit. 10.2-12.2: Similar d in color to olive 12.2-13.2: Similar 10.2-13.2: Similar 10.2-13.5: Similar 10.3-15.5 Similar	Non dark gray bolo plastic to a dark gray ish to above will change with a change will gray thin con land gray ish bolo micaceaus si land gray ish come with a come will graded, show with a come with	2543/1); 2543/1); 2543/1); 2543/1); 2543/1); 2543/1); 2547/2); 2547/2); 2547/3); 2547/3); 2547/3); 2547/3); 2547/3);	

Boring/ Well No.: AEいし					Site/Location: Naval Weapons Station, Seal Beach		
Client: N	Client: NFESC (Naval Facilities Eng. Serv. Ctr.)				Project No.: 50999-56254-6225.001.TK4.EQUIP Page of _		Page of _2
Drill Cor	ntractor: F	RSI Drilling			Drill Method: Hollow Stem Aug	ger	Elevation:
Date Star	rted: 9	111107			Sampling Method: CA-Modified		
Date End	led: 9	112/67			Total Depth:	Remarks:	
On-site (On-site Geologist: Kristeen Bennett				Depth to Water:		
Depth (feet)	Blow Count	Recovery (feet)	PID (ppm)	USCS	Description	on	Well Construction Detail
5	notue	2			3"-othick asphalt co n4'-thick read base	red 18" dian. (sand/grave)	(wixture)
0_			0.2	a	7.5-8: Gravelly 5. 60% f. to c., 5xtos: micaceons sitt; 15. SA gravel; dm. 8-12.5: Clay: Bla plastic clay; 15. organic material	ole micaceons Si Jiwarm casings	70 maderately
5		33"/60"	0.2 54.3 0.2 0.3 0.1	5C	12.5-18: Sandy Cl brown (10/R412) plastic day; 25: powly graded; 11 some organic in nottling especia 18-20: Sandy Clay (10/R812); 65% clay; 30% f. to	10, 65% moderate 10, t. to m. SA to 10% microus silvate rial; Feroxid 11/ near sonds. 11 Very pale bro 10 triable (?) plass 10. poorly graded	tis e
20		48"/00"	5.2 16.0 12.4 0.9	SP (sm	-3" beds of Olivers 4"ilt & 18.5; 20(?) - 23.5: Sindly 80% f. tom. sand 15% micaceous si some fe-oxide 23.5-25: Sundy 5ilt micaceous silt; 2	and 19.8. 4 Silt: Olive (54 4 Silt: 5% plastic Staining (54 413) 5% f. sand, 58, u	(4/4); A toskj clanjonet;
30	1	160	0.0	, Tibe	-25(?)-30: Similar -25(?)-30: Similar -caliche rodule(?)	ar todoore w/n	none

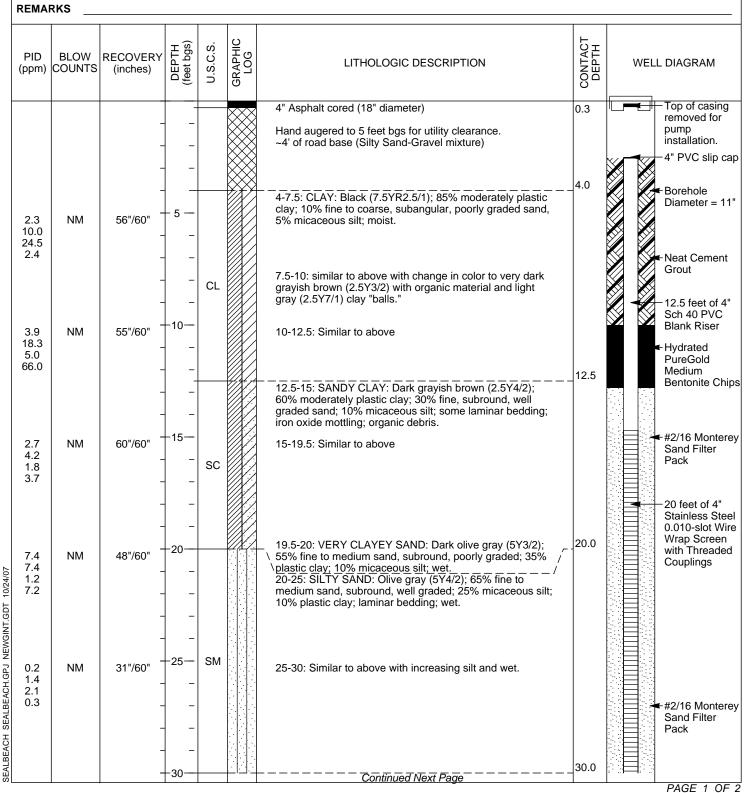
Boring/ Well No.: Aモルン				Site/Location: Naval Weapons Station, Seal Beach		
Client: NFESC (Naval Faciliti	es Eng. S	erv. Ctr.)	Project No.: 50999-56254-6225.001.TK4.EQUIP Page		
Drill Contractor	: RSI Drilling			Drill Method: Hollow Stem Auger	Elevation:	
Date Started:	9/11/0	7		Sampling Method: CA-Modified Split Spoon		
Date Ended:	9/12/0	7		Total Depth: 35 ft.bas Remarks:		
On-site Geologi	st: Kristeen B	ennett		Depth to Water:		
Depth Blov (feet) Cour		PID (ppm)	USCS	Description	Well Construction Detail	
35	120 60 1/00°	3.5 2 2.5 3.3	Su / M	30-35: Similar to above 33-laminar bedding 2034' T.D. of bore hole 35ft bys (w	(81 000)	

Appendix E.2 Phase I Well Logs

6	U	

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Fax. (949) 125-3901	
PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP	BORING/WELL NUMBER AEW1
PROJECT NAME Naval Weapons Station-Seal Beach, Site 70	DATE DRILLED 9/12/07
LOCATION Naval Weapons Station-Seal Beach	CASING TYPE/DIAMETER 4" Schedule 40 PVC
DRILLING METHOD CME 75 Hollow Stem Auger	SCREEN TYPE/SLOT 4" Stainless Steel Wire Wrap/0.010-slot
SAMPLING METHOD 4' Split Spoon-Continuous Core	GRAVEL PACK TYPE #2/16 Monterey Sand
GROUND SURFACE ELEVATION (FT MSL)	GROUT TYPE/QUANTITY Neat Cement Grout / Medium Bentonite Chip
TOP OF CASING ELEVATION (FT MSL)	STATIC WATER LEVEL (FT BELOW TOC)
LOGGED BY Kristeen Bennett	GROUND WATER ELEVATION (FT MSL)
REMARKS	





BORING/WELL CONSTRUCTION LOG

PROJECT NUMBER

SEALBEACH SEALBEACH.GPJ NEWGINT.GDT 10/24/07

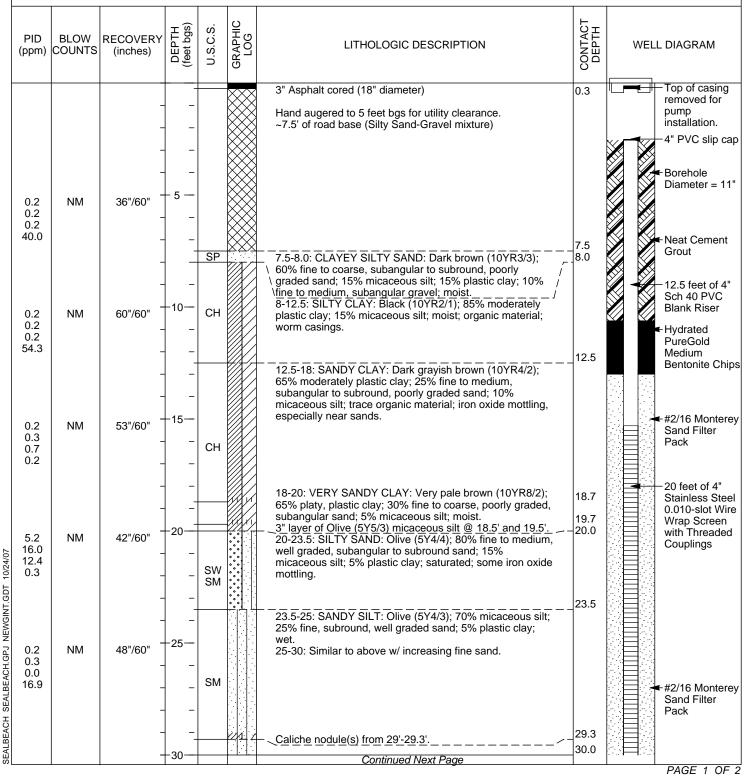
50999-56254-6225.001.TK4.EQUIP

BORING/WELL NUMBER AEW1

PROJECT NAME DATE DRILLED 9/12/07 Naval Weapons Station-Seal Beach, Site 70 Continued from Previous Page GRAPHIC LOG CONTACT DEPTH DEPTH (feet bgs) U.S.C.S. PID BLOW RECOVERY LITHOLOGIC DESCRIPTION WELL DIAGRAM COUNTS (inches) (ppm) 37"/60" 1.3 NM 30-32: No Recovery (Silt?) 8.7 #2/16 Monterey Sand Filter 32.0 Pack 32-34: GRAVELLY CLAY: Olive (5Y4/3); 50% plastic clay; 25% fine to medium, subangular to subround gravel; 15% GC micaceous silt; 10% fine to coarse, subangular to 34.0 subround, poorly graded sand; saturated. 34-35: Similar to 25-30'. SM 35.0 Welded -35 Stainless Steel bottom plate Slough Total Depth of Borehole: 35 feet bgs Total Depth of Well: 34.7 feet bgs 40 45 50 55 60

6	U	

1 dx. (040) 720 0007	
PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP	BORING/WELL NUMBER AEW2
PROJECT NAME Naval Weapons Station-Seal Beach, Site 70	DATE DRILLED 9/11/07
LOCATION Naval Weapons Station-Seal Beach	CASING TYPE/DIAMETER 4" Schedule 40 PVC
DRILLING METHOD	SCREEN TYPE/SLOT 4" Stainless Steel Wire Wrap/0.010-slot
SAMPLING METHOD 4' Split Spoon-Continuous Core	GRAVEL PACK TYPE #2/16 Monterey Sand
GROUND SURFACE ELEVATION (FT MSL)	GROUT TYPE/QUANTITY Neat Cement Grout / Medium Bentonite Chip
TOP OF CASING ELEVATION (FT MSL)	STATIC WATER LEVEL (FT BELOW TOC)
LOGGED BY Kristeen Bennett	GROUND WATER ELEVATION (FT MSL)
REMARKS	





BORING/WELL CONSTRUCTION LOG

PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP

Naval Weapons Station-Seal Beach, Site 70

BORING/WELL NUMBER AEW2 DATE DRILLED

PROJECT NAME 9/11/07 Continued from Previous Page GRAPHIC LOG CONTACT DEPTH DEPTH (feet bgs) U.S.C.S. PID BLOW RECOVERY LITHOLOGIC DESCRIPTION WELL DIAGRAM COUNTS (inches) (ppm) NM 60"/60" 0.5 30-35: Similar to above with laminar bedding from 34'-35'. 0.2 #2/16 Monterey 2.5 3.3 Sand Filter Pack SM 35.0 -35 Slough Welded Stainless Steel bottom plate Total Depth of Boring = 35 feet bgs (with slough) Total Depth of Well = 35.3 feet bgs 40 45 50 55 SEALBEACH SEALBEACH.GPJ NEWGINT.GDT 10/24/07 60 -

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0-slot
<u> Bentonite C</u> hip

PID (ppm)	BLOW COUNTS	RECOVERY (inches)	DEPTH (feet bgs)	U.S.C.S.	GRAPHIC LOG	LITHOLOGIC DESCRIPTION	CONTACT	WEL	L DIAGRAM
0.0/0.0			-			2.5" Asphalt cored (18" diameter)	0.2		Top of casing removed for
		_	-			Hand augered to 5 feet bgs for utility clearance.			pump
		_	_			~4.5' of road base (Silty Sand-Gravel mixture)			installation.
		_	_						⊢4" PVC slip ca
		_	_						≍ -Borehole
					\bowtie		5.0		Diameter = 11
0.0	NM	24"/60"	- 5 — -	CL		5-5.5: SILTY CLAY: Very dark grayish brown (2.5Y3/2);	5.5		
0.2 0.3		-	-	ML		\ 60% plastic to moderately stiff clay; 30% micaceous silt; / \10% fine sand.			
0.0		_	_			5.5-10.2: SILTY CLAY: Very dark gray (2.5Y3/1); 80% moderately stiff clay; 20% micaceous silt.			Neat Cement
		_	_			moderately Still Clay, 20% micaceous Silt.			Grout
		_							12.5 feet of 4
		_	_						Sch 40 PVC
0.3	NM	60"/60"	-10-	CL		10.2-12.2: Similar to above with change in color to Olive			Blank Riser
0.2 0.0		_	-			brown (2.5Y4/3).			
0.0		_	_						Hydrated
		_	_			12.2-13.2: Similar to above with change in color to Very dark grayish brown (2.5Y3/2).			PureGold Medium
		_				13.2-15.5: Similar to above with change in color to Olive			Bentonite Chi
		_				brown (2.5Y4/3) with thin, coarse sand and gravel layers throughout.			
0.3	NM	60"/60"	-15 -				15.5		##2/16 Montere
0.2 8.8		-	-			15.5-18: VERY CLAYEY SAND: Brown (10YR4/3); 60% fine to medium, well graded, subangular to subround			Pack
22.9		-	_	SC		sand; 35% stiff clay; 5% micaceous silt; iron oxide mottling; few worm casings.			
		_	_				18.0		20 feet of
		_		SC		18-18.8: Similar to 12.2 to 13.2 with abrupt physical change.	18.8		4"Stainless Steel 0.010-sl
				CL		18.8-20.5: VERY SANDY CLAY: Light gray (2.5Y7/2); 65% platy, plastic clay; 35% fine to coarse, poorly graded,			Wire Wrap
58.7	NM	60"/60"	-20— -			_ subangular sand; <5% micaceous silt; <1% fine, angular _	20.5		Screen with Threaded
13.7 8.2		_	-	00		\ \frac{\text{gravel; heavily bioturbated.}}{20.5-22.8: CLAYEY SAND: Pale yellow (2.5Y7/3); 50%			Couplings
22.8		_	-	SC		fine to coarse, subangular, poorly graded sand; 30% micaceous silt; 15% platy, plastic clay; 5% fine to medium,			
		_				subangular gravel; moist.	22.8		
		_	_	CL		22.8-24.5: SILTY CLAY: Olive (5Y5/3); 50% plastic clay; 40% micaceous silt; 10% fine to medium, well graded,			
			-	ML SC		- <u>subangular sand.</u> - <u>24.5-25: Similar to 15.5 to 18.</u>	24.5 25.0		
13.1 93.2	NM	46"/60"	-25		T - 1	25-26: No Recovery.	26.0		
7.8		-	_		† m †	26-28.8: SILTY SAND: Olive gray (5Y4/2); 70% fine to			
85.7		-	-	SM		medium, subround, well graded sand; 25% micaceous silt; 5% plastic clay; saturated; some iron oxide mottling.			#2/16 Montere
		-	_	5.**		,			Pack
		_				28.8-35: SANDY SILT: Olive gray (5Y4/2); 60%	28.8		
						micaceous silt; 25% fine, well graded, subround sand;			



BORING/WELL CONSTRUCTION LOG

PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP

BORING/WELL NUMBER AIW1

PROJECT NAME DATE DRILLED Naval Weapons Station-Seal Beach, Site 70 9/10/07 Continued from Previous Page GRAPHIC LOG CONTACT DEPTH DEPTH (feet bgs) U.S.C.S. PID BLOW RECOVERY LITHOLOGIC DESCRIPTION WELL DIAGRAM COUNTS (inches) (ppm) NM 60"/60" 6.5 <5% clay; <5% fine to medium, subangular gravel (as SM 3.2 layers); saturated. #2/16 Monterey 8.0 Sand Filter 2.1 ML Pack 35.0 -35 Welded Stainless Steel bottom plate Total Depth of Boring = 35 feet bgs Total Depth of Well = 35 feet bgs 40 45 50 55 60 -

PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP	BORING/WELL NUMBER AIW2
PROJECT NAME Naval Weapons Station-Seal Beach, Site 70	DATE DRILLED9/11/07
LOCATION Naval Weapons Station-Seal Beach	CASING TYPE/DIAMETER 4" Schedule 40 PVC
DRILLING METHOD CME 75 Hollow Stem Auger	SCREEN TYPE/SLOT 4" Stainless Steel Wire Wrap/0.010-slot
SAMPLING METHOD 4' Split Spoon-Continuous Core	GRAVEL PACK TYPE #2/16 Monterey Sand
GROUND SURFACE ELEVATION (FT MSL)	GROUT TYPE/QUANTITY Neat Cement Grout / Medium Bentonite Chips
TOP OF CASING ELEVATION (FT MSL)	STATIC WATER LEVEL (FT BELOW TOC)
LOGGED BY Kristeen Bennett	GROUND WATER ELEVATION (FT MSL)
REMARKS	

PID (ppm)	BLOW COUNTS	RECOVERY (inches)	DEPTH (feet bgs)	U.S.C.S.	GRAPHIC LOG	LITHOLOGIC DESCRIPTION	CONTACT	WEL	L DIAGRAM
		_	_			3" Asphalt cored (18" diameter)	0.3		Top of casing removed for
		_	-			Hand augered to 6 feet bgs for utility clearance.			pump
		-	_			~5' of road base (Silty Sand-Gravel mixture)			installation. —4" PVC slip ca
		_	_			Hand augered to 8 feet bgs.			
		_	_						■ Borehole Diameter = 11
0.7	NM	20"/60" -	- 5 —						Diameter = 11
1.0 0.7		-	_						
···		_	_			ROAD BASE: GRAVELLY SILTY SAND: Olive brown			Neat Cement
		_	_			(2.5Y5/4); 20% fine to coarse, subangular, poorly graded gravel; 60% fine to coarse, poorly graded, subangular to	8.5		Grout
		_	_			subround sand; 15% micaceous silt; 5% plastic clay. 9-10: SILTY CLAY: Black (2.5Y2.5/1); 70% moderately			-12.5 feet of
0.0	NM	52"/60" –	-10-			plastic clay; 25% micaceous silt; 5% fine to medium, well			4"-diameter S 40 PVC Blank
0.0	INIVI	52 /60 -	_	CL		graded, subangular to subround sand; wet. 10-11.5: Similar to above with change in color to Olive			Riser
0.0 0.4		_	_	ML		brown (2.5Y4/3). 11.5-13: Similar to above with change in color to Very dark			Hydrated
		_				grayish brown (2.5Y3/2).	13.0		PureGold Medium
		_	_			13-15: SANDY SILTY CLAY: Dark yellowish brown (10YR3/4); 50% moderately plastic clay; 30% fine to			Bentonite Chi
		_	-15			coarse, poorly graded, subangular to subround sand; 20% micaceous silt; wet; iron oxide mottling with organic			-#2/16 Montere
0.3 0.7	NM	60"/60"	_	CL		debris. 15-17: Similar to above with more fine to medium sand			Sand Filter Pack
26.1 9.6		_	_			(40%).			1 dok
		_				17-18.5: Similar to above with change in color to Very dark grayish brown (10YR3/2).			-20 feet of
						18.5-20: SILTY SANDY CLAY: Light gray (5Y7/2); 75%	18.5		4"-diameter
		_		CL		plastic, platy clay; 15% fine to medium, subanguilar to subround, well graded sand; 10% micaceous silt; wet.	20.0		Stainless Stee 0.010-slot Win
2.3 4.1	NM	54"/60"	-20			20-21.5: Similar to 15'-17'.	7		Wrap Screen with Threaded
4.7 2.1		_	_	CL		21.5-22.5: Similar to 18.5'-20'.			Couplings
2.1		_				22.5-24.5: CLAYEY SILT: Olive (5Y4/3); 80% micaceous	_22.5		
		_	-	ML		silt; 15% plastic clay; 5% fine, subround, well graded			
		_		CL	HHH	sand; saturated24-24.2: Sandy clay layer similar to 18.5'-20'.	24.2 -24.5		
0.7	NM	44"/60" -	-25 			24.5-28: SILTY SAND: Olive (5Y4/4); 65% fine, well graded, subround sand; 25% micaceous silt; 10% plastic			
0.3 4.1		_	-	SM		clay; saturated.			
		_	_				00.0		#2/16 Montere Sand Filter
		_			$\parallel \parallel $	28-30: SILT: Olive (5Y4/3); 90% micaceous silt; 10%	28.0		Pack
		-	_	ML		plastic clay; saturated.			
		_	-30		μшц	Continued Next Page	30.0		



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BORING/WELL CONSTRUCTION LOG

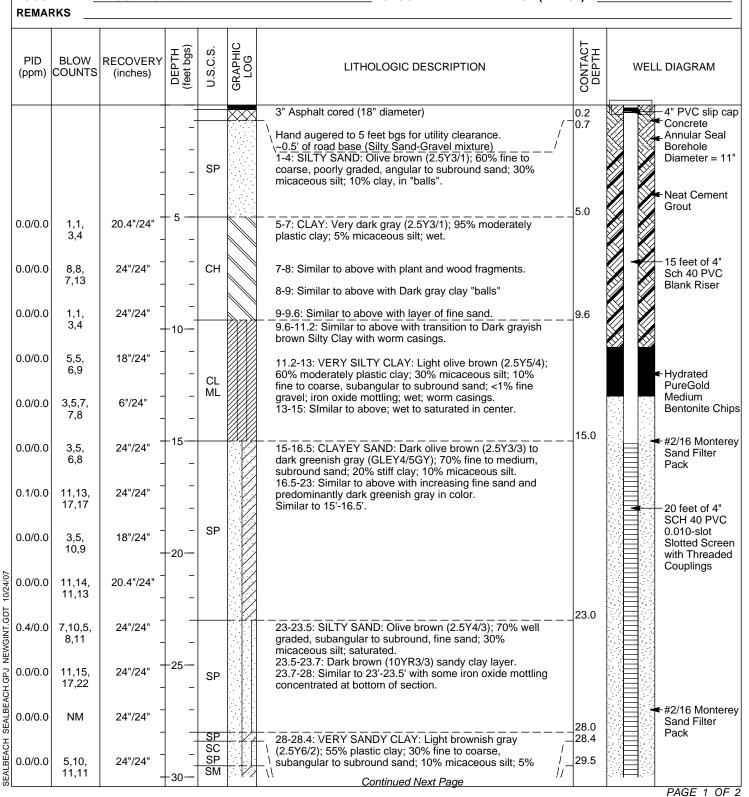
50999-56254-6225.001.TK4.EQUIP **PROJECT NUMBER**

BORING/WELL NUMBER AIW2

PROJECT NAME DATE DRILLED Naval Weapons Station-Seal Beach, Site 70 9/11/07 Continued from Previous Page GRAPHIC LOG CONTACT DEPTH DEPTH (feet bgs) U.S.C.S. PID BLOW RECOVERY LITHOLOGIC DESCRIPTION WELL DIAGRAM COUNTS (inches) (ppm) NM 47"/60" 0.3 30-33: VERY SILTY CLAY: Olive gray (5Y4/2); 60% plastic clay; 40% micaceous silt; saturated; heavy 0.1 #2/16 Monterey 0.0 bioturbation. CL Sand Filter 2.2 ML Pack 33.0 33-35: Similar to 28'-30'. ML 35.0 -35 -Slough Welded Stainless Steel bottom plate Total Depth of Boring = 35.1 feet bgs (with slough) Total Depth of Well = 35.6 feet bgs 40 45 50 55 60

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PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP	BORING/WELL NUMBER AMW1
PROJECT NAME Naval Weapons Station-Seal Beach, Site 70	DATE DRILLED
LOCATION Naval Weapons Station-Seal Beach	CASING TYPE/DIAMETER 4" Schedule 40 PVC
DRILLING METHOD CME 75 Hollow Stem Auger	SCREEN TYPE/SLOT 4" Schedule 40 PVC 0.010-slot Slotted Screen
SAMPLING METHOD 1.5' CA-Modified Split Spoon	GRAVEL PACK TYPE #2/16 Monterey Sand
GROUND SURFACE ELEVATION (FT MSL)	GROUT TYPE/QUANTITY Neat Cement Grout / Medium Bentonite Ch
TOP OF CASING ELEVATION (FT MSL)	STATIC WATER LEVEL (FT BELOW TOC)
LOGGED BY Kristeen Bennett	GROUND WATER ELEVATION (FT MSL)
REMARKS	





BORING/WELL CONSTRUCTION LOG

PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP

PROJECT NAME Naval Weapons Station-Seal Beach, Site 70

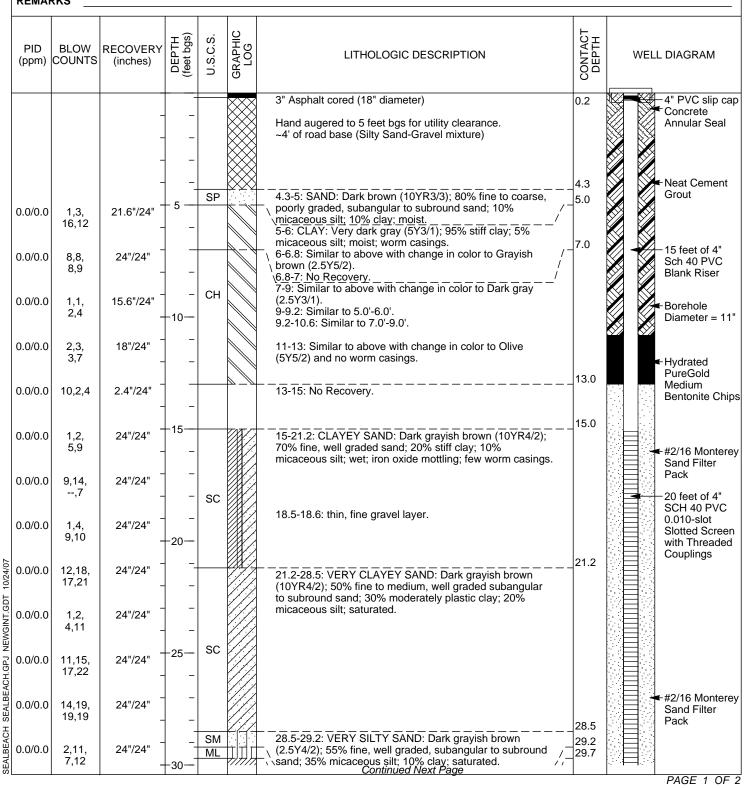
BORING/WELL NUMBER AMW1

DATE DRILLED 9/5/07

11,12, 15,17 4,7, 11,21 4,5, 7,11	RECOVERY (inches) 24"/24" 20.4"/24"	DEPTH (feet bgs)	SP SC SP SM SP ML SP SM	GRAPHIC	LITHOLOGIC DESCRIPTION Subangular, fine gravel; saturated. 1/28.4-29.5: Similar to 23.7'-28'. 31.0 32.0 32.0 33.32: Similar to 23.7'-28'. 32-33: Similar to 29.5'-31'. 33.3 33-34: Similar to 31'-32'. 34.3 34.2 35.1: VERYSANDY SILT: Olive (5Y4/3); 60% 134.2-35.1: VERYSANDY SILT: Olive (5Y4/3); 60% 136.2
4,7, 11,21 4,5,	20.4"/24" -	-35	SC SP SM SP SC SP SM SP ML SP		28.4-29.5: Similar to 23.7'-28'. 31.0 #2/16 Month 29.5-31: Similar to 28'-28.4', but heavilty bioturbated. 32.0 32.0 32.0 32.0 33.3: Similar to 29.5'-31'. 33.0 34.3-34: Similar to 31'-32'. 34.3-34: Similar to above with fine to medium, subangular to subround gravel. 34.2-35.1: VERYSANDY SILT: Olive (5Y4/3); 60% micaceous silt; 35% well graded, subround, fine sand; 25.0
11,21 4,5,	-	35	SC SP SM SP ML SP		33-34: Similar to 31'-32'. 34-34.2: Similar to above with fine to medium, subangular to subround gravel. 34.2-35.1: VERYSANDY SILT: Olive (5Y4/3); 60% micaceous silt; 35% well graded, subround, fine sand; >>10% clav; wet to saturated Cap
4,5, 7,11	24"/24" -	-35 	ML SP		to subround gravel. \[\] 34.2-35.1: VERYSANDY SILT: Olive (5Y4/3); 60% \[\] micaceous silt; 35% well graded, subround, fine sand; \[\] \ \ \ \ \ \ \ \ \ \ \ \ \
		_			35.1-37: Similar to 33'-34' with iron oxide mottling.
		40 -			Total Depth of Boring = 36.5 feet bgs (with slough) Total Depth of Well = 35.1 feet bgs
		- 45 - -			
		- - 50 - -			
		- - 55 - -			
		60 -			
			- 55 - - - - 60 -	- 55 - - - - - 60 - - -	- 55 - - - - - 60 - - -

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- 1 dx. (545) 725 5567	
PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP	BORING/WELL NUMBER AMW2
PROJECT NAME Naval Weapons Station-Seal Beach, Site 70	DATE DRILLED 9/6/07
LOCATION Naval Weapons Station-Seal Beach	CASING TYPE/DIAMETER 4" Schedule 40 PVC
DRILLING METHOD CME 75 Hollow Stem Auger	SCREEN TYPE/SLOT 4" Schedule 40 PVC 0.010-slot Slotted Screen
SAMPLING METHOD 1.5' CA-Modified Split Spoon	GRAVEL PACK TYPE #2/16 Monterey Sand
GROUND SURFACE ELEVATION (FT MSL)	GROUT TYPE/QUANTITY Neat Cement Grout / Medium Bentonite Ch
TOP OF CASING ELEVATION (FT MSL)	STATIC WATER LEVEL (FT BELOW TOC)
LOGGED BY Kristeen Bennett	GROUND WATER ELEVATION (FT MSL)
REMARKS	





BORING/WELL CONSTRUCTION LOG

PROJECT NUMBER

50999-56254-6225.001.TK4.EQUIP

PROJECT NAME Naval Weapons Station-Seal Beach, Site 70

BORING/WELL NUMBER AMW2

DATE DRILLED 9/6/07

						Continued from Previous Page
PID (ppm)	BLOW COUNTS	RECOVERY (inches)	DEPTH (feet bgs)	U.S.C.S.	GRAPHIC LOG	LITHOLOGIC DESCRIPTION LITHOLOGIC DESCRIPTION WELL DIAGRAM
0.0/0.0	7,12, 15,12	20.4"/24"	- - -	ML		_\\\\29.2-29.7:\SANDY\SILT:\Olive\(5\text{Y4/3}\);\tag{70\%}\minicaceous\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\
0.0/0.0	4,11, 12,12	24"/24" –	_ 	CL SM		_ \ \and iron oxide mottling. \\ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \
0.0/0.0	3,5, 7,10	24"/24"	35 -	CL		\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\
			- 40 - -			Total Depth of Boring = 36 feet bgs (with slough) Total Depth of Well = 34.95 feet bgs
			- - -			
			45 - - -			
			- - 50 -			
			- - -			
			- 55 - -			
			- - -			
			60 -			
			- -			

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BORING/WELL CONSTRUCTION LOG

PAGE 1 OF 2

			ı ax. (,							
		'						AMW3			
	CT NAME					al Beach, Site 70 DA					
LOCAT		Naval Weapor					SING TYPE/DIAMETER			-	
						Push / Hollow Stem Auge SC					less Steel Mesh
						ous Core GR					 .
						GR					
						ST					
		Kristeen Ben				GR		•			
REMAI	RKS P	ort designatio	n are la	abeled	count	erclockwise (i.e. Port 1 is still	Port 1, Port 2 is Port 6, Po	rt 3 is Port	<u>5, and</u>	Port 4 is Po	ort 4)
PID	BLOW	RECOVERY	DEPTH (feet bgs)	U.S.C.S.	GRAPHIC LOG	LITHOLOG	GIC DESCRIPTION		CONTACT DEPTH	WELL	_ DIAGRAM
(ppm)	COUNTS	(inches)	DE (feer	S.U	GRA L(CON		
						0-5: No Recovery.					Concrete
		-	_			Hand augered to 5 feet b	gs for utility clearance				Concrete Annular Seal
		_	_								
0.0	NM	21"/36"	5 —			F O. CLAV. Block /EVO F	1) to Very dark grayish bro		5.0		
0.0 0.0 0.0 0.0	INIVI	21/36 - -	-				1) to very dark grayish bro 10% micaceous silt; moist.				-Neat Cement Grout
0.0 0.0 0.0	NM	38"/48" _	- -	СН		8-13.5: Similar to above to Olive brown (2.5Y5/3)	with color change at sample and with increasing silt.	e base			
0.0		<u>-</u>	-10— -	0.1							+Borehole Diameter = 8"
0.0 0.0 0.0	NM	48"/48" -	_ _						13.5		
0.0		<u>-</u>	- -15 	CL ML		brown (10YR4/3); 55% m micaceous silt; 10% fine, moist	CLAY: Grayish brown (2.5) oderately plastic clay; 35% subround, well graded san	nd;	15.5		
0.0 0.0 0.0	NM	48"/48" -	_ _ _			15.5-20: VERY SANDY Colive brown (2.5Y4/3); 55 subround, well sorted sar	CLAY: Dark brown (10YR3/ % stiff clay; 40% fine to mend; 5% micaceous silt; dry;	edium,			
0.0		_	_ _	CL		oxide mottling throughout	i.				PORT 4 (17 to 18 feet bgs)
		, =	-20						20.0		
0.0 0.0 0.0 0.0	NM	43"/48" - -	- -			medium, subangular to s	live brown (2.5Y4/3); 65% ubround, well graded sand; clay; saturated; some iror	; 30%			Hydrated PureGold Medium
0.0 0.0 0.0	NM	47"/48" -	- -25-	SM		24-28.5: Similar to above					PORT 3 (24 to 25 feet bgs) +#2/16 Monterey
0.0		-	- 		V	27-27.2: Light brownish g	ıray clay layer.		27.2		Sand Filter Pack
0.0 0.2 0.0	NM	36"/36"	- - -30—	ML		\ olive brown (2.5Y4/3); 60 \ clay; 5% fine to coarse, s	ILT: Grayish brown (2.5Y5/ % micaceous silt; 35% plas ubangular to subround, po	stíc /	28.5		PORT 2 (28 to 29 feet bgs)
			30-		l	Continu	ued Next Page				



BORING/WELL CONSTRUCTION LOG

PROJECT NUMBER

SEALBEACH SEALBEACH.GPJ NEWGINT.GDT 10/24/07

50999-56254-6225.001.TK4.EQUIP

BORING/WELL NUMBER AMW3

PROJECT NAME DATE DRILLED 9/17/07 Naval Weapons Station-Seal Beach, Site 70 Continued from Previous Page GRAPHIC LOG CONTACT DEPTH DEPTH (feet bgs) U.S.C.S. PID BLOW RECOVERY LITHOLOGIC DESCRIPTION WELL DIAGRAM COUNTS (inches) (ppm) [\]graded sand; saturated._ Hydrated 31.0 PureGold ML 31-31.5: SANDY SILT: Olive gray (5Y4/2); 60% 31.5 0.0 NM 24"/24" Medium micaceous silt; 30% fine, subround, well graded sand; \10% plastic clay; saturated. \31.5-32.5: Similar to 28'-31'. 0.0 ML Bentonite Chips 32.5 0.0 ML 0.6 NM 36"/36" 32.5-34: Similar to 31'-31.5'. PORT 1 (33 to 34.0 0.2 34 feet bgs) 0.0 34-35: Similar to 28'-31'. ML #2/16 Monterey 35.0 0.0 35 Sand Filter 35-36: SILTY CLAY: Olive brown (2.5Y4/3); 80% CL Pack 36.0 moderately plastic clay; 20% micaceous silt; wet. ML Total Depth of Boring = 36.5 feet bgs Total Depth of Well = 35 feet bgs 40 45 50 55 60

SEALBEACH SEALBEACH.GPJ NEWGINT.GDT 10/24/07

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BORING/WELL CONSTRUCTION LOG

PAGE 1 OF 2

PROJE LOCAT DRILLI SAMPL GROUI TOP OI LOGGE REMAR	CT NAME TION! NG METH LING METI ND SURFA F CASING ED BY RKSP	Naval Wayal Wayal Weapor Naval Weapor OD Geopor HOD 4'S ACE ELEVAT ELEVATION Kristeen Ben ort designation	/eaponns Stat robe 66 plit Spo ION (F (FT M nett n are la	is Stati ion-Se 620DT con-Co T MSL ISL)	ion-Se eal Bea Direct ontinuo	BORING/WELL NUMBER AMW4 al Beach, Site 70 DATE DRILLED 9/17/07 ch CASING TYPE/DIAMETER 1.6" So Push / Hollow Stem AugeSCREEN TYPE/SLOT 3 0.38" Hole ous Core GRAVEL PACK TYPE #2/16 Monte GROUT TYPE/QUANTITY Neat C STATIC WATER LEVEL (FT BELOW T GROUND WATER ELEVATION (FT Misserclockwise (i.e. Port 1 is still Port 1, Port 2 is Port 6, Port 3 is Port	s covererey Sar ement (FOC) _ SL) _ t 5, and	red by Stainless Steel Mesh nd Grout / Medium Bentonite Chips
PID (ppm)	BLOW COUNTS	RECOVERY (inches)	DEPTH (feet bgs)	U.S.C.S.	GRAPHIC LOG	LITHOLOGIC DESCRIPTION	CONTACT	WELL DIAGRAM
		- - -	- - -			0-5: No Recovery. Hand augered to 5 feet bgs for utility clearance.	5.0	Concrete Annular Seal
0.0 0.0 0.0	NM	21"/36" - -	- 5 - -	CL		5-8: SILTY CLAY: Dark olive gray (5Y3/2); 80% moderately plastic clay; 15% micaceous silt; 5% fine, subround, well graded sand; moist; worm casings.	_3.0	Neat Cement Grout
0.0 0.0 0.0 0.0	NM	32"/48" - - -	- - -10 -			8-10: Similar to above with change in color to Olive gray (5Y4/2). 10-11.5: SANDY CLAY: Olive gray (5Y5/2); 70% plastic clay; 25% fine, subround, well graded sand; 5% micaceous silt; saturated.	10.0	Borehole Diameter = 8"
0.0 0.0 0.0 0.0	NM	33"/48" - - -	- - -15	CL		11.5-13.5: SImilar to 8'-10'. 13.5-16: VERY SANDY CLAY: Olive brown (2.5Y4/4); 50% stiff clay; 40% fine to medium, subangular to subround, well sorted sand; 10% micaceous silt; moist; iron oxide mottling.	16.0	
NR	NM	43"/48" - - -	- - -	SC		16-20: VERY CLAYEY SAND: Dark yellowish brown (10YR4/4); 60% fine to medium, subangular to subround, well sorted sand; 35% stiff clay; 10% micaceous silt; moist.	20.0	PORT 4 (18 to 19 feet bgs)
NR	NM	48"/48" - -	-20 - - -	SC		20-21: No Recovery. 21-22.5: Similar to 16'-20' with increasing sand and change in color to Olive gray (5Y4/2); iron oxide mottling concentrated at bottom of section. 22.5-24: SILTY CLAY: Olive (5Y4/4) to light gray (5Y7/2);	21.0	
0.0 0.2 0.0 0.0	NM	48"/48" - -	- -25- -	CL ML SM		60% platy, moderately plastic, banded (see colors above) clay; 30% micaceous silt; fine to medium grained, well sorted, subangular to subround sand; wet. 24-27: SILTY SAND: Olive (5Y4/3); 70% fine to medium, subangular to subround, well sorted sand; 30% micaceous silt; saturated.	24.0	PORT 3 (24 to 25 feet bgs) #2/16 Monterey Sand Filter Pack
NR	NM	24"/24" - -	- - -30-	CL		27-28: VERY CLAYEY SILT: Pale olive (5Y6/3); 50% micaceous silt; 40% plastic clay; 10% fine, subround, well sorted sand; wet. 28-29: Similar to 27'-28' with some iron oxide mottling; saturated. Continued Next Page	29.6	PORT 2 (28 to 29 feet bgs)



BORING/WELL CONSTRUCTION LOG

50999-56254-6225 001 TK4 FOLUP

BORING/WELL NUMBER AMW4

	CT NUME					K4.EQUIP BORING/WELL NUMBER AMW4 al Beach, Site 70 DATE DRILLED 9/17/07			
						Continued from Previous Page			
PID (ppm)	BLOW COUNTS	RECOVERY (inches)	DEPTH (feet bgs)	U.S.C.S.	GRAPHIC LOG	LITHOLOGIC DESCRIPTION	CONTACT DEPTH	WEL	L DIAGRAM
NR NR	NM NM	29"/36" - - 36"/36" -	-35-	ML CL SM CL ML		29-32: Similar to 27'-28' with laminar bedding. 29.5-29.6: Fine to medium gravel layer. 32-32.5: SILTY CLAY: Olive gray (5Y5/2); 80% moderately plastic clay; 20% micaceous silt; wet. 32.5-34: SILTY SAND: Olive (5Y5/3); 60% fine to medium, subround, well sorted sand; 30% micaceous silt; 10% plastic clay; saturated. 34-36: Similar to 27'-28'.	32.0 32.5 34.0		 Hydrated PureGold Medium Bentonite Ch PORT 1 (33 and 134 feet bgs) #2/16 Monter Sand Filter Pack
			- 40 45 50 55			Total Depth of Boring = 36 feet bgs Total Depth of Well = 35 feet bgs			
			60 -						PAGE 2 OF

SEALBEACH SEALBEACH.GPJ NEWGINT.GDT 10/24/07

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BORING/WELL CONSTRUCTION LOG

PAGE 1 OF 2

DPO IF		SEP 50000				K4.EQUIP	BORING/WELL NUMBER AMW5		
	CT NAME						DATE DRILLED 9/17/07		
LOCAT		·					CASING TYPE/DIAMETER 1.6" S	olinst Cl	MT Multiport HDPE Tubing
DRILLI	NG METH	OD Geopi	obe 66	320DT	Direct	Push / Hollow Stem Auge	eSCREEN TYPE/SLOT 3 0.38" Hol	es cove	red by Stainless Steel Mesh
SAMPL	ING MET	HOD 4'S	plit Spo	on-Co	ontinuo	us Core	GRAVEL PACK TYPE #2/16 Mont	erey Sa	ınd
							GROUT TYPE/QUANTITY Neat 0		
TOP O							STATIC WATER LEVEL (FT BELOW		
LOGGI		Kristeen Ben					GROUND WATER ELEVATION (FT N		
REMAI	RKS P	ort designatio	n are la	abeled	count	erclockwise (i.e. Port 1 is	still Port 1, Port 2 is Port 6, Port 3 is Po	rt 5, and	I Port 4 is Port 4)
PID (ppm)	BLOW COUNTS	RECOVERY (inches)	DEPTH (feet bgs)	U.S.C.S.	GRAPHIC LOG	LITHO	LOGIC DESCRIPTION	CONTACT	WELL DIAGRAM
						0-8: No Recovery.			
		-	-			Hand augered to 8 fe	et bgs for utility clearance.		Concrete Annular Seal
		_	-						
		_ _	- - 5 —						
		_	_						Neat Cement Grout
		_	- _		7,77,77			_8.0	
9.7 9.2 3.7	NM	30"/48" -	_				ark olive brown (2.5Y3/2.5); 85% ay; 15% micaceous silt; dry.		- Borehole
2.8		_	10-						Diameter = 8"
		_	_	CL					
		_	_						
0.2 0.0	NM	48"/48"				12-13: Similar to above	ve with worm casings.	13.0	
0.0		-	_			yellowish brown (10Y clay; 20% fine to med	: Dark olive brown (2.5Y3/3) to dark R4/6); 70% moderately plastic to stiff lium, subround, well sorted sand;		
		_	-15			10% micaceous silt; o	dry; iron oxide mottling.		
0.0 0.0 0.0	NM	48"/48" -	- -	CL		16-18.5: Similar to ab	ove with increasing sands.		 Hydrated PureGold Medium Bentonite Chips
0.0		_	_						
		-	- -20			change in color to Da 19-20: Similar to 16'-	ove with decreasing sand and rk olive gray (5Y3/2); dry. 18.5'.	20.0	PORT 4 (18 to 19 feet bgs)
0.0 0.4 0.0	NM	36"/36" -	_	sc		20-22.5: CLAYEY SA to medium, subangula 30% stiff clay; 10% m	ND: Olive brown (2.5Y4/4); 60% fine ar to subround, well graded sand; icaceous silt; moist; iron oxide		
		-	_			mottling.		22.5	PORT 3 (22 to
0.0	NM	36"/36"	_	SM			TY SAND: Olive brown (2.5Y4/3); subround, well sorted sand; 25%	23.0	23 feet bgs) #2/16 Monterey
0.0	INIVI	30 /30 -		CL ML		_ \micaceous silt; 15% r	noderately plastic clay; wet / ,	24.0	Sand Filter Pack
0.0 0.0			0.5	SM			Pale olive (5Y6/3) to light gray // tely plastic clay; 25% micaceous silt; //		Fack
5.5		_	-25 -	CL		\ 5% fine to medium, s	ubangular to subround, well graded	25.5 26.0	
		-	_	ML	 	_\ \ <u>sand; wet.</u> \\ 24-25.5: SILTY SAND	D: Olive (5Y4/3); 70% fine to medium, #	120.0	
		-	-	SM			nd, well graded sand; 30%	27.5	PORT 2 (26.5
		_	_	CL		– <u>∖ \25.5-26: Simi</u> lar to 23		28.0	to 27.5 feet bgs)
		_	_	ML ML		\\26-27.5: Similar to 24 -\27.5-28: Similar to 23	<u>'-25.5'</u>	29.0	
			20	CL			SILT: Olive (5Y4/3); 60% micaceous	30.0	
	İ	_	⊢ა∪—		T /-4		ntinued Next Dogo		1



BORING/WELL CONSTRUCTION LOG

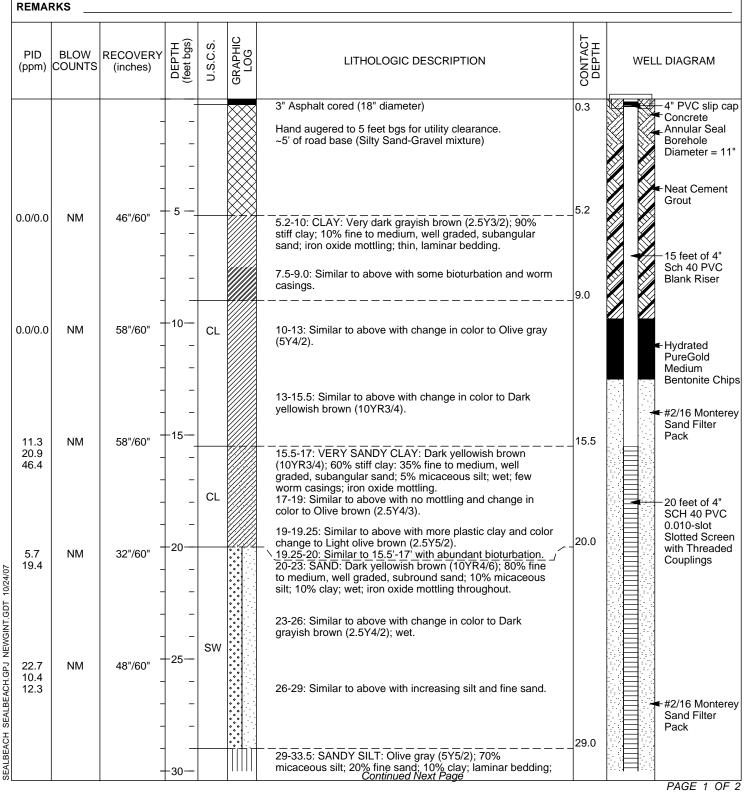
50999-56254-6225.001.TK4.EQUIP

BORING/WELL NUMBER AMW5

					Continued from Previous Page		
PID BLC (ppm) COUI	W RECOVER'	DEPTH (feet bgs)	U.S.C.S.	GRAPHIC LOG	LITHOLOGIC DESCRIPTION	CONTACT	WELL DIAGRAM
			ML ML CL ML		\\\^29-30: SILTY CLAY: Light olive brown (2.5Y5/3); 85% // \\\^moderately plastic clay; 15% micaceous silt; moist. // \\^30-31: Similar to 28'-29'. // \\^31-32: CLAYEY SILT: Grayish brown (2.5Y5/2); 60% // micaceous silt; 30% moderately plastic clay; 10% fine, // \\^32-33.5: Similar to above. // \\^33.5-35: Similar to 28'-29'. // \\^30-30-30-30-30-30-30-30-30-30-30-30-30-3	31.0 33.5 35.0 36.0	PORT 1 (33 34 feet bgs) ##2/16 Monter Sand Filter Pack
		40 - 40 - 45 - 50 - 55 - 60 - 60 -			Total Depth of Boring = 36.4 feet bgs Total Depth of Well = 35 feet bgs		

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PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP	BORING/WELL NUMBER AMW6
PROJECT NAME Naval Weapons Station-Seal Beach, Site 70	DATE DRILLED 9/7/07
LOCATION Naval Weapons Station-Seal Beach	CASING TYPE/DIAMETER 4" Schedule 40 PVC
DRILLING METHOD CME 75 Hollow Stem Auger	SCREEN TYPE/SLOT 4" Schedule 40 PVC 0.010-slot Slotted Screen
SAMPLING METHOD 4' Split Spoon-Continuous Core	GRAVEL PACK TYPE #2/16 Monterey Sand
GROUND SURFACE ELEVATION (FT MSL)	GROUT TYPE/QUANTITY Neat Cement Grout / Medium Bentonite Ch
TOP OF CASING ELEVATION (FT MSL)	STATIC WATER LEVEL (FT BELOW TOC)
LOGGED BY Kristeen Bennett	GROUND WATER ELEVATION (FT MSL)
REMARKS	





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BORING/WELL CONSTRUCTION LOG

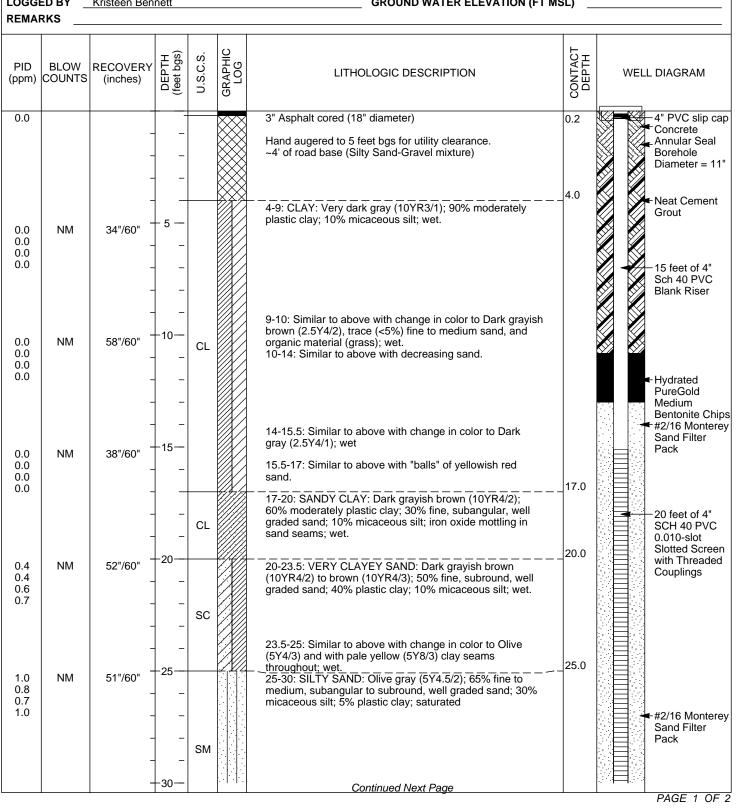
BORING/WELL NUMBER PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP _AMW6 **PROJECT NAME** DATE DRILLED Naval Weapons Station-Seal Beach, Site 70 9/7/07 Continued from Previous Page GRAPHIC LOG CONTACT DEPTH DEPTH (feet bgs) U.S.C.S. PID BLOW RECOVERY LITHOLOGIC DESCRIPTION WELL DIAGRAM COUNTS (inches) (ppm) NM 44"/60" 3.4 wet. 2.1 #2/16 Monterey 0.9 Sand Filter ML Pack 33.5-35: Similar to above with iron oxide mottling. 35.0 -35 Threaded SCH 40 PVC Bottom Cap Total Depth of Boring = 35 feet bgs (with slough) Total Depth of Well = 35.5 feet bgs 40 45 50 55 60

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SEALBEACH SEALBEACH.GPJ NEWGINT.GDT 10/24/07

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1 ax. (343) 723 3307	
PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP	BORING/WELL NUMBER PMW1
PROJECT NAME Naval Weapons Station-Seal Beach, Site 70	DATE DRILLED 9/13/07
LOCATION Naval Weapons Station-Seal Beach	CASING TYPE/DIAMETER 4" Schedule 40 PVC
DRILLING METHOD CME 75 Hollow Stem Auger	SCREEN TYPE/SLOT 4" Schedule 40 PVC 0.010-slot Slotted Screen
SAMPLING METHOD 4' Split Spoon-Continuous Core	GRAVEL PACK TYPE #2/16 Monterey Sand
GROUND SURFACE ELEVATION (FT MSL)	GROUT TYPE/QUANTITY Neat Cement Grout / Medium Bentonite Chip
TOP OF CASING ELEVATION (FT MSL)	STATIC WATER LEVEL (FT BELOW TOC)
LOGGED BY Kristeen Bennett	GROUND WATER ELEVATION (FT MSL)
REMARKS	



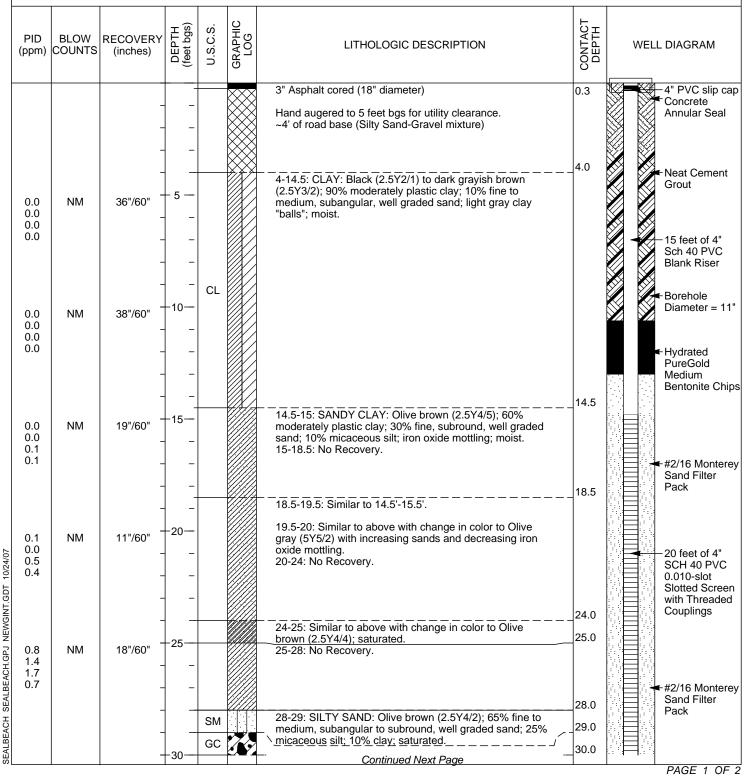


BORING/WELL CONSTRUCTION LOG

BORING/WELL NUMBER PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP PMW1 **PROJECT NAME** DATE DRILLED Naval Weapons Station-Seal Beach, Site 70 9/13/07 Continued from Previous Page GRAPHIC LOG CONTACT DEPTH DEPTH (feet bgs) U.S.C.S. PID BLOW RECOVERY LITHOLOGIC DESCRIPTION WELL DIAGRAM COUNTS (inches) (ppm) 60"/60" 0.0 NM 30-32: Similar to above with increasing silt. 0.2 #2/16 Monterey 6.8 Sand Filter 32.0 4.7 32-34: SILTY CLAY: Olive brown (2.5Y4/4) to light olive brown (2.5Y5/4); 70% stiff clay; 30% micaceous silt; wet; iron oxde mottling throughout. Pack CL ML 34.0 34-35: SILT: Grayish brown (2.5Y5/2); 85% micaceous ML silt; 10% plastic clay; 5% fine, subround sand; wet. 35.0 -35 Threaded SCH 40 PVC Bottom Cap Total Depth of Boring = 35.3 feet bgs Total Depth of Well = 35.3 feet bgs 40 45 50 55 SEALBEACH SEALBEACH.GPJ NEWGINT.GDT 10/24/07 60

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1 dx: (0-1	0) 120 0001		
PROJECT NUMBER 50999-56254-6	225.001.TK4.EQUIP BORING/WEL	L NUMBER	PMW9
PROJECT NAME Naval Weapons S	Station-Seal Beach, Site 70 DATE DRILLE	D 9/12/07	
LOCATION Naval Weapons Station	-Seal Beach CASING TYPI	DIAMETER	4" Schedule 40 PVC
DRILLING METHOD CME 75 Hollow	Stem Auger SCREEN TYP	E/SLOT 4"	Schedule 40 PVC 0.010-slot Slotted Screen
SAMPLING METHOD 4' Split Spoor	n-Continuous Core GRAVEL PAC	K TYPE#2	/16 Monterey Sand
GROUND SURFACE ELEVATION (FT I	MSL) GROUT TYPE	/QUANTITY	Neat Cement Grout / Medium Bentonite Ch
TOP OF CASING ELEVATION (FT MSL	.) STATIC WAT	R LEVEL (FT	BELOW TOC)
LOGGED BY Kristeen Bennett	GROUND WA	TER ELEVATION	ON (FT MSL)
REMARKS			





BORING/WELL CONSTRUCTION LOG

SEALBEACH SEALBEACH.GPJ NEWGINT.GDT 10/24/07

PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP

BORING/WELL NUMBER PMW9

	ECT NUME	•				AL Beach, Site 70 DATE DRILLED 9/12/07
						Continued from Previous Page
PID (ppm)	BLOW COUNTS	RECOVERY (inches)	DEPTH (feet bgs)	U.S.C.S.	GRAPHIC LOG	LITHOLOGIC DESCRIPTION LITHOLOGIC DESCRIPTION WELL DIAGRAM
46.5 9.6 9.9 1.8	NM	38"/60"		ML GC CL ML		29-30: GRAVELLY CLAY: Light yellowish brown (2.5Y63); 60% plastic logy; 30% fine to coarse, subangular gravel; 10% micaceous silt; saturated. 30-32: No Recovery. 32-32.5: VERY SANDY SILT: Light olive brown (2.5Y5/3); 33.0 (1.5% plastic clay; saturated.) 33-33: Maximilar to 29-30. Av. Light olive brown (2.5Y5/3); 36% moderately plastic clay; 25% micaceous silt; 15% fine to medium, subangular to subround gravel; wet. Total Depth of Boring = 35.5 feet bgs (with slough) Total Depth of Well = 34.8 feet bgs

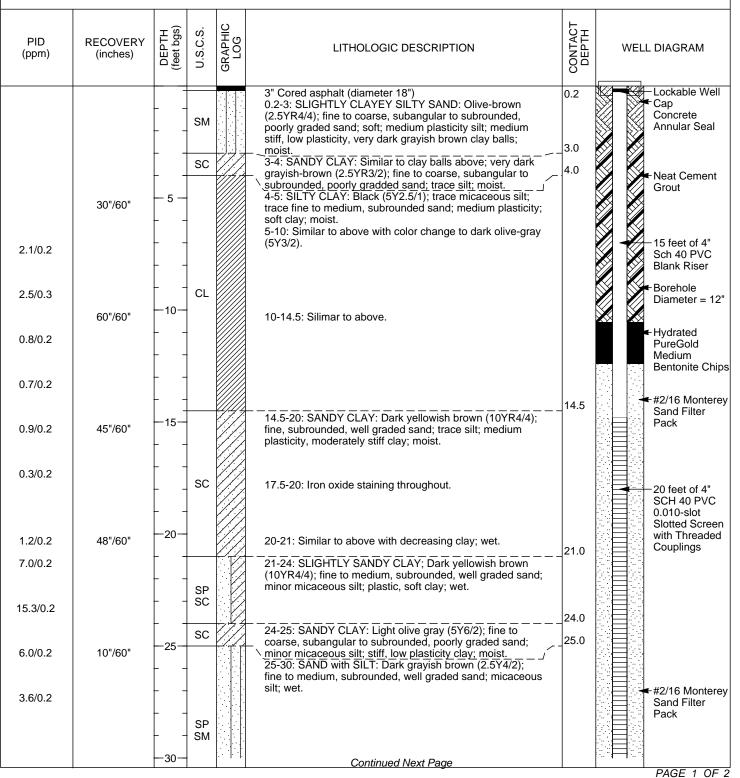
Appendix E.3 Phase II Well Logs

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SEALBEACH SEALBEACH.GPJ NEWGINT.GDT 4/28/08

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1 ax. (343) 123 3301	
PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP	BORING/WELL NUMBER PIW1
PROJECT NAME Naval Weapons Station-Seal Beach, Site 70	DATE DRILLED 3/26/08
LOCATION Naval Weapons Station-Seal Beach	CASING TYPE/DIAMETER 4" Schedule 40 PVC
DRILLING METHOD CME 75 Hollow Stem Auger	SCREEN TYPE/SLOT 4" PVC 0.010 Slot
SAMPLING METHOD 5' CA-Modified Split Spoon	GRAVEL PACK TYPE #2/16 Lapis Lustre Sand
GROUND SURFACE ELEVATION (FT MSL)	GROUT TYPE/QUANTITY Neat Cement Grout / Medium Bentonite Ch
TOP OF CASING ELEVATION (FT MSL)	STATIC WATER LEVEL (FT BELOW TOC)
LOGGED BY Kristeen Bennett	GROUND WATER ELEVATION (FT MSL)
REMARKS	





BORING/WELL CONSTRUCTION LOG

PROJECT NUMBER

50999-56254-6225.001.TK4.EQUIP

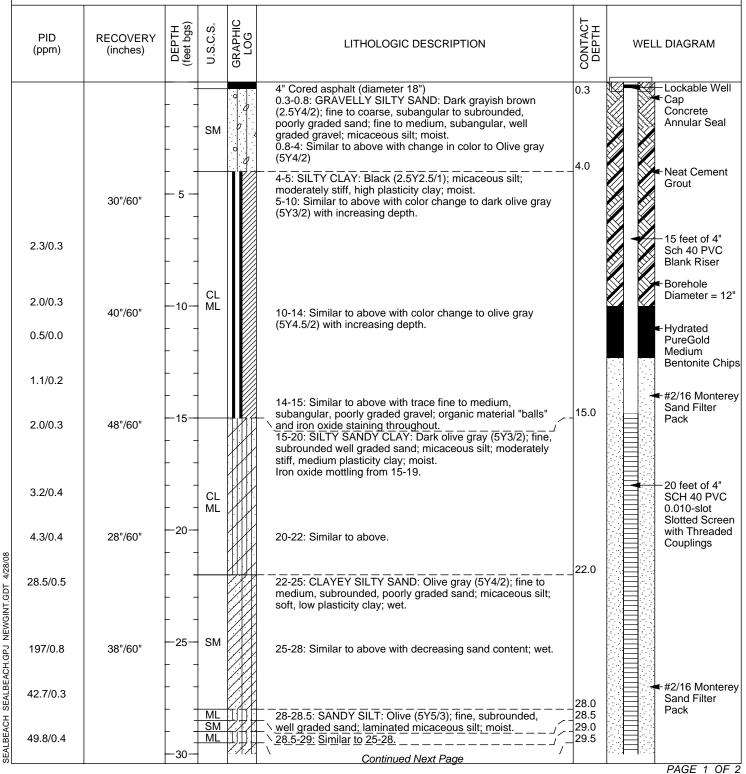
BORING/WELL NUMBER

PIW1 **PROJECT NAME** Naval Weapons Station-Seal Beach, Site 70 **DATE DRILLED** 3/26/08

	1	1			Continued from Previous Page			
PID (ppm)	RECOVERY (inches)	DEPTH (feet bgs)	U.S.C.S.	GRAPHIC LOG	LITHOLOGIC DESCRIPTION	CONTACT	WEL	L DIAGRAM
8.3/2.0	34"/60"	-			30-32: Similar to above with very dark gray clay "balls;" wet.	32.0		+#2/16 Monte Sand Filter
5.2/0.2			ML		32-35: SLIGHTLY SANDY SILT: Dark grayish brown (2.5Y2/4); fine to medium, subrounded, poorly graded sand; micaceous silt; moderately stiff, low plasticity clay; moist.			Pack
32.0/2.0		-35-				35.0		Threaded So
		_			Total Depth of Borehole = 35.5 feet bgs			Сар
		-						
		40 -						
		-						
		45 -						
		-						
		_						
		50 -						
		-						
		-						
		55 -						
		_						
		_						
		60 -						
		-						
		_						

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Fax. (949) 125-3901					
PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP	BORING/WELL NUMBER PIW2				
PROJECT NAME Naval Weapons Station-Seal Beach, Site 70	DATE DRILLED 3/27/08				
LOCATION Naval Weapons Station-Seal Beach	CASING TYPE/DIAMETER 4" Schedule 40 PVC				
DRILLING METHOD CME 75 Hollow Stem Auger	SCREEN TYPE/SLOT 4" PVC 0.010 Slot				
SAMPLING METHOD 5' CA-Modified Split Spoon	GRAVEL PACK TYPE #2/16 Lapis Lustre Sand				
GROUND SURFACE ELEVATION (FT MSL)	GROUT TYPE/QUANTITY Neat Cement Grout / Medium Bentonite Ch				
TOP OF CASING ELEVATION (FT MSL)	STATIC WATER LEVEL (FT BELOW TOC)				
LOGGED BY Kristeen Bennett	GROUND WATER ELEVATION (FT MSL)				
REMARKS					





BORING/WELL CONSTRUCTION LOG

PROJECT NUMBER

SEALBEACH SEALBEACH.GPJ NEWGINT.GDT 4/28/08

50999-56254-6225.001.TK4.EQUIP

BORING/WELL NUMBER PIW2

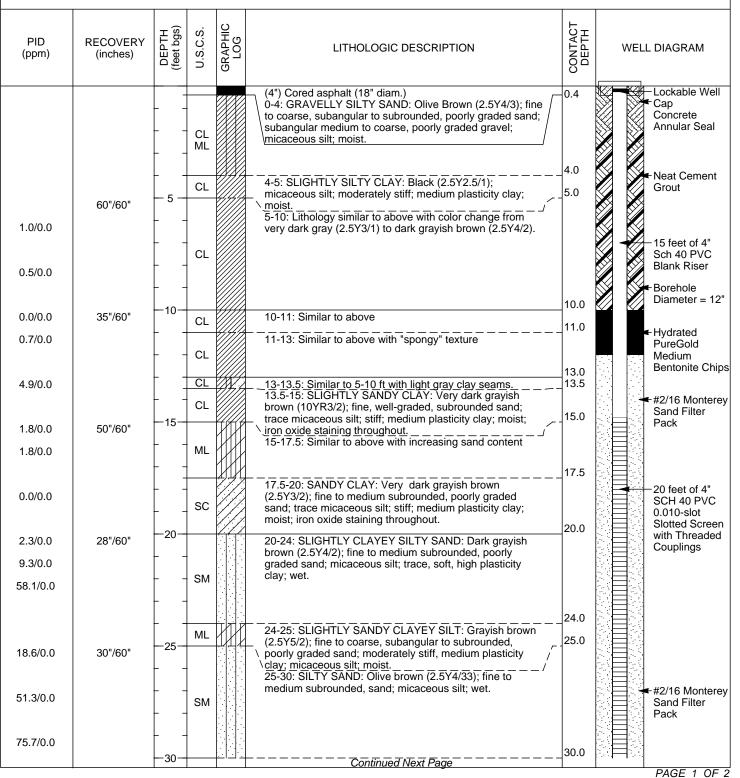
PROJECT NAME DATE DRILLED Naval Weapons Station-Seal Beach, Site 70 3/27/08 Continued from Previous Page GRAPHIC LOG DEPTH (feet bgs) CONTACT DEPTH U.S.C.S. RECOVERY PID LITHOLOGIC DESCRIPTION WELL DIAGRAM (inches) (ppm) 29-29.5: <u>Similar to 28-28.5.</u> 29.5-30: Similar to 28.5-29. 135/0.3 48"/60" #2/16 Monterey SM 30-33: Similar to above. Sand Filter Pack 25.7/0.2 33.0 33-34.5: SILTY CLAY: Olive (5Y4/3); micaceous, laminated silt; stiff, high plasticity clay; moist. 34.5 ML 34.5-35: SANDY SILT: Olive (5Y5/3); fine, subrounded, well graded sand; micaceous, laminated silt; moist. ML 35.0 Threaded SCH 40 PVC Bottom Total Depth of Borehole = 35.5 feet bgs 40 45 50 55 60

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SEALBEACH SEALBEACH.GPJ NEWGINT.GDT 4/28/08

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1 ax. (343) 723 3307	
PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP	BORING/WELL NUMBER PIW3
PROJECT NAME Naval Weapons Station-Seal Beach, Site 70	DATE DRILLED 3/25/08
LOCATION Naval Weapons Station-Seal Beach	CASING TYPE/DIAMETER 4" Schedule 40 PVC
DRILLING METHOD	SCREEN TYPE/SLOT 4" PVC 0.010 Slot
SAMPLING METHOD 5' CA-Modified Split Spoon	GRAVEL PACK TYPE #2/16 Lapis Lustre Sand
GROUND SURFACE ELEVATION (FT MSL)	GROUT TYPE/QUANTITY Neat Cement Grout / Medium Bentonite Ch
TOP OF CASING ELEVATION (FT MSL)	STATIC WATER LEVEL (FT BELOW TOC)
LOGGED BY Kristeen Bennett	GROUND WATER ELEVATION (FT MSL)
REMARKS	





BORING/WELL CONSTRUCTION LOG

PROJECT NUMBER <u>50999-56254-6</u>

50999-56254-6225.001.TK4.EQUIP

BORING/WELL NUMBER PIW3

PROJECT NAME Naval Weapons Station-Seal Beach, Site 70

DATE DRILLED 3/25/08

		_	1		Continued from Previous Page			
PID (ppm)	RECOVERY (inches)	DEPTH (feet bgs)	U.S.C.S.	GRAPHIC LOG	LITHOLOGIC DESCRIPTION	CONTACT	WEL	L DIAGRAM
16.5/0.0	50"/60"		SM ML		30-31: Similar to above; wet. 31-32: CLAYEY SILT; Olive gray (5Y5/2); micaceous silt; stiff; medium plasticity clay; moist. 32-33: Similar to 30-31 ft.	31.0		#2/16 Monter Sand Filter Pack
7.8/0.0			SM ML		33-34: Similar to 31-32 ft.	33.0		1 dok
273/0.0		-35-	SM		34-35: Similar to 32-33 ft. Total Depth of Boring = 35.5 ft bgs	35.0		Threaded S0 40 PVC Bott Cap
		-			Added 5 gallons of water for heaving sands.			
		40 -						
		-						
		-						
		45 -						
		_						
		_						
		50 -						
		-						
		55 -						
		-						
		-						
		60 -						
		-						
		_						PAGE 2 C

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111 Academy, Suite 150 Irvine, CA 92617 Phone: (949) 752-5452 Fax: (949) 725-3907						BORING/WELL CONSTRUCTION LOG			
PROJECT N	JMBER _ 5099	9-56254-	6225.0	01.TI	K4.EQUIP	BORING/WELL NUMBER PMW2			
PROJECT NA	AME Naval V	Veapons	Station	n-Sea	al Beach, Site 70	DATE DRILLED 3/26/08			
LOCATION						CASING TYPE/DIAMETER 4" Sch	edule 40) PVC	
					er		10 Slot		
SAMPLING N	METHOD 5' C	CA-Modifi	ied Spli	it Spo	oon	GRAVEL PACK TYPE #2/16 Lapis	Lustre	Sand	
GROUND SU	IRFACE ELEVAT	TION (FT	MSL)_			GROUT TYPE/QUANTITY Neat C	Cement (<u> Grout / Medium Bentonite Chips</u>	
TOP OF CAS	ING ELEVATION	N (FT MS	6L)			STATIC WATER LEVEL (FT BELOW	TOC) _		
LOGGED BY	Kristeen Ber	nett				GROUND WATER ELEVATION (FT M	ISL) _		
REMARKS									
PID (ppm)	RECOVERY (inches)	DEPTH (feet bgs)	U.S.C.S.	GKAPHIC	LITHO	DLOGIC DESCRIPTION	CONTACT	WELL DIAGRAM	
1.7/0.2	26"/60"	 - 5 - 	SM		(10YR5/4); fine to co poorly graded sand; subangular, poorly g staining.	ameter 18") SILTY SAND: Yellowish brown barse subangular to subrounded, micaceous silt; fine to coarse, iraded gravel; moist; some iron oxide Dark olive gray (5Y3/2) to Olive gray silt; moderately stiff, medium plasticity	0.3	Lockable Well Cap Concrete Annular Seal Borehole Diameter = 12" Neat Cement Grout 15 feet of 4"	
1.8/0.5 1.2/0.2 1.6/0.2	40"/60"		S. C.		10-15: Similar to abo (10YR4/3)	ove with color change at 14.5 to brown	_10.0	Sch 40 PVC Blank Riser - Hydrated PureGold Medium Medium	

SEALBEACH SEALBEACH.GPJ NEWGINT.GDT 4/28/08

2.3/0.4

1.9/0.2

3.0/0.6

2.1/0.6

4.3/0.9

4.2/0.7

9.1/0.0

6.1/0.8

16.4/0.7

40"/60"

48"/60"

30"/60"

CL

CL

-20

-30

24.5-25: SLIGHTLY SANDY CLAYEY SILT: Dark grayish brown (2.5Y4/2); trace, fine to medium, subrounded, poorly graded sand; laminated micaceous silt; low \plasticity, soft clay; moist; heavily bioturbated. 25-26: SLIGHTLY SANDY CLAY: Dark grayish brown \(10YR4/2); trace, fine to coarse, subangular, soft; low soft; low plasticity clay; wet; heavily bioturbated. 25-26: SILTY CLAY; Olive gray (5Y4/2); micaceous silt; Continued Next Page

15-20: SLIGHTLY SILTY SANDY CLAY: Dark yellowish brown (10YR3/4) fine to medium, subrounded, poorly

graded sand; minor micaceous silt; moderately stiff,

20-21.5: Similar to above with decreasing clay and

21.5-24.5: SILTY CLAYEY SAND: Olive brown (2.5Y4/3); fine to medium subrounded, poorly graded sand; micaceous silt; medium plasticity; soft; clay; wet.

medium plasticity clay; moist.

increasing sand contents.

#2/16 Monterey

Sand Filter

Pack

Bentonite Chips #2/16 Monterey Sand Filter Pack

20 feet of 4" SCH 40 PVC

with Threaded

0.010-slot Slotted Screen

Couplings

15.0

20.0

21.5

24.5 25.0

26.0 26.5

30.0



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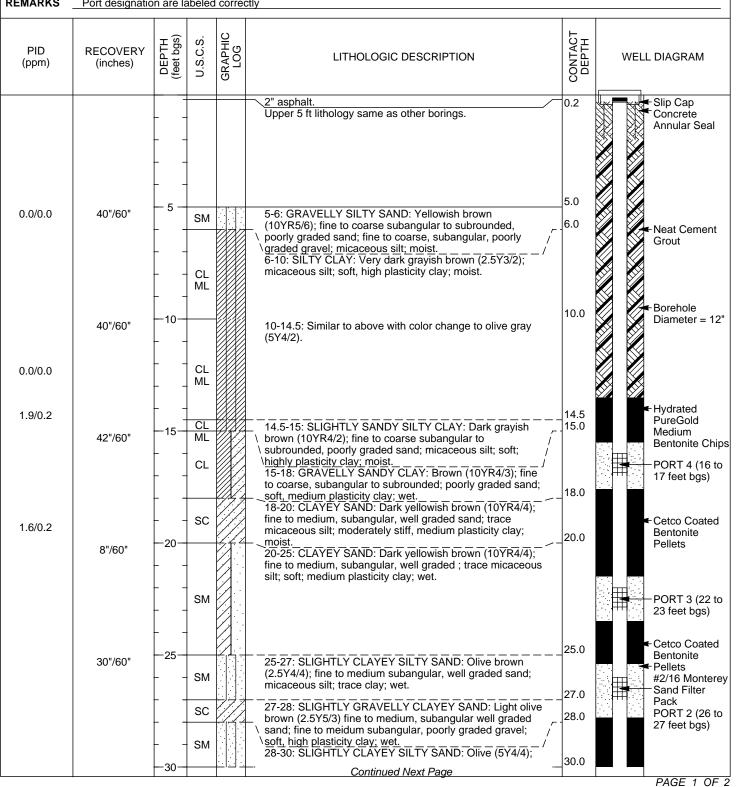
BORING/WELL CONSTRUCTION LOG

PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP **BORING/WELL NUMBER** PMW2 **PROJECT NAME** DATE DRILLED Naval Weapons Station-Seal Beach, Site 70 3/26/08 Continued from Previous Page GRAPHIC LOG CONTACT DEPTH DEPTH (feet bgs) U.S.C.S. PID RECOVERY LITHOLOGIC DESCRIPTION WELL DIAGRAM (ppm) (inches) 52"/60" fine to medium subrounded, poorly graded sand; wet. 31.0 30-31: CLAYEY SILT: Light olive brown (2.5 Y5/3); 33.1/0.6 alternating laminations of clay and silt; micaceous silt; soft; / high plasticity clay moist
31-32.5: SILTY CLAY: Olive (5Y5/3); micaceous silt;
moderately stiff, medium plasticity clay; moist some #2/16 Monterey 32.5 Sand Filter Pack \iron-oxide staining
32.5-34: Similar to 30-31 with trace subangular to
subrounded, poorly graded, fine to medium gravel. 27.0/0.8 34.0 35.0 34-35: CLAYEY SANDY SILT: Olive gray (5Y4/2); fine, Threaded SCH -35 29.7/0.9 subrounded sand; soft, high plasticity clay; micaceous silt; 40 PVC Bottom moist. Cap Total depth of bore hole is 35.5 ft bgs 40 45 50 55 60

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BORING/WELL NUMBER PMW3
DATE DRILLED
CASING TYPE/DIAMETER 4" Schedule 40 PVC
SCREEN TYPE/SLOT 4" PVC 0.010 Slot
GRAVEL PACK TYPE #2/16 Lapis Lustre Sand
GROUT TYPE/QUANTITY Neat Cement Grout / Medium Bentonite Ch
STATIC WATER LEVEL (FT BELOW TOC)
GROUND WATER ELEVATION (FT MSL)





BORING/WELL CONSTRUCTION LOG

PROJECT NUMBER

SEALBEACH SEALBEACH.GPJ NEWGINT.GDT 4/28/08

50999-56254-6225.001.TK4.EQUIP

BORING/WELL NUMBER PMW3

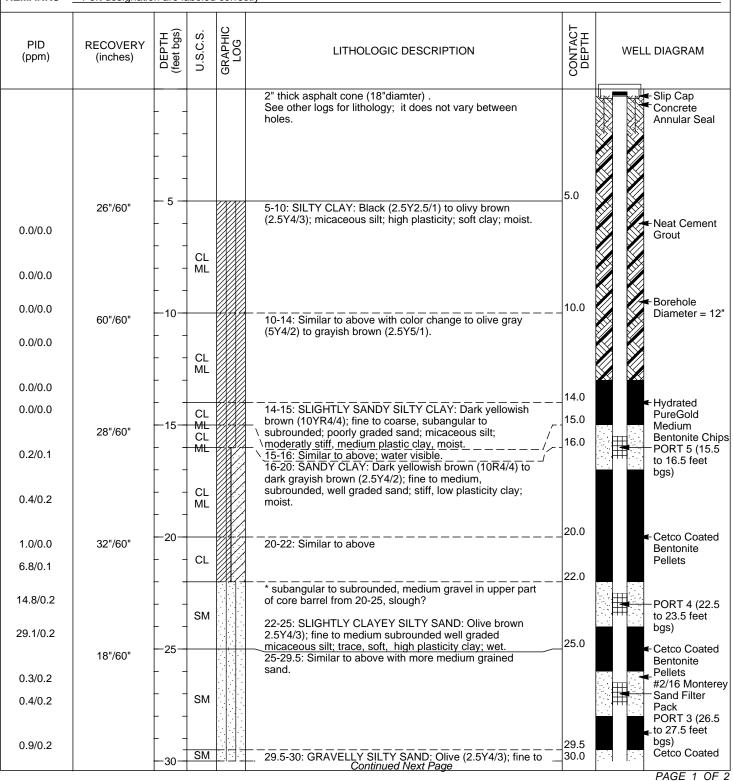
PROJECT NAME DATE DRILLED Naval Weapons Station-Seal Beach, Site 70 3/31/08 Continued from Previous Page GRAPHIC LOG DEPTH (feet bgs) CONTACT DEPTH U.S.C.S. PID **RECOVERY** LITHOLOGIC DESCRIPTION WELL DIAGRAM (inches) (ppm) 58"/60" fine to medium subangular, well graded sand; micaceous SC Cetco Coated 31.0 silt; trace clay; wet. **Bentonite** 30-31: Similar to 27-28'.
31-33: SILTY CLAY: Olive (5Y4/3); micaceous silt; moderately stiff, high plasticity clay; moist. Pellets CL ML 33.0 33.5 ML 33-33.5: SILT: Olive (5Y5/4); laminated micaceous silt; moist.
33.5-34: Similar to 31-33'; dry to moist.
34-35: SLIGHTLY SANDY SILT: Olive (5Y4/4); fine, well CL 34.0 ML PORT 1 (34 to 35.0 35 35 feet bgs) graded, subangular sand; laminated micaceous silt; moist. #2/16 Monterey Sand Filter Total Depth of Boring = 36 ft bgs. Pack 40 45 50 55 60

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PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP	BORING/WELL NUMBER PMW4
PROJECT NAME Naval Weapons Station-Seal Beach, Site 70	DATE DRILLED 4/1/08
LOCATION Naval Weapons Station-Seal Beach	CASING TYPE/DIAMETER 4" Schedule 40 PVC
DRILLING METHOD	SCREEN TYPE/SLOT 4" PVC 0.010 Slot
SAMPLING METHOD 5' CA-Modified Split Spoon	GRAVEL PACK TYPE #2/16 Lapis Lustre Sand
GROUND SURFACE ELEVATION (FT MSL)	GROUT TYPE/QUANTITY Neat Cement Grout / Medium Bentonite Ch
TOP OF CASING ELEVATION (FT MSL)	STATIC WATER LEVEL (FT BELOW TOC)
LOGGED BY Kristeen Bennett	GROUND WATER ELEVATION (FT MSL)
REMARKS Port designation are labeled correctly	





BORING/WELL CONSTRUCTION LOG

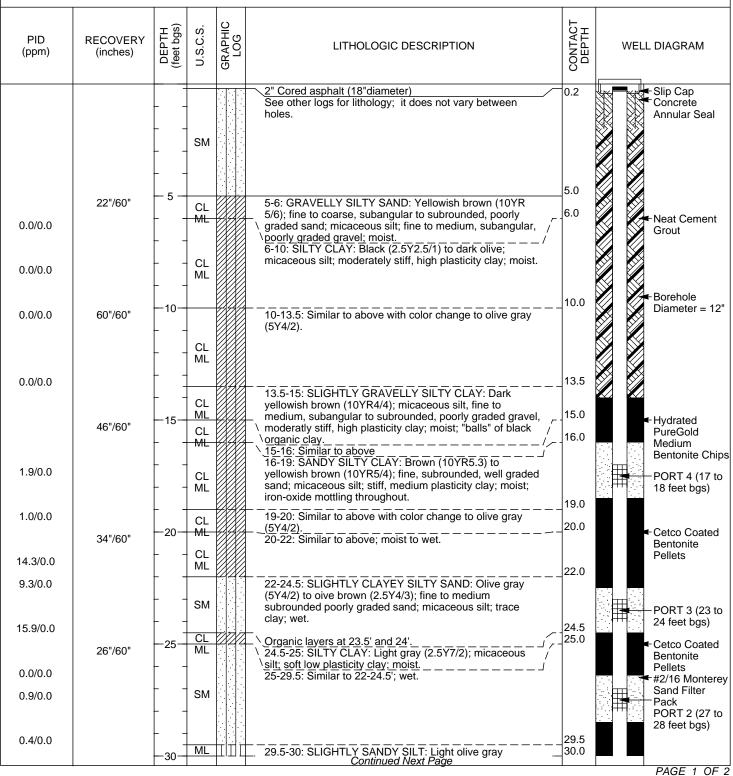
SEALBEACH SEALBEACH.GPJ NEWGINT.GDT 4/28/08

PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP **BORING/WELL NUMBER** PMW4 **PROJECT NAME DATE DRILLED** Naval Weapons Station-Seal Beach, Site 70 4/1/08 Continued from Previous Page GRAPHIC LOG CONTACT DEPTH DEPTH (feet bgs) U.S.C.S. PID **RECOVERY** LITHOLOGIC DESCRIPTION WELL DIAGRAM (ppm) (inches) Bentonite 60"/60 SM coarse, subrounded, poorly graded sand; fine to medium, 30.5 Pellets subangular, poorly graded gravel; micaceous silt; wet. _ _/ ML 30-30.5: Similar to above. 30.5-31.5: SLIGHTLY SANDY SILT: Light olive gray (5Y6/2); fine, subrounded well graded sand; laminated PORT 2 (30 to 31.5 31 feet bgs) CL ML 33.0 \micaceous silt; moist. 31.5-33: SILTY CLAY: Light gray (2.5Y7/2); micaceous silt; moderately stiff, medium plasticity clay; moist. 33-34.5: SANDY SILT: Olive 2/5Y4/3); fine, subrounded ML PORT 1 (33.5 34.5 to 34.5 feet 35.0 well graded sand; laminated micaceous silt; moist. -35 bgs) ML 34.5-35: Similar to 31.5-33'.
Total Depth of Boring = 36 ft bgs. #Ž/16 Monterey Sand Filter Pack 40 45 50 55 60

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PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP	BORING/WELL NUMBER PMW5	
PROJECT NAME Naval Weapons Station-Seal Beach, Site 70	DATE DRILLED	
LOCATION Naval Weapons Station-Seal Beach	CASING TYPE/DIAMETER 4" Schedule 40 PVC	
DRILLING METHOD CME 75 Hollow Stem Auger	SCREEN TYPE/SLOT 4" PVC 0.010 Slot	
SAMPLING METHOD 5' CA-Modified Split Spoon	GRAVEL PACK TYPE #2/16 Lapis Lustre Sand	
GROUND SURFACE ELEVATION (FT MSL)	GROUT TYPE/QUANTITY Neat Cement Grout / Medium Bentonite Chip	
TOP OF CASING ELEVATION (FT MSL)	STATIC WATER LEVEL (FT BELOW TOC)	
LOGGED BY Kristeen Bennett	GROUND WATER ELEVATION (FT MSL)	
REMARKS Port designation are labeled correctly		





BORING/WELL CONSTRUCTION LOG

SEALBEACH SEALBEACH.GPJ NEWGINT.GDT 4/28/08

PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP

BORING/WELL NUMBER PMW5

Continued from Previous Page							
PID (ppm)	RECOVERY (inches)	DEPTH (feet bgs)	U.S.C.S.	GRAPHIC LOG	LITHOLOGIC DESCRIPTION	CONTACT DEPTH	WELL DIAGRAM
19.7/0.0 87.4/0.0 14.5/0.0	54"/60"	40 - 45 - 50 - 60 - 60 - 60 - 60 - 60 - 60 - 6	SM ML CL ML ML CL ML ML ML ML ML ML ML		(5Y6/2); fine, subrounded well graded sand; micaceous /silt; moist. \\\\30.30.5: Similar to 25-29.5; wet. \\\\\30.5-31: Similar to 29.5-30. \\\\31.32.5: Similar to 24.5-25. \\\32.5-33.5: SILIGHTLY SANDY SILT: Olive gray (5Y5/2); laminated micaceous silt; trace; fine, well graded, subrounded sand; moist. \\\\\\33.5-34.5: SANDY SILT: Olive (2.5) Y4/3); fine, well graded, subrounded sand; laminated micaceous silt; moist. \\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\	30.5 31.0 32.5 33.5 34.5 35.0	PORT 1 (33. to 34.5 feet bgs) #2/16 Monte Sand Filter Pack

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BORING/WELL CONSTRUCTION LOG

PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP	BORING/WELL NUMBER PMW6	
PROJECT NAME Naval Weapons Station-Seal Beach, Site 70	DATE DRILLED	
LOCATION Naval Weapons Station-Seal Beach	CASING TYPE/DIAMETER 4" Schedule 40 PVC	
DRILLING METHOD	SCREEN TYPE/SLOT 4" PVC 0.010 Slot	
SAMPLING METHOD 5' CA-Modified Split Spoon	GRAVEL PACK TYPE #2/16 Lapis Lustre Sand	
GROUND SURFACE ELEVATION (FT MSL)	GROUT TYPE/QUANTITY Neat Cement Grout / Medium Bentonite Ch	
TOP OF CASING ELEVATION (FT MSL)	STATIC WATER LEVEL (FT BELOW TOC)	
LOGGED BY Kristeen Bennett	GROUND WATER ELEVATION (FT MSL)	
REMARKS		

GRAPHIC LOG CONTACT DEPTH DEPTH (feet bgs) U.S.C.S. PID **RECOVERY** LITHOLOGIC DESCRIPTION WELL DIAGRAM (ppm) (inches) 4" Cored asphalt (18" diam.). 0.3-0.8: GRAVELLY SILTY SAND: Brown (10YR4/3); fine Lockable Well 0.2 Cap 0.8 to coarse, subangular to subrounded, poorly graded sand; / Concrete tine to medium subangular gravel; micaceous silt; moist. __′ 0.8-4: SLIGHTLY CLAYEY SILTY SAND: Olive gray Annular Seal Borehole CL Diameter = 12" (5Y5/2); fine to coarse, subangular to subrounded, poorly ML graded sand; micaceous silt; soft, medium plasticity clay; 4.0 **Neat Cement** 4-5: SILTY CLAY: Black (2.5Y2.5/1); micaceous silt; soft, CL Grout 5.0 high plasticity clay; moist.
5-10: Similar to above with color change to very dark gray 5 ML 26"/60" (5Y4/1) with depth. 1.5/0.6 15 feet of 4" Sch 40 PVC CL ML Blank Riser 2.6/0.9 10.0 50"/60" 10-14: Similar to above with color change to dark gray (5Y4/1) with depth. Hydrated 3.8/0.8 PureGold CL Medium ML Bentonite Chips #2/16 Monterey 3.1/0.8 Sand Filter 14.0 Pack 14-15: SLIGHTLY SANDY CLAY: Brown (10YR4/3); fine, CL subrounded, well graded sand; trace micaceous silt; 15.0 3.8/0.4 42"/60" moderately stiff, high plasticity clay; moist; "balls" of \organic rich clay. 15-19: Similar to above with increasing sand with depth and color banding of brown and dark olive gray. CL 3.3/0.3 20 feet of 4" SCH 40 PVC 19.0 0.010-slot 19-20: CLAYEY SAND: Dark olive gray (5Y3/2); fine, SC Slotted Screen 20.0 subrounded, well graded sand; trace micaceous silt; -20 with Threaded moderately stiff, medium plasticity clay; moist.
20-22: SLIGHTLY CLAYEY SILTY SAND: Olive gray 5.4/0.5 20"/60" Couplings SM (5Y4/2); fine to medium, subrounded, well graded sand; 22.0 micaceous silt; moderately stiff, medium plasticity clay; 8.4/0.6 22-25: SILTY SAND: Olive gray (5Y4/2); fine to medium, SM subrounded, well graded sand; micaceous silt; wet. 25.0 25 36.4/0.0 30"/60" 25-30: Similar to above. #2/16 Monterey 37.1/0.5 SM Sand Filter Pack 24.7/0.1 30.0 -30 Continued Next Page PAGE 1 OF 2



BORING/WELL CONSTRUCTION LOG

PROJECT NUMBER

SEALBEACH SEALBEACH.GPJ NEWGINT.GDT 4/28/08

50999-56254-6225.001.TK4.EQUIP

BORING/WELL NUMBER PMW6

PROJECT NAME DATE DRILLED Naval Weapons Station-Seal Beach, Site 70 3/27/08 Continued from Previous Page GRAPHIC LOG DEPTH (feet bgs) CONTACT DEPTH U.S.C.S. PID **RECOVERY** LITHOLOGIC DESCRIPTION WELL DIAGRAM (inches) (ppm) 48"/60" 30-32: Similar to above. SM 15.5/0.2 32.0 #2/16 Monterey 32-32.5: SILTY CLAY: Olive (5Y5/3): micaceous silt; moderately stiff, high plasticity clay; moist.
32.5-34.5: SLIGHTLY SANDY SILT: Light olive brown 32.5 Sand Filter ML Pack 25.6/0.3 ML (2.5Y5/3); fine, well graded, subrounded sand; laminated micaceous silt; trace high plasticity, soft clay; moist.

34.5-35: SANDY SILT: Olive (5Y4/4); fine, well graded, 34.5 ML 35.0 Threaded SCH 86.3/0.4 subrounded sand; micaceous silt; moist. 40 PVC Bottom Total Depth of Boring = 35.5 ft bgs. 40 45 50 55 60

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BORING/WELL CONSTRUCTION LOG

1 ax. (949) 125-5901		
PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP	BORING/WELL NUMBER PMW7	
PROJECT NAME Naval Weapons Station-Seal Beach, Site 70	DATE DRILLED 3/27/08	
LOCATION Naval Weapons Station-Seal Beach	CASING TYPE/DIAMETER 4" Schedule 40 PVC	
DRILLING METHOD CME 75 Hollow Stem Auger	SCREEN TYPE/SLOT 4" PVC 0.010 Slot	
SAMPLING METHOD 5' CA-Modified Split Spoon	GRAVEL PACK TYPE #2/16 Lapis Lustre Sand	
GROUND SURFACE ELEVATION (FT MSL)	GROUT TYPE/QUANTITY Neat Cement Grout / Medium Bentonite Ch	
TOP OF CASING ELEVATION (FT MSL)	STATIC WATER LEVEL (FT BELOW TOC)	
LOGGED BY Kristeen Bennett	GROUND WATER ELEVATION (FT MSL)	
REMARKS		

GRAPHIC LOG DEPTH (feet bgs) U.S.C.S. PID **RECOVERY** LITHOLOGIC DESCRIPTION WELL DIAGRAM (ppm) (inches) Lockable Well 4" Cored asphalt (18" diam.). Cap 0-4: GRAVELLY SILTY SAND: Olive-brown (2.5Y 4/3); Concrete subangular to subrounded, poorly graded sand; medium Annular Seal SM to coarse, subangular, poorly graded gravel; micaceous Borehole GM Diameter = 12" silt; moist. 4.0 **Neat Cement** 4-5: SLIGHTLY SILTY CLAY: Black (2.5Y5/1); micaceous silt; moderately stiff, medium plasticity clay; moist CL Grout 5.0 5 42"/60" 5-10: Similar to above with color change to very dark grayish brown (2.5Y3/2) with depth. 0.0/0.0 15 feet of 4" CL Sch 40 PVC Blank Riser 0.0/0.0 10.0 42"/60" 10-14.5: Similar to above with color change to olive brown (2.5Y4/4) at base; trace fine to medium, subangular, Hydrated poorly graded gravel from 13-14.5'. PureGold Medium CL Bentonite Chips #2/16 Monterey Sand Filter Pack 14.5 CL 14.5-15: SANDY CLAY: Dark yellowish brown (10YR4/4); 15.0 0.7/0.1 58"/60" trace, subrounded, well graded sand; micaceous silt; stiff, medium plasticity clay; moist, iron oxide staining and \"balls" organic-rich clay. 15-20: Similar to above with no "balls" of clay; alternating bands of dark yellowish brown/olive brown; moist CL SC 20 feet of 4" SCH 40 PVC 0.010-slot Slotted Screen 20.0 -20 with Threaded 4.6/0.1 20"/60" 20-23: Similar to above with increasing sand content. Couplings SEALBEACH SEALBEACH.GPJ NEWGINT.GDT 4/28/08 CL SC 23.0 23-25: SLIGHTLY CLAYEY SILTY SAND; Yellowish brown (10YR 5/4); fine to medium, well graded, SM subrounded sand; micaceous silt; soft, low plasticity clay; 25.0 25 97.3/0.1 24"/60" 25-30: SLIGHTLY SILTY SAND: Olive brown (10YR 4/4); medium, subrounded, well graded sand; minor micaceous silt; wet. #2/16 Monterey SW Sand Filter Pack 30.0 -30 Continued Next Page PAGE 1 OF 2



BORING/WELL CONSTRUCTION LOG

PROJECT NUMBER

SEALBEACH SEALBEACH.GPJ NEWGINT.GDT 4/28/08

50999-56254-6225.001.TK4.EQUIP

BORING/WELL NUMBER PMW7

PROJECT NAME DATE DRILLED Naval Weapons Station-Seal Beach, Site 70 3/27/08 Continued from Previous Page GRAPHIC LOG DEPTH (feet bgs) CONTACT DEPTH U.S.C.S. RECOVERY PID LITHOLOGIC DESCRIPTION WELL DIAGRAM (inches) (ppm) 140.0/0.1 30"/60" 30-33: Similar to above; wet. SW #2/16 Monterey Sand Filter 33.0 Pack 33-33.5: SILT: Olive (5Y4/3); laminated micaceous silt; 33.5 ML <u>moist</u> CL 33.5-34.5: SILTY CLAY: Olive (5Y /4); laminated micaceous silt and stiff, high plasticity clay; moist 34.5-35: SLIGHTLY SANDY SILT: Olive (5Y 4/4); fine, 34.5 ML 35.0 Threaded SCH 35 ML 152/0.2 40 PVC Bottom subrounded, well graded sand; laminated micaceous silt; Cap moist to wet. Total Depth of Boring = 35.5 ft bgs. 40 45 50 55 60

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BORING/WELL CONSTRUCTION LOG

PROJECT NUMBER 50999-56254-6225.001.TK4.EQUIP	BORING/WELL NUMBER PMW8	
PROJECT NAME Naval Weapons Station-Seal Beach, Site 70	DATE DRILLED	
LOCATION Naval Weapons Station-Seal Beach	CASING TYPE/DIAMETER 4" Schedule 40 PVC	
DRILLING METHOD	SCREEN TYPE/SLOT 4" PVC 0.010 Slot	
SAMPLING METHOD 5' CA-Modified Split Spoon	GRAVEL PACK TYPE #2/16 Lapis Lustre Sand	
GROUND SURFACE ELEVATION (FT MSL)	GROUT TYPE/QUANTITY Neat Cement Grout / Medium Bentonite Ch	
TOP OF CASING ELEVATION (FT MSL)	STATIC WATER LEVEL (FT BELOW TOC)	
LOGGED BY Kristeen Bennett	GROUND WATER ELEVATION (FT MSL)	
REMARKS		

GRAPHIC LOG DEPTH (feet bgs) U.S.C.S. PID **RECOVERY** LITHOLOGIC DESCRIPTION WELL DIAGRAM (ppm) (inches) Lockable Well 4" Cored asphalt (18" diam.). Cap 0-4: GRAVELLY SILTY SAND: Olive-brown (2.5Y 4/3); Concrete fine to coarse, subangular to subrounded, poorly graded Annular Seal SM sand; medium to coarse, poorly graded gravel; micaceous silt; moist. 4.0 **Neat Cement** 4-5: SLIGHTLY SILTY CLAY: Black (2.5Y2.5/1); micaceous silt; medium stiff, medium plastic clay; moist. CL Grout 5.0 5 35"/60" 5-10: Similar to above with color change to dark olive gray (5Y3/2) with increasing depth. 3.2/0.0 15 feet of 4" Sch 40 PVC CL Blank Riser Borehole 1.9/0.0 Diameter = 12" 10.0 4.0/0.0 48"/60" 10-14: Similar to above. Hydrated PureGold CL Medium Bentonite Chips 9.9/0.0 14.0 #2/16 Monterey 14-15: SLIGHTLY SANDY CLAY: Dark yellowish brown CL Sand Filter (10YR4/4); fine, subrounded, well graded sand; stiff, low 15.0 15 Pack plasticity clay; trace micaceous silt; moist; some iron oxide 3.9/0.0 52"/60" 15-18: Similar to above, organic clay balls from 15 to CL 15.25'. 3.5/0.0 18.0 20 feet of 4" 18-20: Similar to above with no iron oxide staining. SCH 40 PVC CL 0.010-slot Slotted Screen 20.0 -20 with Threaded 12.7/0.0 30"/60" 20-22: CLAYEY SAND: Dark olive gray (5Y3/2); fine to Couplings medium, subrounded, well graded sand; soft, medium SC plasticity clay; trace micaceous silt; wet. 22.0 22.5/0.0 22-23: Similar to above with decreasing clay content. SC 23.0 23-25: SLIGHTLY CLAYEY SILTY SAND: Olive gray (5Y4/2); fine to medium, subrounded, well graded sand; SM 63.2/0.0 micaceous silt; trace clay; wet. 25.0 25 25-29: Similar to above. 30"/60" SM #2/16 Monterey 55.3/0.0 Sand Filter Pack 29.0 29-30: Similar to above with increasing silt content. SM 30.0 -30 Continued Next Page PAGE 1 OF 2



BORING/WELL CONSTRUCTION LOG

PROJECT NUMBER

SEALBEACH SEALBEACH.GPJ NEWGINT.GDT 4/28/08

50999-56254-6225.001.TK4.EQUIP

BORING/WELL NUMBER PMW8

PROJECT NAME DATE DRILLED Naval Weapons Station-Seal Beach, Site 70 3/26/08 Continued from Previous Page GRAPHIC LOG CONTACT DEPTH DEPTH (feet bgs) U.S.C.S. RECOVERY PID LITHOLOGIC DESCRIPTION WELL DIAGRAM (inches) (ppm) 110/0.0 20"/60" 30-32: SILTY SAND: Olive gray (5Y4/2.5); fine to medium, subrounded, well graded, sand; micaceous silt; wet. #2/16 Monterey SM Sand Filter 32.0 Pack 32-34.5: SLIGHTLY SANDY SILT: Light olive brown (2.5Y5/3); fine, subrounded, well graded sand; laminated 18.3/0.0 ML 166/0.0 micaceous silt; moist. SM 34.5 34.5-35: Similar to above with increasing sand content Total Depth of Boring = 35.5 ft bgs. 35.0 Threaded SCH -35 SM 116/0.0 40 PVC Bottom 40 45 50 55 60

Appendix E.4 Well Development Forms

Well No.:	EW1		Site: 500/	Beach			Date: 9/	20/07	7
Client:				Project Number	· -			 	
Well Casing	Diameter (ii	nches):	1''	Well Casing Ma	terial: (P	/c) ss	Other:		
Well Heads	pace:	PID (ppm):	0.0				<u>.</u>		
Samplers: (Lad Ma	arvin	with CDM				ب	with Blain	s-Tech-
Total Depth	of Well (feet	t):	35.0	2" - 0.16					
Depth to Wa	ater (feet):			(X) 4" - 0.65 G	Gal/ft. = <u> </u>	.52			
Water Colu	mn Height (fe	eet):	17.73	6" - 1.47				Minimu	m purge volume (gallons)
Well Refere			15.2						(galions)
PURGE ME	THOD:	Submersible	e pump 🔏	Bladder pump		osable bal	<u></u>		
	Model: 2" C		6 7	Depth of pump					
1	ment decon			Container type:	55 ga	1/00 D	rum	i	1110 "
Purge/deco	n water cont						s Pungi	2d - /	40 gallons
	Start Time:	0950) 		Flow Rate:			-	
Time	Gallons	Temp.	рН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments
1011	≈ 3	2254	6.54	4739	>1000	3.99	193.6		swab & Bail
1019	≈ 8	22.49	6.50	5076	> 1000	3.52	162.8		//
1027	× 20	22.67	6.53	5414	> 1000	4.22	163.5		<i>U</i>
1036	≈25	22.59	6.60	5696	> 1000	4.42	162.5		
								stwe	
1045	~	Begin	to Pump	p well				17.31	Pumping
	21.5 grm	23,28	6.38	5641	545	2.13	114.2	20.48	11
1055	≈20	23.41	6.40	5914	221	0.89	86.0	21.96	ž/
1100	≈30	2336	6.40	6001	890	0.70	46.0	21.95	
1105	21.7gpm			6009	695	0.69	49.2	21.96	11
1110	×40	23.41	ł .	6117	122	0.66	27.2	31.97	
1115		23.43	6.40	6129	183		27.0	21.97	//
1120	≈60	23,38	1	6/23	56.7	0.64	48-3	21.94	(1
1125	~ 1.79pm		6.40	6/60	45.7	0.67	60.4	22.10	11
1130	70	23.40	6.40	6/56	520	0.66	42.0	28-10	
				Chemets DO (r	ng/L):		-		
			Analyzed?	EPA Method	(Container 7	Type/Volume)	Preservative
Sa	mple Analys	ses:							
	,								
		····							
	llection Meth	iod: 🖌					0		
Pump: X	Flow Rate: Stainles: Type: cli spoe	s steel	Sample ID:				Sample Tin		
		able '	Duplicate ID:			··	Sample Tin Sample Tin		
	Desc.:		Equip. blank		CLI NIDC	E ANIE C	<u> </u>		
	MC		МО	NITORING W	ELL PURG	E AND SA	AMPLING F	OKM	

Well No.: A E VV	1	Site: Seal	Beach		<u> </u>	Date: 9	121/0	7
Client:			Project Number	:				
Well Casing Diamete	er (inches):		Well Casing Ma	iterial: P	vc ss	Other:		
Well Headspace:	PID (ppm):							
Samplers:		with CDM	·····				with Blain	e Tech
Total Depth of Well (feet):		2" - 0.16					
Depth to Water (feet):		(X) 4" - 0.65 (Gal/ft. =				
Water Column Heigh	nt (feet):		6" - 1.47				Minimu	ım purge volume
Well Reference Poin	t:		" - 					(gallons)
PURGE METHOD:	Submersible	e pump 🔲	Bladder pump	Disp	osable ba	<u> </u>		
Pump Make/Model:	2" Grundfos Re	diflo	Depth of pump	intake (feet	:):			-
Purge equipment de	contaminated?	YUNU	Container type:					
Purge/decon water o	containerized?	YUNU	Volume:					
Start Tir	ne:			Flow Rate:			_	
Time Gallon	Temp.	рН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments
1140 2 1.5	ym 23.94	6.44	6230	176	1.50	87.6	20.54	Pumping
1200 229	m 23.61	6.39	6110	189	0.71	93.6	I .	Pump & Sun
1270 229	m 23.53	6.40	6/15	148	1.07	174.2	2301	11
1231 =90	23.21	6.38	3315	85.4	1.08	1125	23.41	
1246 229	m 23.26	6.39	5830	119	1.04	105.7	23.91	(1)
1250 2 29		6.42	6641	183	0.84	108.7	I .	After Surge
1300 215		6.40	6389	83.2	0.71	1026	24.11	/ *
1308 22gg		6.43	7/80	290	0.56	115.1	21.38	After suage
1320 229		6.40	6451	104	0.76	165.2	24.27	Durin g
1326 219	5 23.37	6.41	6287	103	1.05	97.9	I	Affer surge
1341 2291	on 23.40	6.40	6416	<i>ω</i> 5.3	0.75	93.6	24,23	During
1345 22	0 23.68	6.46	6753	89.2	ł .	111.60	20_62	After surge
1400 729	on 23.41	6.40	6355	125	0.96	110.4	23.92	
1430 623		6.40	6467	45.2	0.85	112.2	24.18	During
1455 2270	23.40	Le. 39	6415	39-9	0.86	115.2	24.04	During
			Chemets DO (r	ng/L):		_		
		Analyzed?	EPA Method	(Container ⁻	Type/Volume	e	Preservative
Sample An	alvene:							
Sample An	aryses.							
Sample Collection N	∕lethod: 🖌					T		
Pump: Flow Rat	e:	Sample ID:				Sample Tin	ne:	
Bailer: Type: dis	posable	Duplicate ID:				Sample Tin	ne:	
Other: Desc.:		Equip. blank	ID:		<u>.</u>	Sample Tin	ne:	
CDM		МС	NITORING W	ELL PURG	E AND SA	AMPLING I	FORM	

Client: Project Number: Well Casing Material: PVC SS Other: Well Hoadspace: PID (ppm): with CDM with Blaine Tech Samplers: with CDM 2" - 0.16	Well No.: AEW1	ite: Sea	1 Beach			Date:		
Weel Headspapes: PID(ppm): Samplers: with CDM Wath Blaine Tech		F	Project Number:					
Well Hoadspace: PID (ppm): with CDM	Well Casing Diameter (inches):	١	Well Casing Mat	erial: PV	C SS (Other:		
Samplers:	Well Headspace: PID (ppm):							
Depth to Water (feet): (X) 4" - 0.65 Gal/ft. = (X) 3 =		vith CDM					with Blaine	Tech
Water Column Height (feet): 6" · 1.47 Minimum purge volume (gallons) Well Reference Point: Bladder pump □ Disposable bal □ Disposable	Total Depth of Well (feet):							
Water Column Height (feet): 6° · 1.47 Minimum purge volume (gallons) Well Reference Point: PURGE METHOD: Submersible pump	Depth to Water (feet):		(X) 4" - 0.65 G	al/ft. =	(X) 3 =		
Well Reference Point:			6" - 1.47			•	Minimur	
Purp MakeModel: 2 Grundios Redillo Purge equipment decontaminated? Y □ N □ Container type: Purge/decon water containerized? Y □ N □ Volume: Start Time:	Y .							(gallerie)
Purge equipment decontaminated? Y N Otomes: Purge/decon water containerized? Y N Otomes: Start Time: Flow Rate: Flow Rate:	PURGE METHOD: Submersible	pump 🗌	Bladder pump	Dispo	osable ba	<u></u>		
Purge/decon water containerized? Y N Volume: Flow Rate: Flow Flow Rate:	Pump Make/Model: 2" Grundfos Red	iflo	Depth of pump i	ntake (feet)):			
Start Time: Flow Rate:			Container type:					
Time Gallons (G°F) pH (Conductivity (µmhos/cm) (NTUs) (mg/L) (mg/L) (H TOC) (Comments (1507 ≈ 2gpm 23.52 to.41 to 20.3 54.2 fo.7 125.3 20.20 After surge (1520 × 2gpm 23.54) to.41 to 20.3 54.2 fo.99 (18.1 24.12) Puring (1538 ≈ 32.0 23.73 to.39 to.47 to.29 42.7 0.99 (18.1 24.12) Puring (1538 ≈ 32.0 23.73 to.39 to.39 57.5 to.1.58 (10.5 2.2347 Puring (1559 ≈ 335 23.44 to.39 to.39 55.9 to.90 (10.6 23.68 Puring (16.00 steep Pumping for 70 to.2) (10.00 steep Pumping for 70 to.2) (10.00 to.3) (10.00 steep Pumping for 70 to.2) (10.00 to.3)	Purge/decon water containerized?	Y L N L	Volume:					
Time Gallons (GPP) pH (umhos/cm) (NTUs) (mg/L) (mV) (11TOC) Ostillations (GPP) pH (umhos/cm) (MTUs) (mg/L) (mV) (11TOC) Ostillations (GPP) pH (umhos/cm) (MTUs) (mg/L) (mV) (11TOC) Ostillations (GPP) pH (umhos/cm) (mg/L) (mV) (11S S 23.97 puring (GPP) puring (MTUs) (mg/L) (mV) (mg/L) (m	Start Time:			Flow Rate:			-	
11me Gallots (Gre) Fire (jumhos/cm) (ints)	Ti Callara Temp.	nН				0.00		Comments
15 0	Time Gallons (Cy°F)	·						1010- 0
1520	1507 22gpm 2352				The state of the s			3
1558 \$\times 335 23.47 6.39 6535 59.4 0.90 10.6 23.68 puring 1600			6725					During
1600			6771		And the same of th			
1920	1559 2335 23.44	6.39			0.90	100.0	77.60	puring
1 1 25 25 23 39 6 35 5 23 9 2.7 2.53 151.6 19.77 During 19.75 29.83 0.35 5 5 5 8 1.13 126.5 22.83 During 19.78 29.87 23.08 6.51 5 5 26 2944 3.30 138.1 29.17 After surge 1000 20.57 23.41 6.42 60.13 41.6 1.00 120.5 24.25 During 15.10 23.45 6.42 60.85 21.2 0.99 120.6 23.29 During 15.30 22.95 23.31 6.46 70.14 2.8 0.82 127.3 21.02 After surge 15.30 22.95 23.31 6.44 60.76 15.6 0.87 115.9 23.88 During 15.95 24.25 24		Pumpin	y for To	day			stul	16.88'
1445			ump -		2 53	11.1		D
1948 2960 23.68 6.51 5.52 6 244 3.30 138.1 39.17 After surger 1000 20.5 20.25 During 15.10 20.415 23.45 6.42 60.85 21.2 0.99 120.6 23.29 During 15.20 415 23.77 6.46 70.14 36.8 0.82 127.3 21.00 After surger 15.30 229 23.31 6.40 6.272 21.5 0.77 12.0 23.88 During 15.45 × 460 23.33 6.41 60.76 15.6 0.87 11.8 23.83 During 16.00 × 475 23.31 6.41 60.76 15.6 0.87 11.8 23.29 During 16.00 × 475 23.31 6.41 60.76 15.6 0.87 11.8 23.29 During 16.00 × 475 23.31 6.41 60.76 15.6 0.87 11.8 23.29 During 16.00 × 475 23.31 6.41 60.76 15.6 0.87 11.8 23.29 During 16.00 × 475 23.31 6.41 60.76 15.6 0.87 11.8 0.33.29 During 16.00 × 475 23.31 6.41 60.76 15.6 0.87 11.8 0.33.29 During 16.00 × 475 23.31 6.41 60.76 15.6 0.87 11.8 0.33.29 During 16.00 × 475 23.31 6.41 60.76 15.6 60.87 11.8 0.33.29 During 16.00 × 475 23.31 6.41 60.76 60.87 11.8 0.87 0	1425 × 1.5 gpm 2339				500 700			
1000	1 2 2 2			- (
1510 x 3415 53.45 6.42 6.85 21.2 0.99 130.6 25.77 130.0 25.77 130.0 25.77 130.0 25.77 130.0 25.77 15.0 25.80 25.77 15.0 25.80 25.77 15.0 25.80 25.77 15.0 25.80 25.77 15.0 25.80 25.77 15.0 25.80 25.77 15.0 25.80 25.77 25.80 25.77 25.80 25.77 25.80 25.77 25.80 25.80 25.77 25.80		4		244	100	150.1	200	ATTER SUITE
15.20						120.5	2 2 20	paring
1530	1512 × 1415 23.45	6,42	6085	21.2		120.6	23.27	1.Ci
15 45				36.8		127.3	27,00	Du siange
Sample Analyses: Analyzed? EPA Method Container Type/Volume Preservative Sample Collection Method: Pump: Flow Rate: Sample ID: Sample Time: Bailer: Type: disposable Duplicate ID: Sample Time: Other: Desc.: Equip. blank ID: Sample Time:		0.1						During
Sample Analyses: Analyzed? EPA Method Container Type/Volume Preservative Sample Collection Method: Pump: Flow Rate: Sample ID: Sample Time: Bailer: Type: disposable Duplicate ID: Sample Time: Other: Desc.: Equip. blank ID: Sample Time:			-					During
Sample Analyses: Sample Collection Method: Pump: ☐ Flow Rate: Sample ID: Sample Time: Bailer: ☐ Type: disposable Duplicate ID: Sample Time: Other: ☐ Desc.: Equip. blank ID: Sample Time:		7			0.07	118.0	27,2	Dario
Sample Analyses: Sample Collection Method: Pump: Flow Rate: Sample ID: Sample Time: Bailer: Type: disposable Duplicate ID: Sample Time: Other: Desc.: Equip. blank ID: Sample Time:	1600	b bumbind	Chemets DO (mg/L):		_		
Sample Collection Method: Pump: Flow Rate: Sample ID: Sample Time: Bailer: Type: disposable Duplicate ID: Sample Time: Other: Desc.: Equip. blank ID: Sample Time:		Analyzed?	EPA Method		Container	Type/Volum	e	Preservative
Sample Collection Method: Pump: ☐ Flow Rate: Sample ID: Sample Time: Bailer: ☐ Type: disposable Duplicate ID: Sample Time: Other: ☐ Desc.: Equip. blank ID: Sample Time:								
Sample Collection Method: Pump: ☐ Flow Rate: Sample ID: Sample Time: Bailer: ☐ Type: disposable Duplicate ID: Sample Time: Other: ☐ Desc.: Equip. blank ID: Sample Time:								
Pump: Flow Rate: Sample ID: Sample Time: Bailer: Type: disposable Duplicate ID: Sample Time: Other: Desc.: Equip. blank ID: Sample Time:	Sample Analyses:							
Pump: Flow Rate: Sample ID: Sample Time: Bailer: Type: disposable Duplicate ID: Sample Time: Other: Desc.: Equip. blank ID: Sample Time:								
Pump: Flow Rate: Sample ID: Sample Time: Bailer: Type: disposable Duplicate ID: Sample Time: Other: Desc.: Equip. blank ID: Sample Time:								
Pump: Flow Rate: Sample ID: Sample Time: Bailer: Type: disposable Duplicate ID: Sample Time: Other: Desc.: Equip. blank ID: Sample Time:	Sample Collection Method: 🖌							
Bailer: Type: disposable Duplicate ID: Sample Time: Other: Desc.: Equip. blank ID: Sample Time:		Sample ID:				Sample Tir	me:	
Other: Desc.: Equip. blank ID: Sample Time:		-):			_		
THE PROPERTY OF THE PROPERTY O		Equip. blank	k ID:			Sample Ti	me:	
	CDM	M	ONITORING W	ELL PURC	GE AND S	AMPLING	FORM	

4 0-1 4

Well No.:	AEWI		Site:	Sent	Bench	NWS	Date:	9/25	167	
Time	Gallons	Temp.	рН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments	
1:50		Begin	ioum	in				16.95		
7.51	~ Jypin	3310	6.44	5245	24.9	1.91	94.1	19.52	After sta	7
7:59	25gcl	23.23	6.44	5971	2 6 33.8	1.04	240.6	29/40	_ , _ <i> ,</i> _	imping
8:16	50gr 1	23.31	6.44	6083	44.6	1.07	155.1	22,48 22.00	Ducing	
8:36	0+50	23.30	6.43	5900	35.5	0.88	134.7	20.63	After Sura	
2.52	75gal	23.30	1 42	6079	41.2	1.16	136.1	22.46	During	
9.06	100gal	23.52	6.38	6060	21.5	0.99	109.8	22.20	Ancion	
4.97	117 90	23.30	6.43	\$5980	215	0.80	129.7	22.65	During 1	
9.35	125901	23.28	6.45	605 5	14.1	0.96	(30.7	21.21	After sug	a
9.42	140gml	23.32	6.42	6016	27.1	0.77	126.7	23.00	During	
10:15	150 ga	22.22		6451	15.7	1.27	127.3	24.25	During	
10:24	180001	23.53	(.52	7568	11.8	1.50	125.0	20.65	After Sur	ŋ-e
10:30	19500	23.33	6.44	6070	19.0	1.03	120.7		Aring.	1
10:46	21597	23.37	6.44	6256	\$27,7	0.99	118.7	23.20	Daring	
([:35	52		- +	200	H2 W	ntil 5		otracted	b	
11:37	Begin ≈215ac	23.23	7 43	6122	9.51	1,35	132.3	19,30		1
11:45		<u>ව</u> ා. 0 ව	0.					22.11		
11:47	~ 232 (1)	23,36	6.42	6103	28.7	1.65	192.4	33.5		
11:50			0.216		. 0		1246	23.15		
11:65	~ 249 (34)	23.36	6.49	6495	11.9	0.99	134.9	23.16		
12:00	~ D (50)	53.36	6.45	6553	10.2	1.15	1305	23.16		
12:07	Fnd	50	1 Ta	7	10.00		1302	2,7,70		
End	AE	7	velopm	ent						
Total	740 gal									
		ļ ————				<u> </u>				1
								· · · · · · · · · · · · · · · · · · ·		
				 						
							<u> </u>			
							-	-	<u> </u>	-
			 				 			1
										-
			<u> </u>							
-	 									
							1	<u></u>		

	Well No.: /	EW 2		Site: Sea	eal Beach Date: 1/20					7	
	Client:				Project Numbe	:					
	Well Casing	Diameter (i	nches):	40	Well Casing Ma	terial: (P	vc) ss	Other:			
	Well Heads	pace:	PID (ppm):	0.0							
	Samplers:	chad M	arvin	with CDM				-with Blaine Tech			
	Total Depth	of Well (fee		<u>35.0</u>		•	0 - (
	Depth to Wa	ater (feet):			(X) 4" - 0.65 (Gal/ft. = <u>1</u>	7.06				
	Water Colu	mn Height (f	eet):	18.55	6" - 1.47				Minimu	m purge volume (gallons)	
	Well Refere			No.						(galloris)	
	PURGE ME		Submersible		Bladder pump		osable ba				
		e/Model: 2" C			Depth of pump						
		ment decon			Container type:	<u>55</u>	gallon	drum	1 - 110	1 ()	
	Purge/deco	n water cont			Volume: Bai				ed = 110	10 gallins	
		Start Time:	14:25			Flow Rate:			<u> </u>		
	Time	Gallons	Temp. (°C)°F)	рН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments	
	1440	≈ 9	23.69	6.54	6601	>1000	3.32	188.2		smab ¢ Bail	
	1452	≈13	22.98	6.59	6740	>1000	3,55	116-1		.,	
	1456	716	2271	6.61	7346	>1000	4.01	112,2		"	
	1502	=19	22.84	6-56	7793	>1000	3.67	118.5		٠,	
	1508	×25	22.69	6.62	8360	>1000	4.48	139.2			
	1515	229	22.57	6.5b	8736	7/000	4.06			"	
	15/9	234	22.44	6.58	8794	>1000	4.18	134.3		',	
	1530	241	22.43	4.62	8854	> 100v	4.75	148.6		·	
101	A 10.1.	<u> </u>							11 22		
9/21	0 840	Beg	in to	Punx	well.	C1 0	2 24	122 E	16.32		
	08 95	* 2 gpm	23.00	6.49	8857 8800	308 208	2.34	1225	19.96	pumping	
	0854		l	6.48 6.44	9207	178	1.35	113,8	21.04	t1	
	0900	1 2.5gpm	22,99	6.42	9744	123	0.48		21.42	11	
	0910	≈ 50 ≈ 60	22.96	6.42	9733	176	0.37	112.0	21.52		
	0 110	,, ,,	7.0.270	<u> </u>	Chemets DO (r			-			
				Analyzed?	EPA Method	(Container T	ype/Volume)	Preservative	
	50	mple Analys	.oc: — >								
	Sa	mple Analys	es. 						, .		
	Sample Col	llection Meth	od: 🖌								
	Pump: I	Flow Rate:	16 (1 00 1	Sample ID:				Sample Tim			
		\$ Fain/6 Type: dispos	able 1661	Duplicate ID:				Sample Tim			
		Desc.:		Equip. blank	ID:			Sample Tim	ne:		
	CI	MC		МО	NITORING W	ELL PURG	E AND SA	MPLING F	ORM		

Well No.: AEW?	Site:				Date: 9	171/07		
Client:		Project Number:						
Well Casing Diameter (inches)		Well Casing Mat	erial: PV	C SS	Other:	<u></u>		
Well Headspace: PID (p								
Samplers:	with CDM					with Blaine	Tech	
Total Depth of Well (feet):		2" - 0.16						
Depth to Water (feet):		(X) 4" - 0.65 G	al/ft. =	(X) 3 =			
Water Column Height (feet):		6" - 1.47				Minimur	n purge volume (gallons)	
Well Reference Point:		и		 _			(gailons)	
PURGE METHOD: Subm	ersible pump	Bladder pump	Disp	osable bal	_}·			
Pump Make/Model: 2" Grundfo	os Rediflo	Depth of pump i	ntake (feet):				
Purge equipment decontamina	F 1 1 1	Container type:						
Purge/decon water containeriz	red? Y N N	Volume:						
Start Time:			Flow Rate:					
Time College Ter	mp. pH	Conductivity	Turbidity	DO	ORP	DTW (ft TOC)	Comments	
	PF)	(µmhos/cm)	(NTUs)	(mg/L)	(mV)		9 46	
0915 × 2 gpm 22		9792		0.34	111.7	21.62	Pumping	
0925 285 22	1 .	9822	51.3			22.03	ir	
0935 2105 22		9826		0.56	111.4	22.19	11	
0950 2 gpm 22	.93 6.43		280.1		111.5		11	
1000 × 140 22		9787		0.63		i i	· · ·	
1010 × 1.5gpm 22			10.4	0.45		21.34	10	
1020 2175 23			6.86	1 -	•		11	
1030 & 1.5gpm 28			5.81				1/	
1035 2200 23	5+ op P	1010 M						
	31 OF 1	They				57WL 16.15		
12:55 Begin	pumpin	Q 20	10 H2					
12.59 23	.96 6.56	1)	76.2	1.81	88.7	18.15	During pu	
	3.13 6.53	7236	59.6	1.30	112.5	19.00	During	
13:20 = 25 gal -			37,9		 	19.00	During	
12.5		Chemets DO (mg/L):		_		J	
	Analyzed '	? EPA Method		Container	Type/Volun	ne	Preservative	
	Analyzou							
Sample Analyses:								
.								
Sample Collection Method:	/							
Pump: Flow Rate:	Sample ID:				Sample T			
Bailer: Type: disposable	Duplicate l					Sample Time:		
Other: Desc.:	Equip. blar	nk ID:			Sample T	ime:		

Vell No.:	AENZ		Site:	Seal Be	ach N	WS	Date:	9/25/0	7	
Time	Gallons	Temp. (°C/°F)	pН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments	
131.29	=30 cm	23.02	6.47	9031	50.2	0.99	128.8	19.70	After sure	e
1235	340 30 Sol	2 3.05	6.47	9139	74.7	1.02	126.0	20,57	Arin	
12.5)	= 65 gal	23.03	6.46	9487	53.8	0.90	125.0	20.65	During	
14.00	=82.1	23.08	6.46	9518	46-6	0.80	121,3	20.86	Dur: Za	
14121	=105gal	23.01	6.45	9723	36.8	0.45	1219	20,83	Daring	
14'29	Begin	5-0	1	Test	@ 20	oflz	,			
111: 701			gal					12,17		
14:35	=120gal	23.01	6.45	9806	246	0.48	121,8	20.05		
14:40	100%							20.68		
14:45	~135 gal)	23.05	6-46	9707	365	0,47	119.0	21.03		-
14:50				701				21,20		·
14:57	[2000]	23,05	6.45	9854	123	0,12	109.6	21.50	Possible err	1
- 1 1 4 - 1	(20gal	Tect				<u> </u>				
15:04	Resu	10-16-21	mping							
15.17	187 901	23.03	6.45)	9954	57.6	0.43	109.3	20.83	During	ļ
15:27	200 go 1	23.03	6.45	9852	12.3	0.69	119.3	20.91	Ancia	
15:36	205ga	23.28	6.50	9551	7.38	0.60	135.2	19.25	After Janes	Ļ
15:45	22096	23.05	6.45	9812	14.0	0.72	122.0	20.78	Our	
	End	1	ping	for	today					
9/26	Buir	Pon	ר מנמי	10 Line	is be	- 0	20 g. /	ans co	way pung	1
1244	220	22.87	6.46	9601	15.7	1.20	19.1	A 0	By Draw	
2754	~240	2304	4,52	8559	61.0	139	158.8	20,66	0 / 3	
7804	~255	23.0 j	6.48	9893	10,5	0.70	142.3	21,22		
7814	200	22.98	7.00	9287	117	0.92	135.2	16.20	Diggina	 -
0826	~278	23.02	6,50	9621	38,8	0,50	117.2	16.20	Dirga risuma	11/
२८ ४ २८	~290	23.06	6.48	9653	43. 2.	1,10	1547	22,1		
9539	Rino	5 Loppe	1					100.0		
2844	300	22.75	6.49	9741	13.8	0.56	142.6	19108	Pumping 1	ķ,
<u> </u>	320	23 00	6.48	96 96 18	18.2	0.93	125.4	23,7		\
0356	325							1	Purga pausi	ľ
190 i		22,95	6,52	7673	10.8	1.05	137.5	19,25	Kesum prom	P
916	345	22.97	6.48	9686	23.7	1.11	142.0	21.00	pr/32 09/	8
921	350	22.93	6.49	9733	10,9	0.76	138.Z	20,30	Resume	
0931	365	28.02	I	9746	109	0,74	124,0	21.08	pers -09.	18
1936	376	22.70	1	9833	14.4	0,42	125,8	20.65	STOP. 0941	
757	370	22.95		9725	43.7	4.72	123.1	18,30	Resum	
1007	390	23.04	6.48	9827	20.8	1.26	163.2	70.65	Ocusa 1009	
D12	355	22.98	6.49	9724	10.4	0.71	1575			
1024	415	13.01	6.48	9158	56,1	276	133,9	31,6	Revse 10	4
,	 	 	<u> </u>	1			1			1
					į.	<u> </u>	1	<u></u>		-11

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Vell No.:	AEWZ		Site:				Date: 9/			
Time	Gallons	Temp. (°C/°F)	pН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments	
1057	7	22.96	6.48	9849	41.5	0.34	130.4	19.37		
107	25	23 10	6.48	9713	26,5	1.09	215.0	20.90		
117	50	23.09	6.47	7846	47,5	0.30	140.7	21.46		
177	70	22.99	6,47,	9913	98,4	0.18	117.8	22,41		
1137	85	23.08	6,47	9917	137	0.60	95.8	22.24		
1147	100	23.09	6.48	9892	184	2.67	96.1	12.48		
1157	120	23.08	6.48	9875	65,5	0.67	96,5	22,54		
1207	140	22.97	6.48	9878	372	0.72	99,1	22,98		
12/7	160	23.13	6,48	9801	72.3	1.01	107.5	23.19		
227	180	23,14	6.42	785-6	21,4	0.72	111.1	22.56		
1237	200	23.16	6.47	9800	8.27	0.77	112.4	22.52		
1277	220	23.05	6.47	9896	9.17	0.91	114.0	23.03		
1257	240	23.14	6,48	9892	26,2	0.71	113, 6	23.02		
1307	266	23.15	6.48	9921	50.7	0:63	114, 4	23.02		
1317	280	23.04	6.47	9330	6.04	0.77	114,1	23./6		
327	300	33.08	6.48	13 8%	7.46	0.69	113.6	23.01		
1337	320_	23.00	6.48	9997	3.76	0.93	114.9	22.84		
1347	340	23.13	6.48	9705	11.2	0.68	112.7	23.33		
1357	360	23.12	6.48	99,05	4.40	0.72	112.1	23,10		
1410	330	23.10	6,49	7836	2.20	0.72	111.0	23.06	Pump Stop	nd
1420								15.95	5tali	
1453	385	23,50		9832	29,3	1.81		19.65	Resim &	Vingoir
1440	405	23.28	6.48	9820	15.3	1,19	194.3	20,88	/	82
	425	23,21	6.47	9711	20,4	0,52	145.6			14
1800	446	23.14	6.46	9819	31.3	0.36	136.0	21,91	Plynp stop	w is
1510	450	23.10	6.50	9763	14.7	0.28	135.8	1	By n Pump	'n
1520	470	3.12	6.46	9863	18.0	0.89	129,1	21.17		,
1530	490	23.08		9775		0.51	128.5	2,93		
<u>1540</u>	510	23.03	6,47	9877		0.38	121.4	22.05	Prmping S	201921
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1430

2.08 - Vellbux No toc

Well No.: ۽	ATW!		Site: Sex 1	Buch		<u>-</u>	Date: 9/37/67			
Client:				Project Number	er:			/		
Well Casin	g Diameter (inches): 4	<i>((</i>)	Well Casing M	aterial: 🛭	vc ss	Other:			
Well Heads	space: <i>O. O</i>	PID (ppm):								
Samplers:	·—		with CDM					with Blair	ne Tech	
Total Depti	n of Well (fee	et):	Be 31,1			. 0.4				
Depth to W	/ater (feet):			(X) 4" - 0.65	Gal/ft. = 🧏	7 U	(X) 3 =		-	
Water Colu	ımn Height (•	13.79	6" - 1.47	,			Minimu	ım purge volume	
Well Refere	ence Point:		706		-				(gallons)	
PURGE MI	ETHOD:	Submersib	le pump 🔼	Bladder pump	L Disp	oosable ba				
Pump Mak	e/Model: 2"	Grundfos Re	- 107 1	Depth of pump						
	pment decoi			Container type						
Purge/deco	on water con		YMNU	Volume: Bar				d = 329	gallons	
	Start Time:	0353			Flow Rate	: ~ Q E	1,5 gpm			
Time	Gallons	Temp.	рН	Conductivity	Turbidity	DO	ORP	DTW	Comments	
		(°C/°F)		(µmhos/cm)	(NTUs)	(mg/L)	(mV)	(ft TOC)		
08/7	10	2235	7,43	16662	6>999		142.5		Sweb+bail	
0823	20	22.07	7.43	17438	>999	5.93	135.7			
5.56 E					1000			19.0-	<u>.</u>	
1	35	22.25	7.15	20017	2999	1.55	53.7	19.97	Begin pumpin	
0720	40	20,50	7.15	19775	2999	1.79	5-4,1	20.75		
0930	45	22,45	7,16	20041	1000	262	58.3	21.23	150hz	
0946	55	31.88	7,14	17586	1000	3.78	58.1	24,12		
0950	65	22.00	7.15	19892	0000	3.10	51.3		Fine send in p	
1000	75	21,79	7,10	19951	997	1.77	37,9	26.75	extr	
1010	85	01,59	7.07	20289	221	1.44	48.8	27/1	"	
1020	95	91. 31	7,09	20329	217	213	56,1	27.68		
1030	108	21,27	7.09	20423	72.3	208	40,1	<i>3</i> 2.75	Purg, my passil Resum@ 150	
1100	105	22.74	7.13	20325	115	2.20			l II	
1110	115	37.84	7.09	19501	590	3,08		22,63		
1159	125	22.73	7,10	19,255	1000	2.91	18:2	25.58		
				Chemets DO (r	mg/L):		_			
			Analyzed?	EPA Method		Container ⁻	Гуре/Volum	е	Preservative	
C.	ample Analys	100' - h						. (
36	imple Analys	ses.								
						/ · · · · · ·				
									·	
Sample Co	llection Meth	nod: 🖌								
Pump:	Flow Rate:		Sample ID:			Sample Time:				
	Type: dispos	able	Duplicate ID:			Sample Time:				
	Desc.:		Equip. blank	ID:			Sample Tin			
	MC		MO	NITORING W	ELL PURG	E AND SA	MPLING	ORM		
5	-1 71			OIII 10 71)		OF		J 44174		

ell No.:	AIWI		Site:				Date:		
Time	Gallons	Temp.	рН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments
1130	135	33.05	70	1067	763	2.42	45,9	28.32	Still his Six
140	146	22,97	7.07	19865	31.2	2.14	43.3	28.22	i)
150	185	23.11	7.06	17905	131	2.15	29,8	28.60	1)
200	165	23.10	7.06	19910	165	2.23	34,1	28.53	Priva puging
230 130	165	270	7.08	20520	26.4	0.66	41,6	1202	Resim O K
240	175	22.80	7.06	19874	265	2.17	45.4	22,91	
250	185	23.01	7.06	19430	31 1	3.61	75.7	21.01	
1300	195	23.00	7.05	19685	263	2.24	53.2	27.37	
310	205	22.97	7,05	19837	155	1,88	47.6	28.08	
320	215	22.96	706	19863	36,1	2.20	57.9	28.55	
330 330	225	23.00	706	19891	29.2	2.13	56.8	28.27	Rigin As po
400	225	22.64	7.07	20530	10.8	0.75	31, 2	19.08	Resum po
1110	235	22.94	7,08	19428	235	3.87	65,0	24.09	
1420	245	22.99	7.07	17453	303	3.36	79,2	26.01	
1430	255	23.00	7.05	19694	124	2.21	66.8	26,95	
440	215	23.03	706	19817	325	1.75	61.2	27.07	
450	275	2252	207	19847	15.9	183	54.3	21,48	
80D	285	22.54	207	2000 7	15.6	2.00	53.5	27.96	
1210	295	2,57	7,07	20030	10.0	2.10	56.8	28,14	
1820	305	22.66	707	19993	5.86	2.14	55.7	28.21	
1530	315	22.67	7.07	19989	6.14	208	55, 3	28.34	
1540	325	22.68	7,07	19979	5.68	2.17	56.0	28.34	
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Pg 1 0+2

2" - 0.16 2" - 0.65 (2) 6" - 1.47 —" - Depth of pump Container type:	aterial: PV ID (ppm): N Gal/ft. = 3	/A	3=	vith-Blaine I	reeh- n purge volume (gallons)
FCDM 5 2" - 0.16 2 9 (X) 4" - 0.65 (C) 6" - 1.47 " - " - Depth of pump Container type:	ID (ppm): N Gal/ft. = $\frac{13}{2}$. Dispos	//A // (X)	3 =		n purge volume
2" - 0.16 2" - 0.65 (2) 6" - 1.47 —" - Depth of pump Container type:	Gal/ft. = $\frac{13}{100}$.	0 (x)	3=		n purge volume
5 2" - 0.16 2 9 (X) 4" - 0.65 (2) 6" - 1.47"	Dispos		3=		m purge volum
5 2" - 0.16 2 9 (X) 4" - 0.65 (2) 6" - 1.47"	Dispos		7	 Minimur	
(X) 4" - 0.65 (C) 6" - 1.47 —" Depth of pump N	Dispos		7	 Minimur	
Bladder pump Depth of pump Container type:	Dispos		7	✓ Minimur	
Bladder pump Depth of pump Container type:		able bailer	7	,	(gallons
Bladder pump Depth of pump Container type:		able bailer	1		
Depth of pump		able baller L	1		
N Container type:	IIIIake (leet)				
N L Container type.		•			
IN DIVISION ROLL	10 d - 014	11000	Pune	d - 410	an Iluns
		•	100/17		90011000
		and the second second second			
Conductivity					Comments
(µmnos/cm)	^	! 1 :	(1117)	(11.100)	
		71	1770		Swab & bail
				2681	sumb t bai
			•	•	il
	-	7.87	9 1 • 1	2/11	
				:704	0
o Pump ive		- 24	02 1		Pumping
	•	-		The state of the s	Pumping
				7	"
		· ·			Removed
mping to my	ch sedi	ment @	bot tin	4	Removed 2 45 gc
Trab & B	ail —			STWL	
67 10899	71000	6.75	246.3	15.23	shub & Ba
69 16630	>1000	7.70	213.3		Swab & Ray
W/ Smab	Bail			egwi	
	1011 -			15.94	Pumping
	(mg/L):				
		Containor T	ivno/\/olume		Preservative
Method		Container	ype/ volume		11000110011
			T		
mple ID:					
plicate ID:					
uip. blank ID:			Sample Ti	me:	
	Conductivity (µmhos/cm) fo Swab 62 5925 54 6100 58 6212 Swab & Ba 9 10580 31 9349 23 10950 mfing to ma 54 B 67 10899 69 16630 W/ Swab & B Chemets DO Method	Flow Rate:	Flow Rate: Conductivity	Flow Rate:	Flow Rate:

				2.5					
1002									
					, , ,	,		1600	(11)
		pa	8012130	1121	-igning	80+5		र्व्ध ३१६ %	5260
IVING CT	no"16	C 801	15E	8.6	80191	1.12	1000	NIB112	0660
72 De 57	70.66		172	6.9	18651	213	78 CC	W61.13	5060
1000	20.56	£.7P	80 h	82	ptini	1,12	08 CC	106 OK.E.	2060
solar ?	88.66	C 36		0.5	61091	51°2	68.66	4.161-1 %	5680
14m 25	6770	726	84.8	6.61	56651	212	ZLEC	w85112	5880
1447 51	E0:50	F. 8P	3.41		78691	11.7	1966	1260182	5680
rul ws	1978	5801	08 %	F. 8	801.71	61. L	51-CC	nd5/12	0180
300005	56.26	5'201	3.30	0.81	-		187°C	udb1-12	5560
10/w 5/	487C	L.C31	66%	L.11.	[868]	88.6	29.66	w/6/12	91.40
bujdunj	79.77	0.996	16.2	333	27561	1.1.7	01600	1	OhLO
	PE SI			19 M	dund	94	. 8	126066	0101
			Today		prignil				0101
in oh	36 7C	2/48	678	800	108 \$1	612	21 66	POLE ~	
N-1051	6838	800	652	the	8/851	512	() (1061-1 X	SK
50/4 56	157E	L 78	69.6	L h	661.51	1.12	80€€	U16112	sest
sows!	16 60	118	87€	30.3	186.51	112	50.EC.	10h086	5/51
14/45	1000	29.3	150	820	et151	812	18:00	w/61.1 %	
rvin 58	3670		25€	6 ° h	04151	51-6	118C	20/6/12	5661
5 min 21			686	8-01	96861	CIZ	21.86	watt 1.1 x	5881
3000	E1:50	3 58	18.6	87	8661	512	8126	486171 ×	sehl
	10°CC		19°C	8.5	76861	31.7	6/26	106581	011.1
10,0057	E1 70			776	80[6]	21.7	1.6.50	48/1/2	OOEI
10,0051	1.8°h€		C5 C		CZOKI	1/2	6080	-db11 %	0581
svins	88.00		10.8	10.1	198941	01.7	8180	w/61-1 1/2	08.81
solmot	56 5C		C76	C8-6		11:1	6186	P651.1%	5/ 21
inin?	LCCC			01.01	7051.1	11.7	06.56		1300
rywsh	157C	6.78	89.€	85.0	556 hl	61.7	OCEC	w/61-1x	Shel
544 05	85.5€	5,28	876	5.01	EOFKI	8/2	1-6-49		0861
501051	EL "CC		3.38	Chl	44981	312	16.86	10-	5111
of 1628	16 L. 91	1876	10.4	6.18	084 61	08.2	PC.EC	1	
Fuldring of t	A 100 P		180	78 %	368 61	0C.T	HI'EC		0/01
bull ynd	6650		26.0		58611	162	C1. Ec	1 100	
Fish of E	ht 50			58.6	825 61	16.7	C186	w/6/1/2	
	EZ he		59°C	4	676 61	61.7	13.69	w/66.1 %	,
11	€0°C€			001	CLIKI	he'L	1050	496 C. 12	
	36 ZC				56cs1	LCL	LLECC	"186 Str	0 h 0
"			66-6		76061		83.66	mare 1 -1	
и	Ch 20		61.5	1668	1059:		CC 26	was 1:	o col
6.16.100	78.61		40	- 6 6			Ltiet		\$ 5000
<i>bidung</i>	,87-61	9 hh			nmhos/cm)	1/		Sallons January Art	
Comments	(TOC)	RAO (Vm)	(mâ\r) DO	Turbidity (NTUs)	onductivity		Temp.	auolle 5)	
	Z 0/				Beach	16: 50 m/	!S	CMI	:.oV lle
	_ V/	1/ 9/ .540			, 0				

Cto CbJ

Pg 1 of 3

Well No.: AMW 1	9	Site: Sea	1 Beach			Date: [0]	12/07	
Client:			Project Numbe	er:				
Well Casing Diameter (inch	nes):	11	Well Casing M	laterial: (F\	/c) ss o	ther:		
	D (ppm):	0.0		FID (ppm): N	VA			
Samplers: Chad Ma		with CDM					with Blaine	F ech-
Total Depth of Well (feet):		35,10	<u></u>					
Depth to Water (feet):	- -	18.88	- (X) 4" - 0.65	Gal/ft. = 10	<u>, 54</u> (x)	3 =		
Water Column Height (feet	t):	16.22	6" - 1.47	·		•	Minimu	m purge volume
Well Reference Point: TOO		F-71			Г	7		(gallons)
		1	Bladder pump		sable bailer L	<u> </u>		···
Pump Make/Model: 2" Gru	_		Depth of pum):			
Purge equipment decontar	minated? `	<u>HuAy</u>	Container type		<i>C</i> 13		1.1. 7.	201 11 1
Purge/decon water contair	nerized? `	ר א ואלו א	Volume: 🛭 🛭 a) gallun	s fun	40d= 30	gallons
Start Time:				Flow Rate: _				
Time Gallons	Temp. (°C/°F)	рН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments
0950	Begin	to 5	urge	* Bai	,		18.88	
	23.84	7.72	15982	>1000	5, 32	128-4		swab ¢
	22.90	7.77	15643	>1000	5.47	111.4		Bail
1040 240	22.64	7.70	19098	>1000	4.65	83.7		"
1050 245	Finish		swab	& Ba				
	Begin	/,,	imp we	11			20.61	Pumping
	23.84	7.52	17210	7/000		24.0	3006	Pumping Pumping Pumping
	23.54	7.46	18440	>/000	10.90	56.3	31.73	Pumping
1130	STOP	Pumpi	ng 18+	re// R	ecover		5746	
	uné	Pun	ping-				25.26	
1230 =1 gpm =	2 3. 35	7.03	16203	784	6.40	74.7		Pumping
	24-17		17279	402	7.65	69.7		20 mins
	25.18	7.29	16977	87	7.72	632	3136	35 mins
	25.22	7.29	16526	71	8.14	le le-le	31.70	45 min
1340 2 0.9 mm	23.98	7.15	16214	>1000	10.60	39.7	29.45	10 miss
			Chemets DO	(mg/L):				
		Ме	ethod		Container T	ype/Volume		Preservative
Sample Analyses	a							
Sample Analyses	J. –							
Sample Collection Me	ethod:							
•						r		
Pump: Flow Rate:		Sample ID:	· · · · · · · · · · · · · · · · · · ·			Sample Tin	ne:	
Bailer: Type: disposal	ole	Duplicate I	D:			Sample Tin		
Other: Desc.:		Equip. blar	ık ID:			Sample Tir	ne:	
CDM		1	MONITORIN	G WELL PU	RGE AND S	AMPLING	FORM	

Pg 20+3

W	/ell No.: /	4nw1	1	Site: So	eal Bea	ch		Date:	10/2/0	7
F	Time	Gallons	Temp. (°Ĉ) °F)	рН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments
1	1350	× 0-legin		7.24	17477	261	9.19	54.9	32.27	20 mins
\parallel	1400	- Jugin	- well	went					72.2	_
	1425	≈ 0.6ggn	انسمما	7.17	16312	101	8.83	70.6	31.89	5 mins
\parallel	1535		23.14	7.08	15291	45	9.12	86.5	30.10	5 2/2
F	15 15	~ a. leg pm		7.16	16777	410	8.29	48.7	32.63	
\parallel	1555	Malican	27.68	7.11	16892	207	8.01	563		25 mins
	600	21109	41. STOP	pumpi		Today			57606	
71	0740		Begin		mp ive	, ,			19.04	
	0745	~ 0.89ar	22.86	7.09	15788	438	6.00	2284	24.21	5 mins
н	7755	2 0.7gpm	22.68	7.13	16777	830	6.01	208.9	26.27	15 mins
11	805	20.70am	2 3.06	7.19	16884	497	5.81	1552	27.39	25 mins
Ħ	0815	20,6900	23,23	7.12	16258	788	6.15	172.4	28.86	
- 13	0830	20.59pm	23.45	7.19	16650	137	6.76	l	30.00	· ·
	0845	20.59pm	23.17	7.76	17687	87	6.91	776.	32.85	65 mins
11	0855		t. 1200 //	went					STWL	
4	0975		Resu	me Pu	mping	.00			29.68	
1	0930	20.5gpm	22.85		14915	189	7. y8	205.2	31.72	5 mins
- 11	0940	20,5gg~		7.24	16584	332	7.30	1654	33.24	15m/ns
41	950		2387	7.20	15855	5/2	6.49	167.7	33.89	25 mins
41	7 955		- well		Dry -				stut	
F	1025	24 . 12	Resun	e Pun	ping -	1 - 0	702	177 /~	29.11	Pre- Fan
ŀ	1030		23.28	7.09	15056	102	7.83	1775	31.09	#5 nw
ŀ	10 40	2 0. Sopr	24.05	7.14	15913	53 259	7.35	217.6	32.74	
H	055	× 0. Sapon	74.71	7.15	18881				33.48	
H	105	~ 0.5 gpm	24.68	7.16	16069	395 48		93.5	33.91	
H	1125	20.5 pp	24.87	7.13	15814	48	9.81	142.6	33.93	60 mins
H	1130		well	went					STW L	
H	1700		- Resu	7,03	Pumping	109	10.89	193.1	29.78	10 mins
1	710	~ 0.89pm	23.96	7,03	14610° 15407	111	11.10	153.2	31.13	20 mins
\mathbb{H}	1220	~ 0.5 ym	23.98	i	15731	17	12.13	1272	3206	Yomins
\parallel	1240	≈ U.Sgpm	2461	7.08	15532	39	11.91	142.1	31.99	55 mins
-	1255	=0.5gpm	24.92		D (V			(101)		65 mins
1	1305		Nell Resun	o Pin	ping -				574137	
\vdash	1330	208	24.47	7.05	15017	127	14.23	120.6	30.19	10 mins
\parallel	1350	20.5gpm = 20.5gpm	24.32	7.04	14881	91	13.82	73.4	30.47	20 mins
1	1400	~ 0.6gun		7.08	14999	204	13.36	129-8	32.81	30mins
\parallel	1410	Ad) ysted Slightyng	W		, ,	C Y -				
\parallel	1430	> lightly ME	Resu		mping				29.86	
\parallel	1435	20.5 ypm	23.69	7.06	14877	76	15,40	1787	3073	5 mins
\parallel	·	~ 0. Sam		7.07	15620	41	16.16	404.6	31-97	20 mins
1	1770	-9///								

Pg 3 of 3

		·						19	
Well No.: /	AMN 1		Site:	seal Br	ach		Date:	10/3/0	7
Time	Gallons	Temp.	pН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments
1510	≈ 0.5gpm	24.28	7.13	15955	784	15. 98	398,2	33.86	Yomins
1320	~ 0.5gpm	24.57	7.12	15831	453	15.41	372.9	33.88	50 mins
1535	20.4gra		7.10	15675	8/	14.85	346.2	33,90	65 mins
1345		Well		1	, ——			stuck	75 mins
1550			me	Pumping	Well	-		30.01	
1555	20. ygm	23.98	7.08	15526	199	17.17	485.7	3327	5 mins
1600	120 gal.		fungi.	1	Today -				10 mins
0725		- Beg	i ' ' ' '	Pump				57602	
0730	20.5gan		7.09	15927	152	5.97	508.7	22.51	Smins
0740	~ 0.5gpm	· _	7.07	15331	593	6.01	513.0	25.44	15 mins
0750	≈0.5 gpm		7.11	16043	359	5.95	443.9	27.44	25 mins
0805	≈0.5gpm		7.07	15578	91	8-16	310.9	29.38	Homins
0820	~ 0.5gpm	. ,	7.13	15902	35	10.40	157.7	31.02	55 mins
0840	20.5gpm	i _	7.15	15430	90	12-64	159.8	33.86	75 mins
0850		well	went	Dry -					85 mine.
0920		Resu	me Pu	mping V	re/1 -			28.96	~ ~~~
0935	20.5gpm	22.78	7.17	14811	83	8.84	80.4	29.67	5 mins
0940	2 0. Sapor	23.89	7.09	14844	32	13.78	106.5	30.63	20 mins
0955	~ 0.5gfm	14	7.10	15533	18	13.96	495.2	32.31	35 mins
1005	~0.5gm	23.98	7.13	15657	409	12.49	16.5	33.75	45 mins
1010		well	went						
1030		Resur	e Pur	ping w	e1/			29.96	
1035	=0.5gpm	24.08	7.67	14682	26	15.54	49.2	30.25	Smins
1045	245 0A2	24.11	7.05	14453	13	15.78	118.6	30,99	15 mins
1050	~ as ofm	24.30	7.64	14523	8	16.10	132.4	31.07	20 mins
1100	~ Sapo	24.36	7.05	1450) mping Go	4-9	16.14	136-2	31.26	30 mins
1100	70 gal.	<u> </u>	top Pu	mping Go	+ to Ne	xt we	//		
									,
				<u> </u>					
								-	
							<u> </u>		

Pg 1 of 2

Well No.:	4mw2		Site: 50	al Beac	ለ		Date: [0]	14107	<i></i>
Client:				Project Number	er:				
Well Casing	Diameter (i	nches):	1 ''	Well Casing M	laterial: (P\	vc) ss c	ther:		
Well Heads	pace:	PID (ppm):	0-088	m	FID (ppm): 1	V/A			
Samplers:	Chad A	narvin	with CDM	des (1000 100 100 100 100 100 100 100 100 1				with Blaine	Tech
Total Depth	of Well (feet	t):	35	2" - 0.16	1.0	10			
Depth to Wa	ater (feet):			(X) 4" - 0.65	Gal/ft. = <u>/ / /</u>	<u>', / / (</u> x			
1	mn Height (fe		15.68	_ 6" - 1.47				Minimu	m purge volume (gallons)
Well Refere	nce Point: T	THE RESERVE OF THE PARTY OF THE	<u> </u>	"				0.0 1.000 //	(galloris)
PURGE ME	THOD:	Submersible		Bladder pump		sable bailer			1000
	e/Model: 2" G			Depth of pum		:):			
	ment decon			Container type	,	<i>F</i> ° 11	0	; F1	<i>r</i>
Purge/deco	n water cont	ainerized?	Y N N N	Volume: Ba		> gallor	s fung	pe d= 59.	9alluns
	Start Time:				Flow Rate: _			and the constant	and our management of the same
Time	Gallons	Temp.	рН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments
1130	_	Bægi	n to	Surge	& Ba	11 -			surge &
1205	4	23.82	7.36	19221	>1000	8.26	2241		Bail
1210	20 -	-well	Went	Dry-		N	0.011 /		surge &
1230	40	23.24	7.38	19384	>1000	18.59	234.6		Bail
1235	43-	- well		Dry		12 10	70/7		Surge &
1310	48	22.59	7,24	19503	> 1000	13.10	326.2		Bail
1320	65	Finisi				ail -			
1345		Begin		Pump		270	F 0	20.10	0
1350	20,7 gpm	22.59	7.10	20190	71000	2.78	-5.2	22.18	Pumping
1400	20.6gpm	22.87	7.10	19006	>1000	5.75	70.6	23.23	15min
1415	2 0.6gpm	22.96	7.05	18702	71000	7.74	101.1	24.10	30 mins
1430	20.6gpm	22.99	7.03	19098	882	6.35	774	1110	45 mins leoning
1445	~ 0.6gpm		7.05	18677	791	7.01	24-1	2432	75 mins
1515	2 0.6 gpm	23.09	7.05	19256	554	7.06	98.1	27.49	90 mins
() ()	10.09pm	71.01	7.0	Chemets DO			1001	, , , ,	1 - 11/13
			Me	ethod		Container T	ype/Volume		Preservative
Sa	mple Analys	es: →							
	0 - 11 11 1	M - 411.							
Sample	e Collection I	vietnoa:							
	4 .		0 1 15		l		Consula T'		~~~
	Flow Rate:	-1-1-	Sample ID:				Sample Tin		
	Type: dispos	able	Duplicate II Equip. blan				Sample Tin		
	Desc.:				TAIRLE DIT	OCE AND C			
			Ŋ	MONITORING	J WELL PUR	GE AND S	AMITLING		

Pg 2 of 2

										000
	Well No.: A	MW2		Site: Sea	al Beach				5/07	
	Time	Gallons	Temp. (*Ĉ)°F)	pН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments
	1530	≈ 0.8gpm	23.02	7.07	19457	253	4.10	100-le	28.11	105 mins
	1545	20.88pm	22.97	7.06	19240	284	3.75	69.1	28.19	120mins
	11000	~ 0.8gm	22.88	7.09	19910	183	4.11	123,8	30.21	135 mins
	1600	140901	Stop	rumpi	ng for	Today				
	0800		Begin	+0	Pump	Well-			5tal 19.45	
10/5	0805	~0.7gpm	22.53	7.12	19285	174	5.54	481.7	23.65	5 mins
	0815		22.45	7.08	19310	43	4.40	407.8	25.02	15 mi~
	0830	~ 0.7gpm ~ 0.8gpm		7.06	19053	389	3.46	341.7	25.98	30 mins
	0875	20. 9gpm	22-64	7.06	19146	105	3,60	139-3	26.78	45 mins
	0900		99 - 6	7.07	19/68	111	8.66	133.1	26.88	60 mins
	0915	7	22.65	7.10	19777	260	3.04	74.9	30.04	75min
	0930		, , , , , , , , , , , , , , , , , , ,	7.09	19560	182	3.12	79.8	31.17	90 mins
	1000	~ 1.2gp	22 75	7.09	19388	14	3.46	1220	31.42	120 mins
	1005		Stop	purpin		WP 11 R	ecover			
	1030	155gal		une F	unping	Well			21.07	<u> </u>
	1035	~1.2gap		708	19105	74	3.71	11/0.1	23.98	5 mins
	1040	=1.29Ab		7.05	18685	108	4.80	125.1	25.38	10 mins
	1050	~ 1.2gpm	22 74	7.04	18849	110	3 03	113.9	26.84	20 mins
	1105	~ 1.2 gpm	22.80	7.06	19214	23	4.85	124.6	27/04	35mins
	1120	7.12 gpm	22.76	706	19138	(0	3.17	119.1	2769	50mins
	1130	71.79pm		7.06	19111	38	4.01	108.2	27.82	Comins
	1140	240 901		- Stop	Pumpin		Toda			
10/8	0720	J 10 961	Begin	. ,	Pump	Well		/	19.50	
107 0		1-7 gpn	22.50	7.14	20,398	39	4.80	475,2	23.34	Smins
	0735		1000	 ' ' ' 	19943	11	3.95	469.7		Ismins
		~1.2 gpm	1	7.08	19719	13	3.94	414.0	27.61	30 mins
	0805	~1-2 gpm	I	7.07	19407	5.3	3.7/	385.0	28.29	45 mins
		~1.29Pm		7.07	19177	2.le	3.29		28.57	60 mins
	1	- 11 April	Stop			well re	·			
	0820			n P		we 11 -			20.96	
	0840	× 1 2 au	22.78	7.05	18926	16.1	3.37	353.8	24.14	5 mins
	lt .	21.29pm	22.79	1	18768		3.80	339.6	T	15 mins
	0855	1. 19pm	22 81	7.05		8.1	2.79	280.7		
	0910		22.81	7.06e		5.32	3.01	244.9		45 mins
	0925	21.2 gpm				- ,	3.08	7		60 mins
	0 4 10	165 "	22.82	I Pu	rping W					
	0 770	165 gall.	<i> - -\\\\</i>	le a var	rying V	\ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \ \				
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Pg 1 0 + 2

	4mw3	<u> </u>	Site: Sea	1 Beach			Date: [0	<u> </u>	
Client:	· · · · · · · · · · · · · · · · · · ·			Project Numb	er:				
Well Casing	Diameter (i	nches):		Well Casing N	Material: P\	/C SS E	ther Pol	у	
Well Heads	pace:	PID (ppm):	0.0		FID (ppm): 1	N/A			
Samplers:	Chadn	narvin	with CDM					with Blaine	Tech
Total Depth	of Well (fee	t):	<u> 35</u>	_ 2" - 0.16					
Depth to Wa	ater (feet):		19.87	(X) 4" - 0.65	Gal/ft. =	(X) 3 =		
Water Colur	nn Height (f	eet):		_ 6" - 1.47			•	Minimu	m purge volun
Well Refere	nce Point: T	ос			·	· · · · · · · · · · · · · · · · · · ·			(gallon
PURGE ME	THOD:	Submersible	pump 🗌	Bladder pump	Dispos	sable bailer			
Pump Make	/Model: 2" G	rundfos Red	diflo	Depth of pum	p intake (feet):			
Purge equip	ment decon	taminated?	$_{Y} \square_{N} \square$	Container type					
Purge/deco	n water cont	ainerized?	$\nabla \square $ \square	Volume: ρ	eristalt,	ic Pum	ping = 1	1 gallor	5
	Start Time:				Flow Rate: _		<i></i>		
T :	Callana	Temp.	n L J	Conductivity	Turbidity	DO	ORP	DTW	Comments
Time	Gailons	(C/°F)	рН	(µmhos/cm)	(NTUs)	(mg/L)	(mV)	(ft TOC)	Comments
1445		Begio	n to	Pump-				19.87	
1450	* Igal	26.57	7-24	20284	>1000	6.69	131.7	cant	Pumping
1455	= 1.3gal	25.61	7.16	20781	388	<u>4.82</u>	50.7	measure	le .
1500	2	24.61	7.18	20502	421	5.01	29.2	measure went fit while	· · · · · · · · · · · · · · · · · · ·
1505	~	24.17	7.16	20501	509	3.18	21.0	Pumping	- 11
1510		2433	7.16		269	2.71	-3,9	NA	
1515		24.09	7.18	21320	17.3	3.74	3.8	"	
1520	~ 2.5gal	23.88	7.15	20622	196	3.95	-34.9	,, ,,	
1535		23.56	7. al	21539	51.8	<u>5.14</u>	55_1		
1545	≈ 5 gal	23.37	7.16	21576	80	4.16	50.3	· · · · · · · · · · · · · · · · · · ·	
1555		23,28	7,20	21755	38 Y	4.10	-4.2		
1600	~ legal	Stop		g for	Today			STUL	
0725		Begin	to P	unp we				319.98	
0730		21.03	7.25	21946	64	<u> 8.21</u>	25.3		Pumping
0740		20.49	7.24	22 435	139	8.46	-7.2		·····
				Chemets DO	(mg/L):				
			Me	ethod		Container T	ype/Volume	· · · · · · · · · · · · · · · · · · ·	Preservative
									,
Sa	mple Analys	es:							
Sample	Collection I	Method:							
	¥								
Pump: F	low Rate:		Sample ID:				Sample Tim	ie:	
		abla	Duplicate II):			Sample Tim		
Bailer: 1	ľype: dispos	abie	Dapiloato it						

Pg 2 of 2

Well No.: A	mw3	71	Site: Seal	Beach			Date: 107		
Time	Gallons	Temp.	рН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments
0750	× 2 00/	22.08	7.20	21891 22006 22179 22481 22501 umping	47	4.90	-7.2	,	Pumping
0800	≈ 2.5 est	22.04	7.24	22006	66	4.62	-7.3		
0810		21.33	7,22	22/79	18	3.55	-15.9		"
0820		21.09	7.21	22/8/	4.92	3.86e	-18.0		
0825	≈ 5 gal	21.16	7.22	22501	0.91	3.91	-19.2		
0825		- Film	shed P	umping	Zone #	1 Deve	loped		
					· · · · · · · · · · · · · · · · · · ·				
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			<u> </u>]	<u></u>		<u> </u>	

Pg Jof Well No.: AMW3 23 Beach Seal Site: Project Number: Client: Well Casing Material: PVC SS Well Casing Diameter (inches): PID (ppm): (), () Well Headspace: with Blaine Tech Marvin with CDM Samplers: Chad 2" - 0.16 Total Depth of Well (feet): O. O (X) 4" - 0.65 Gal/ft. = _____(X) 3 = _ Depth to Water (feet): Minimum purge volume Water Column Height (feet): 6" - 1.47 (gallons) Well Reference Point: Disposable bailer PURGE METHOD: Submersible pump Bladder pump ____ Depth of pump intake (feet): Pump Make/Model: 2" Grundfos Rediflo Purge equipment decontaminated? Y 🗓 N Container type: Purge/decon water containerized? Y 🗵 N Volume: ~ Flow Rate: Start Time: ORP DTW Turbidity DO Temp. Conductivity Comments Time Gallons рΗ (NTUs) (mg/L) (mV) (ft TOC) (°C/°F) (µmhos/cm) Chemets DO (mg/L): Container Type/Volume Preservative Analyzed? **EPA Method** Sample Analyses: Sample Collection Method: 🖌 Sample Time: Pump: Flow Rate: Sample ID: Sample Time: Bailer: Type: disposable Duplicate ID: Sample Time: Other: Desc.: Equip. blank ID: **CDM** MONITORING WELL PURGE AND SAMPLING FORM

	01
Well No.: Anw3 23 Site: Seal Beach Date: 10/10/07	
Client: Project Number:	
Well Casing Diameter (inches): Well Casing Material: PVC SS Other 101 /	
Well Headspace: PID (ppm): O · O	
Samplers: Chad Marvin with CDM with Bleine Tech	
Total Depth of Well (feet): 2" - 0.16	
Depth to Water (feet): (X) 4" - 0.65 Gal/ft. = (X) 3 = (X) 3 = (X) 3	
Water Column Height (feet): 6" - 1.47 Minimum purge	volume gallons)
Well Reference Point: "	
PURGE METHOD: Submersible pump Bladder pump Disposable bailer	
Pump Make/Model: 2" Grundfos Rediflo Depth of pump intake (feet):	
Purge equipment decontaminated? Y N U Container type:	
Purge/decon water containerized? Y N Volume:	
Start Time: Flow Rate:	
Time Gallons Temp. (°C/°F) pH Conductivity Turbidity DO ORP DTW (μmhos/cm) (NTUs) (mg/L) (mV) (ft TOC)	nents
1/0 Water sone Dry	
Chemets DO (mg/L):	
Analyzed ? EPA Method Container Type/Volume Preser	vative
Sample Analyses:	
Sample Analyses.	
Sample Collection Method: ✔	
Pump: Flow Rate: Sample ID: Sample Time:	
Bailer: Type: disposable Duplicate ID: Sample Time:	
Other: Desc.: Equip. blank ID: Sample Time:	

CDM

MONITORING WELL PURGE AND SAMPLING FORM

Well No.:	Anu3	2	Site: Sea/	Beach	·		Date: [0]	110/07	7
Client:				Project Numb	er:				-
Well Casing	g Diameter (i	nches): C/	nt well	Well Casing N	Material: P	vc ss	Other: Po	/y	
Well Heads	space:	PID (ppm):	0.0		FID (ppm):	N/A		, 	
Samplers:	Chadn	arvin	with CDM					with Blaine	Tech
Total Depth	of Well (fee	t):	35	-					
Depth to W	ater (feet):		17.38	<u>(</u> X) 4" - 0.65	Gal/ft. =	(>	() 3 =		
Water Colu	mn Height (f	eet):		_ 6" - 1.47				Minimu	m purge volume
Well Refere	ence Point: T	oc		¹¹		· · · · · · · · · · · · · · · · · · ·			(gallons)
PURGE ME	THOD:	Submersible	e pump	Bladder pump	Dispo	sable bailer			
Pump Make	e/Model: 2" C	Grundfos Re	diflo	Depth of pum	p intake (fee	t):			
Purge equi	pment decon	taminated?	YXNL	Container typ					
Purge/deco	n water cont	ainerized?	YXNL	Volume: ₽ e l	ristaltic	Purp=	= 4 ga	10n	**************************************
	Start Time:				Flow Rate: _				
Time	Gallons	Temp.	рН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments
0025			- to P	ump Z		4	(,		
0925		23.40	774	14406	>1000	12 21	65.5	17.38 BTOP.	Pumping
ii -	~ 4 gal	well	i	+ Dry		W. 8 1			1 Con prog
0931		00 6 11	1000	y					
							,		
	1		l	Chemets DO	(mg/L):		<u>.</u>		
			Me	ethod		Container T	ype/Volume		Preservative
Sa	ample Analys	ses: →							
				1					
									
Sample	e Collection I	Method:						··· · · · · · · · · · · · · · · · · ·	
	7								
Pump:	Flow Rate:		Sample ID:				Sample Tir	ne:	
	Type: dispos	able	Duplicate II	D:			Sample Tir		
	Desc.:		Equip. blan				Sample Tir		
CI	MC		N	MONITORING	G WELL PUI	RGE AND S	AMPLING	FORM	
							····		

Pg 10+2 Well No .: AMWY Z 1 Site: Seal Beach 10/10/07 Date: Client: Project Number: Well Casing Diameter (inches): Well Casing Material: PVC SS Other) Well Headspace: PID (ppm): 0.0 Samplers: Chad Marvin with CDM with Blaine Tech 35 Total Depth of Well (feet): 2" - 0.16 <u>15.70</u> (X) 4" - 0.65 Gal/ft. = _____(X) 3 = _ Depth to Water (feet): Minimum purge volume Water Column Height (feet): 6" - 1.47 (gallons) Well Reference Point: PURGE METHOD: Submersible pump Disposable ba Bladder pump Pump Make/Model: 2" Grundfos Rediflo Depth of pump intake (feet): Purge equipment decontaminated? Y 🗵 N 🛭 Container type: Volume: Peristaltic Pump = 23.5 gallons Purge/decon water containerized? Y 🔀 N Start Time: Flow Rate: Turbidity Temp. Conductivity DO ORP DTW Gallons Time pΗ Comments (C)F) (NTUs) (µmhos/cm) (mg/L) (mV) (ft TOC) 5 Tal 15.70 Begin to Pump 0955 1000 22.6/ 7.36 2.51 36.8 21304 >1000 1005 × 1 gal 22.35 7.31 2/054 484 1.77 7.1 ~ 1.5 gal 22.49 7. 3<u>2</u> 10/0 20967 780 2.08 -13.0 22 gal 22.64 20809 478 1020 3.90 -4.1 7.30 1030 22.48 21045 9.0 638 4,13 23.08 409 1045 7.26 21129 4.06 24.2 11 1100 23.86 7.20 2.79 20966 218 50.8 ≈7.5gal 23.68 1115 7.20 21234 113 2.70 55.10 11 ~ 9 gal 23.45 21086 1, 1130 7.21 201 2.72 60.0 210gal 23.46 7.20 2/108 1145 84 2.57 57,2 1155 - Stop Pumping well 1215 Resure Pumping 320 Purping 1225 212 gal 22.90 7.21 81.2 21128 =14 gal 23.03 1245 7.19 21116 148 3,26 78.0 Chemets DO (mg/L): Analyzed? **EPA Method** Container Type/Volume Preservative Sample Analyses: Sample Collection Method: <a>K Pump: Flow Rate: Sample ID: Sample Time: Bailer: Type: disposable Duplicate ID: Sample Time: Other: Desc.: Equip. blank ID: Sample Time:

CDM

MONITORING WELL PURGE AND SAMPLING FORM

Vell No.:	gmwy	21	Site: Sea	1 Beach				11010	<u> </u>
Time	Gallons	Temp.	рН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments
1300	£15.5 gcl	23 14	7.18	21207	167	2.98	46.2		Pumping
12/5	× 17 500	22.44	7.20	2/128	105	2.90	50.2		1,
1230	~ 19501	2287	721	2/1/0	44	3.04	46.8		11
17 45	~ 10.7gal	72 05	720	21142	32.1	3.20	5.5,5	_	11
1 2 7 2	~ 17.5gal ~ 18.5gal ~ 20 gal	22 01V	772	21170	15.6	3.62	55.0		11
900	-025	22.07	774	21142 21179 21138	8.1	3.78		_	81
1715	rdd gal	177/	7 14	2/127	2.13				1/
1425	.325	2775	7.27	2/137	2.81	3.69	38.7		11
1750	~ 0.75 /	201)	51.0	Pumpin	71	F. //.	Doing lac	0-1-	
1430	~ 23.5gal		STOP	Pumpin	921	1 4/1 4	Deverop	<u> </u>	
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Pg 10+2

	Well No.: 🗡	AMW4	72	Site: Sea	al Beach			Date: 10	110/07	
	Client:				Project Numb	er:				
	Well Casing	Diameter (i	nches):		Well Casing N	//aterial: P\	vc ss o	ther Pol	<u> </u>	
	Well Heads	pace:	PID (ppm):	0.0		FID (ppm): !	N/A			
i	Samplers:	ChadA	narvin	with CDM	-				with Blaine	Tuch:
	Total Depth	of Well (feet	t):		_ 2" - 0.16					
	Depth to W	ater (feet):		19.29	<u>(</u> X) 4" - 0.65	Gal/ft. =	(X			
Y		mn Height (fe			_ 6" - 1.47				Minimu	m purge volume
	Well Refere	ence Point: T	ОС		<u> </u>					(gallons)
	PURGE ME	THOD:	Submersible	e pump 📙	Bladder pump	Dispo:	sable bailer			
		e/Model: 2" G			Depth of pum	p intake (feet):			
		oment decon		YMNU	Container typ		<u>0</u>	ar I was	-	
1	Purge/deco	n water cont	ainerized?	Y FAN L	Volume: Per	istalti	c rung	0 - 17.3	gallo	۸۲
		Start Time:				Flow Rate: _				
	Time	Gallons	Temp. (ⓒ/°F)	рН	Conductivity	Turbidity	DO (marila)	ORP	DTW	Comments
				•	(µmhos/cm)	(NTUs)	(mg/L)	(mV)	(ft TOC)	
	1440		Begin 23.27		20648	72-	4.02	79.4	19.29 870	i
	1445	.201		7.17	2/025	> 1000 331	4 00	70.1		Pumping
		:2.2 gal	23.22	7.16	21008	140	3.75	87.9		//
	1515	^ F /	23.16	7.14	21040	149	3.68	98.2		1/
	1530 1545	25 gal	23.09	7.20	20476	257	3.94	106.5		//
	1600		2288	,	21062	108	3.96	112.7		11
	1605	~7.5gd		Stop	Pumpin		Today			
		1.090	*		umping	2	,,,,			
10/11	0720		21.31	7.23	20294	172	4.16	155,0		Pumpina
	0730		21.88	7.19	20702	57.2	3.92	1279	_	,,,,,,
		~ 2.5gal	_	7.17	20826	47.3	3.57	101.0	_	/2
	0800	= 4gal	22.17	7.18	20944	31.2	3.05	97.9		"/
	0815	≈ 5.5 gel		7.18	19890	17.4	316	98.2) /
	0830		22,33	7-19		10-1	3.29	97.0		"
					Chemets DO	(mg/L):				
		- , 		Me	ethod		Container T	vpe/Volume		Preservative
								7.1		
	Sa	imple Analys	es:>							
									······································	
	Sample	e Collection l	Method:							
		`\								
	Pump:	Flow Rate:		Sample ID:				Sample Tir	ne:	
	Bailer:	Type: dispos	able	Duplicate II	D:			Sample Tin	ne:	
	Other:	Desc.:		Equip. blan	ık ID:	· · · · · · · · · · · · · · · · · · ·		Sample Tin	ne:	
	CI	MC		Ŋ	MONITORIN	G WELL PUR	RGE AND S	AMPLING	FORM	

Pg 2 of 2

VEILING.	MNYZ	72	Site: Se	al Beach)		Date: 10	/ 	
Time	Gallons	Temp.	pН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments
0845	~ 900l	22.55	7.17	20964	5.02	3.19	107.2		Punping
0900	= 10gal	22,47	7.17	20896	1.24	3.26	105.9	:	1,
0900 -		· sto	P Pun	op Zun	e 2 De	wlope	4		
Total =	17.5ga	llons		,					
	,					· · · · · · · · · · · · · · · · · · ·			
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Pg 10f1

Well No.: /	anny	2 3]	Site: Sea	1 Beach			Date: 10	/11/0	Z
Client:				Project Number		. <u>.</u>			
Well Casing	Diameter (ir	nches):		Well Casing Ma	terial: P\	vc ss	Other Pa	<i>') y</i>	
Well Heads	pace:	PID (ppm):	0.0						· · · · · · · · · · · · · · · · · · ·
Samplers:	Chad M		with CDM					with Blaim	: Tec h
Total Depth	of Well (feet	•	<u> 35</u>						
Depth to Wa	ater (feet):		19.27	(X) 4" - 0.65 G	Gal/ft. =	((X) 3 =		
Water Colu	mn Height (fe	eet):		6" - 1.47				Minimu	m purge volume (gallons)
Well Refere	nce Point:			# <u></u>					(gallorio)
PURGE ME	PURGE METHOD: Submersible pump Bladder pump Disposable ba								
Pump Make/Model: 2" Grundfos Rediflo Depth of pump intake (feet):									
Purge equip	ment decon	taminated?	YXNU	Container type:					
Purge/deco	n water cont	ainerized?	A MN M	Volume: Per	istaltic	Pun	P= 9 g	a/1003	
	Start Time:				Flow Rate:			_	
Time	Gallons	Temp.	рН	Conductivity	Turbidity	DO	ORP	DTW	Comments
	Gallons	(CG/°F)		(µmhos/cm)	(NTUs)	(mg/L)	(mV)	(ft TOC)	
0915		Begin	to Pu			r ra	200 7	19.27	A .
0920		23.14	7.58	22086	> 1000	5.50			Pumping "
0930	~ lgal	23.41	7.44	21991	302	4.88	The state of the s		
0945		23-46	7.49	22253	46	4.96	37. D		· /
1000	22.5g al	23.58	7.45	22248	27.1	5,03	1		//
1015		23./4	7.38	22050	47.6	4.36	36.7		<i>)</i> ;
1030		23,29	7.39	2/890	32.8	4-17	2)-/		71
1045	25gal		7.38	2/8/0	19.3	7.24			î,
1100		22.89	7.35	2/169/	13.4	3.18	22.0		11
1115	A. 3m F	22.71	7.36	20908	22.9	3.76			ř'
1130	27.5gd		7.33	2/2002	6.89	Į			17
1145	201	23,10	7.34	21300	5.01	3.34 2.29	26.0		i,
1200	~ 9 gal	23.15	7.33	21096 ing Zon	1-/0	3.37	124/04	0 1 -	
1200		5 top	pump	ing zon	<u>e 5 F</u>	u II y	verery	<i>t</i> 4	
	<u> </u>	<u> </u>		Chemets DO (r	na/!):	L	l	1	· · · · · · · · · · · · · · · · · · ·
				Ouemers DO (i			- 		
			Analyzed?	EPA Method	(Container T	Type/Volume)	Preservative
Sa	ample Analys	ses:							
	llection Meth	nod: ⊭		<u> </u>			0		
	Flow Rate:		Sample ID:				Sample Tin		
	Type: dispos	able	Duplicate ID:				Sample Tin		
	Desc.:		Equip. blank				<u> </u>		
C			MO	NITORING W	ELL PURG	E AND SA	AMPLING I	OKM	

Pg 10f 1

Well No.: A	MWY Z	7 4	Site: Sea/	Beach	- <u>-</u>		Date: /	0/11/0	7
Client:				Project Numbe	r:				
Well Casing	j Diameter (in	nches):		Well Casing Ma	aterial: P	vc ss	Other	Poly	
Well Heads	pace: I	PID (ppm):	0.0						
Samplers:	chad M.	arvin	with CDM					-with Blain	o Teel n
Total Depth	of Well (feet)):	35	2" - 0.16					
Depth to W	ater (feet):		18.10	(X) 4" - 0.65 (Gal/ft. =		(X) 3 =		
Water Colu	mn Height (fe	et):		6" - 1.47				Minimu	m purge volume
Well Refere	nce Point:								(gallons)
PURGE ME	THOD:	Submersibl	e pump	Bladder pump	Disp	osable ba			
Pump Make	e/Model: 2" G	rundfos Re	diflo	Depth of pump	intake (fee	t):			
Purge equip	ment decont	aminated?	YXND	Container type:					
Purge/deco	n water conta	ainerized?	YXN	Volume: Per	istaltic	Pun	0 = G	OZ	
	Start Time: _				Flow Rate	,			
Time	Gallons	Temp. (°C/°F)	рН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments
1215		Begin	to Pu	mp Zon	e 4 -			- STWL,	
1218	boz	Well	Went	mp Zon			<u> </u>		
		•							
									<u> </u>
									· • • • • • • • • • • • • • • • • • • •
				Chemets DO (r	ng/L):		_		
			Analyzed?	EPA Method	(Container [*]	Type/Volu	me	Preservative
Sa	mple Analyse	es: →				· -			
	,,								
						· · · · · · · · · · · · · · · · · · ·			
	· · · · · · · · · · · · · · · · · · ·								
Sample Col	lection Metho	od: ⊭∕					1		
Pump: F	Flow Rate:		Sample ID:				Sample 1	ime:	
Bailer: 7	Type: disposa	ble	Duplicate ID:				Sample 7		
Other: [Desc.:		Equip. blank	ID:			Sample 7	ime:	
CI	MC		МО	NITORING WI	ELL PURG	E AND SA	AMPLING	FORM	

Pg lof 2

Well No.: Am	WS Z	1	Site: 5ea	1 Beach			Date: 10,	111107	
Client:				Project Numbe	r:				
Well Casing Diar	neter (inches)	:		Well Casing Ma	aterial: PV	SS Oth	ner:) Pol	, y	
Well Headspace:	PID (p	pm):	0.0						
Samplers: Ch.	ad Marv	10	with CDM				•	with Blaine	•
Total Depth of W	ell (feet):		35	2" - 0.16					
Depth to Water (f	feet):		18,25	(X) 4" - 0.65 (Gal/ft. =	(X) :			
Water Column H	eight (feet):			6" - 1.47				Minimu	
Well Reference F	Point:			" -					
PURGE METHO	D: Subm	ersible	e pump 🔲	Bladder pump	Dispo	sable bailer			
Pump Make/Mod	lel: 2" Grundfo	s Re	diflo	Depth of pump					
Purge equipment				Container type:			F 2		
Purge/decon wat	er containeriz	ed?	YLNL	Volume: 101		rung	= 53 g	allons	
Start	Time:			· -	Flow Rate: _		·		
Time Ga	llons Ten	np.	рН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	
1075			1 0		one 1	(Hig/L)	(1114)	- 18.25 - 18.25	,
1235 -	I -	e 911	7.05	13278	21000	7.32	-63.3		Purp
1250	24.0		7.18	11748	989	3.21	-51.9		, ,
	.5gd 22		7.16	11025	>1000	4.06	-53.6		
1315			7.23	10158	731	2.90	-50.		
11									
ll	0,5gy 23.		7.26	9857	363	3.08	-44.3		
	9.5,123		7.26	9597	305	3-30	-43.4		
	0.8, 12.		7.37	10186	192	3.80	-40.9		
II ' ' _ I	12.5, 22.		7.25	9487	127	3.46	- 4a7		
	15gal 22.		7.25	967/	70.5	3.40	-40.6		
	7gal 22.		7.28	8603	55.0	3,75	-42.1		
	8.5ga 22.	~ ^	7.28	9136	36,2	3.72	-42.1		
1530 ≈2	0 gal 22-	53	7.29	9382	193	3-81	- 37.6		
1545 ^	22.	19	7.28	9100	172	3.99	-42.2		
				Chemets DO (r	mg/L):				
			Analyzed?	EPA Method		Container T	ype/Volume		
			<u> </u>						
0	A b								
Sample	Analyses:	→							
Sample Collection	n Method: 🖌						r		
Pump: Flow	Rate:		Sample ID:				Sample Tim	ne:	
Bailer: Type:	disposable		Duplicate ID:				Sample Tim		
Other: Desc.	•		Equip. blank	ID:			Sample Tim	ie:	
CDM			N	MONITORING	WELL PURC	GE AND SA	MPLING FO	ORM	

Pg 2 of 2

	Well No.: A	nw5	21	Site: Sea	1 Beach			Date: /	0/11/07	,
	Time	Gallons	Temp.	pН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments
	1600		22.54	7.30	9013	84.7	4.58	- 79.3		
	1610	24gal		Stop	pumpin	g for	Todo	<i>y</i> —		
10/17	0725		Begin	to P	imp Z	one 1	2 51	12.2		
·	0730		21.81	7.19	10149	369	6.51	-43.3 -58.4		Pumping
	0745	~ 2.5gal	21.91	7.27	9739	109	5.52			1/
1	0815	75 gal	21.87	7.28	9/54	134	5.12	-42.8		"
	0830	≈ 9gal		7.26	9074	98	4.01	-42.7		1/
	0845	~ 10gal	22.07	7.29	90/2	226	4.36.	- 738		<i>i</i>)
	0900	~11.8gal	22.21	7.29	8958	172	4.61	-41.3		11
	09/5	~13.5 yal	22.23	7.31	8734	29.2	4.11	-356		· ·
		= 15.5 gal		7.29		18.9	4.19	-34.1 -36.0		11
		≈17.5gal		7.28		49.9	5.72	- 313		è
	1000	≈20gal ≈27gal	22.24	7.30		20.8	4.60	-332	,	- 11
	1030	27 ygal	22.43	7.29	8864	16.7	4.70	-33.3		!/
	1045	25 gal	2230	7.31	8794	8.21	5.26	- 322		"
	1100	,	22.92	7.30		<u>4.79</u>	,	-32.7		<i>P</i> (
	11/5	~29gal		7.30		1.3/	4.87	- 3/.5		
-	1//5			top pu	meing	Zone 1	Perelo	pea -		
	· · ·	<u></u>								
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					· · · · · · · · · · · · · · · · · ·					
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										<u> </u>

lg 1 of 2

Well No.: AMW 5 72 Site: Seal Beach Date: 10/12/07									•
Client:				Project Number	:				
Well Casing	Diameter (i	nches):		Well Casing Ma	terial: PVC	ss Oth	her) Pol	<u>y</u>	
Well Heads	pace:	PID (ppm):	0.0						
Samplers: Chad Markin with CDM with Blaine									•
Total Depth of Well (feet): 2" - 0.16									
Depth to Water (feet):(X) 4" - 0.65 Gal/ft. =(X) 3 =									
Water Column Height (feet): 6" - 1.47 Minimu									ı
Well Refere	nce Point:			_"-					
PURGE METHOD: Submersible pump Bladder pump Disposable bailer									
Pump Make/Model: 2" Grundfos Rediflo Depth of pump intake (feet):									
Purge equip	oment decon	taminated?	$Y \boxtimes N \square$	Container type:					
Purge/deco	n water cont	ainerized?	YZN	Volume: (୧୯)	staltic	Pump	<u>= 39</u>	gallons	
	Start Time:				Flow Rate: _				
Time	Gallons	Temp.	рН	Conductivity	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	
		(℃ /°F)	1 0	(μmhos/cm)	2 —	(mg/L)	(1117)	5746	•
1130			to fun	1 7 .		145	-18	18.09	pumping
1135		23.53	7.39	18900	>1000	6.45	-1.8		
1145		2369	7.17	17571	890 980	5.42	40.6 56.3		*
1200		23.70		17430		5.14	43.0		ir
12/5		23.45	6.99	17063	9/2	5.34			11
	≈5 ga l		7.01	17246	899		58.7		11
1245	. ~	23.57	7.00	16827	-880 -799	4.33	60.0		
1300	27. Sgal		6.97	16757		7.72	58./ 59.0		11
1315	~ 9.5gd		6.99	16585	8 45	7,79			
1330	A 11 F	23.51	6.98	16395	702	5.28	61.0		4
1345	2/1.5ga	23.72		16344	771	5.69	620		4
1400	~ 1.1. <i>T</i>	23.41	6.96	16278	590		43.9		Zi
1 415	214-5ga		6.95	16223		5,41	67.8	ļ,	iv
1430	217.5	23.29	6.96	16431	647	5.50	74.7 72.3		1
1445	· 17.3	70.11	0.77			7.07	7 2 3		
				Chemets DO (n					
			Analyzed?	EPA Method		Container T	ype/Volume		
Sa	mple Analys	es:							
Sample Co	llection Meth	ıod: ≰				·			
	Flow Rate:		Sample ID:	<u> </u>			Sample Tim		í
	riow Hate: Type: dispos	ahla	Duplicate ID:				Sample Tim		
	Type: dispos Desc.:	ผมเธ	Equip. blank		·		Sample Tim		
			· · · · · · · · · · · · · · · · · · ·	ONITORING	WELL DID	TE AND SA			
_ G			N	TOMITORING	WELL FUKC	AL AIND 3A	MI PHACL	~ 1.11.11 ~ 1.11.11	ı

Pg 2 of 2

We	ell No.: /	Inw5	77	Site: Sea	1 Beach			Date: 0/	112/07	
	Time	Gallons	Temp. ((Ĉ) F)	pН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments
	1500	≈ 19 gal	23.26	6.94	16032	372	4.80	68.2		Pumping
1	515	- 	22.79	6.96	16090	499	5.87	70.1		22
11		222 gal	22.70	6.97	16157	368	6.02	71.2		11
11-	1545		22,54	6.92	15951	339	7.8/	73.7		
1	600		22.41	6.94	12881	286	7.06	70.1		11
	605	≈ 24gal	<u> </u>	top Pu	impiny Z	2 for	tiday	,	*****	
	750		Begin		ing Z	2-			57WL, 18.38	
0	755		21.31	7.04	16291	18/		129.5		Pumping
- (1	18/5		21.09	7.01	16/35	122	6.07	126.2		
13 -	830	24 gal	21.46	6,96	16264	53.8	5.60	157.1	,	
- 11	845		2180	6.96	16025	30.5	609	94.8		
15	- •	≈ 7.5gal	21.88	6.96	15818	16.7	4.18	87.1	,	′,
li I		= 10 gal	21.52	6.97	15789	30.0	6.01	81.3		,
317	930		21.80	6.93	15570		5.80	77.8	-	- ' '
C	7945	2 /3ga/	2189	6.92	15618	ľ	7.49	77.9		//
	000	~ 15gal	21.88	6.93	15599	3,8/	4.98	77.7	-	()
	000		Stop	pump	ed Zon	e 2 D	evelobe	d		
										
				····						
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Pg 10+1 Well No.: AMW5 Z3 Site: Seal Beach Client: Project Number: **Other** Well Casing Material: PVC SS Well Casing Diameter (inches): Well Headspace: PID (ppm): 0.0 Samplers: chad Marvin with CDM -with Blaine 35 2" - 0.16 Total Depth of Well (feet): 19.73 (X) 4" - 0.65 Gal/ft. = _____ (X) 3 = _ Depth to Water (feet): Minimu Minimu 6" - 1.47 Water Column Height (feet): n _ Well Reference Point: PURGE METHOD: Submersible pump Bladder pump Disposable bailer L Depth of pump intake (feet): Pump Make/Model: 2" Grundfos Rediflo Purge equipment decontaminated? Y L N Container type: Volume: Peristaltic Pump= 4 pallon Purge/decon water containerized? Y Jи Flow Rate: Start Time: Turbidity DO ORP DTW Conductivity Temp. рΗ Time Gallons (NTUs) (°C/°F) (µmhos/cm) (mg/L) (mV) (ft TOC) Begin to Pump Zone 5 1010 1015 7.65 15430 8.26 763 20.64 >1000 Well 1020 Went Dry-Chemets DO (mg/L):_ Analyzed? **EPA Method** Container Type/Volume Sample Analyses: -> Sample Collection Method: 🖌 Sample Time: Pump: Flow Rate: Sample ID: Bailer: Type: disposable Duplicate ID: Sample Time: Other: Desc.: Equip. blank ID: Sample Time: CDM MONITORING WELL PURGE AND SAMPLING FORM

Site: Seal Beach Date: Well No.: AMN5 24 Project Number: (Other) Well Casing Material: PVC SS Well Casing Diameter (inches): PID (ppm): 0.0 Samplers: chad Marvin with CDM -with-Blaine 35 2" - 0.16 Total Depth of Well (feet): \sqrt{g} (X) 4" - 0.65 Gal/ft. = ____ (X) 3 = _ Depth to Water (feet): Minimu 6" - 1.47 Water Column Height (feet): Well Reference Point: Submersible pump Bladder pump Disposable bailer ___ PURGE METHOD: Depth of pump intake (feet): Pump Make/Model: 2" Grundfos Rediflo Purge equipment decontaminated? Y 🖾 N Container type: Pump = 3 gallon Volume: Peristaltic Purge/decon water containerized? Y 📈 N Flow Rate: Start Time: _ DO ORP DTW Turbidity Conductivity Temp. рΗ Gallons (NTUs) (mg/L) (mV) (ft TOC) **(*C)**°F) (µmhos/cm) 5 TWL 18-18 Begin Zone to pun 9.02 834 7.80 14007 21.48 >1000 Punping Well Went Chemets DO (mg/L):_ Container Type/Volume **EPA Method** Analyzed? Sample Analyses: Sample Collection Method: 🖌 Sample Time: Flow Rate: Sample ID: Sample Time: Type: disposable Duplicate ID: Sample Time: Equip. blank ID:

Desc.:

Pump:

Bailer:

Other: L

Client:

Well Headspace:

Time

1040

1045

1050

MONITORING WELL PURGE AND SAMPLING FORM

							19	1043		
Well No.: ,	Annle		Site: Sen	1 Beach Date: 9/21/07						
Client:				Project Number:						
Well Casin	g Diameter ((inches):	y "	Well Casing M	aterial:	vc) ss	Other:			
Well Heads	space:	PID (ppm):	0.0							
Samplers:	chad 1	Marvin	with CDM					-with Blair	re Tech	
Total Depth	of Well (fee	et):		2" - 0.16						
Depth to W	ater (feet):		18.86	(X) 4" - 0.65	Gal/ft. = <u></u>	2.49	(X) 3 =		-	
Water Colu	mn Height (feet):	16.14	6" - 1.47	•			Minimu	ım purge volum	
Well Refere				" " " " " " " " " " " " " " " " " " "					(gallons	
PURGE ME	ETHOD:	Submersib	le pump 🏻	Bladder pump	Disp	osable ba	<u> </u>			
		Grundfos Re		Depth of pump		•				
Purge equi	oment decor	ntaminated?	YMNU	Container type	: 55 9	allon o	drum	•		
Purge/deco	n water con	tainerized?	Y X N 🗌	Volume: Bajl	ed = 7	4 ga/10	ins Pu	mped=	260 gallons	
	Start Time:	0730)		Flow Rate					
Time	Gallons	Temp.	рН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments	
0755	4	21.33	7.36	6571	> 1000	4.14	179.2)	Swab # Bail	
0800	9	21.80	7.37	7346	> 1000	4.15	166.6		"	
0807	17	2156	7.11	9564	>1000	4.49	161.5		4	
0815	29	21.59	7.24	10544	> 1000	4.76	160.2		ti	
	Be	gin to	Punp	well				16.32		
		J								
			,							
				Chemets DO (n	ng/L):		•		i	
			Analyzed?	EPA Method	C	ontainer T	ype/Volun	ne	Preservative	
Sa	mple Analys	.ac								
Odi	iipic / tilaiyo									
			:							
Sample Col	ection Meth	od: 📈				·				
Pump: F	low Rate:		Sample ID:				Sample Ti	me:		
Bailer: 🔀 T	ype: disposi	ss steel	Duplicate ID:				Sample Ti	me:		
Other: D	esc.:		Equip. blank l	D:	Sample Time:					
CE	M		MO	NITORING WE	ELL PURGE	AND SA	MPLING	FORM		

fg 2 of 3

Well No.:	Vell No.: AMW & Site: Seal Beach Date: 10/8/07									
Client:				Project Numb						
Well Casing	g Diameter (i	inches):	4"	Well Casing N	Material:	vc)ss c	Other:			
Well Heads	pace:	PID (ppm):	0.0		FID (ppm):	N/A				
Samplers:	chadn	arvin	with CDM				· ·	with Blaine	Jech -	
Total Depth	otal Depth of Well (feet): 2" - 0.16									
Depth to W	Pepth to Water (feet):(X) 4" - 0.65 Gal/ft. =(X) 3 =									
Water Colu	Vater Column Height (feet): 6" - 1.47 Minimum purge volume									
Well Refere	ence Point: T	OC		" _	·				(gallons)	
PURGE ME	URGE METHOD: Submersible pump Bladder pump Disposable bailer									
Pump Make	e/Model: 2" C	<u>∃rundfos Re</u>	diflo	Depth of pum	ıp intake (feet	t):				
Purge equip	pment decon	ntaminated?	YMNL	Container typ	e:					
Purge/deco	n water cont	tainerized?	YXN	Volume:	<u></u>	·		-		
	Start Time:			•	Flow Rate: _					
Time	Gallons	Temp.	рН	Conductivity	Turbidity	DO (ma/l)	ORP	DTW	Comments	
1005		\		(µmhos/cm)	(NTUs)	(mg/L)	(mV)	(ft TOC)		
1040	15	23,20	7.18	11327	* Bail		274.5		surge &	
1050	22	23.28	7.37	12096	>1000	5.64	248.2		SUIGE \$	
1050	40	well			71000	U+1>	-11.4	7	Ball	
(105	27	22,42		ľ / l	1000	1.16	^ ET 2		surge \$	
- ,	30 -		7-18	12065	>/000	6.60	257.3		Bail	
1108	34	2227	Went	12135	1000	/ 02	2.11.2		- Surge a	
1130	45	2237	7./9	1	>/000	6.92	241.2		surge &	
1130	7.5	22.03 F:	7.15		908	9.02	249-1		Bail	
1200		_	ished		urge \$	Bail		STWL	<i>v o</i> ;	
1205	5	Begin 23.72	b.85	12784	562	4.17	159.2	20.27	Pumping	
1215		23.54		12872	280	4.51		23.98		
1270	2 Lypn			12777	204	7.28	89.5	28. 13 33. 12	15 mins	
1235	~19pm	23.11	12011	Went	1 1	/, 0	107-1	32,10	30 mins	
1300	≈ 35gcl	Resu	200	,	pry			STIL		
1,000		1/C3/M		Chemets DO	(ma/l):			20.68		
			1	₁	ſ					
			Me	ethod		Container Ty	/pe/Volume		Preservative	
Sai	mple Analys	es: →								
			ļ							
	 		<u> </u>							
• .	- ·· ·· •	· ·							-	
Sample	Collection N	/lethod:								
-										
	Flow Rate:		Sample ID:		-		Sample Tim	e:		
	Type: disposa		Duplicate ID				Sample Tim	e:		
Other: D	Desc.:		Equip. blank	k ID:			Sample Tim	e:		
CE	MC		N	MONITORING	WELL PUR	GE AND SA	AMPLING F	ORM		

Ī	Well No.:	Amul	0	Site: 500	al Beach			Date: 10	ORP (mV) (ft TOC) Common (mV) (ft TOC) Common (mV) (ft TOC) Common (mV) (ft TOC) Common (mV) (ft TOC) Common (mV) (ft TOC) Common (mV) (ft TOC) Common (mV) (ft TOC) Common (mV) (ft TOC) (ft TO	
1				T	Conductivity	Turbidity	DO			
	Time	Gallons	Temp.	pН	(µmhos/cm)	(NTUs)	(mg/L)		,	
10/6	1305	≈ 1gpm	23.50		13083	169	5.27	, , , , , , , , , , , , , , , , , , ,		
10/ 5		~ 1ggs	23.71	6.84	12999	176	4.83	90.7		15 mins
	1330	~ 19Pm	23.88	6.85	12705	54	6.40			30 mins
	1345	2/gpm	23,83	6.87	12634	27_	6.92	99.1	33.60	45min
	1355		- Well	Wes	+ Pc	y —			stut	
	1420		Kes	, , , , , , , , , , , , , , , , , , , ,	Pung	T 7.	/	116.0	20.44	
	1425	xlypn		6.82	12518	3/	3.76			1
	1440	nigen	23.79	6.84	12601	10	3.82			20 mins
	1500	~ lgfn	23.87	6.85	12 444	38	5.26	97.0	33.39	40 mins
	1500		we		ent d	(X -				
	1515		Resu	ne Pu		re1/-	4 6 0			
	1520	2 lapm	23.57	6-81	12981	35	6.89			5 mini
	1530	~ Igpn	33.48	6.83	12418	28	6.93			15 mins
	1545	~ 1gpm	23.90	6.86	12425	11	8-59			30 mins
ĺ	1600	× 1gpm	23.98	6.83	12349	8	9.04	111.3	33.70	45 mins
	1605	125gal	Stop		Ping Fo		hy —			
10/9	0735	<u> </u>	Begin			e11 =	11 21	12/0)
	0740	~1gfm	23.02	6.82	12.870	82	4.31		V 10 Z Z	5 mins
	0750	~19pm	23.37	6.83	13140	349	5.77			15 mins
	0805	=19pm	23.38	6.85	12887	73	6.47			30 mins
	0840	~ lgpm	23.56	6.85	12744		8.00			65mins
	0855	~ lgrm	2350	6.84	12390	6.8	8.73	110.2	33.70	80 mins
	0857		Well - Pas	Wen		1000	11		SIWL	
	0920	· · ·		une	Pungi.	10	3.06	110 ;		<i>F</i>
	0925	2/gpm	23.42	6.81 6.83	12823		4,83		1 1	- 16
	0940	2/gpm	2354	6.85	12365	6.8	6.82			
	1000	~ 1 ypm	23.57	6.87	, 12340	5.2	7.60		771.8	
ı	193050	21gpm	0.51	TO PUMP	12340 un ping	we/1.	7.00	113,0	37.00	10 CVIV 0 01
ľ		21.00	23.49	6.81	12893	8.1	8.73	113.6		Soning
ŀ	1055	2/gpm	23.57	6.80	12909	14.9	8.70			
ŀ		2/gpm	23.76			4.2	9.58			
ŀ	1135	~ 1gpm ~1.gpm		6.82	12306		9.51			- 11
1	1135	135	Einid	ad Pu	mping	Well	Develo	7		
	1132	133	1-17127	<i>y</i> 0 <i>y</i> 00	INPING	70 (17	7000	<i>/</i> · · · · · · · · · · · · · · · · · · ·		
1	<u>·</u>									
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i.										

4" PVC PID= aopen Pg. 1 of 1

Well No.:	mw1		Site: Se	al Beach			Date: 10,	15/07	7
Time	Gallons	Temp.	pН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments
1400		Begin	to s	urge	and B	9/1-		19.82	
1415	4gal	21.27	6.97	7469	>1000	11.47	301.6		surge \$ Bail
1420	17901	21.59	6.92	11729	71000	14,28	270.8		(1
1425	26gal	21.35	6.86	11656	> 1000	15.51	244.0		
1430	30gal			ed W/s	urge \$	Bail-		stul	
1445			ginto	Pump	well:	/ ~-		20.30'	Purping
1450	≈ 1gpm	21.95	6.71	11967	>1000	6.77	66.7	22.26	5 mins
1455	algpm	22.08	6.73	11999	> 1000	7.8/	82.8	23.04	10 mins
1500	~ lagar	22.43	6.77	12034	>1000	6.97	84.5 95.8	23.85	15 mins
15/5	~19pm	22,44	6.75	12225	582	6.38		27. 37	75 mins
1530	x/gpm	22.51	6.74	12/04	491	6.08	103.8	24.76	60 mins
1545	~)gpm	22.35	6.73	12/06	78	6.30	107.26		75 mins
1600	~ 1gpm	22.42	6.74				10,00	5,0 2	7 27477
1600	85 gat			top P	umping				
	05	· Page	Remo	iced				7	
	85g) Kerry	<u>V. C. O.</u>	Suba	ecsible	Pump	/	
					[-0.61]				
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Y"PVC PID=0.0ppm fg lof)

	<u> </u>			110-0.0	PIN			1.5	7
Well No.:	PMW9		Site: Sec	.1 Beach				115/07	<u>/</u>
Time	Gallons	Temp.	рH	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments
1130		Begin	to 5	urge a	nd Ba	il w	e/1-	20.04	
1155	24	21.72	7.04		> 1000	l.	160.2		suige &
1201	212	21-94	7.10	7058	>1000	12.81	141.0		Ü
1204	220	21.60		6902	> 1000	14.88	135.9		ir
	* * *			1 Surge		ail -			
1205				unp W	· · · · · · · · · · · · · · · · · · ·			21.20	Punping
1230	5 1	Begir	6.85	7,7.29	>1000	12.40	89.6	24.37	5 mins
1235	5 gal	2200	6.84	8049	>1000	11.55	77.7	28.27	10 mins
1240		22.99	6.87	9/49	> 1000	11.78	60.3	29.46	15 mins
1245	2/gpm	23.50	6.86	9485	898	11.90	69.2	29.99	30 mins
1300	19pm	23.31	6.86	9433	383	12.39	73.2	30.88	45 mins
1315	2/gpm		6.86	9471	106	12.43	75.8	3123	
1330	x /gpm	23.26			,,,,,				7.17.17
1330	45galler	<u> </u>	HOP TU	mping-					
	15 110	3 Rem	2100 1						
 	7>gallo1	JAN CON	by ta	Toub.	ersible	Punt		<u> </u>	
				1300	ersibix	1 ~ ~ ~	[/		
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Appendix E.5 AEW-1 Development

1 of

Well No.:	4EW1		Site: Sea	1 Beach			Date: 9/	120/0	7	
Client:			-	Project Number	er:					
Well Casin	g Diameter (inches):	4"	Well Casing M	aterial: (F	evc) ss	Other:			
Well Heads	space:	PID (ppm)	0.0							
Samplers:	Chad M	arvin	with CDM	Marie 1				with Blair	e Tech	
Total Depth	of Well (fee	et):	35.0	2" - 0.16					74. 7.5	
Depth to W	ater (feet):		17.27	(X) 4" - 0.65	Gal/ft. = _	1.52	(X) 3 =			
Water Colu	mn Height (feet):	17.73	6" - 1.47				Minimu	ım purge volun	
Well Refere	ence Point:			<u> </u>					(gallon	
PURGE M	ETHOD:	Submersib	le pump 🗶	Bladder pump	Dis	posable b	a			
Pump Make	e/Model: 2" (Grundfos Re	ediflo	Depth of pump	intake (fee	et):	· · · · · · · · · · · · · · · · · · ·			
Purge equi	oment decor	ntaminated?	YXNU	Container type	: 55 9	allon D	rum			
Purge/deco	n water con	tainerized?	YXN	Volume:			jk	Fire		
	Start Time:	0950	0		Flow Rate	:		_		
Time	Gallons	Temp.	рН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments	
1011	≈3	22.54	6.54	4739	>1000	3.99	193.6		swab & Bail	
1019	~ 8	22.49	6.50	5076	> 1000	3.52	162.8		//	
1027	~ 20	22.67	6.53	5414	>1000	4.22	163.5			
1036	≈25	22.59	6.60	5696	> 1000	4.42	162.5		"	
1045	~1.5 gpm	Begin 23.28	to Punj 6-38	5641	545	2.13	114.2	17.31 20.48	Pumping	
1055	≈20	23.41	6.40	5914	221	0.89	86.0	21.96		
1100	£30	2336	6-40	6001	890	46.0	21.95	21		
1105	21.79pm	23.36	6-39	6009	695	0.70		21.96	11	
1110	240	23.41	6.40	6117	122	0.66	27.2	21.97	/1	
1115	21.6gpm	23.43	6.40	6129	183		27.0	21.97	"	
1120	≈60	23.38	6.39	6123	56.7	0.64	48-3	21.94	"	
1125	~ 1.7gpm	23.39	6.40	6160	45.7	0.64	60.4	2211	- "	
1130	70	23.40	6.40	6156	520	0.66	63.8	22-10	"	
				Chemets DO (m	ng/L):		址			
			Analyzed?	EPA Method	c	ontainer 1	ype/Volume		Preservative	
Sai	mple Analyse	es: →						77		
							76			
		1.45								
Sample Coll	ection Metho	od: 📈		838,4				H		
rump:X F	low Rate:		Sample ID:				Sample Tim	e:		
Bailer: 🔽 T	stainless ype: dispose	stee 1	Duplicate ID:				Sample Tim	No. of Street,	11 241-1	
	esc.:		Equip. blank II	D:		77.4	Sample Time:			
CD	AA.		MON	NITORING WE	II PURGE					

2 of

Well No.:	AEW1	-	Site:				Date: 9	121/0	7		
Client:				Project Number	er:						
Well Casin	ng Diameter (inches):		Well Casing M	aterial: F	vc ss	Other:				
Well Head	space:	PID (ppm)	- 7 <u>.</u> 1 E								
Samplers:	bring a		with CDM		4			with Blair	ne Tech		
Total Dept	h of Well (fee	et):		2" - 0.16	3						
Depth to W	Vater (feet):			(X) 4" - 0.65	Gal/ft. =		(X) 3 =				
Water Colu	umn Height (f	feet):		6" - 1.47				Minim	um purge volui		
Well Refer	ence Point:			н_					(gallor		
PURGE M	ETHOD:	Submersib	le pump	Bladder pump	Dis	posable ba		7.7	172		
Pump Mak	e/Model: 2" (Grundfos Re	ediflo	Depth of pump	intake (fee	et):					
Purge equi	ipment decor	ntaminated?	Y N	Container type							
Purge/deco	on water con	tainerized?	Y N	Volume:		4			S 22 S		
	Start Time:		-417		Flow Rate				2 7 PL		
Time	Gallons	Temp.	рН	Conductivity	Turbidity	DO	ORP	DTW	Commercial		
		(°C)°F)		(µmhos/cm)	(NTUs)	(mg/L)	(mV)	(ft TOC)	Comments		
1140		23.94	6.44	6230	176	1.50	87.6	20.54			
1200	~ 2 gpm	23.61	6.39	6110	189	0.71	93.6	22.13	Pump \$ 50		
1220	~ 29pm	23.53	6.40	6/15	148	1.07	1742	23.01	61		
1231	≈ 90	23.21	6.38	3315	85.4	1.08	1125	23.41	"		
1246	2 29pm			5830	119	1.04	105.4	23.91	u		
1250	≈ 2gpm	23.19	6.42	6641	183	0.84	108.7	22.4	After sur		
1300	≈150	23.34	6.40	6389	83.2	0.71	1026	24.11	11		
1308	22 gpm	23.33	6.43	7/80	290	0.56	115.1	21.38	After surge		
1320	2 2 9pm	23.33	6.40	6451	104	0.76	165.2	24.27			
1326	2195	23.37	6.41	6287	103	1.05	97.9		After surg		
1341	x 2gpm	23.40		6416	65.3	0.75	93.6	24.23	During		
1345				6753	89.2	2.32	111.60	20_62	After sur		
1400		23.41		6355	125	0.96	110.4		During		
1430	£235	23.42	6.40	6467	45.2	0.85	112.2	24.18	During		
1455	270	23.40	Le. 39	6415	39-9	0-86	115.2	24.04	During		
				Chemets DO (n	ng/L):		- 4-8		6.		
			Analyzed?	EPA Method	C	ontainer T	ype/Volume	1	Preservative		
90	mple Analys	00: -					-				
38	imple Analys	es. 						<i>i</i>			
Sample Co	llection Metho	od: 🖌				-4-	25.3				
Pump: I	Flow Rate:		Sample ID:				Sample Tim	e:			
Bailer:	Гуре: disposa	able	Duplicate ID:				Sample Tim	ample Time:			
Other: 🔲 🛭	Desc.:		Equip. blank l	D:			Sample Tim	e:			

3 of

Well No.:	AEW:	1	Site:	17 , 17			Date:		1.
Client:				Project Number	er:				
Well Casing	Diameter (inches):		Well Casing M	aterial: P	VC SS	Other:		
Well Heads	pace:	PID (ppm)							
Samplers:			with CDM					with Blair	e Tech
Total Depth	of Well (fee	et):		2" - 0.16					
Depth to Wa	ater (feet):			(X) 4" - 0.65	Gal/ft. =		(X) 3 =		
Water Colu	mn Height (f	eet):		6" - 1.47				Minimu	ım purge volun
Well Refere	nce Point:			" <u> </u>					(gallon
PURGE ME		Submersib		Bladder pump	Disp	osable ba			
	e/Model: 2" (Depth of pump		t):	****		
	ment decor	7		Container type	:				
Purge/deco	n water cont	tainerized?	YUNU	Volume:					
Code	Start Time:				Flow Rate:				
Time	Gallons	Temp.	рН	Conductivity (µmhos/cm)	Turbidity (NTUs)	DO (mg/L)	ORP (mV)	DTW (ft TOC)	Comments
1507	~ 2gpm	23.52	6.41	6203	54.2	1.07	125.3	22,22	After suc
1520	x 29pm	23.40		6725	42.7	0.99	118.1		
1538	2320			6227	95.6		105.9	23.47	During During During
1559	~ 335	23.44 stop		6535 g for To	59.4	0.90	100.6	23.68	During
				Chemets DO (r	ng/L):				
			Analysis of O	EDA Made ed			0 (- 1		D
Sar	mple Analys	es: →	Analyzed ?	EPA Method		ontainer 1	ype/Volume		Preservative
Sample Coll	ection Meth	od: 🖌	gr8**						
Pump: F	low Rate:		Sample ID:				Sample Tim	e:	
	ype: disposa	able	Duplicate ID:				Sample Tim		
	esc.:		Equip. blank	D:			Sample Tim	**************************************	

Appendix F Injection Details

		NI	AVAL WI							
		IN/	AVAL WI							
Injection Event	Well ID	Injection Date Range	Volume Water Injected	Volume 60% Sodium Lactate Injected	Sodium Lactate Conc.	Volume Lactate Injected ¹	Lactate Injection Conc.	Total Injection Time (Hours)	Average Lactate Injection Flowrate (gpm)	Comments
			(gallons)	(gallons)	(%)	(gallons)	(%)	riiiio (riodio)	(92)	
	AIW-1	4/23-4/24/08	445							
1	AIW-2	4/23-4/24/09	505	27	5.3%	13	2.5%	10.0	0.8	
	TOTAL	4/23-4/24/10	950	50	5.3%	24	2.5%	10.0	1.6	
	AIW-1	7/16-18/08	408	22	5.5%	11	2.6%	16.0	0.7	
2	AIW-2	7/16-18/08	577	28	4.8%	13	2.3%	16.0	0.7	
	TOTAL	7/16-18/08	985	50 24	5.1%	24 12	2.4%	16.0	1.0 0.7	A stirry sounds are more for AC become following a injuration
3	AIW-1 AIW-2	10/17-21/2008	734 800	24	3.3% 3.2%	12	1.6% 1.5%	17.5 17.5		Active system run for 16 hours following injection
3	TOTAL	10/17-21/2008 10/17-21/2008	1,534	0.6 1.5	Active system run for 16 hours following injection					
	AIW-1	1/6-8/09	756	50 26	3.3% 3.5%	24 13	1.6% 1.7%	17.5 17.0	0.7	
4	AIW-1	1/6-8/09	625	21	3.4%	10	1.6%	17.0	0.7	
4	TOTAL	1/6-8/09	1,381	48	3.4%	23	1.7%	17.0	1.4	
PRE-CON		IG TOTALS	1,001	40	0.470		1.1 70	17.0	1.4	
		AIW-1	2,343	96	4.1%	46	2.0%	60.5	0.6	
		AIW-2	2,507	101	4.0%	49	1.9%	60.5	0.7	
		TOTAL	4,850	198	4.1%	95	2.0%	60.5	1.3	
						BIOAL	JGMENTATIO	NC		
	AIW-1	1/30/09	648	12.9	2.0%	6.2	1.0%	9.5	1.1	Switched to weekly injections
1	AIW-2	1/30/09	593	11.9	2.0%	5.7	1.0%	9.5	1.0	See separate spreadsheet for details:
	TOTAL	1/30/09	1,241	24.8	2.0%	11.9	1.0%	9.5	2.2	(0109 Active Injection Log.xls)
	AIW-1	2/5/09	428	6.9	1.6%	3.3	0.8%	6.5	1.1	See separate spreadsheet for details:
2	AIW-2	2/5/09	376	6.0	1.6%	2.9	0.8%	6.5	1.0	(0209 Active Injection Log.xls)
	TOTAL	2/5/09	804 337	12.9	1.6%	6.2	0.8%	6.5	2.1	
3	AIW-1 AIW-2	2/13/09 2/13/09	337	6.2 6.3	1.8% 1.8%	3.0	0.9% 0.9%	6.0 6.0	0.9 1.0	
3	TOTAL	2/13/09 2/13/09	682	12.5	1.8%	6.0	0.9%	6.0	1.9	
	AIW-1	2/20/09	359	6.1	1.7%	2.9	0.8%	7.5	0.8	
4	AIW-1	2/20/09	394	6.6	1.7%	3.2	0.8%	7.5	0.9	
"	TOTAL	2/20/09	753	12.7	1.7%	6.1	0.8%	7.5	1.7	
	AIW-1	2/27/09	391	5.6	1.4%	2.7	0.7%	7.8	0.8	
5	AIW-2	2/27/09	485	6.9	1.4%	3.3	0.7%	7.8	1.0	
	TOTAL	2/27/09	876	12.5	1.4%	6.0	0.7%	7.8	1.9	
FE	BRUARY	TOTALS	4,356	75	1.7%	36	0.8%	37	1.9	
	AIW-1	3/5/09	319	5.7	1.8%	2.8	0.9%	6.5	0.8	See separate spreadsheet for details:
6	AIW-2	3/5/09	382	6.9	1.8%	3.3	0.9%	6.5	1.0	(0309 Active Injection Log.xls)
	TOTAL	3/5/09	701	12.6	1.8%	6.0	0.9%	6.5	1.8	
	AIW-1	3/13/09	385	5.8	1.5%	2.8	0.7%	7.0	0.9	
7	AIW-2	3/13/09	451	6.9	1.5%	3.3	0.7%	7.0	1.1	
	TOTAL	3/13/09	836	12.7	1.5%	6.1	0.7%	7.0	2.0	

ESTCP Project ER-0513, NAVAL WEAPONS STATION SEAL BEACH, SITE 70 ACTIVE CELL INJECTION SUMMARY

			710	TIVE OLLE II	10EO HOI	I OOIVIIVI/ (I	V 1			
Injection	Well ID	Injection Date	Volume	Volume 60%	Sodium	Volume	Lactate	Total	Average Lactate	Comments
Évent		Range	Water	Sodium Lactate	Lactate	Lactate	Injection	Injection	Injection Flowrate	
		3.	Injected	Injected	Conc.	Injected ¹	Conc.	Time (Hours)	(gpm)	
			(gallons)	(gallons)	(%)	(gallons)	(%)	((9P)	
		2/22/22				,				
	AIW-1	3/20/09	456	6.9	1.5%	3.3	0.7%	7.0	1.1	
8	AIW-2	3/20/09	377	5.7	1.5%	2.7	0.7%	7.0	0.9	
	TOTAL	3/20/09	833	12.6	1.5%	6.0	0.7%	7.0	2.0	
	AIW-1	3/27/09	419	5.6	1.3%	2.7	0.6%	7.0	1.0	
9	AIW-2	3/27/09	495	6.7	1.3%	3.2	0.6%	7.0	1.2	
	TOTAL	3/27/09	914	12.3	1.3%	5.9	0.6%	7.0	2.2	
N	ARCH T	OTALS	3,284	50	1.5%	24	0.7%	28	2.0	
	AIW-1	4/2/09	285	7.1	2.5%	3.4	1.2%	7.3	0.6	See separate spreadsheet for details:
10	AIW-2	4/2/09	228	5.6	2.5%	2.7	1.2%	7.3	0.5	(0409 Active Injection Log.xls)
	TOTAL	4/2/09	513	12.7	2.5%	6.1	1.2%	7.3	1.2	(
	AIW-1	4/8/09	383	6.8	1.8%	3.3	0.9%	5.7	1.1	
11	AIW-2	4/8/09	327	5.8	1.8%	2.8	0.9%	5.7	1.0	
1 ''	TOTAL	4/8/09	710	12.6	1.8%	6.0	0.9%	5.7	2.1	
		4/18/09	333							
	AIW-1			6.2	1.9%	3.0	0.9%	6.0	0.9	
12	AIW-2	4/18/09	331	6.2	1.9%	3.0	0.9%	6.0	0.9	
	TOTAL	4/18/09	664	12.4	1.9%	6.0	0.9%	6.0	1.8	
	AIW-1	4/24/09	396	6.6	1.7%	3.1	0.8%	6.0	1.1	
13	AIW-2	4/24/09	341	5.6	1.7%	2.7	0.8%	6.0	0.9	
	TOTAL	4/24/09	737	12.2	1.7%	5.9	0.8%	6.0	2.0	
	APRIL TO	OTALS	2,624	50	1.9%	24	0.9%	25	1.8	
	AIW-1	5/1/09	398	6.6	1.6%	3.2	0.8%	6.0	1.1	See separate spreadsheet for details:
14	AIW-2	5/1/09	360	5.9	1.6%	2.8	0.8%	6.0	1.0	(0509 Active Injection Log.xls)
	TOTAL	5/1/09	758	12.5	1.6%	6.0	0.8%	6.0	2.1	, ,
	AIW-1	5/7/09	463	7.3	1.6%	3.5	0.8%	7.0	1.1	
15	AIW-2	5/7/09	325	5.2	1.6%	2.5	0.8%	7.0	0.8	
13	TOTAL	5/7/09	788	12.5	1.6%	6.0	0.8%	7.0	1.9	
	AIW-1	5/15/09	458	6.6	1.4%	3.2	0.7%	7.0	1.1	
16	AIW-1	5/15/09	420	6.1	1.4%	2.9	0.7%	7.0	1.0	
16										
	TOTAL	5/15/09	878	12.7	1.4%	6.1	0.7%	7.0	2.1	
	AIW-1	5/22/09	444	6.3	1.4%	3.0	0.7%	9.0	0.8	
17	AIW-2	5/22/09	450	6.3	1.4%	3.0	0.7%	9.0	0.8	
	TOTAL	5/22/09	894	12.6	1.4%	6.0	0.7%	9.0	1.7	
	AIW-1	5/29/09	442	7.0	1.6%	3.3	0.8%	7.8	0.9	
18	AIW-2	5/29/09	352	5.5	1.6%	2.7	0.8%	7.8	0.7	
	TOTAL	5/29/09	794	12.5	1.6%	6.0	0.8%	7.8	1.7	
	MAY TO	TALS	4,112	63	1.5%	30	0.7%	37	1.9	
		6/3/09	421	6.9	1.6%	3.3	0.8%	6.0	1.2	See separate spreadsheet for details:
	AIW-1									
19	AIW-1	6/3/09	356	5.8	1.6%	2.8	0.8%	6.0	1.0	(0609 Active Injection Log.xls)
19			356	5.8 12.7		2.8 6.1			1.0 2.2	(0609 Active Injection Log.xls)
19	AIW-2 TOTAL	6/3/09 6/3/09	356 777	12.7	1.6%	6.1	0.8%	6.0	2.2	(0609 Active Injection Log.xls)
19	AIW-2	6/3/09	356							(0609 Active Injection Log.xls)

ESTCP Project ER-0513, NAVAL WEAPONS STATION SEAL BEACH, SITE 70 ACTIVE CELL INJECTION SUMMARY

			710	TIVE CELE II	10EOTION	i COIVIIVI/ (I	V 1			
	Well ID	Injection Date	Volume	Volume 60%	Sodium	Volume	Lactate	Total	Average Lactate	
Event		Range	Water	Sodium Lactate	Lactate	Lactate	Injection	Injection	Injection Flowrate	;
			Injected	Injected	Conc.	Injected 1	Conc.	Time (Hours)	(gpm)	
			(gallons)	(gallons)	(%)	(gallons)	(%)			
	TOTAL	6/9/09	637	12.4	1.9%	6.0	0.9%	5.0	2.1	
	AIW-1	6/20/09	529	25.9	4.9%	12.4	2.3%	7.0	1.3	**Injection volume changed to 50 gallons
21	AIW-2	6/20/09	491	24.0	4.9%	11.5	2.3%	7.0	1.2	
	TOTAL	6/20/09	1,020	49.9	4.9%	24.0	2.3%	7.0	2.4	
	AIW-1	6/26/09	378	24.7	6.5%	11.8	3.1%	5.0	1.3	
22	AIW-2	6/26/09	391	25.5	6.5%	12.3	3.1%	5.0	1.3	
	TOTAL	6/26/09	769	50.2	6.5%	24.1	3.1%	5.0	2.6	
	JUNE TO	TALS	3,203	125	3.8%	60	1.8%	23	2.3	
	AIW-1	7/2/09	484	23.8	4.9%	11.4	2.4%	15.2	0.5	See separate spreadsheet for details:
23	AIW-2	7/2/09	541	26.5	4.9%	12.7	2.4%	15.2	0.6	(0709 Active Injection Log.xls)
	TOTAL	7/2/09	1,025	50.3	4.9%	24.1	2.4%	15.2	1.1	
	AIW-1	7/9/09	521	27.1	5.2%	13.0	2.5%	13.8	0.6	_
24	AIW-2	7/9/09	446	23.2	5.2%	11.1	2.5%	13.8	0.5	
	TOTAL	7/9/09	967	50.3	5.2%	24.1	2.5%	13.8	1.2	
	AIW-1	7/17/09	498	25.1	5.0%	12.0	2.4%	9.8	0.8	
25	AIW-2	7/17/09	505	25.4	5.0%	12.2	2.4%	9.8	0.9	
	TOTAL	7/17/09	1,003	50.5	5.0%	24.2	2.4%	9.8	1.7	
	AIW-1	7/24/09	361	25.9	7.2%	12.4	3.4%	11.0	0.5	
26	AIW-2	7/24/09	343	24.6	7.2%	11.8	3.4%	11.0	0.5	
	TOTAL	7/24/09	704	50.5	7.2%	24.2	3.4%	11.0	1.1	
	AIW-1	7/29/09	436	26.2	6.0%	12.6	2.9%	11.0	0.7	
27	AIW-2	7/29/09	405	24.3	6.0%	11.7	2.9%	11.0	0.6	
	TOTAL	7/29/09	841	50.5	6.0%	24.2	2.9%	11.0	1.3	
	JULY TO	TALS	4,540	252	5.7%	121	2.7%	61	1.3	
	AIW-1	8/7/09	346	22.4	6.5%	10.8	3.1%	10.3	0.6	See separate spreadsheet for details:
28	AIW-2	8/7/09	434	28.1	6.5%	13.5	3.1%	10.3	0.7	(0809 Active Injection Log.xls)
	TOTAL	8/7/09	780	50.5	6.5%	24.2	3.1%	10.3	1.3	
	AIW-1	8/14/09	458	24.3	5.3%	11.7	2.5%	16.0	0.5	
29	AIW-2	8/14/09	483	25.6	5.3%	12.3	2.5%	16.0	0.5	
	TOTAL	8/14/09	941	49.9	5.3%	24.0	2.5%	16.0	1.0	
	AIW-1	8/21/09	469	24.1	5.1%	11.6	2.5%	12.2	0.6	
30	AIW-2	8/21/09	507	26.0	5.1%	12.5	2.5%	12.2	0.7	
	TOTAL	8/21/09	976	50.1	5.1%	24.0	2.5%	12.2	1.3	
	AIW-1	8/28/09	426	22.4	5.2%	10.7	2.5%	14.0	0.5	_
31	AIW-2	8/28/09	521	27.3	5.2%	13.1	2.5%	14.0	0.6	
	TOTAL	8/28/09	947	49.7	5.2%	23.9	2.5%	14.0	1.1	
Α	UGUST T	OTALS	3,644	200	5.5%	96	2.7%	53	1.2	
	AIW-1	9/3/09	536	31.4	5.9%	15.1	2.8%	9.0	1.0	See separate spreadsheet for details:
32	AIW-2	9/3/09	319	18.7	5.9%	9.0	2.8%	9.0	0.6	(0909 Active Injection Log.xls)
	TOTAL	9/3/09	855	50.1	5.9%	24.0	2.8%	9.0	1.6	
	AIW-1	9/11/09	-	-	-	-	-	2.0	-	System power outage. System restarted on 9/11/09

		NI	^\/^I \//I	ESTCP PI EAPONS STA	-		I CITE 70			
		IN/		TIVE CELL IN			•	1		
Injection Event	Well ID	Injection Date Range	Volume Water Injected (gallons)	Volume 60% Sodium Lactate Injected (gallons)	Sodium Lactate Conc. (%)	Volume Lactate Injected ¹ (gallons)	Lactate Injection Conc. (%)	Total Injection Time (Hours)	Average Lactate Injection Flowrate (gpm)	Comments
33	AIW-2	9/11/09	-	-	-	-	-	2.0	-	temporarily but was again shut down after not working
	TOTAL	9/11/09	-	9.8	-	4.7	-	2.0	-	properly. System fixed and restarted on 9/17/09.
	AIW-1	9/17/09	470	25.9	5.5%	12.4	2.6%	8.0	1.0	
34	AIW-2	9/17/09	277	15.2	5.5%	7.3	2.6%	8.0	0.6	
	TOTAL	9/17/09	747	41.1	5.5%	19.7	2.6%	8.0	1.6	
	AIW-1	9/18/09	571	32.3	5.7%	15.5	2.7%	10.0	1.0	
35	AIW-2	9/18/09	324	18.3	5.7%	8.8	2.7%	10.0	0.5	
	TOTAL	9/18/09	895	50.6	5.7%	24.3	2.7%	10.0	1.5	
	AIW-1	9/25/09	483	30.2	6.2%	14.5	3.0%	14.0	0.6	
36	AIW-2	9/25/09	326	20.3	6.2%	9.8	3.0%	14.0	0.4	
0==	TOTAL	9/25/09	809	50.5	6.2%	24.2	3.0%	14.0	1.0	
SEF		R TOTALS	3,306	202	5.8%	92	2.8%	41	1.4	
07	AIW-1	10/2/09 10/2/09	321 374	19.7	6.1% 6.1%	9.4 11.0	2.9% 2.9%	9.0 9.0	0.6 0.7	
37	AIW-2	10/2/09	695	22.9 42.6	6.1%	20.4	2.9% 2.9%	9.0	1.3	
0(CTOBER		695	42.6	6.1%	20.4	2.9%	9.0	1.3	
	JIOBER	IOTALS	13,974	797	5.7%	378	2.7%	175	1.3	
DOCT I		INACNITATION		_	5.7%	378	2.1%	175	1.3	
PUS1-1	BIOAGU	JMENTATION								
		AIW-1	15,389	547	3.6%	262	1.7%	313	0.8	
		AIW-2	14,375	504	3.5%	242	1.7%	313	0.8	
		TOTAL	29,764	1,061	3.6%	504	1.7%	313	1.6	
OVERA	LL TO	TALS (PRE-C	ONDITIO	NING & POST	-BIOAUG	MENTATIO	ON)			
			17,732 16,882	643 605	3.6%	309 290	1.7% 1.7%	373 373	0.8	
			34,614	1,258	3.6%	599	1.7%	373	1.5	

ESTCP Project ER-0513, NAVAL WEAPONS STATION SEAL BEACH, SITE 70 PASSIVE CELL INJECTION SUMMARY Injection Well ID Injection Date Volume Volume 60% Sodium Volume Lactate Total Average Lactate Comments Event Range Water Sodium Lactate Lactate Lactate Injection Injection Injection Injected Injected Conc. Conc. Time (Hours) Flowrate Injected 1 (gallons) (gallons) (gpm) (%) (%) (gallons) Pre-Conditioning 8/7-8/8/08 0.9% PIW-1 924 16.5 1.8% 7.9 14.5 1.1 Tracer performed in PIW-1 during this injection. PIW-2 8/7-8/8/08 1,066 17.0 1.6% 8.2 0.8% 12.3 1.5 1 PIW-3 8/7-8/8/08 1.066 17.0 1.6% 8.2 0.8% 12.3 1.5 TOTAL 8/7-8/8/08 3.057 51 1.7% 24 0.8% 14.5 3.5 PIW-1 9/8-9/9/08 1,067 17.0 1.6% 8.1 0.8% 16.2 1.1 9/8-9/9/08 17.0 1.6% 0.8% 13.3 1.3 PIW-2 1,071 8.2 2 PIW-3 9/8-9/9/08 1,067 17.0 1.6% 8.1 0.8% 13.3 1.3 51 TOTAL 9/8-9/9/08 3.205 1.6% 24 0.8% 16.2 3.3 PIW-1 10/21-22/08 1,067 17 1.6% 8 0.8% 18.0 1.0 PIW-2 10/21-22/08 1.066 17 1.6% 8 0.8% 18.0 1.0 3 PIW-3 10/21-22/08 17 1.6% 0.8% 18.0 1.0 1,066 8 52 25 **TOTAL** 10/21-22/08 3.199 1.6% 0.8% 18.0 3.0 PIW-1 1/6-8/09 953 15.8 1.7% 7.6 0.8% 15.9 1.0 1/6-8/09 1.7% 0.8% 15.9 PIW-2 954 15.8 7.6 1.0 4 PIW-3 1/6-8/09 952 15.8 1.7% 7.6 0.8% 15.9 1.0 TOTAL 1/6-8/09 2,859 48 1.7% 23 0.8% 15.9 3.0 PIW-1 8/7/08-1/12/09 4.011 67 1.7% 32 0.8% 65 1.0 Pre-PIW-2 32 1.2 8/7/08-1/12/09 4.156 67 1.6% 0.8% 60 Conditio ning PIW-3 32 8/7/08-1/12/09 4.151 67 1.6% 0.8% 60 1.2 Totals **TOTAL** 8/7/08-1/12/09 12.319 201 1.6% 96 0.8% 65 3.2 POST-BIOAUGMENTATION 2/4-2/6/09 PIW-1 1,001 16.7 1.7% 8.0 0.8% 17.2 1.0 PIW-2 2/4-2/6/09 16.7 1.7% 0.8% 17.2 1.001 8.0 1.0 PIW-3 2/4-2/6/09 1.000 16.7 1.7% 8.0 0.8% 17.2 1.0 TOTAL 2/4-2/6/09 50 1.7% 24.0 17.2 3,002 0.8% 2.9 PIW-1 3/2-3/5/09 1.000 16.6 1.7% 8.0 0.8% 16.8 1.0 PIW-2 3/2-3/5/09 1,006 16.7 1.7% 8.0 0.8% 16.8 1.0 2 PIW-3 3/2-3/5/09 1.007 16.7 1.7% 8.0 0.8% 16.8 1.0 50 **TOTAL** 3/2-3/5/09 3,013 1.7% 24.0 0.8% 16.8 3.0 PIW-1 4/1-4/2/09 1,000 16.6 1.7% 8.0 0.8% 17.7 0.9 PIW-2 4/1-4/2/09 1,002 16.7 1.7% 8.0 0.8% 17.7 0.9 3 4/1-4/2/09 17.7 PIW-3 1.005 16.7 1.7% 8.0 0.8% 0.9 4/1-4/2/09 50 24.0 0.8% 17.7 **TOTAL** 3.007 1.7% 2.8 5/5-5/7/09 1.000 16.7 1.7% 8.0 0.8% 16.6 1.0 PIW-1 PIW-2 5/5-5/7/09 1.000 16.7 1.7% 8.0 0.8% 16.6 1.0 4 PIW-3 5/5-5/7/09 1.001 16.7 1.7% 8.0 0.8% 16.6 1.0 **TOTAL** 5/5-5/7/09 3,001 50 1.7% 24.0 0.8% 16.6 3.0 PIW-1 6/1-6/3/09 1.000 16.6 1.7% 8.0 0.8% 18.0 0.9

		NA	VAL WE							
Injection Event	Well ID	Injection Date Range	Volume Water Injected (gallons)	Volume 60% Sodium Lactate Injected (gallons)	Sodium Lactate Conc. (%)	Volume Lactate Injected ¹ (gallons)	Lactate Injection Conc. (%)	Total Injection Time (Hours)	Average Lactate Injection Flowrate (gpm)	Comments
5	PIW-2 PIW-3	6/1-6/3/09 6/1-6/3/09 6/1-6/3/09	1,001 1,003 3,004	16.7 16.7 50	1.7% 1.7% 1.7%	8.0 8.0 24.0	0.8% 0.8% 0.8%	18.0 18.0 18.0	0.9 0.9 2.8	
6	PIW-1 PIW-2 PIW-3	6/30-7/2/09 6/30-7/2/09 6/30-7/2/09 6/30-7/2/09	1,197 1,152 1,161 3,510	1.0 0.9 0.9						
7	PIW-1 PIW-2 PIW-3 TOTAL	8/19-8/21/09 8/19-8/21/09 8/19-8/21/09 8/19-8/21/09	1,117 1,158 1,172 3,447	19.4 20.2 20.4 60	1.7% 1.7% 1.7% 1.7%	9.3 9.3 9.3 28.8	0.8% 0.8% 0.8% 0.8%	18.5 18.5 18.5 18.5	1.0 1.0 1.1 3.1	
8	PIW-1 PIW-2 PIW-3 TOTAL	9/1-9/3/09 9/1-9/3/09 9/1-9/3/09 9/1-9/3/09	1,166 1,200 1,200 3,566	19.6 20.2 20.2 60	1.7% 1.7% 1.7% 1.7%	9.4 9.4 9.4 28.8	0.8% 0.8% 0.8% 0.8%	18.0 18.0 18.0	1.1 1.1 1.1 3.3	
Post-B		entation Tota								
		PIW-1 PIW-2 PIW-3	8,481 8,519 8,549	143 143 144	1.7% 1.7% 1.7%	69 69 69	0.8% 0.8% 0.8%	143 143 143	1.0 1.0 1.0	
		TOTAL	25,549	430	1.7%	206	0.8%	143	3.0	
Overall	Totals	(Pre-Condition	oning & F	ost-Bioaugm	entation)					
		PIW-1 PIW-2 PIW-3	1.0 1.0 1.0							
		TOTAL	12,701 37,868	211 631	1.7% 1.7%	101 303	0.8% 0.8%	203 208	3.0	

Appendix G Sampling Methods Supplemental Information and Quality Assurance Information

Appendix G – Quality Assurance and Quality Control Procedures

G.1 Calibration Procedures, Quality Control Checks, and Corrective Action

The purpose of this section is to provide a summary of the specific maintenance/calibration procedures for all equipment related to the collection of data either in the field or through laboratory analysis of samples during completion of the project.

G.1.1 Laboratory Equipment Calibration

Calibration procedures for laboratory instruments are found in each laboratory's QA Manual. Calibration for analyses performed by offsite laboratories were defined by the analytical methods. Data reduction and validation for the laboratory data and for the final reporting were described in the laboratory's QA Manual.

G.1.2 Field Instrumentation

Field instrumentation was used to provide data concerning health and safety considerations and as a method for field screening samples.

G.1.2.1 Photoionization Detector

Calibration of the instrument was performed with a factory supplied calibration kit according to the manufacturer's specifications. Calibration was performed daily as a part of routine instrument maintenance, with a calibration record being maintained in the field manager's logbook.

G.1.2.2 HACH Kits

HACH kits were used to measure concentrations of specific parameters in the field. Vendor instructions for use of these kits were followed and documented; kits were calibrated by the vendor and do not require calibration by the user. This includes the operation of the HACH DR2000 spectrometer.

G.1.2.3 Multi-parameter Water Quality Instrument

The mutli-parameter Water Quality Instrument is a specially designed vessel that allows simultaneous measurement of water quality parameters as fresh flowing water is passed through the cell. For this field work, the instrument was used to measure temperature, conductivity, pH, redox potential (Eh), and DO. Calibration was performed in accordance with instrument procedures requiring fresh calibration solutions. Instruments were rented for this demonstration project, and were properly calibrated by the vendor. However, field calibration was performed as necessary when parameter drift or malfunction was noted. Field calibration was recorded in the field logbook.

G.2 Quality Assurance Sampling

G.2.1 Accuracy

For this demonstration, accuracy of laboratory results was assessed using the analytical results of method-defined surrogates, laboratory control samples, matrix spikes, and calibration standards. The percent recovery (%R) was calculated using the following equations:

$$\%R = \frac{A - B}{C} \times 100$$

where: A = Analyte concentration determined experimentally in the spiked sample; B = Analyte concentration determined by a separate analysis of the unspiked sample; and, C = Concentration of Spiked analyte.

The only parameters that required matrix spikes are the VOC samples sent to an offsite laboratory. The accuracy goal for these samples was a percent recovery of 70-130%. The accuracy goal for all field and trip blanks was no detections of analytes in these samples.

G.2.2 Precision

Precision was assessed by calculating the relative percent difference (RPD) between the field duplicate samples. The RPD was calculated for each pair of duplicates using the following equation:

$$\% RPD = \frac{S - D}{(S + D)/2} \times 100$$

where: S = First sample value D = Second sample value (duplicate value)

The precision goal for this project for sample pairs whose values are both greater than 10X the MDL limit was an RPD \leq 25%. For sample pairs that have one or both values less than 10X the MDL, the precision goal was RPD \leq 50%. Sample pairs that have one or both values that are less than the MDL did not have RPDs calculated.

G.2.3 Completeness

Completeness of data was assessed as the percentage amount of valid data compared to the total amount of expected data using the following equation:

$$\% \, Completeness = \frac{Valid \ Data \ Obtained}{Total \ Data \ Planned} \times 100$$

The completeness goal for this project was 90% of all planned samples, as defined in the Demonstration Plan. Completeness was tracked both for individual sampling rounds and cumulatively over the course of the demonstration.

G.2.4 Representativeness

Representativeness expresses the degree to which data accurately and precisely represent a characteristic of a population and parameter variations at a sampling point, a process condition, or an environmental condition. Representativeness is a qualitative parameter that is dependent on the proper design of the sampling program and proper laboratory protocol. The sampling program was described in Section 3.7.6 of the Demonstration Plan.

Representativeness of the data was assessed by the Project Manager and the QA Coordinator through review and comparison of the applicable data (field and laboratory duplicates, spikes, blanks) and by verifying that the sampling and analysis plan/design set forth in the Demonstration Plan was followed for all data generated during the project activities.

G.2.5 Comparability

Comparability expresses the confidence with which one data set can be compared with another. The extent of comparability between existing and planned analytical data depends in part on the similarity of sampling and analytical methods. The procedures used to obtain the planned analytical data, as documented in the QAPP, were expected to provide comparable data for these project activities.

G.3 Equipment Decontamination

Equipment decontamination was performed for all intrusive instruments that were not dedicated equipment. Decontamination of drilling equipment, including steam cleaning, was performed during well installation. Additionally, decontamination of field instruments that were not dedicated to the wells was performed in between wells utilizing Alconox and distilled water.

G.4 Documentation of Sample Collection

All sample collection was documented as described in the QAPP. The following information, as applicable, was recorded.

- Custody and Document Control
- Chain-of-custody from field to laboratory
- Laboratory custody through designated laboratory-sample custodian
- Sample designation number(s)
- Identity of sampler
- Date of sample collection, shipping, and laboratory analysis
- Physical Data Elements
- Sampling date and time
- Sampling location and description
- Sample collection technique
- Field preparation techniques (e.g., filtering, sieving, compositing)
- Visual classification of sample using an accepted classification system
- A description of the sampling methodology used

Appendix H Active Cell Concentration Trends

ACTIVE	CELL					1		1		1														
	ing Data	aue	4)											Oxygen	s e	- Se	S	Se						
Summa	-	Fetrachloroethene	ene	ethene	eue	<u>e</u>								X yg	ojd Ojd) jo	ojde) ji				_	_	
NAVFA	-	õ	eth	the	- the	oric								Ô	oo 4	Ö	Ö	Ö				vity	ron	छ
Weapor		皇)ro(-i-	1,2- roet	S.	Φ	Φ	ne	<u>i</u>	4)	o)	qe	r ca	alococ	50	9	00				jo j	- SI	nen
	- Site 70	trac	Trichloroeth	cis-1,2- Dichloroe	trans-1, Dichlor	/inyl (ner	lan	tha	i <u>i</u>	rate	ulfate	Chloride	Chemical Demand	hal T	hal A	hal Y	hal		۵		ndt	JO .	Ë
Seal Be	ach, CA	Tel	H ri	cis Dic	traı Dic	≥	Eth	苗	Me	Alk	Z.	nS	чɔ	C P	Dehi 16S	De	De	De	Hd	OR	oa	Co	Fe	O
	Units:	μg/L	μg/L	μg/L	μg/L	μg/L	μg/L	μg/L	μg/L	mg/L	mg/L	mg/L	mg/L	mg/L		gene c	opies/L			mV	mg/L	µmhos/cm	mg/L	
	4/9/2008 -PP	20 U	2100	83	25	20 U	5 U	5 U	5 U	860	0.89	8700	1400	34	ND	ND	ND	ND	7.38	115	2.87	17496	0	
	4/9/2008 - BP	17 U	1800	85	25	17 U	5 U	5 U	5 U	NS	NS	NS	NS	NS	NS	NS	NS	NS						
	5/14/08	4.7 J	2000	140	49	10 U	5 U	5 U	5 U	910	0.94	8100	1200	26	ND	NS	NS	NS	7.49	66.8	4.04	16072	0	
	5/14/2008 - K	4.6 J	1900	140	44	8.3 U	5 U	5 U	5 U	910	1.2	7900	1200	26	NS	NS	NS	NS	0.40	242.0	4.40	1100	0	
_	9/3/08 11/5/08	50 U 50 U	8800 8100	480 990	22 J 29 J	50 U 50 U	5 U	5 U 5 U	5 U 5 U	680 660	0.57 0.5 U	5200 5400	1700 1800	28 32	ND ND	ND ND	ND ND	ND ND	6.18 6.86	212.6 134.9	4.19 1.31	1169 12010	0 NM	
\$	1/29/09	50 U	7200	1200	30 J	50 U	5 U	5 U	5 U	670	0.5 U	6200	1900	38	2.56E+03	1.81E+03	ND	1.55E+03	6.79	-22.9	3.04	15450	0	
₽	2/24/09	50 U	38 J	7000	28 J	50 U	5 U	5 U	6	950	0.5 U	6200	2000	120	5.11E+05	1.54E+06	1.12E+05*	4.49E+05	6.61	-92.5	0.68	13311	2.6	
	3/31/09	50 U	120	6300	27 J	320	5 U	5 U	11	980	1 U	7300	1800	68	2.09E+08	7.55E+08	1.17E+08		6.64	-183.9	0.08	16260	2.7	
	4/29/09	14 J	880	3000	22 J	1100	17	5 U	6	900	1 U	8100	1700	49	2.70E+08		2.30E+08		6.71	-50.7	1.49	16300	3.3	
	5/28/09	3.4 J	1400	2100	24	2000	16	5 U	8	870	1 U	8100	1700	44	4.08E+08	3.57E+08		1.20E+07	6.58	-13.6	0.08	15920	0	
	6/23/09	17 U	1900	1200	25	2800	22	5 U	8	790	1 U	7800	1600	87	7.27E+07		1.70E+07		7.05	-110.9	1.88	15690	0	
	10/16/09	17 U	220	520	29	2000	660	5 U	7	1400	0.5 U	3700	1600	180	540000000	5.32E+08	6.50E+06		5.24	-225.5	0.76	10980	>3.3	
	4/8/2008 -PP	25 U	3400	630	17 J	25 U	5 U	5 U	5 U	780	0.5 U	7400	2700	38	ND	ND	ND	ND	6.92	442.6	1.15	18554	0	
	4/8/2008 - K	31 U	3500	630	12 J	31 U	5 U	5 U	5 U	780 NS	0.5 U	7400 NS	2700 NS	42 NS	NS NS	NS NS	NS NS	NS NS					0 NM	
	4/9/2008 -BP 5/14/08	31 U 11 J	3300 10000	630 1400	16 J 41 J	31 U 21 J	5 U	5 U 5 U	5 U 13	NS 790	NS 0.25 U	NS 5000	NS 2100	NS 44	NS ND	NS NS	NS NS	NS NS	6.7	-56.8	0.26	14500	0.125	
	9/3/08	11 J	6900	4000	25 J	31 U	5 U	5 U	6	670	0.25 U	2100	2000	78	ND	ND	ND	ND	5.83	-197.9	1.95	8146	3.3	
	11/5/08	5.6 J	1300	8400	21	35	5 U	5 U	6	660	0.25 U	2600	2100	47	3.36E+03	2.36E+03	4.60E+02	ND	6.55	-58.5	0.66	9340	>3.3	
N-2	1/29/09	71 U	650	11000	26 J	250	5 U	5 U	12	1400	0.5 U	2900	2000	580	2.01E+06		7.92E+04		6.38	-159.5	0.56	11980	2.98	
Ş	2/24/09	71 U	97	9500	19 J	940	5 U	5 U	14	1100	0.5 U	3400	2000	160	3.44E+08	8.72E+08	3.07E+08	2.10E+06	6.46	-214.2	1.31	9967	NM	
•	3/31/09	36 U	230	5400	21 J	4300	5 U	5 U	5 U	980	0.25 U	3200	1700	61	4.05E+08		4.71E+08		6.55	-194.7	0.25	10710	>3.3	
	4/29/09	21 J	540	1700	27 J	6900	16	1 J	15	1000	0.5 U	3400	1900	63	3.17E+08		4.60E+08		6.47	-203.9	10.76	10750	3.3	
	5/28/09	20 J	210	780	27 J	7500	27	1 J	16	760	0.25 U	2300	1600	42	3.61E+09		6.40E+08		6.34	-108	-0.06	7985	2.97	
	6/24/09	50 U	280	440	50 U	6400	33	1 J	15	750	0.5 U	2100	1500	49	9.09E+07		1.70E+07		6.8	-135.2	1.35	8015	>3.3	
	6/24/2009-K 10/16/09	50 U 25 U	290 390	430 730	12 J 29	6500 4200	34 140	1 J 1 J	15 17	740 1900	0.5 U 0.03 J	2100 540	1500 1500	47	1.23E+08 6900000	1.60E+08 6.54E+07	3.60E+06	1.30E+05 4.50E+05	5.85	-318.2	1.43	7178	0	
-	4/8/08	10 U	1200	32	10 U	10 U	5 U	5 U	20	560	0.03 J	7900	4000	60	ND	ND	ND	ND	6.95	195.7	0.78	22651	0	
	5/15/08	3.5 J	2800	440	16 J	9 J	5 U	5 U	17	580	0.21 J	7300	3600	60	ND	NS	NS	NS	7.12	105.8	0.35	21781	0	
	9/3/08	63 U	8100	1700	20 J	63 U	5 U	5 U	6	730	0.5 U	5900	2700	47	ND	ND	ND	ND	6.77	170.9	3.86	1519	0	
e 1	11/5/08	50 U	8000	1900	30 J	50 U	5 U	5 U	9	710	0.5 U	6000	2900	57	ND	ND	ND	ND	6.85	478.3	1.31	15530	0	
Zone	1/29/09	50 U	9100	1400	21 J	50 U	5 U	5 U	10	650	0.5 U	5200	2700	47	1.36E+04	2.79E+03	3.02E+02		6.77	63.3	2.59	15310	0	
	2/24/09	50 U	6000	2300	20 J	210	5 U	5 U	11	630	0.5 U	4000	2300	42	4.40E+06	1.40E+07	4.72E+06		6.76	74.3	3.25	10911	0.95	
AMW-3	3/31/09 4/29/09	31 U 31 U	3900 1200	3700 2700	27 J 23 J	2400 3900	15 58	5 U 1 J	12 16	870 980	0.5 U 0.5 U	3700 3700	2200 2100	40	5.19E+05 3.08E+05	2.06E+06	5.43E+05 3.70E+05		6.44	-17.9 -31.8	0.63 1.58	12540 11370	0.13	
⋖	5/28/09	31 U	930	1500	32	7500	120	5 U	16	930	0.5 U	3300	1900	38	1.90E+06		7.10E+05		6.24	20.7	0.77	10140	0.13	
	6/24/09	63 U	580	690	30 J	7000	200	1 J	18	890	0.5 U	3100	1800	44	2.80E+05		3.30E+04		6.59	53.4	0.84	9798	0	
	10/16/09	25 U	63	150	26	3900	680	0.7 J	7	1500	0.25 U	1100	1700	420	48800000	3.55E+07	2.90E+06	5.20E+06	5.22	13.4	3.5	9137	0	
ი ⊿	11/6/08	50 U	7700	1300	23 J	50 U	5 U	5 U	8	660	0.5 U	4900	2500	40	ND	ND	ND	ND	6.83	497.4	1.63	14250	0	
ĕĕ	4/29/09	50 U	780	6500	24 J	1900	13	5 U	11	830	0.5 U	3200	2000	51	1.66E+08		2.20E+08		6.44	-14.9	2.16	9682	3.3	
AMW- Zone	6/24/09	50 U	520	1400	23 J	5500	77	5 U	13	700	0.5 U	2500	1800	40	1.29E+08	1.20E+08		1.60E+06	6.69	-31.8	1.58	9212	>3.3	
-	10/16/09 11/6/08	3.3 J 71 U	31 8900	57 780	29 18 J	1500 71 U	1500 5 U	0.9 J 5 U	9 13	1500 530	0.25 U 0.5 U	170 3700	1500 2300	540 38	378000000 ND	2.98E+08 ND	2.60E+06 ND	6.00E+07 ND	5.26 6.78	-111.9 80.6	2.79 1.39	7568 13270	>3.3	
V-3	4/29/09	71 U	130	11000	19 J	190	5 U	1 J	18	830	0.5 U	2100	1600	59	1.82E+07		2.60E+07		6.38	-95.6	2.47	9428	3.3	
AMW	6/24/09		87	2000	21 J	5300	27	2 J	22	670	0.25 U		1600			1.80E+08				-72.6		7396	2.28	
	10/16/09	42 U	43	65	24 J	6000	59	1 J	12	870	0.25 U	1400	1600	41	39300000	1.96E+07	2.90E+06	1.40E+05	5.19	-71.1	1.51	6537	>3.3	
6-1	4/8/08	1.3 J	96	5.2	0.9 J	2.5 U	5 U	5 U	5 U	NS	NS	NS	NS	NS	NS	NS	NS	NS	7.2	144.3	0.78	29500	NM	
AMW-3 Zone 4																								
ΚN	4/0/00	40.11	4000	00	701	40.11			0.1	500	0.44.1	0000	2022	40	NID	N'D	ND	L ND	0.00	404.5	0.54	40040	0.47	
	4/8/08 5/15/08	13 U	1800 7000	86 1500	7.9 J 39	13 U 24 J	5 U	5 U 5 U	21 g	560 720	0.14 J 1 U	6300 4800	3600 2500	48 120	ND ND	ND NS	ND NS	ND NS	6.86 6.96	161.5 -56.4	0.54	18849 14952	0.17	
	9/3/08	6.6 J 50 U	8100	1600	33 J	24 J 15 J	5 U	5 U	7	720	0.5 U		2600	51	ND ND	ND ND	ND ND	ND ND	6.45	22.4	36.76*	1359	3.16	
- m	11/6/08	71 U	8600	1300	29 J	71 U	5 U	5 U	9	710	0.5 U		2800	47	2.07E+03				6.84	-22.11	1.26	14850	3.10	
Zone	1/29/09		3000	7600	26	240	5 U	5 U	10	800	0.5 U	4300	2500	55	7.70E+05			4.98E+04	6.75	-63.8	2.03	14560	2.18	
4 Z	2/24/09	50 U	310	6700	34 J	2900	15	5 U	22	970	0.5 U	3500	2300	85	4.46E+08	1.18E+09	3.72E+08	4.23E+07	6.35	-90.6	1.25	11132	2.6	
AMW-4	3/31/09	50 U	180	6500	32 J	3800	49	5 U	12	950	0.25 U		1800	57	5.13E+08				6.31	-113.3	0.96	9975	2.2	
¥	4/29/09		250	4100	33 J	5400	59	1 J	14	860	0.5 U		1800	49		7.80E+08			6.26	-134.6	3.66	8503	3.3	
	5/28/09	36 U	220	900	30 J	9200	110	5 U	19	830	0.18 J		1700	44	1.37E+08			3.00E+05	6.17	-95.9	0.15	7791	>3.3	
	6/24/09	63 U	150	380	26 J	7000	110	1 J	16	720	0.5 U	2200	1700	49	7.48E+07			6.30E+05	6.65	-98.4	1.16	7784	3.07	
<u> </u>	10/16/09	25 U	63	150	26	3900	680	0.7 J	7	1500	0.25 U		1700	420	48800000		2.90E+06		5.41	-256.1	1.38	6.698	>3.3	
)e 2	4/8/08		610	8.8	2.4 J	5 U	5 U	5 U	41	630	0.13 J	6900	3200	44	ND	ND	ND	ND	6.83	156.5	0.67	18399	0	
Zone	11/6/08	71 U	10000	1700	71 U	71 U	5 U	5 U	10	580	0.25 U	3300	2200	36	ND	ND	ND	ND	6.69	-51.1	1.08	10440	1.42	
4	4/29/09	71 U	81	5500	26 J	8100	12	5 U	12	790	0.25 U	1800	1600	53	1.43E+09		2.00E+09		6.2	-88.9	2.61	7541	3.3	
AMW-4	6/24/09	71 U	44 J	62 J	71 U	7600	65	1 J	14	650	0.25 U	1700	1600	44	1.91E+08	1.60E+08	2.80E+07	3.60E+05	6.61	-66.5	1.08	6755	2.94	
∢	10/16/09																		5.43	-245.6	1.26	6.084	>3.3	

ACTIVE C		e												_	Ġ	i,	'n	· ·						
Monitorin	g Data	chloroethene	e	Ф	Φ									(ygen	de	de	des	de						
Summary		et	Jer	nen	nen	ide								δ	. <u>io</u>	. <u>.</u>		. <u>ē</u>				>-	_	
NAVFAC I	Naval	orc	et	eth	et .	lon			40	_					ŏ≤	Ö	Ö	Ö				ivity	<u>e</u>	ηts
Weapons		동	orc	2- oro	7,7 oro	<u>ა</u>	Φ	Φ	ane.	Ę.	d)	Φ	ge	l ië P	N 00	00	00	00				rot	S	Je
Station - S	Site 70	Гетга	Trichlor	is-1,2- ichlor	-su SHC	/inyl	Jen	Jan	Ĕ	i <u>i</u>	rat	ulfate	Chloride	men.	behal 6S ri	ha A	hal cA	hal Y	_	₽		Ď.	Ē	Ę
Seal Beac	h, CA	<u>e</u>	Ë	cis Dic	tra	Ξ	盂	古	Me	₹	ž	Su	်	Chemical Demand	De 16	tce De	De De	Vcr	H	ORF	DG	ပိ	Fe	ပိ
	Units:	μg/L	μg/L	μg/L	μg/L	μg/L	μg/L	μg/L	μg/L	mg/L	mg/L	mg/L	mg/L	mg/L		gene c	opies/L			mV	mg/L	µmhos/cm	mg/L	
က	4/8/08	10 U	1200	49	4.1 J	10 U	5 U	5 U	19	640	0.5 U	7000	2900	38	ND	ND	ND	ND	7	93	3.3	18109	0	
e e																								
Zor	11/6/08	36 U	7900	1100	24 J	36 U	5 U	5 U	5 U	600	0.25 U	3300	2100	38	ND	ND	ND	ND	6.78	-2.1	1.47	10310	0.37	
4 2	4/29/09	50 U	4200	7400	26 J	2000	5 U	5 U	6	880	0.07 J	2400	1800	40	1.08E+06	6.10E+06	2.50E+06	7.90E+02	6.26	14.5	1.5	7898	1.02	
⋚	6/24/09	36 U	1900	3400	22 J	4600	11	5 U	8	800	0.5 U	2100	1600	40	1.39E+07	1.50E+07	2.50E+06	1.20E+04	6.77	12.4	1.86	8306	0.94	
AMW-																								
	10/16/09	36 U	540	410	24 J	5200	47	0.4 J	5	1100	0.25 U	1300	1500	30	15000000	1.06E+07	1.30E+06	9.90E+04	5.05	-78.7	1.57	7090	0.25	Wall Daweters d
V-4 e 4	4/8/08	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	7.33	141.5	1.17	24495	NM	Well Dewatered
AMW. Zone																								
∢ N																								
	4/9/08	1.7 J	710	14	4.2 J	5 U	5 U	5 U	28	450	0.16 J	3600	2900	42	ND	ND	ND	ND	6.97	-82.9	0.85	13510	0.24	
	5/15/08	20 U	2900	200	11 J	20 U	5 U	5 U	19	490	1 U	4300	3200	48	1.83E+02	4.10E+01	3.80E+01	ND	7.03	-112.2	0.31	15720	0	
_	9/3/08	14 J	4600	560	13 J	42 U	5 U	5 U	10	580	0.5 U	4700	3000	47	1.04E+05	2.43E+04	ND	ND	6.09	14.9	5.51	1478	0	
<u>e</u>	11/6/08	25 U	5200	650	18 J	25 U	5 U	5 U	14	590	0.5 U	4900	3100	47	ND	ND	ND	ND	6.9	19.6	0.98	14720	0	
Zor	1/29/09	36 U	6400	1500	19 J	36 U	5 U	5 U	11	640	0.5 U	4800	2900	51	2.26E+03	3.75E+02	2.39E+01	9.80E+01	6.83	-43.9	2.11	15340	0	
-52	2/24/09	36 U	5800	2800	22 J	57	5 U	5 U	17	730	0.5 U	5200	2900	49	1.30E+05	3.97E+05	1.11E+05	2.04E+03*	6.62	-93.2	1.27	13373	0	
≥	3/31/09	36 U	3000	4700	24 J	1200	8	5 U	14	980	0.5 U	4200	2400	36	3.16E+06	1.10E+07	3.50E+06	4.04E+04	6.42	-85.4	0.79	13140	0.06	
AM	4/29/09	36 U	2600	5500	29 J	3500	19	5 U	17	1000	1 U	4300	2300	44	3.63E+06	1.45E+07	5.90E+06	3.40E+03	6.41	-10.9	1.78	12130	0	
	5/28/09	36 U	1300	3100	35 J	6100	61	5	18	1000	0.5 U	3800	2100	59	3.54E+07	3.13E+07	5.50E+06	1.60E+05	6.25	10.2	0.13	10690	0	
	6/24/09	36 U	1100	1500	20 J	6000	86	5 U	19	940	0.5 U	3600	1900	83	3.96E+06	4.45E+06	7.80E+05	1.20E+04	6.7	-27.8	1.1	10620	0	
	10/16/09	36 U	940	1400	27 J	5400	160	1 J	15	870	0.5 U	3700	2000	50	1.55E+03*	ND	2.6E+01*	ND	5.21	-104.5	1.54	10230	0	
2	4/9/08	2.1 J	1100	21	4.2 J	8.3 U	5 U	5 U	48	630	0.18 J	7100	3100	40	ND	ND	ND	ND	6.83	15.3	0.71	18118	0	
ne	11/6/08	31 U	5700	5200	55	18 J	5 U	5 U	13	710	0.5 U	3800	2300	42	ND	ND	ND	ND	6.68	-20	1.23	12550	3.23	
Zoı																								
-5	4/29/09	36 U	81	9000	28 J	5200	40	1 J	15	820	0.5 U	2000	1600	70	9.54E+08	3.81E+09	1.60E+09	3.10E+07	6.39	-90.1	2.01	8896	NM	
	6/24/09	50 U	91	290	50 U	6900	63	5 U	14	660	0.25 U	1900	1500	55	3.90E+08	4.90E+08	8.60E+07	6.10E+06	6.74	-49.1	1.66	7485	3.18	
A	10/16/09	25 U	27	54	25 U	2600	950	0.3 J	9	1400	0.25 U	1200	1500	350	369000000	2.54E+08	1.00E+07	7.20E+06	5.42	-184.2	1.02	6912	>3.3	
ròω	4/9/08	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	7.35	73.3	2.53	18656	NM	Well Dewatered
-W	11/6/08	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	6.89	19.1	1.55	16080	NM	Well Dewatered
AMW. Zone	4/29/09	36 U	1900	4200	17 J	170	5 U	5 U	3 J	NS	NS	NS	NS	NS	NS	NS	NS	NS	6.51	99.2	4	15420	NM	Well dewatered
τċ 4	4/9/08	1.3 U	170	7	1.3	1.3 U	5 U	5 U	5 U	NS	NS	NS	NS	NS	NS	NS	NS	NS	7.46	87.1	1.17	21518	NM	Well Dewatered
- Š	11/6/08	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NA	NA	NA	NA	NM	Well Dewatered
AMW. Zone	4/29/09	NS	NS	NS	NS	NS			NS			NS		NS	NS	NS	NS	NS	6.9	49.9	7.49	20180	NM	Well Dewatered
	4/9/08	1000 U	140000	660 J	1000 U	1000 U	5 U	5 U	40	600	0.35 J	3300	2900	58	ND	ND	ND	ND	6.65	111.8	1.04	12926	0	vveli Dewalered
		420 U		790																				
	5/14/08		150000		420 U	420 U	5 U	5 U	47	570	0.12 J	3000	2200	42	ND 209	NS ND	NS ND	NS	6.93	46.1	0.8	11.829	0	
	9/3/08 9/3/2008 - K	1300 U 1300 U	190000 190000	1300 U 450 J	1300 U 1300 U	1300 U	5 U	5U	25 23	530 530	0.25 U 0.25 U	3200	2200 2200	55 51	ND	ND ND	ND	ND ND	5.98	255.1	2.51	9607	U	
						1300 U	5 U	5 U		540		3200		47	ND				6.70	200.0	0.0	0011	0	
9-	11/5/08 1/29/09	1000 U	120000 160000	710 J 1200	1000 U 1000 U	1000 U	5 U	5 U	33 47	580	0.1 U 0.5 U	3300 3300	2200 1900	51	ND	ND ND	ND ND	ND ND	6.72 6.73	200.8 -35.2	0.9 1.49	9811 10990	0	
						1000 U	5 U										ND							
₹	2/24/09	1000 U	130000	840 J	1000 U	1000 U	5 U	5	62 52	520 630	0.5 U	3700	2200	55	1.57E+04	1 34E±04		ND 8 26F±02*	6.71	17.4	0.65	10332	0	
	3/31/09 4/29/09	500 U 500 U	77000 70000	840 1100	500 U 500 U	500 U	5 U	5 U 3 J	52 57	630 630	0.26 J	3700 4300	2100 2400	42 49	2.41E+03* ND	1.34E+04 2.15E+02	ND ND	8.26E+02* ND	6.86 6.6	-10.7	5.09 0.2	0.042	0	
						500 U	5 U				0.56 J									78.7		12650		
	5/28/09	500 U 420 U	52000	1500 3600	500 U 420 U	500 U	5 U	2 J 5 U	49	650	0.5 U	4300 4600	2500 2300	55 51	1.52E+04 3.86E+03*	7.72E+03	4.20E+03	7.20E+02 ND	6.48 6.85	6 50	0.12 0.53	12250	0	
	6/24/09 10/16/09		53000 30000	4900	200 U	310 J 4100	5 J 15	0.3 J	28 21	720 720	1 U 0.5 U	3800	2000	63		1.8E+03*	4.1E+02* ND	ND ND	5.2	-33	2.29	12990 10440	NM	
-	4/8/08	200 U 20 J	10000	1900	44 J	4100 48 J	5 U	8	140	560	0.5 U 0.14 J	1600	1800	28	ND 448	ND 1.84E+02	1.40E+02	ND ND	6.71	225.5	128.1*	8060	0	Composite of AEW 1 & 2; DO out of range
	5/14/08	100 U	30000	2000	62 J		5 U	6	92			1500	1700	28	272	1.84E+02 1.10E+02	4.00E+01	ND ND	7.07	140.6	1.42	8060		Extraction system running. Grab sample. Composite of AEW 1 & 2.
			9000	420		27 J	5 U			910	0.77			34	3.06E+04		4.00E+01 2.08E+03		5.43			7913		Extraction system running. Grab sample. Composite of AEW 1 & 2. Only AEW2 running.
nt)	9/3/08	20 J			71 U	71 U		5 U	60	540 570	0.25 U	1800	2100			3.40E+03		ND ND		204.3	3.83		0 NM	Only Acviz Turring.
effluent)	11/5/08	28 J	35000	1700	47 J	29 J	5 U	6	100	570	0.25 U	1500	1100	34	1.60E+04	ND 1 10E + 0E	ND 2.00E+04		NA	NA	NA	NA	INIVI	
eff[11/5/2008 - K	86 J	33000	2200	200 U	200 U	5 U	6	100	560	0.1 U	1800	2200	36	1.11E+05	1.10E+05	2.00E+04	ND	NA	NI A	NIA	NI A	NIP 4	
<u>ā</u>	1/27/09	200 U	22000	930	200 U	200 U	5 U	8	160	530	0.1 U	1700	1800	34	2.17E+04 2.27E+04	2.17E+04	2.75E+03		NA	NA	NA	NA	NM	
× .	1/27/09 - K	200 U	23000	980	200 U	200 U	5 U	6	120	550	0.1 U	1700	1700	30			2.32E+03	ND	0.44	44.0	4.00	2000	•	
tior	2/24/09	200 U	21000	1100	200 U	200 U	5 U	7	120	510	0.25 U	1700	1700	30	2.36E+04	5.27E+04	2.06E+04		6.44	44.3	1.66	6806	0	
ad	2/24/2009 - K	200 U	27000	1500	200 U	200 U	5 U	8	120	530	0.29	1800	1700	30	2.42E+04	9.65E+04	1.76E+04	ND 5.005 : 04*	0.00	07.0	4.00	7007	0.01	
×tr	3/31/09	200 U	22000	1000	200 U	200 U	5 U	10	140	520	0.61	1700	1500	30	2.18E+03*	6.48E+03	1.95E+03*		6.68	-27.3	4.96	7267	0.01	
ğ	3/31/2009 - K	200 U	23000	1000	200 U	200 U	5 U	10	140	510	0.63	1800	1500	34	7.60E+02*	3.15E+03*	7.77E+02*	ND	0.45	4.40.5	0.00	7400	0	
ine	4/28/09	39 J	6500	330	11 J	13 J	5 U	5 J	54	550	0.52	1800	1600	30	ND	ND	ND	ND	6.45	140.5	2.86	7136	0	
(combined	4/28/09-K	200 U	21000	950	200 U	200 U	5 U	5	67	550	0.41	1900	1700	32	ND 4.405+04	ND	ND 0.705 : 00	ND	0.40	04.0	0.01	0700	0	
, (S)	5/28/09	50 U	9000	730	15 J	20 J	5 U	13	150	550	0.48	1800	1600	19	1.19E+04	3.88E+03	2.70E+03	ND	6.48	24.2	3.01	6762	0	
	5/28/2009-K	100 U	17000	1500	100 U	23 J	5 U	5 J	59	560	0.63	1800	1500	23	8.23E+03		3.10E+03		0.55	45.5		00:5	0.5-	
AEW	6/24/09	50 U	8000	920	17 J	23 J	5 U	5	50	520	0.62	1800	1500	61	2.79E+02*	4.10E+02	5.3E+01*	ND	6.86	15.6	1.92	6646	0.25	
	6/24/09-K	170 U	16000	1900	170 U	43 J	5 U	13	130	530	0.72	1700	1400	68	ND	1.2E+02*	5.8E+01*	ND		40 =	0 ==	2055		
	10/15/09	50 U	16000	2100	35 J	510	2 J	5	72	560	0.24 J	2000	1600	35	6.82E+03	2.73E+03*	4.6E+02*	2.0E+01*	5.54	16.7	3.73	6352	0	
	10/15/2009-K	170 U	15000	2200	42 J	510	2 J	5	70	550	0.27	1700	1400	35	3.71E+03*	1.39E+03*	2.8E+02*	3.7E+01*						
Notes: K - Duplicate	comple					micromhos pe	r centimeter																	
v - nanilcate	adilluit				μg/L - mcrog	ams per liter																		

K - Duplicate sample
J - estimated value
U - nondetect (detection limit is indicated)

NA - not analyzed ND - not detected NS - not sampled

* - indicates that the value presented is below the reporting limit.

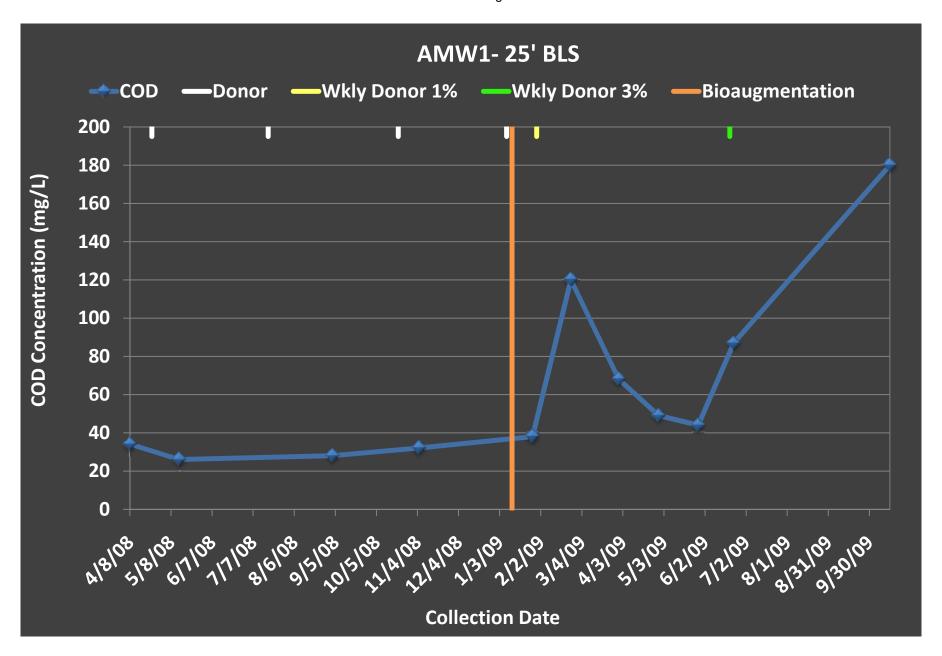
umnos/cm - micromnos per centimeter
µg/L - mcrograms per liter
mg/L - milligrams per liter
mV - millivolts
ORP - oxidation reduction potential
DO - dissolved oxygen
> - greater than
NM - not measured

February 2009	TCE (ppb)	$\delta^{13}C$	cDCE	$\delta^{13}C$	VC	$\delta^{13}C$	Ethene	δ^{13} C eth	2-butanone	chloroform	chlor/DCE
SB-AEW	21000	-24.3	1100	-26.8							
SB-AMW1-25'	27000		7000	-24.6							
SB-AMW2-25'	97		9500	-22.9	940	-43.0				75	0.01
SB-AMW3-Z1	6000	-24.5	2300	-24.5	210					55	0.02
SB-AMW4-Z1	310	-17.9	6700	-36.5	2900		59				
SB-AMW5-Z1	5800	-23.7	2800	-25.0	57					100	0.04
SB-AMW6-25'	130000	-23.9		-31.7		-28.5					
SB-PIW1-25'	42		4	-27.0						120	32.43
SB-PIW2-25'	12	-26.1	3	-27.9	23				110	3	1.00
SB-PIW3-25'	2	-23.5	1	-25.4	5				160	-	
SB-PMW1-25'	1700	-28.0	79		_					74	0.94
SB-PMW2-25'	1800	20.0	43								0.0 .
SB-PMW3-Z1	41000	-23.3									
SB-PMW4-Z1	41000	-23.3									
SB-PMW5-Z1	40000	-23.5									
			900	10.2	E /					200	0.25
SB-PMW6-25'	2100	-23.3	800	-19.2	54					280	0.35
SB-PMW7-25'	6500	20.0	4500	25.2						690	0.01
SB-PMW8-25'	1100	-20.0	4500	-25.2						30	0.01
SB-PMW9-25'	96	-22.6	6								
Average of all		-23.5		-26.4		25.0					
Average of all		-23.5 -22.9		-20.4 -27.4		-35.8 -35.8					
AEW/AMW Ave						-33.6					
PIW/PMW Ave		-23.9		-24.9							
April 2009	TCE (ppb)	$\delta^{13}C$	cDCE	δ ¹³ C	VC	$\delta^{13}C$	Ethene	δ^{13} C eth	2-butanone	chloroform	chlor/DCE
SB-AEW	6500	-22.7	330	-24.9							
SB-AMW1-25'	880	-22.2	3000	-15.4	1100	-35.4	17				
SB-AMW2-25'	540	-9.9	1700		6900	-30.2	16				
					0000	00.0					
SB-AMW3-Z1	1200	-22.5	2700	-15.9	3900	-26.2	58				
SB-AMW3-Z1 SB-AMW3-Z2	1200 780	-22.5 -22.6	2700 6500	-15.9 -13.9	1900	-26.2 -40.2	58 13			99	0.02
										99	0.02
SB-AMW3-Z2	780		6500		1900	-40.2				99	0.02
SB-AMW3-Z2 SB-AMW3-Z3	780 130	-22.6	6500 11000	-13.9	1900 190	-40.2 -47.0	13			99	0.02
SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1	780 130 250	-22.6	6500 11000 4100	-13.9 -9.6	1900 190 5400	-40.2 -47.0 -32.5	13 14			99	0.02
SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z2	780 130 250 81	-22.6 -16.8	6500 11000 4100 5500	-13.9 -9.6 -6.0	1900 190 5400 8100	-40.2 -47.0 -32.5 -33.0	13 14			99 72	0.02
SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z2 SB-AMW4-Z3	780 130 250 81 4200	-22.6 -16.8 -21.9	6500 11000 4100 5500 7400	-13.9 -9.6 -6.0 -19.4	1900 190 5400 8100 2000	-40.2 -47.0 -32.5 -33.0 -37.8	13 14 12				
SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z2 SB-AMW4-Z3 SB-AMW5-Z1	780 130 250 81 4200 2600	-22.6 -16.8 -21.9	6500 11000 4100 5500 7400 5500	-9.6 -6.0 -19.4 -19.3	1900 190 5400 8100 2000 3500	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8	13 14 12 19				
SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z2 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z2	780 130 250 81 4200 2600 81	-22.6 -16.8 -21.9 -21.5	6500 11000 4100 5500 7400 5500 9000	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7	1900 190 5400 8100 2000 3500 5200	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0	13 14 12 19			72	0.01
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SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW6-25'	780 130 250 81 4200 2600 81 1900 70000	-22.6 -16.8 -21.9 -21.5	6500 11000 4100 5500 7400 5500 9000 4200 1100	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8	1900 190 5400 8100 2000 3500 5200 170	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0	13 14 12 19	-17.2	180	72 77 120	0.01 0.02 0.11
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SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW6-25' SB-PIW1-25' SB-PIW2-25' SB-PIW2-25' SB-PIW3-25' SB-PMW3-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z4 SB-PMW5-Z1	780 130 250 81 4200 2600 81 1900 70000 26 44 1400 280 45000 42000 3400 7900 44000	-22.6 -16.8 -21.9 -21.5 -22.9 -23.1 -28.3 -21.7 -21.9 -22.0 -23.4 -23.5 -24.1 -24.0	6500 11000 4100 5500 7400 5500 9000 4200 1100 23 1 65 7 260 520 41 57	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8	1900 190 5400 8100 2000 3500 5200 170	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0 -39.9	13 14 12 19	-17.2		72 77 120 11 58 140 48 220 5600 3300 130 2900	0.01 0.02 0.11 0.48 0.89 21.54 0.18 0.42 136.59 57.89
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SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z3 SB-PIW1-25 SB-PIW2-25 SB-PIW2-25 SB-PIW2-25 SB-PMW1-25 SB-PMW1-25 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW6-Z5 SB-PMW6-Z5 SB-PMW6-Z5 SB-PMW6-Z5 SB-PMW6-Z5 SB-PMW6-Z5 SB-PMW6-Z5 SB-PMW6-Z5 SB-PMW6-Z5	780 130 250 81 4200 2600 81 1900 70000 26 44 1400 280 45000 42000 3400 7900 44000 7100 6900 740 5800	-22.6 -16.8 -21.9 -21.5 -21.5 -22.9 -23.1 -28.3 -21.7 -21.9 -22.0 -23.4 -23.5 -24.1 -24.0 -23.5 -17.9 -22.0	6500 11000 4100 5500 7400 5500 9000 4200 1100 23 1 65 7 260 520 41 57	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8 -25.1	1900 1900 5400 8100 2000 3500 5200 170 4 6 6	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0 -39.9	13 14 12 19 40	-27.5	170	72 77 120 11 58 140 48 220 5600 3300 130 2900 3800 230 570	0.01 0.02 0.11 0.48 0.89 21.54 0.18 0.42 136.59 57.89 51.79 56.72 0.64
SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z3 SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PMW1-25' SB-PMW1-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW6-25' SB-PMW7-25' SB-PMW7-25' SB-PMW7-25' SB-PMW7-25' SB-PMW7-25' SB-PMW7-25'	780 130 250 81 4200 2600 81 1900 70000 26 44 1400 280 45000 42000 3400 7900 44000 7100 6900 740 5800 1500	-22.6 -16.8 -21.9 -21.5 -22.9 -23.1 -28.3 -21.7 -21.9 -22.0 -23.5 -24.1 -24.0 -23.5 -17.9 -22.0 -15.5	6500 11000 4100 5500 7400 5500 9000 4200 1100 23 1 65 7 260 520 41 57 56 67 360 1700 470	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8 -25.1	1900 190 5400 8100 2000 3500 5200 170 4 6 6	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0 -39.9 -28.0	13 14 12 19 40	-27.5	170	72 77 120 11 58 140 48 220 5600 3300 130 2900 3800 230 570	0.01 0.02 0.11 0.48 0.89 21.54 0.18 0.42 136.59 57.89 51.79 56.72 0.64 0.34
SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-PW5-Z3 SB-PW1-25' SB-PIW1-25' SB-PIW2-25' SB-PIW2-25' SB-PMW3-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z3 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW6-25' SB-PMW6-25' SB-PMW6-25' SB-PMW6-25' SB-PMW9-25'	780 130 250 81 4200 2600 81 1900 70000 26 44 1400 280 45000 42000 3400 7900 44000 7100 6900 740 5800 1500 110	-22.6 -16.8 -21.9 -21.5 -21.5 -22.9 -23.1 -28.3 -21.7 -21.9 -22.0 -23.4 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -22.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.0 -23.5 -24.0 -23.5 -24.0 -23.5 -24.0 -23.5 -24.0 -24.0 -25.0 -27.0 -2	6500 11000 4100 5500 7400 5500 9000 4200 1100 23 1 65 7 260 520 41 57 56 67 360 1700 470	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8 -25.1 -23.8 -9.5 -22.8	1900 190 5400 8100 2000 3500 5200 170 4 6 6	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0 -39.9 -28.0	13 14 12 19 40	-27.5	170	72 77 120 11 58 140 48 220 5600 3300 130 2900 3800 230 570	0.01 0.02 0.11 0.48 0.89 21.54 0.18 0.42 136.59 57.89 51.79 56.72 0.64 0.34
SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z2 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-PW5-Z3 SB-PW2-25 SB-PIW2-25 SB-PIW3-25 SB-PIW3-25 SB-PMW1-25 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z5 SB-PMW5-Z5 SB-PMW8-Z5 SB-PMW8-25 SB-PMW9-25	780 130 250 81 4200 2600 81 1900 70000 26 44 1400 280 45000 42000 3400 7900 44000 7100 6900 740 5800 1500 110	-22.6 -16.8 -21.9 -21.5 -22.9 -23.1 -28.3 -21.7 -21.9 -22.0 -23.5 -24.1 -24.0 -23.5 -17.9 -22.0 -23.5 -17.9 -22.0	6500 11000 4100 5500 7400 5500 9000 4200 1100 23 1 65 7 260 520 41 57 56 67 360 1700 470	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8 -25.1 -23.8 -9.5 -22.8 -17.1 -16.6	1900 190 5400 8100 2000 3500 5200 170 4 6 6	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0 -39.9 -28.0	13 14 12 19 40	-27.5	170	72 77 120 11 58 140 48 220 5600 3300 130 2900 3800 230 570	0.01 0.02 0.11 0.48 0.89 21.54 0.18 0.42 136.59 57.89 51.79 56.72 0.64 0.34
SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-PW5-Z3 SB-PW1-25' SB-PIW1-25' SB-PIW2-25' SB-PIW2-25' SB-PMW3-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z3 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW6-25' SB-PMW6-25' SB-PMW6-25' SB-PMW6-25' SB-PMW9-25'	780 130 250 81 4200 2600 81 1900 70000 26 44 1400 280 45000 42000 3400 7900 44000 7100 6900 740 5800 1500 110	-22.6 -16.8 -21.9 -21.5 -21.5 -22.9 -23.1 -28.3 -21.7 -21.9 -22.0 -23.4 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -22.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.0 -23.5 -24.0 -23.5 -24.0 -23.5 -24.0 -23.5 -24.0 -24.0 -25.0 -27.0 -2	6500 11000 4100 5500 7400 5500 9000 4200 1100 23 1 65 7 260 520 41 57 56 67 360 1700 470	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8 -25.1 -23.8 -9.5 -22.8	1900 190 5400 8100 2000 3500 5200 170 4 6 6	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0 -39.9 -28.0	13 14 12 19 40	-27.5	170	72 77 120 11 58 140 48 220 5600 3300 130 2900 3800 230 570	0.01 0.02 0.11 0.48 0.89 21.54 0.18 0.42 136.59 57.89 51.79 56.72 0.64 0.34

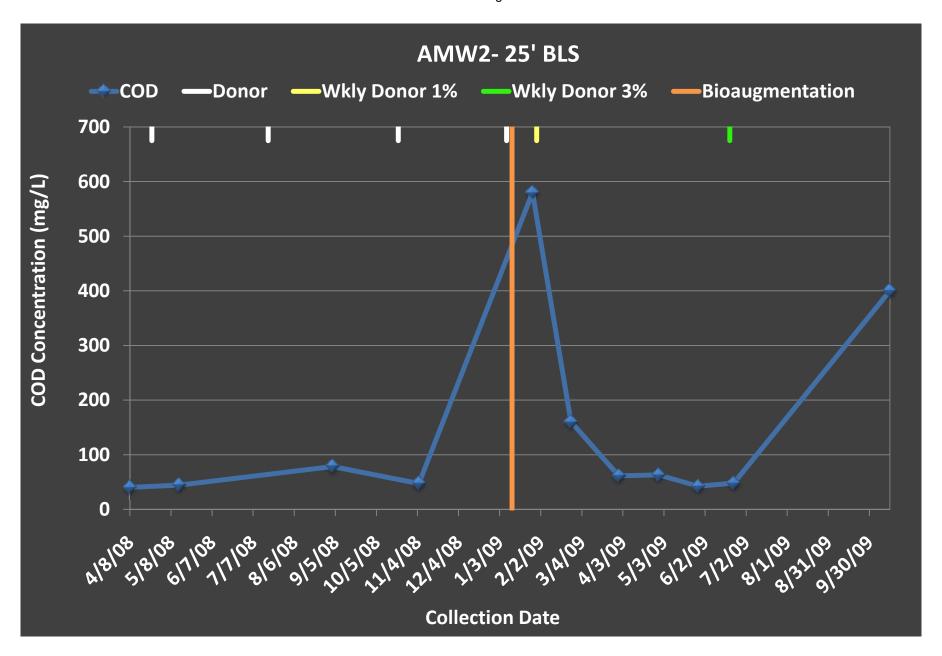
luna 2000	TCE (nnh)	$\delta^{13}C$	•DCE	$\delta^{13}C$	VC	δ^{13} C	Ethene	S ¹³ C oth	2-butanone chloroform	oblor/DCE
June 2009	TCE (ppb)		cDCE		VC		Ethene	o Celli	2-butanone chlorolom	CHIOI/DCE
SB-AEW	8000	-24.4	920	-25.6		-37.2				
SB-AEW K	16000	-23.9	1900	-23.0						
SB-AMW1-25'	1900	-23.2	1200	-14.5	2800	-27.8	22			
SB-AMW2-25'	280	-5.2	440	6.8	6400	-26.8	33			
SB-AMW2-25' K	290	-4.3	430	6.6	6500	-26.1	34			
SB-AMW3-Z1	580	-21.9	690	-14.7	7000	-22.9	200			
SB-AMW3-Z2	520	-20.0	1400	-11.7	5500	-26.4	77			
SB-AMW3-Z3	87	-18.6	2000	-8.2	5300	-26.9	27			
SB-AMW4-Z1	150	-18.2	380	-4.8	7000	-26.9	110			
SB-AMW4-Z2					7600	-26.3	65			
SB-AMW4-Z3	1900	-22.2	3400	-16.4	4600	-27.4	11			
SB-AMW5-Z1	1100	-22.3	1500	-15.5	6000	-25.2	86			
SB-AMW5-Z2	91	-8.6	290	4.0	6900	-27.0	63			
			3600		310	-27.3	05			
SB-AMW6-25'	53000	-23.5	3000	-19.8	310	-27.5				
SB-PIW1-25'	13		9		19	-19.3		-12.6		
						-19.3		-12.0		
SB-PIW2-25'	17		3		8					
SB-PIW3-25'	54		2		4					
SB-PMW1-25'	1400	-28.0	69	-30.5			6			
SB-PMW2-25'	4400	-23.4								
SB-PMW3-Z1	47000	-24.3	190		310					
SB-PMW3-Z2	1400	-24.1	11							
SB-PMW4-Z1	30000	-23.3	4500	-25.2						
SB-PMW4-Z3	2000	-24.2	22							
SB-PMW4-Z4	6700	-24.3	59							
SB-PMW5-Z1	39000	-24.6	380		310					
SB-PMW5-Z2	4600	-23.9	69							
SB-PMW5-Z3	5600	-24.6	53							
SB-PMW6-25'	790	-16.2	460	-20.5	120	-18.5	190	-21.0		
SB-PMW6-25' K		-16.9	490	-20.3	130	-20.6	170	-21.7		
SB-PMW7-25'	190	-8.7	96	1.3	590	-18.3	28	-34.1		
SB-PMW8-25'	710	-12.6	180	-7.3	250	-8.8	350	-26.5		
SB-PMW9-25'	19	-18.8	19		37	-4.2	110	-23.4		
Average of all		10.0		10.6		22.4		22.2		
Average of all		-19.8		-12.6		-23.4		-23.2		
AEW/AMW Ave		-18.2		-10.5		-27.2				
PIW/PMW Ave		-21.2		-17.1		-15.0		-23.2		
		40		40		40		40		
October 2009	TCE (ppb)	$\delta^{13}C$	cDCE	$\delta^{13}C$	VC	$\delta^{13}C$	Ethene	δ^{13} C eth	2-butanone chloroform	chlor/DCE
October 2009 SB-AEW	TCE (ppb)	δ ¹³ C	cDCE	δ ¹³ C -19.8	VC	δ ¹³ C -30.0	Ethene	δ ¹³ C eth	2-butanone chloroform	chlor/DCE
SB-AEW	TCE (ppb)	-24.2	cDCE	-19.8	VC	-30.0	Ethene		2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25'	TCE (ppb)	-24.2 -20.6	cDCE	-19.8 -18.2	VC	-30.0 -19.6	Ethene	δ ¹³ C eth	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25'	TCE (ppb)	-24.2	cDCE	-19.8	VC	-30.0 -19.6 -24.5	Ethene		2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1	TCE (ppb)	-24.2 -20.6	cDCE	-19.8 -18.2	VC	-30.0 -19.6	Ethene		2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2	TCE (ppb)	-24.2 -20.6	cDCE	-19.8 -18.2	VC	-30.0 -19.6 -24.5 -21.4	Ethene	-44.2	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW3-Z3	TCE (ppb)	-24.2 -20.6	cDCE	-19.8 -18.2	VC	-30.0 -19.6 -24.5 -21.4	Ethene	-44.2 -28.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1	TCE (ppb)	-24.2 -20.6	cDCE	-19.8 -18.2	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5	Ethene	-44.2	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z2	TCE (ppb)	-24.2 -20.6	cDCE	-19.8 -18.2	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3	Ethene	-44.2 -28.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW4-Z3 SB-AMW4-Z2 SB-AMW4-Z2 SB-AMW4-Z3	TCE (ppb)	-24.2 -20.6 -6.9	cDCE	-19.8 -18.2 -21.2	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1	Ethene	-44.2 -28.7 -40.4	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW3-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW4-Z3 SB-AMW5-Z1	TCE (ppb)	-24.2 -20.6	cDCE	-19.8 -18.2	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3	Ethene	-44.2 -28.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2	TCE (ppb)	-24.2 -20.6 -6.9	cDCE	-19.8 -18.2 -21.2	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1	Ethene	-44.2 -28.7 -40.4	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z3	TCE (ppb)	-24.2 -20.6 -6.9	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2	TCE (ppb)	-24.2 -20.6 -6.9	cDCE	-19.8 -18.2 -21.2	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1	Ethene	-44.2 -28.7 -40.4	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW3-21 SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW6-25'	TCE (ppb)	-24.2 -20.6 -6.9 -23.4	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z2 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW6-25' SB-PIW1-25'	TCE (ppb)	-24.2 -20.6 -6.9	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z5' SB-PIW1-25' SB-PIW1-25' SB-PIW2-25'	TCE (ppb)	-24.2 -20.6 -6.9 -23.4	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z5' SB-PIW1-25' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25'	TCE (ppb)	-24.2 -20.6 -6.9 -23.4	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z5' SB-PIW1-25' SB-PIW1-25' SB-PIW2-25'	TCE (ppb)	-24.2 -20.6 -6.9 -23.4	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z5' SB-PIW1-25' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25'	TCE (ppb)	-24.2 -20.6 -6.9 -23.4	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z2 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z5' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25'	TCE (ppb)	-24.2 -20.6 -6.9 -23.4 -23.9 -22.9	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW3-Z1 SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z3 SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PMW1-25' SB-PMW1-25' SB-PMW1-25' SB-PMW3-25' SB-PMW1-25' SB-PMW1-25'	TCE (ppb)	-24.2 -20.6 -6.9 -23.4 -23.9 -22.9	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z5' SB-PIW1-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z5'	TCE (ppb)	-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z5 SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PMW1-25'	TCE (ppb)	-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.2	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW3-Z1 SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z5' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z5'	TCE (ppb)	-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.2 -24.4	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z5' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z3 SB-PMW3-Z3 SB-PMW3-Z3 SB-PMW3-Z3 SB-PMW3-Z3	TCE (ppb)	-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.2 -24.4 -23.3	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z5' SB-PIW3-Z5' SB-PIW3-Z5' SB-PIW3-Z5' SB-PIW3-Z5' SB-PIW3-Z5' SB-PIW3-Z5' SB-PIW3-Z5' SB-PMW1-Z5' SB-PMW1-Z5' SB-PMW1-Z5' SB-PMW1-Z5' SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z3	TCE (ppb)	-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.2 -24.4 -23.3 -24.3	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z5' SB-PIW1-25' SB-PIW2-25' SB-PIW3-Z5' SB-PIW3-Z1 SB-PMW3-Z1 SB-PMW4-Z1 SB-PMW4-Z3 SB-PMW4-Z3 SB-PMW4-Z3	TCE (ppb)	-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.2 -24.4 -23.3 -24.3 -24.3	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW3-21 SB-AMW3-21 SB-AMW3-23 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z5' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25'		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -24.4 -23.2 -24.4 -23.3 -24.3 -24.7 -23.3	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8 -27.5	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z2 SB-AMW5-Z2 SB-PW5-Z5' SB-PIW1-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z5' SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW5-Z1		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.3 -24.3 -24.3 -23.3 -24.3 -23.3 -23.5	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z3 SB-AMW5-Z5' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW4-Z4 SB-PMW4-Z4 SB-PMW5-Z1		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.3 -24.3 -24.3 -24.3 -23.5 -23.5 -23.1	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8 -27.5	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z5' SB-PIW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW4-Z4 SB-PMW4-Z4 SB-PMW4-Z4 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW5-Z1		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.2 -24.4 -23.3 -24.7 -23.3 -24.7 -23.5 -23.1 -23.2	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8 -22.8 -22.2 -23.1	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8 -27.5	Ethene	-44.2 -28.7 -40.4 -48.7 -42.7 -45.8 -53.6	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z5' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-Z5' SB-PIW3-Z5' SB-PIW3-Z5' SB-PIW3-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.3 -24.3 -24.3 -24.3 -23.5 -23.5 -23.1	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8 -27.5	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW3-Z1 SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z3 SB-PIW1-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW6-Z5' SB-PMW6-Z5' SB-PMW6-Z5'		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.3 -24.7 -23.3 -24.7 -23.3 -24.7 -23.5 -24.7	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8 -22.8 -22.2 -23.1	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8 -27.5	Ethene	-44.2 -28.7 -40.4 -48.7 -42.7 -45.8 -53.6	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z5' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW4-Z1 SB-PMW4-Z3 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z5' SB-PMW5-Z5 SB-PMW5-Z5 SB-PMW5-Z5 SB-PMW7-25' SB-PMW7-25' SB-PMW7-25'		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.2 -24.4 -23.3 -24.7 -23.3 -24.7 -23.5 -23.1 -23.2	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8 -22.8 -22.2 -23.1	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8 -27.5	Ethene	-44.2 -28.7 -40.4 -48.7 -42.7 -45.8 -53.6	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW3-Z1 SB-AMW3-Z1 SB-AMW3-Z1 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z3 SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW6-Z5' SB-PMW6-Z5' SB-PMW6-Z5'		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.3 -24.7 -23.3 -24.7 -23.3 -24.7 -23.5 -24.7	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8 -22.8 -22.2 -23.1	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8 -27.5	Ethene	-44.2 -28.7 -40.4 -48.7 -42.7 -45.8 -53.6	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z5' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW4-Z1 SB-PMW4-Z3 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z5' SB-PMW5-Z5 SB-PMW5-Z5 SB-PMW5-Z5 SB-PMW7-25' SB-PMW7-25' SB-PMW7-25'		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.3 -24.7 -23.3 -24.7 -23.3 -23.5 -23.1 -23.5 -23.1	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8 -22.8 -22.2 -23.1	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8 -27.5	Ethene	-44.2 -28.7 -40.4 -48.7 -42.7 -45.8 -53.6	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z5' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW4-Z1 SB-PMW4-Z3 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z5' SB-PMW5-Z5 SB-PMW5-Z5 SB-PMW5-Z5 SB-PMW7-25' SB-PMW7-25' SB-PMW7-25'		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.3 -24.7 -23.3 -24.7 -23.3 -23.5 -23.1 -23.5 -23.1	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8 -22.8 -22.2 -23.1	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8 -27.5	Ethene	-44.2 -28.7 -40.4 -48.7 -42.7 -45.8 -53.6	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-Z5' SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW4-Z4 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z5' SB-PMW6-25'		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.2 -24.4 -23.3 -24.7 -23.3 -24.7 -23.3 -23.5 -24.7 -23.3 -24.7 -23.3 -23.5 -24.7 -23.9 -24.7 -25.7 -	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8 -22.8 -22.2 -23.1 -18.4	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8 -27.5	Ethene	-44.2 -28.7 -40.4 -48.7 -42.7 -45.8 -53.6	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z5' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z5' SB-PIW3-Z5' SB-PIW3-Z5' SB-PIW3-Z5' SB-PIW3-Z5' SB-PIW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW6-25' SB-PMW6-25' SB-PMW6-25' SB-PMW9-25'		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.3 -24.7 -23.3 -24.7 -23.3 -24.7 -23.3 -24.7 -23.3 -24.7 -23.9 -24.7 -23.9 -24.7 -23.9 -24.7 -23.9 -24.7 -23.9 -24.7 -25.7 -25.7 -26.7 -27.7 -	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8 -22.8 -22.2 -23.1 -18.4	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8 -27.5	Ethene	-44.2 -28.7 -40.4 -48.7 -42.7 -45.8 -53.6 -43.3 -36.2 -40.3 -42.4	2-butanone chloroform	chlor/DCE

COD

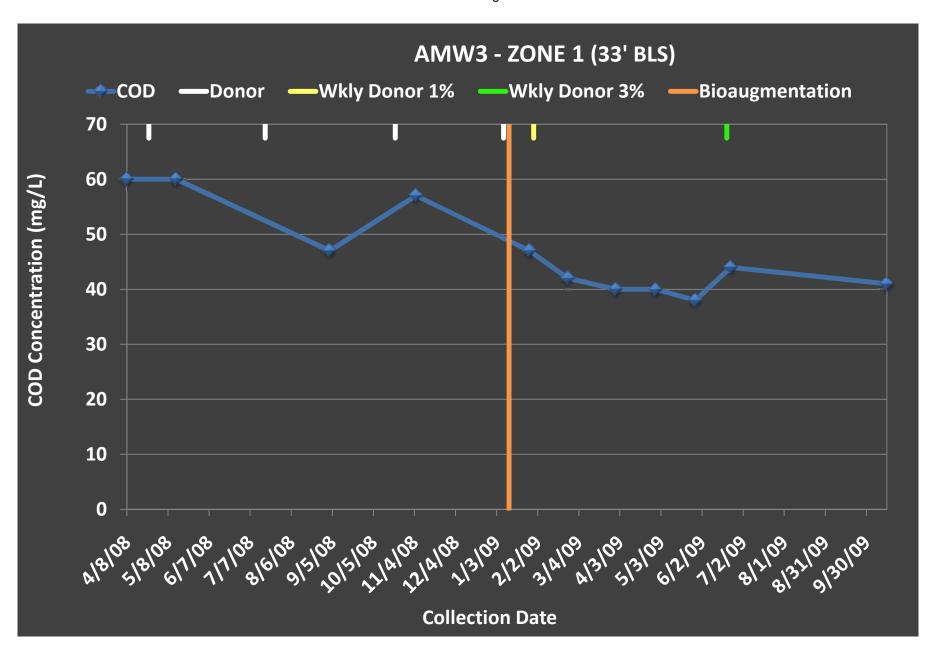
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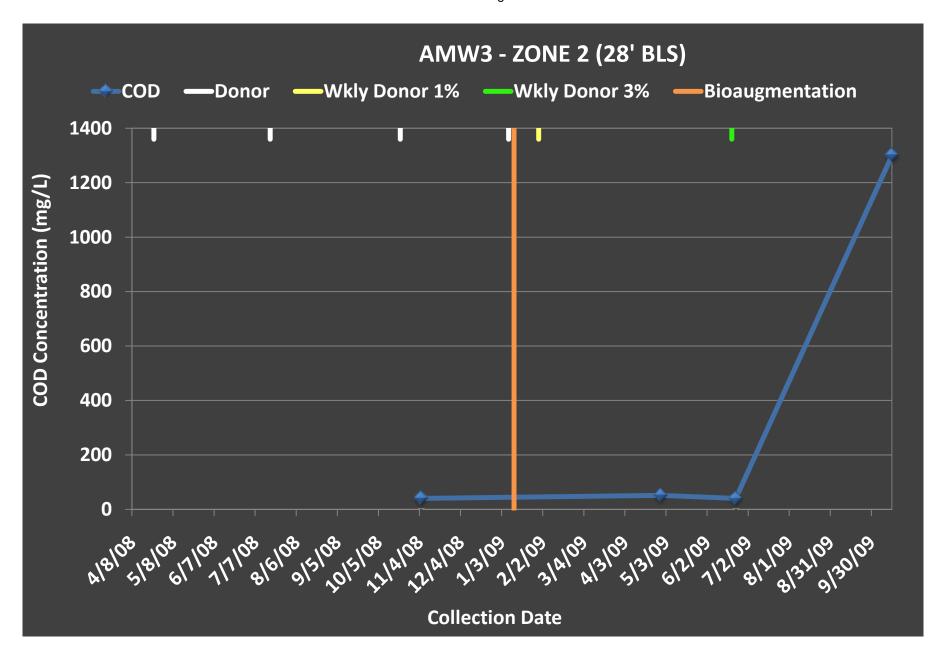
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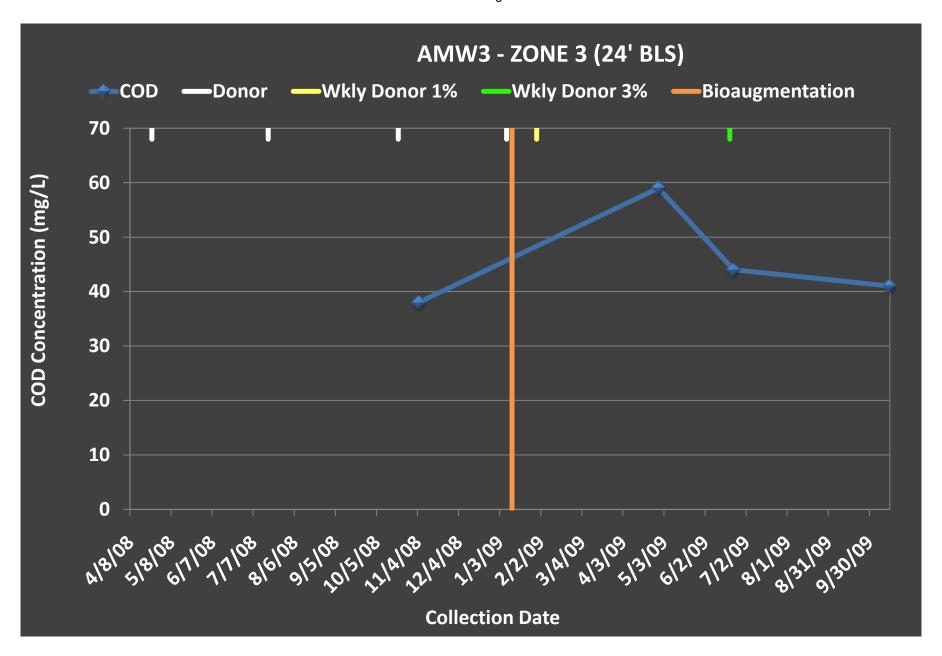
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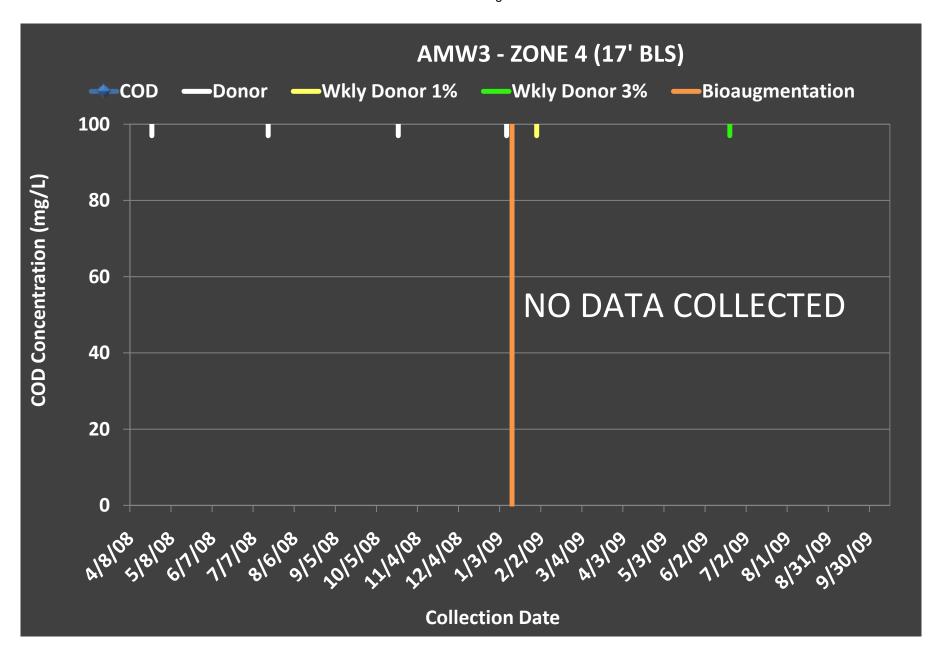
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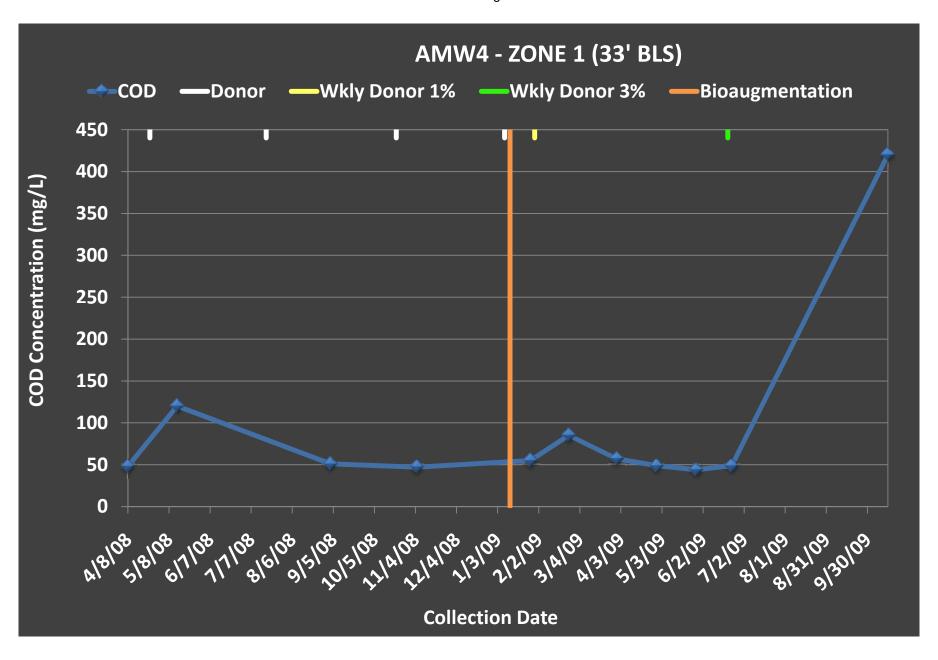
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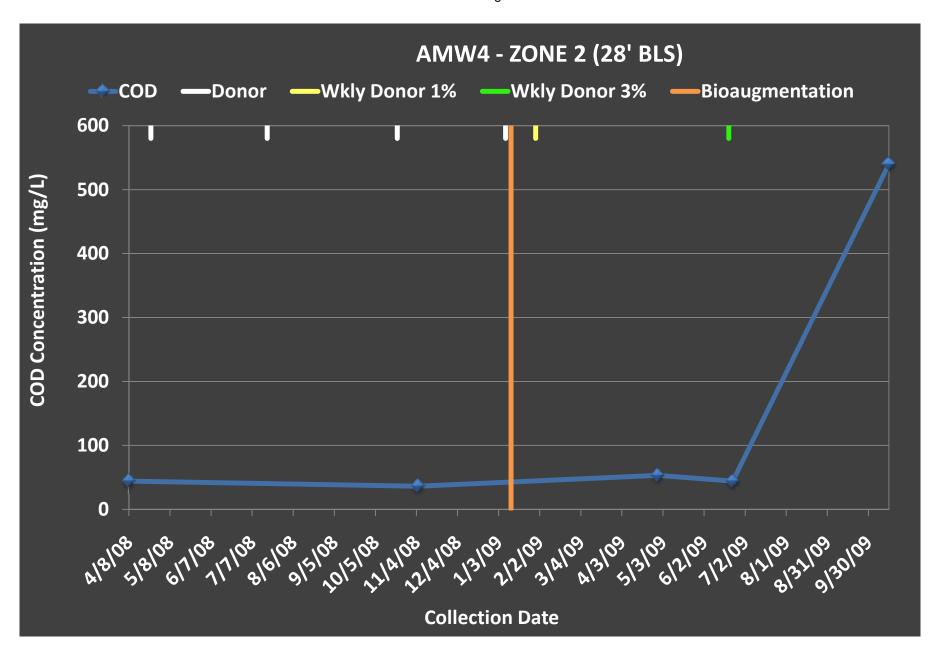
Seal Beach Groundwater Bioaugmentation



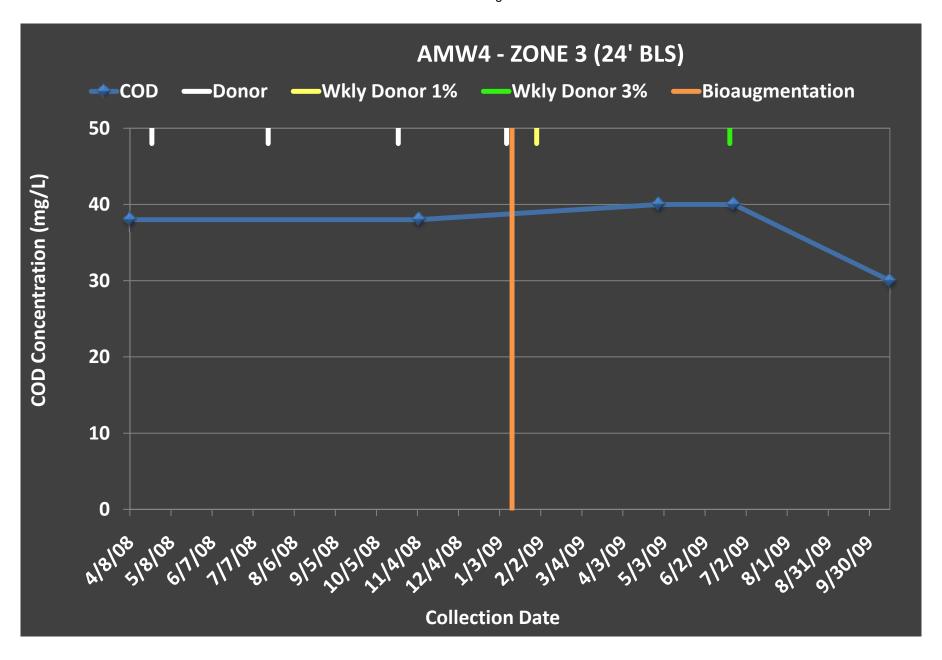
Seal Beach
Groundwater Bioaugmentation



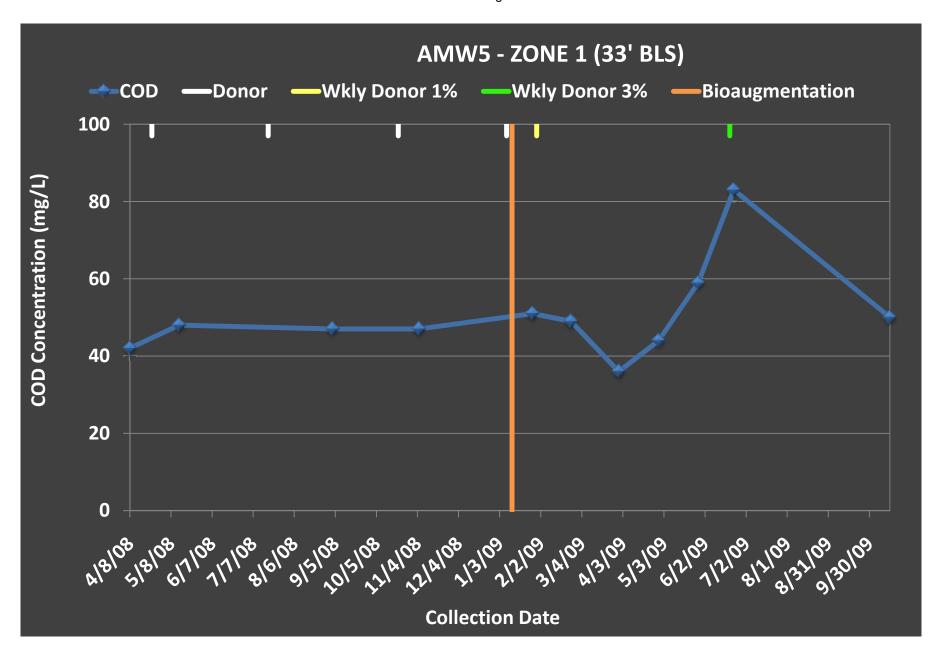
Seal Beach
Groundwater Bioaugmentation



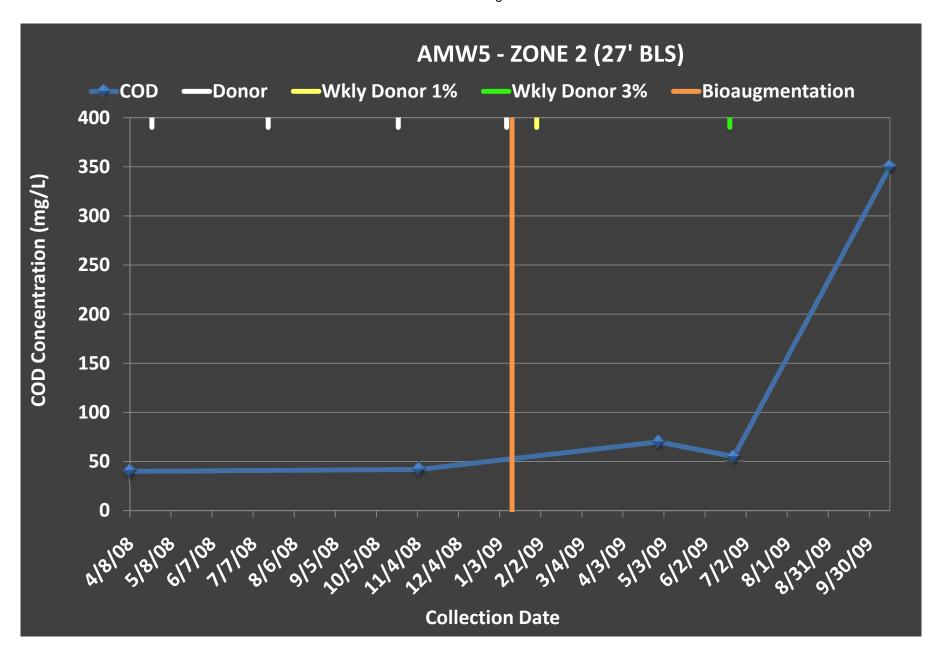
Seal Beach
Groundwater Bioaugmentation



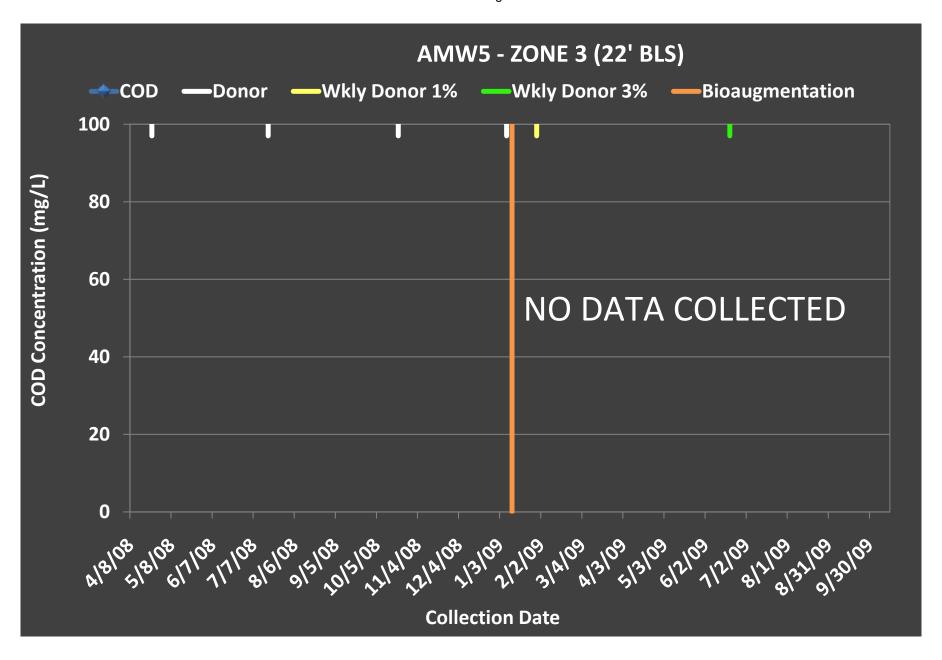
Seal Beach
Groundwater Bioaugmentation



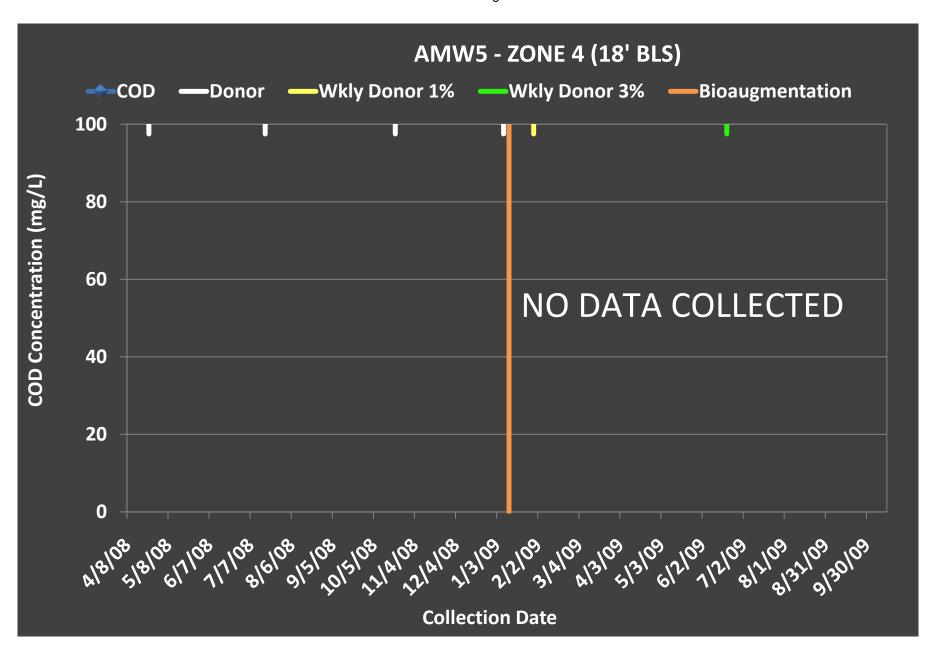
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Groundwater Bioaugmentation



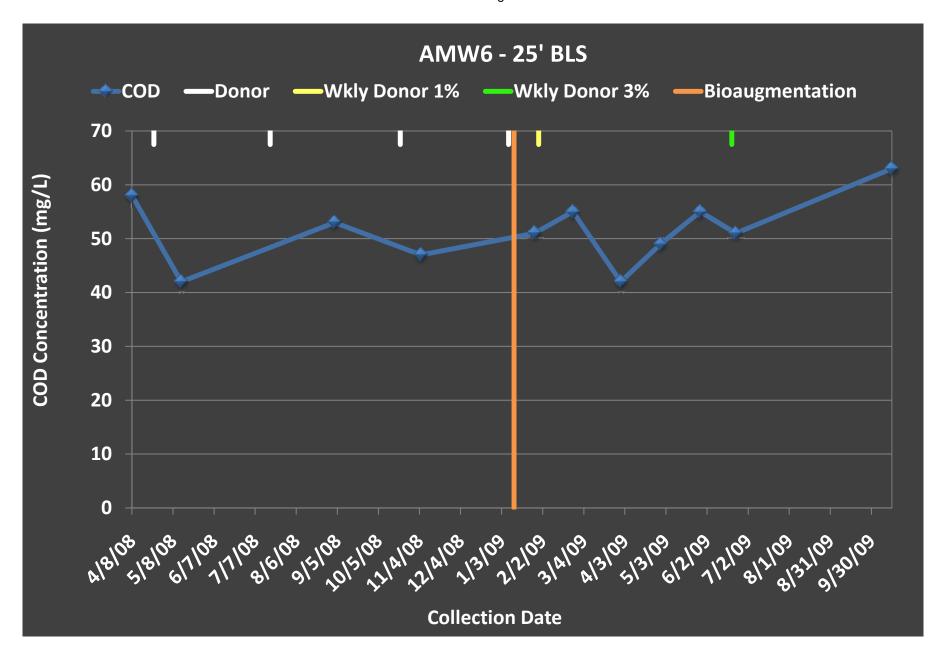
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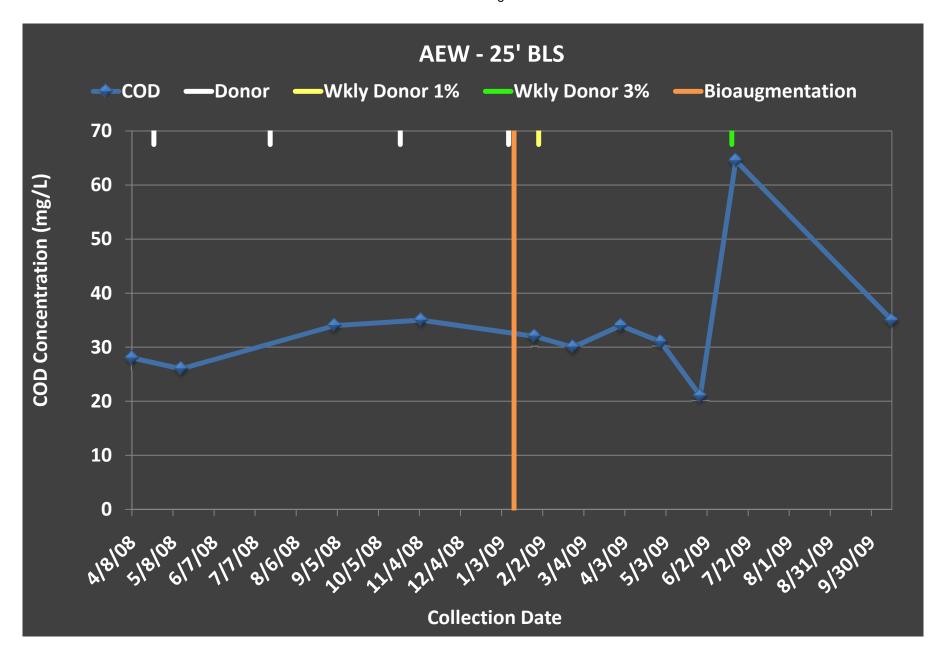
Seal Beach Groundwater Bioaugmentation



Seal Beach
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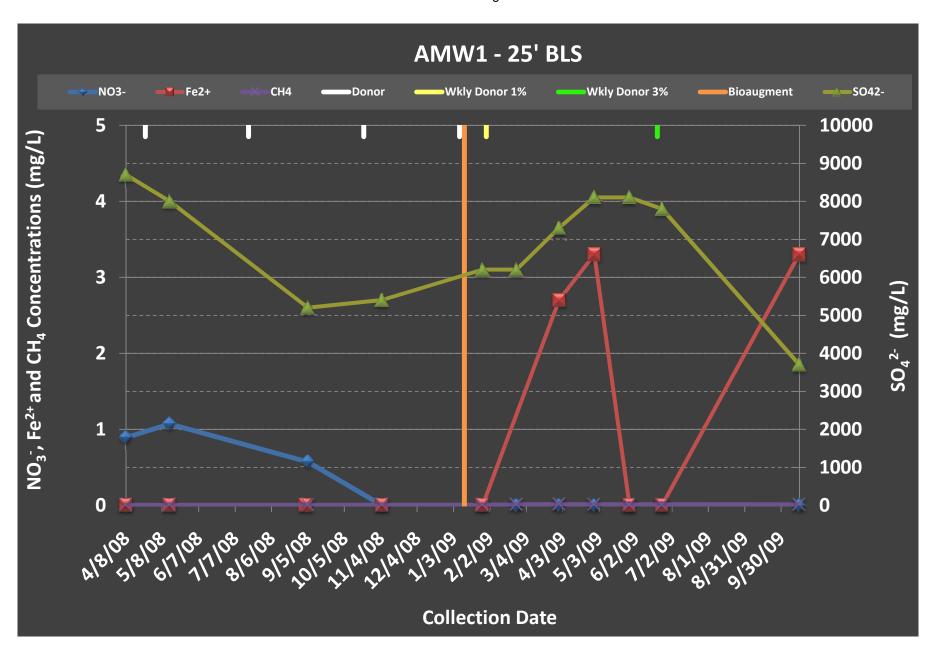


Seal Beach
Groundwater Bioaugmentation

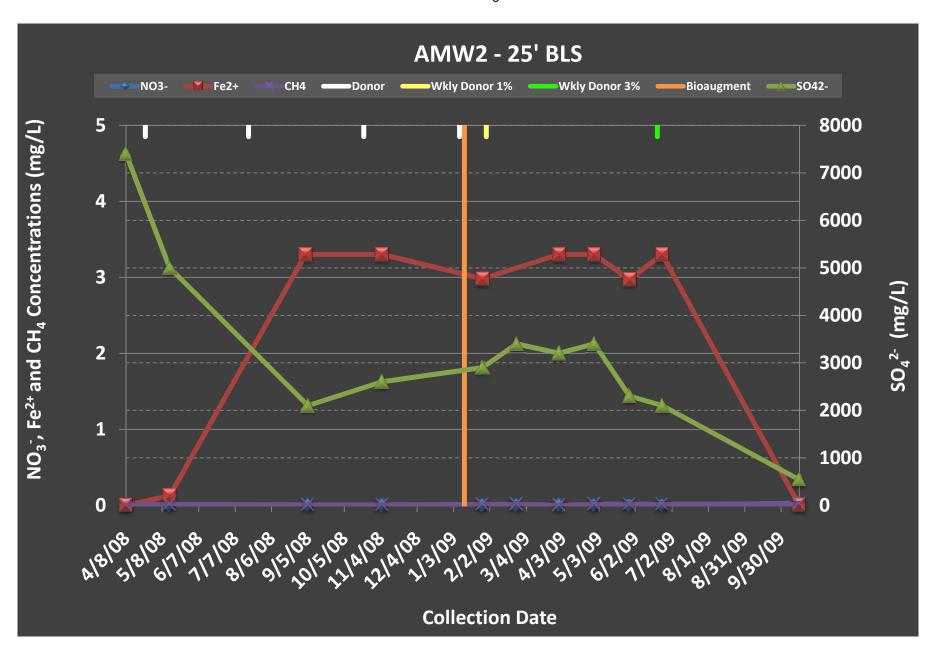


Electron Acceptors

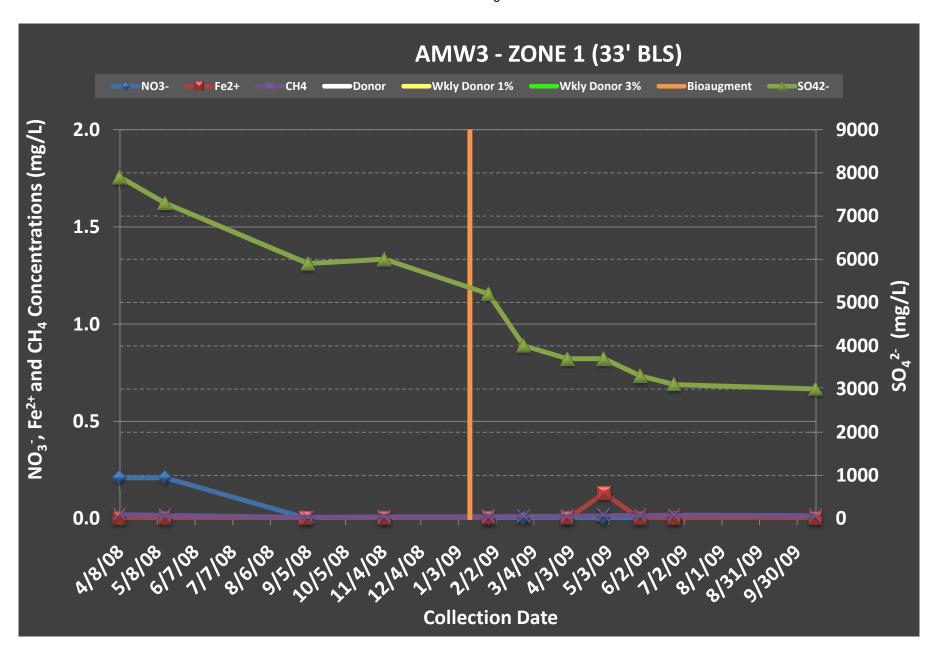
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Groundwater Bioaugmentation



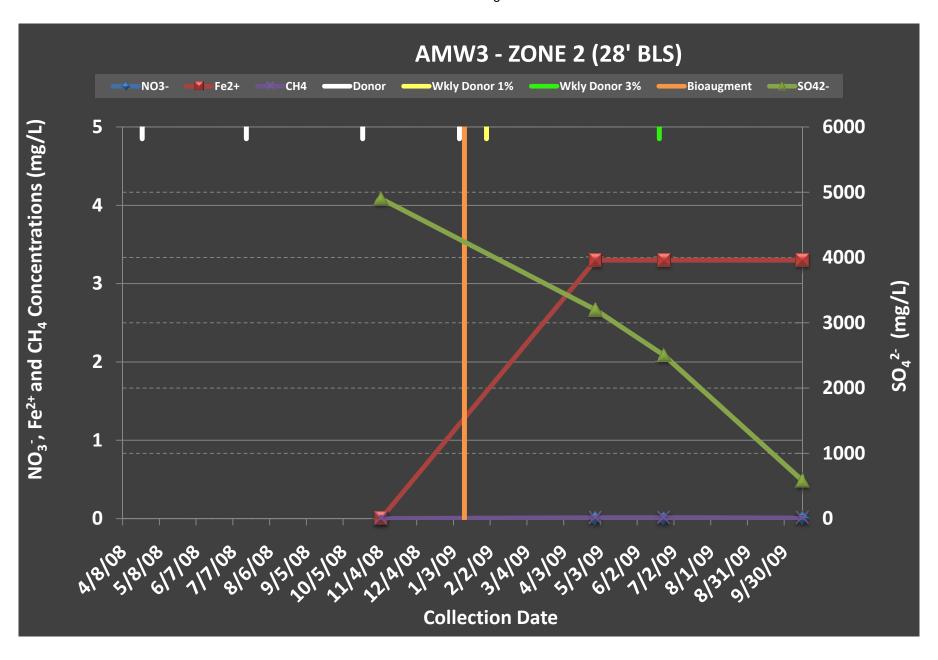
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Groundwater Bioaugmentation



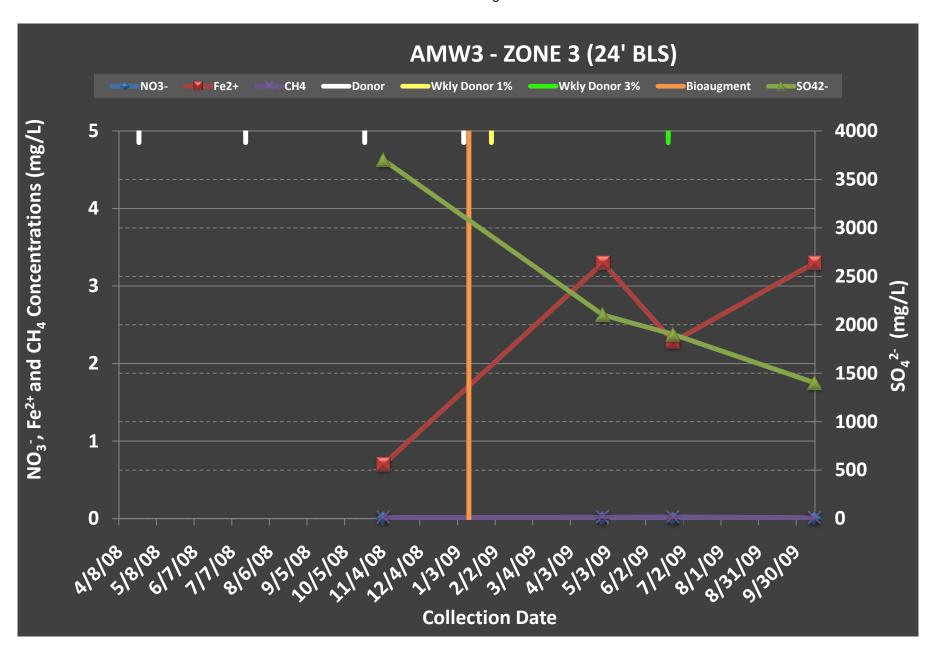
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Groundwater Bioaugmentation



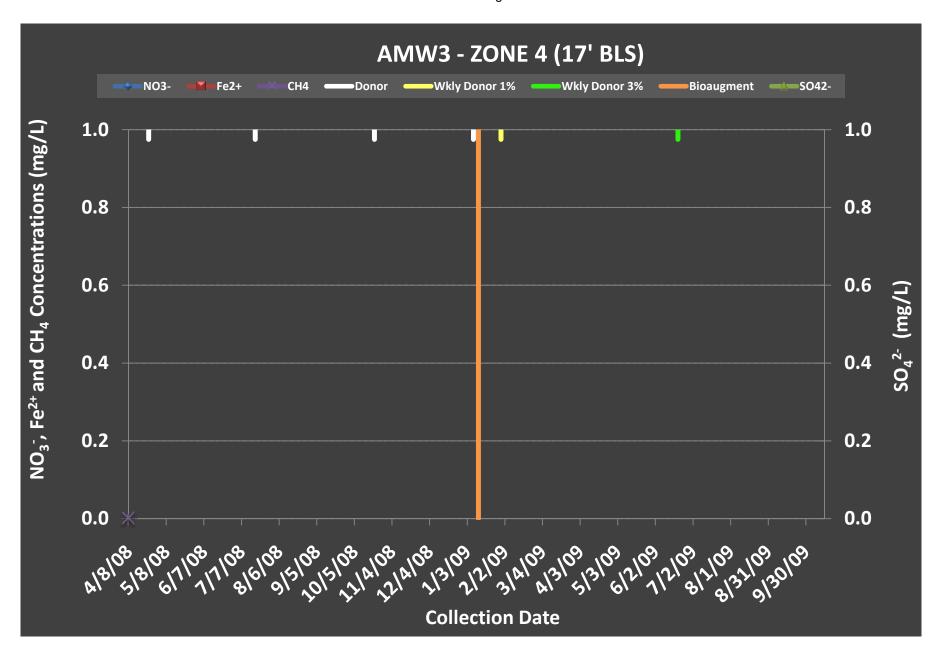
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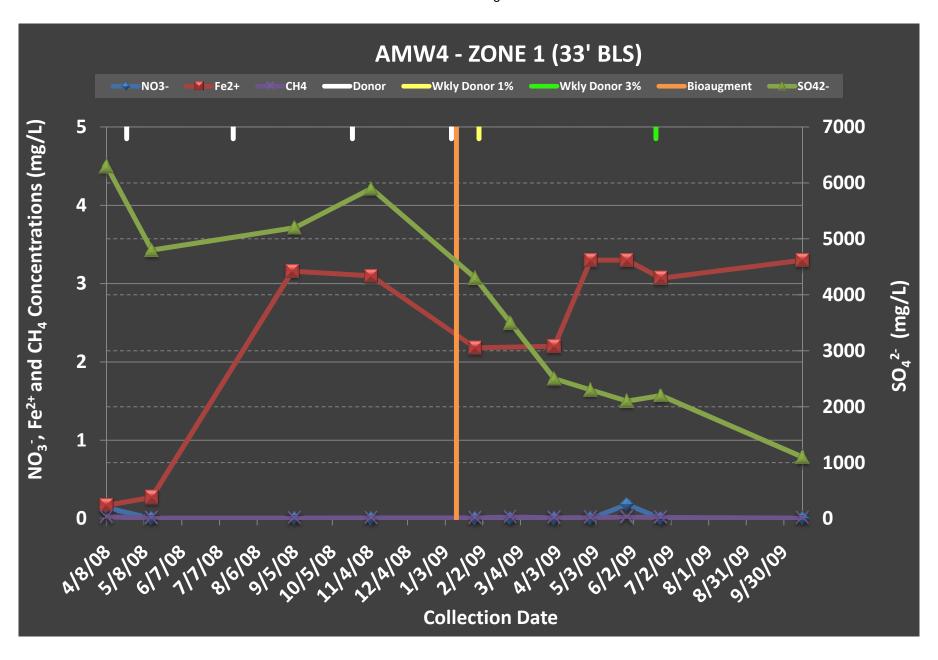
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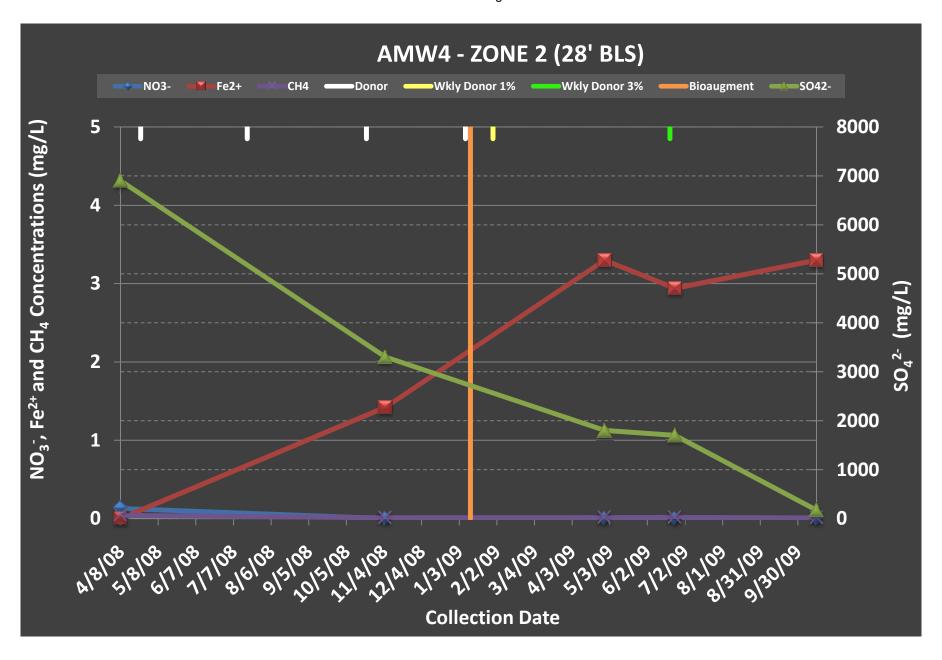
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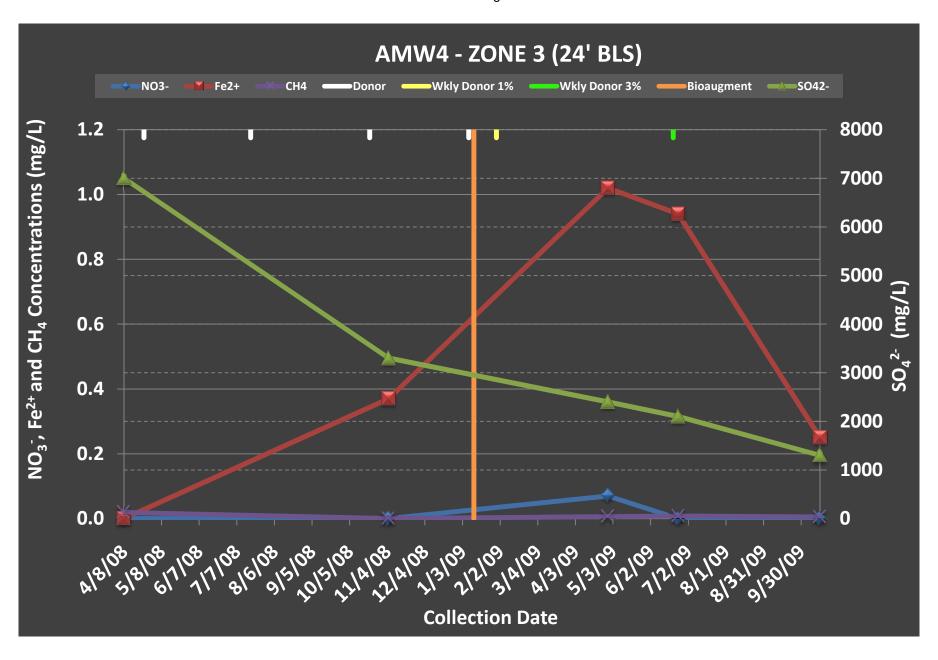
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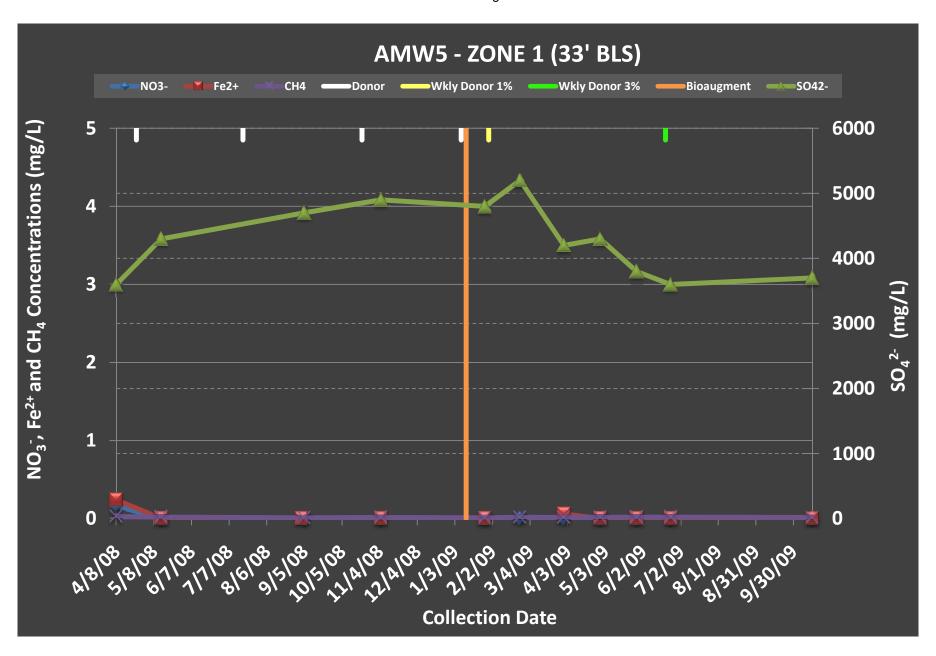
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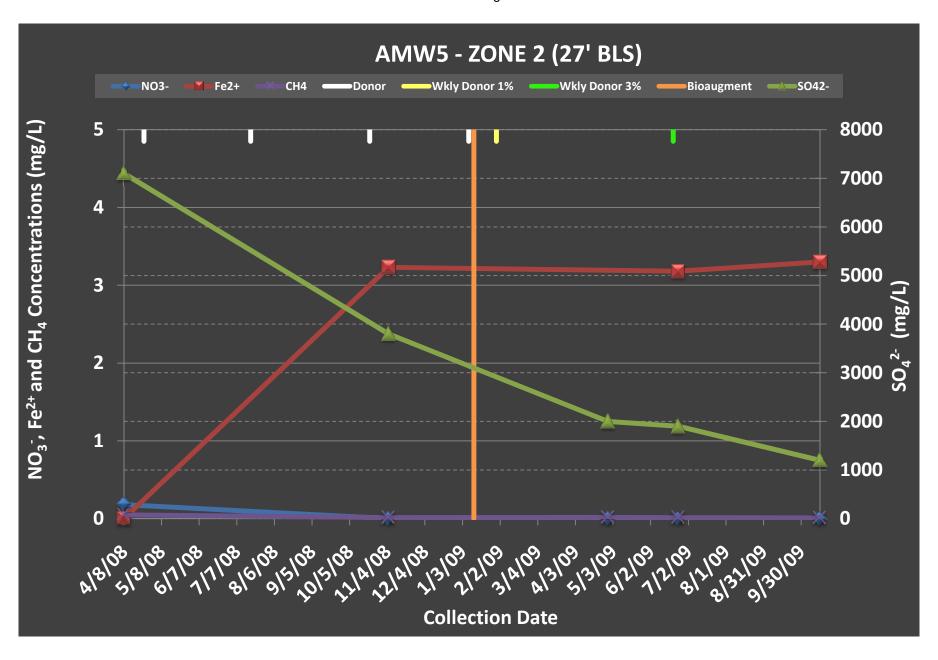
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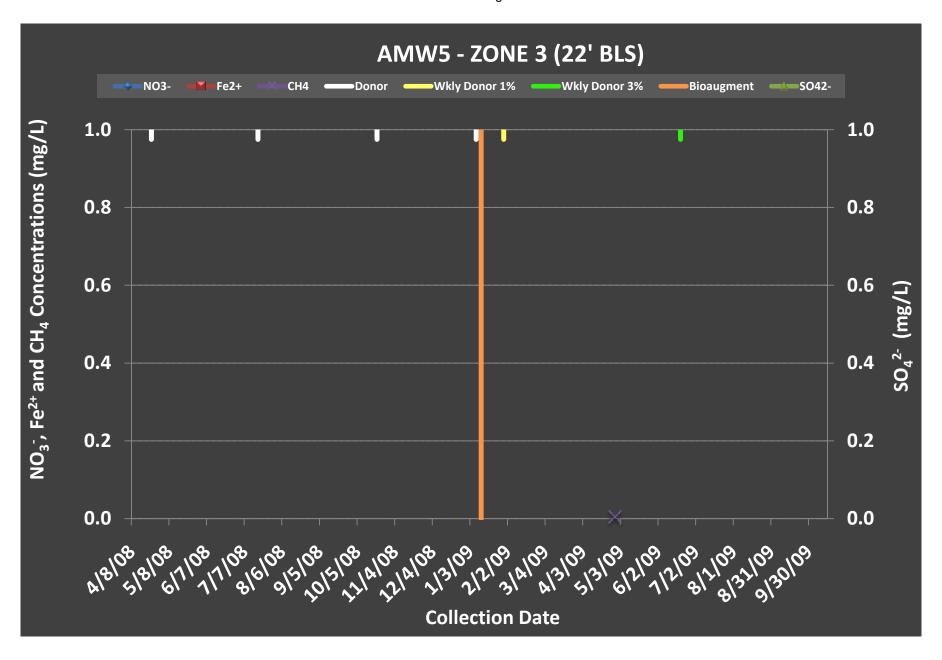
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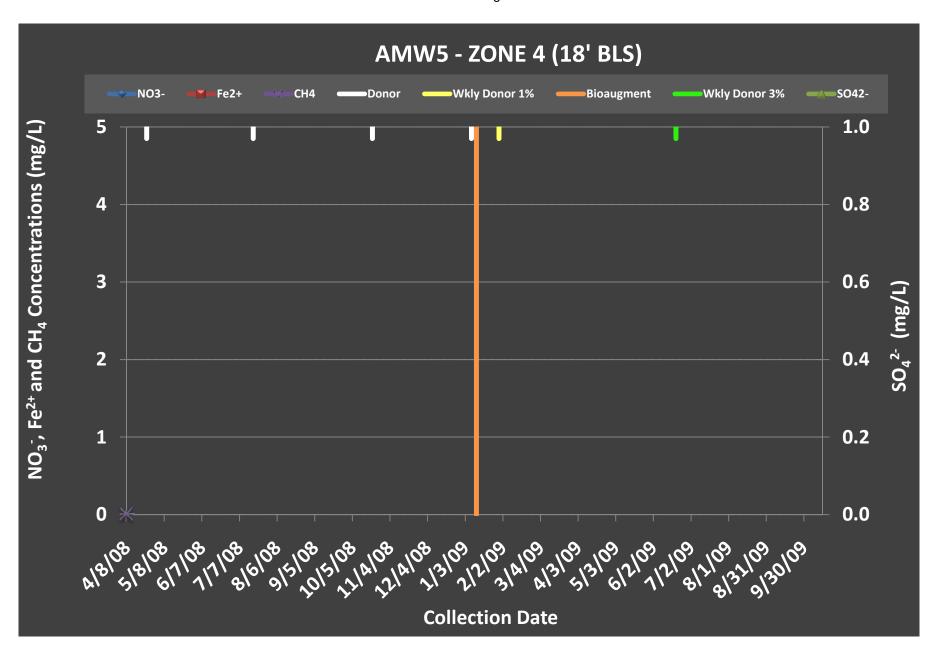
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Groundwater Bioaugmentation



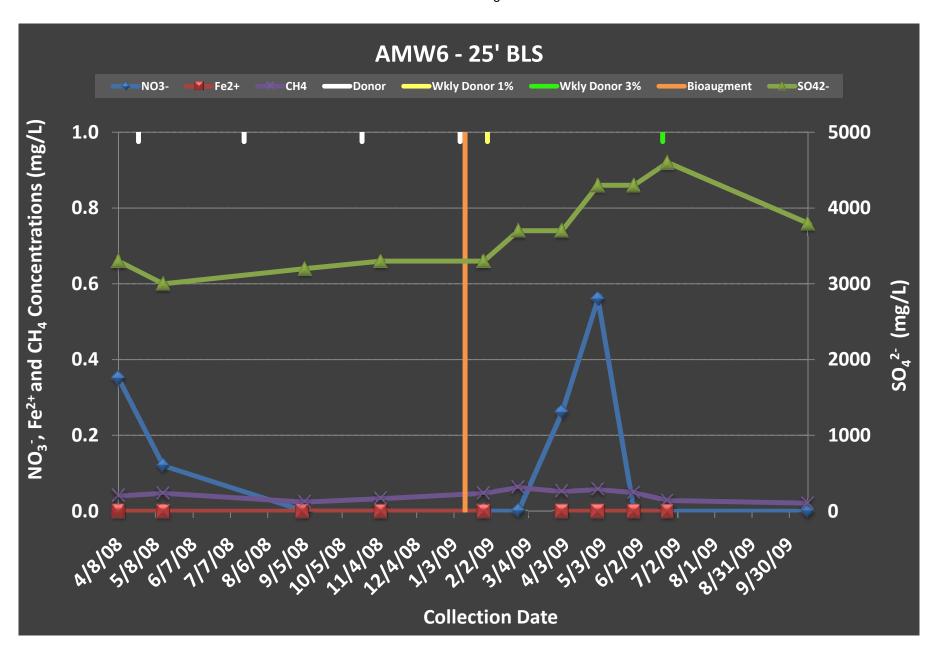
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Groundwater Bioaugmentation



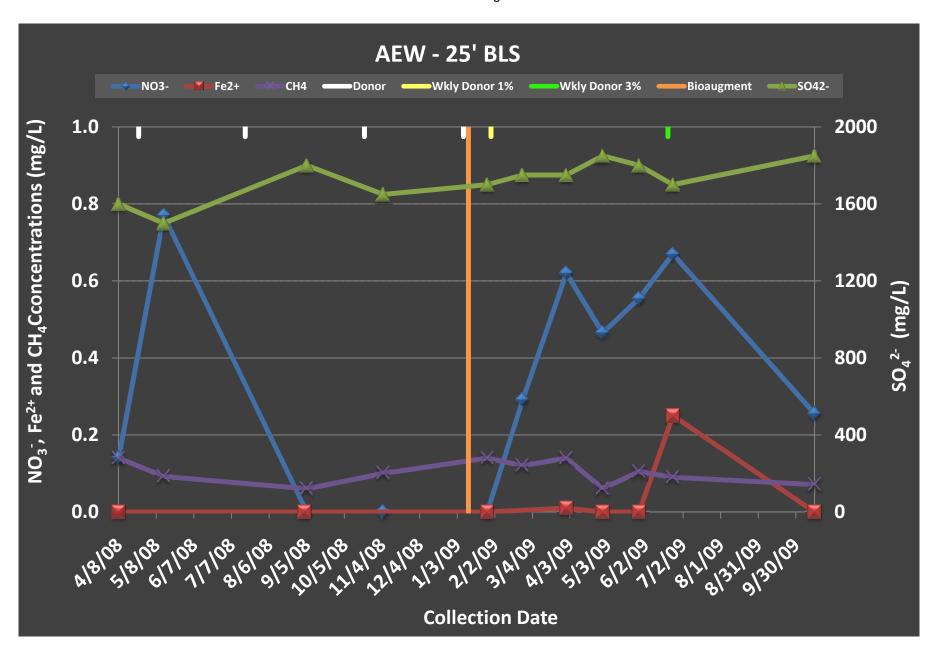
Seal Beach
Groundwater Bioaugmentation



Seal Beach
Groundwater Bioaugmentation

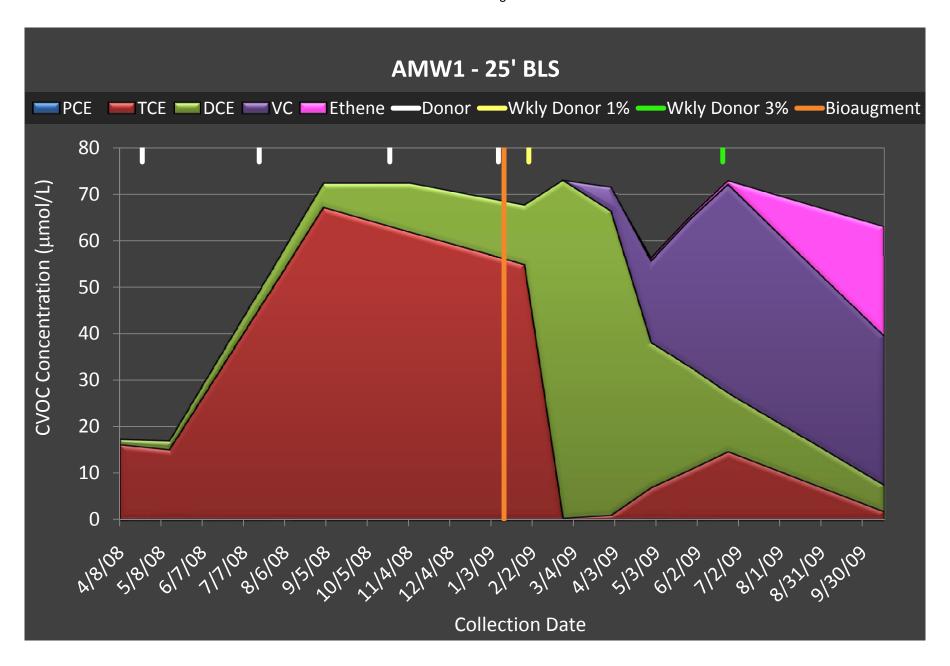


Seal Beach
Groundwater Bioaugmentation

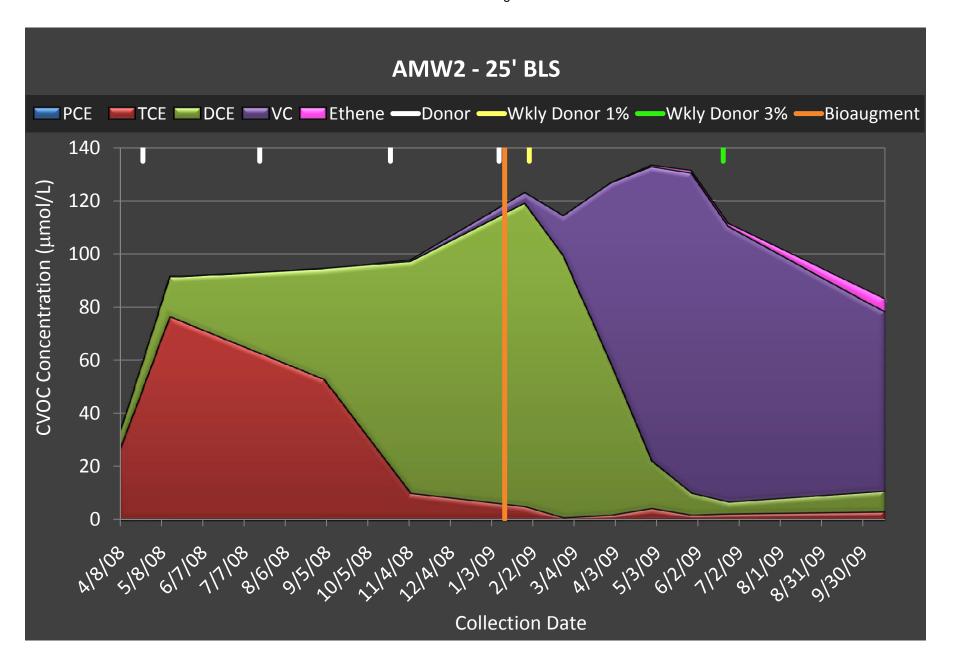


CVOCs Molar Concentrations

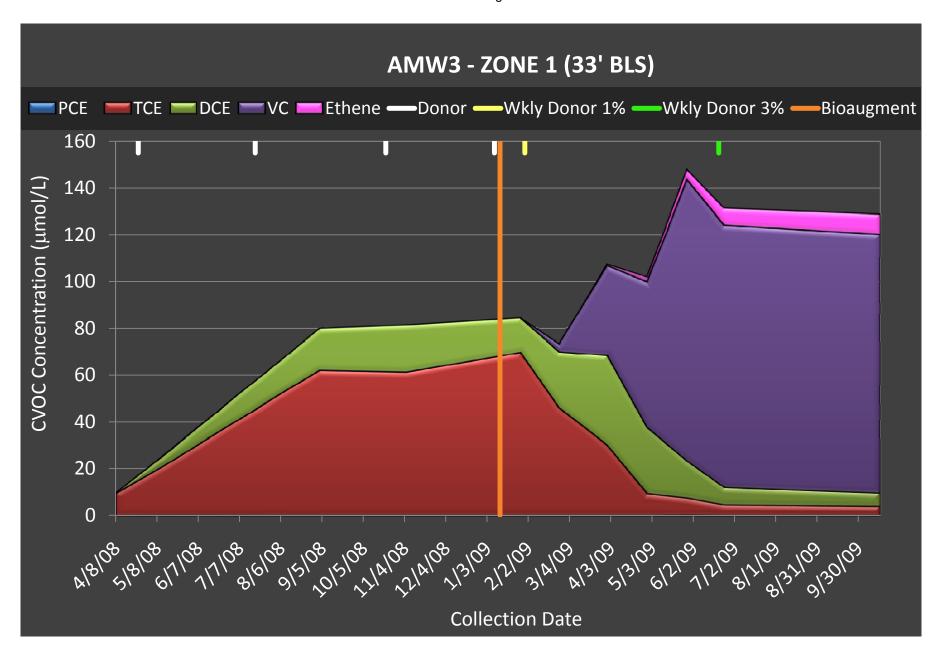
Seal Beach
Groundwater Bioaugmentation



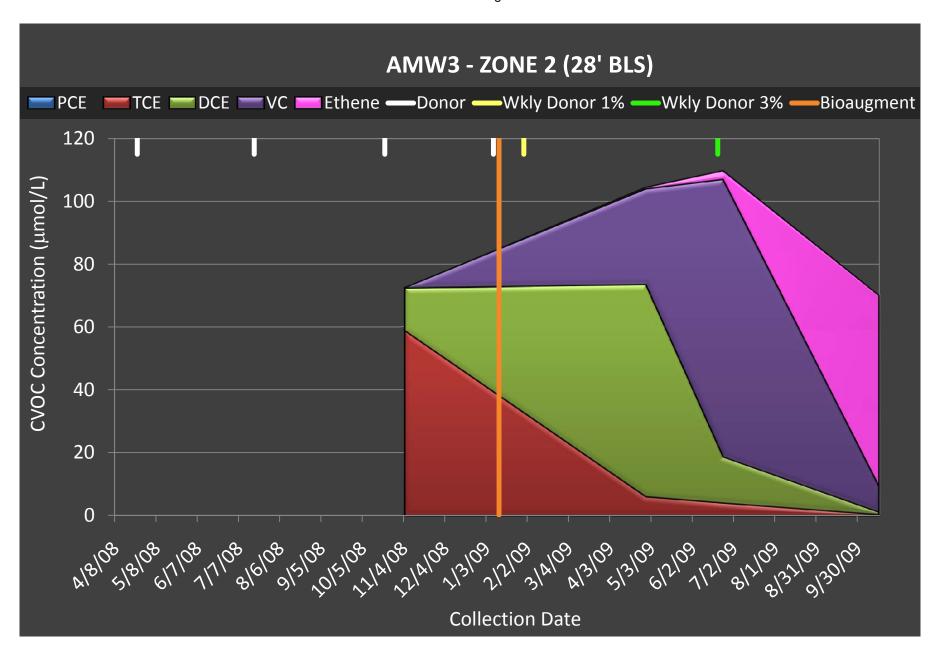
Seal Beach
Groundwater Bioaugmentation



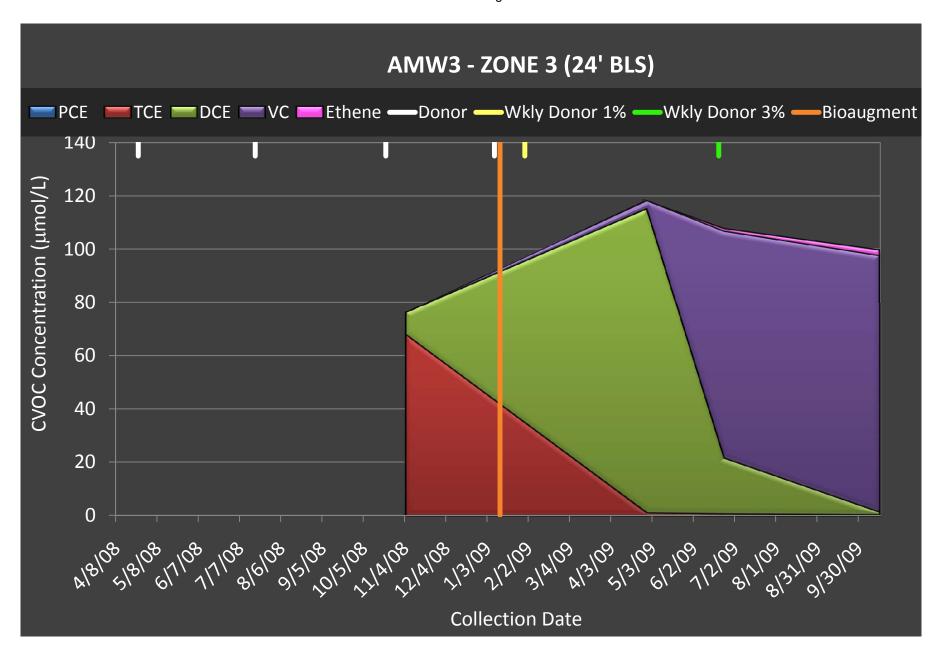
Seal Beach
Groundwater Bioaugmentation



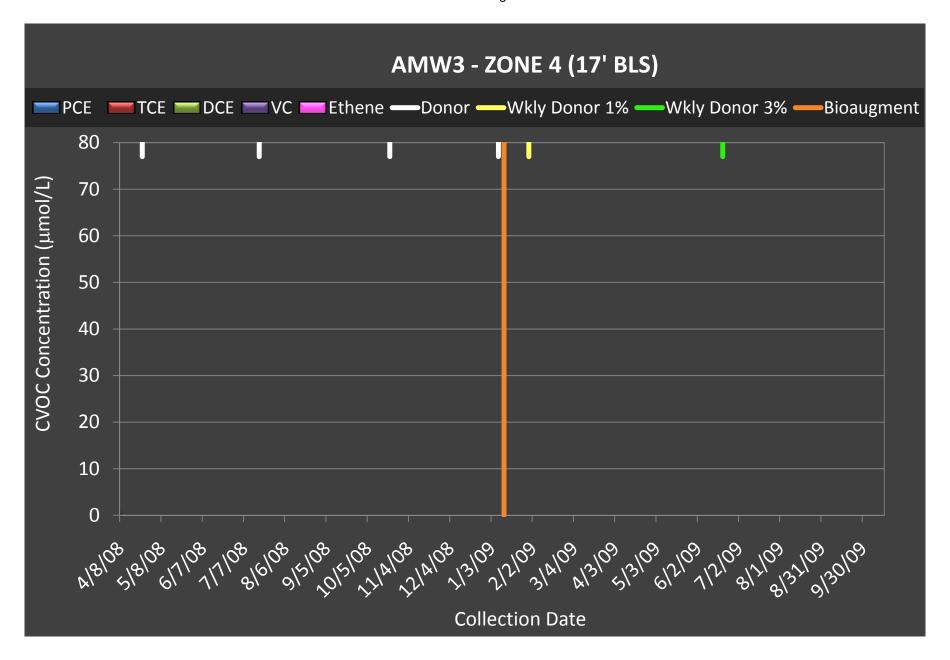
Seal Beach
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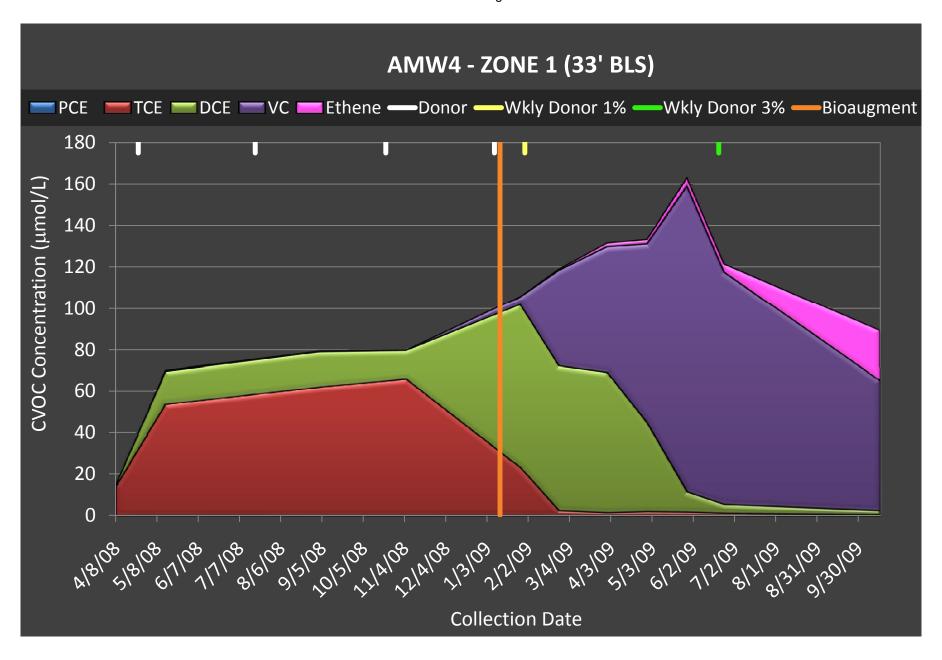
Seal Beach
Groundwater Bioaugmentation



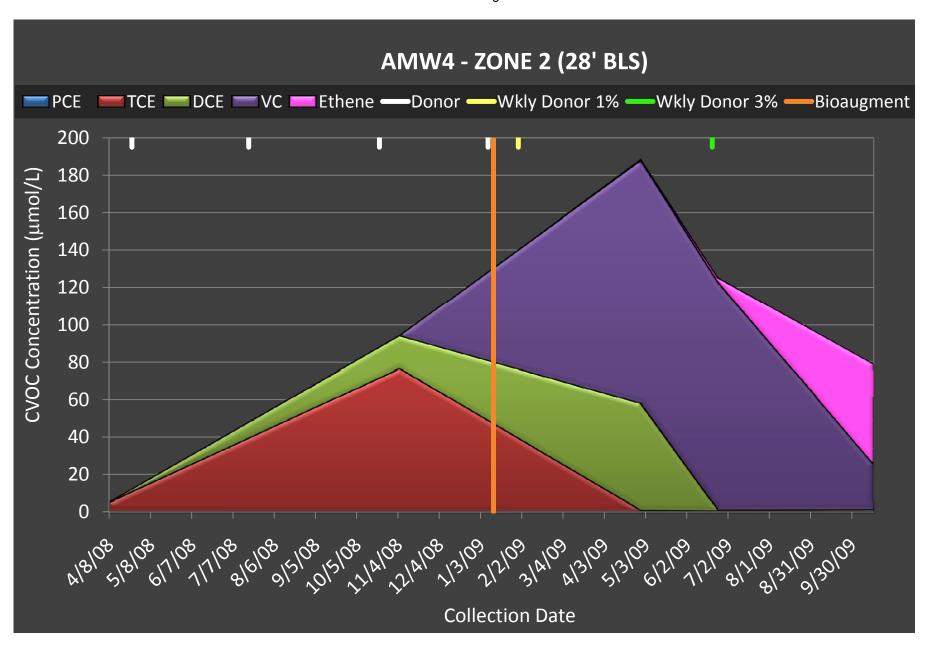
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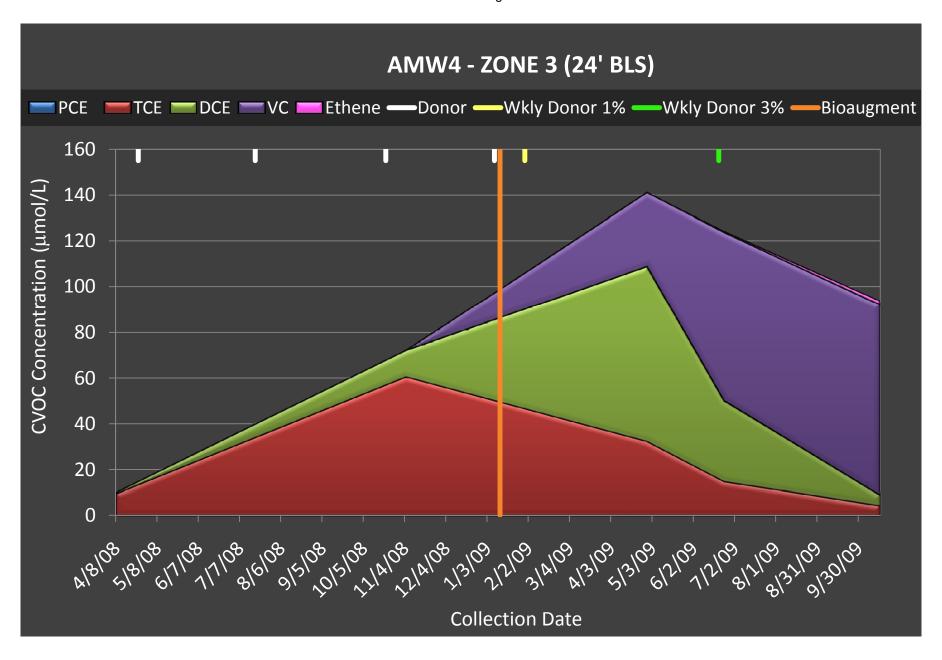
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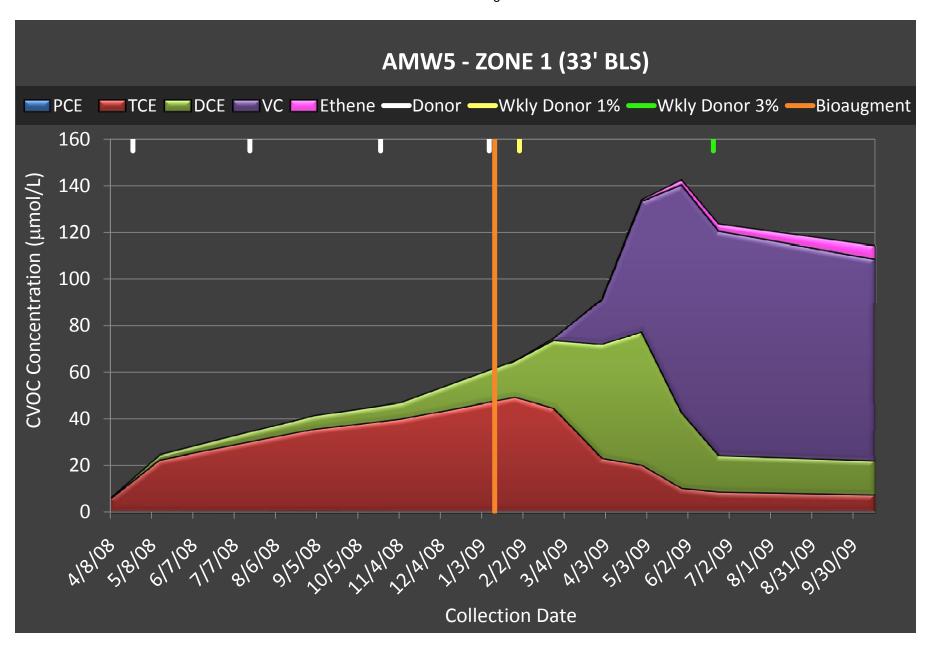
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Groundwater Bioaugmentation



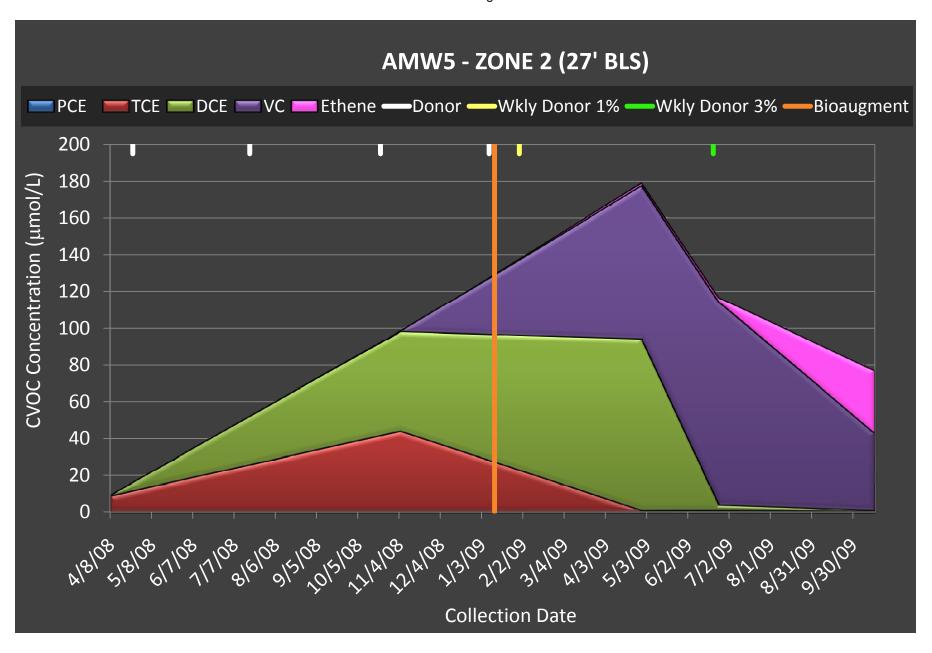
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Groundwater Bioaugmentation



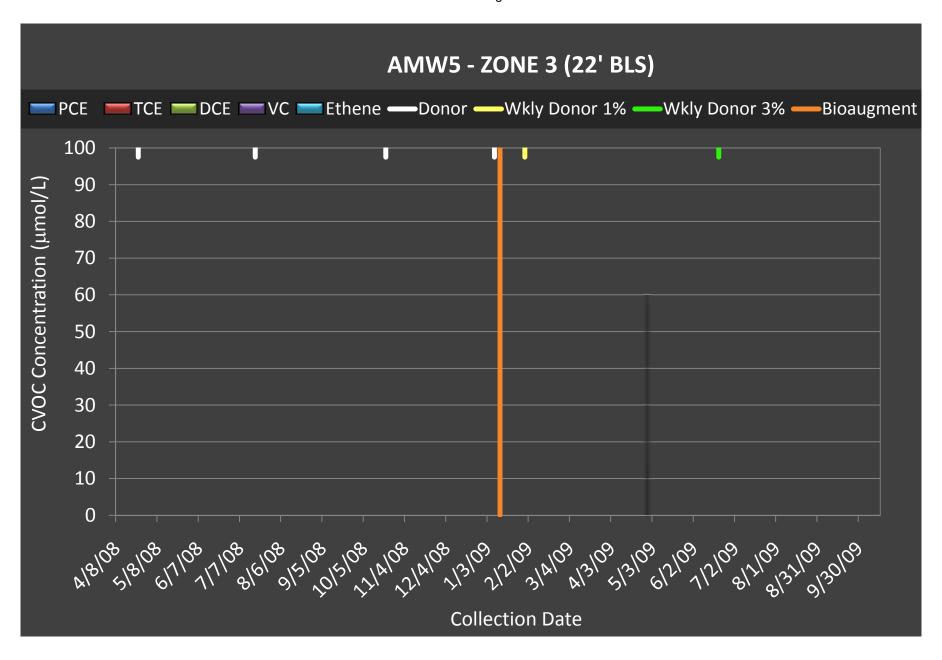
Seal Beach
Groundwater Bioaugmentation



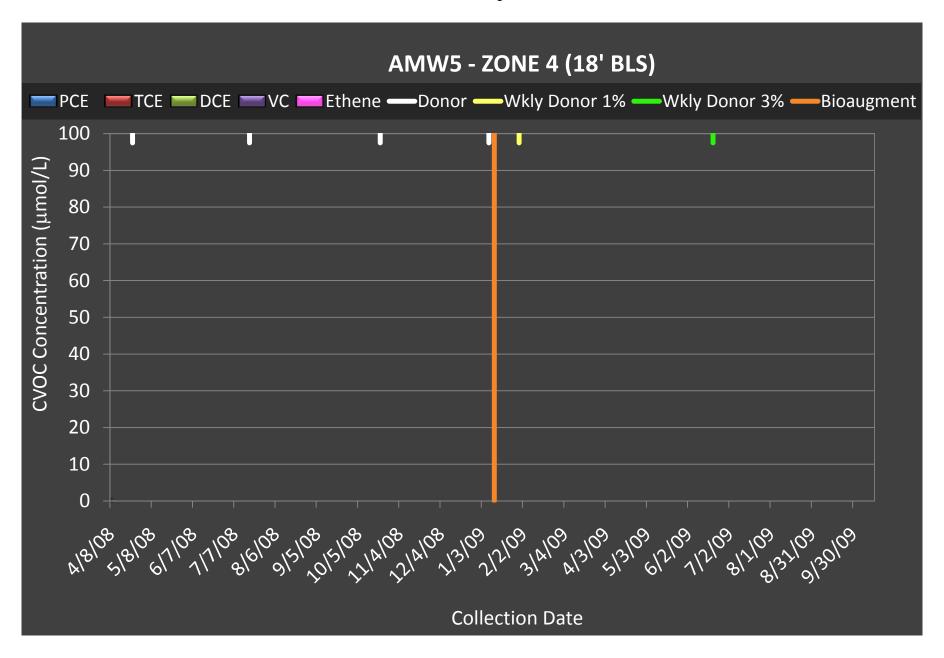
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Groundwater Bioaugmentation



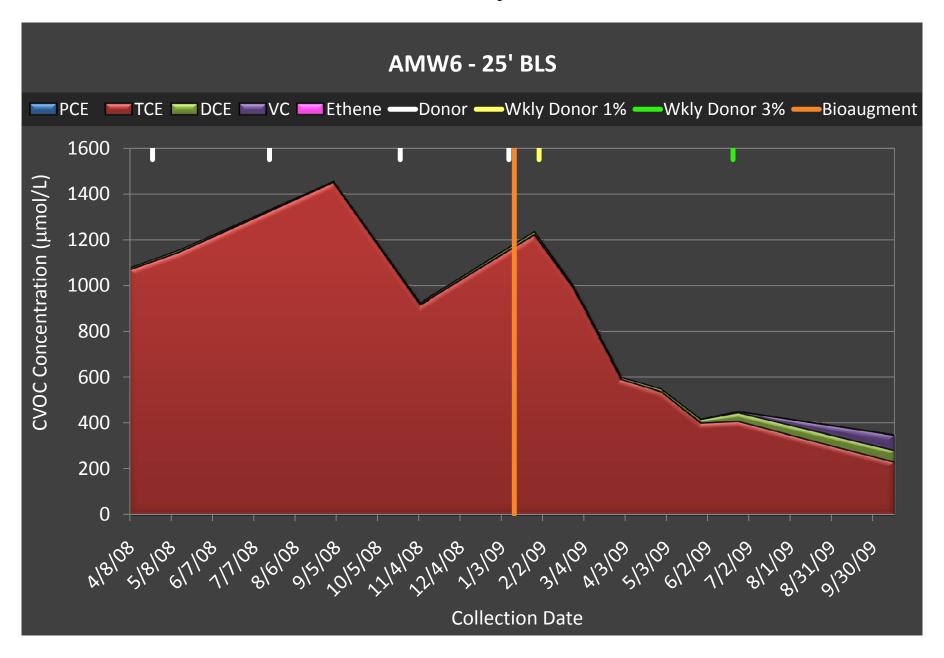
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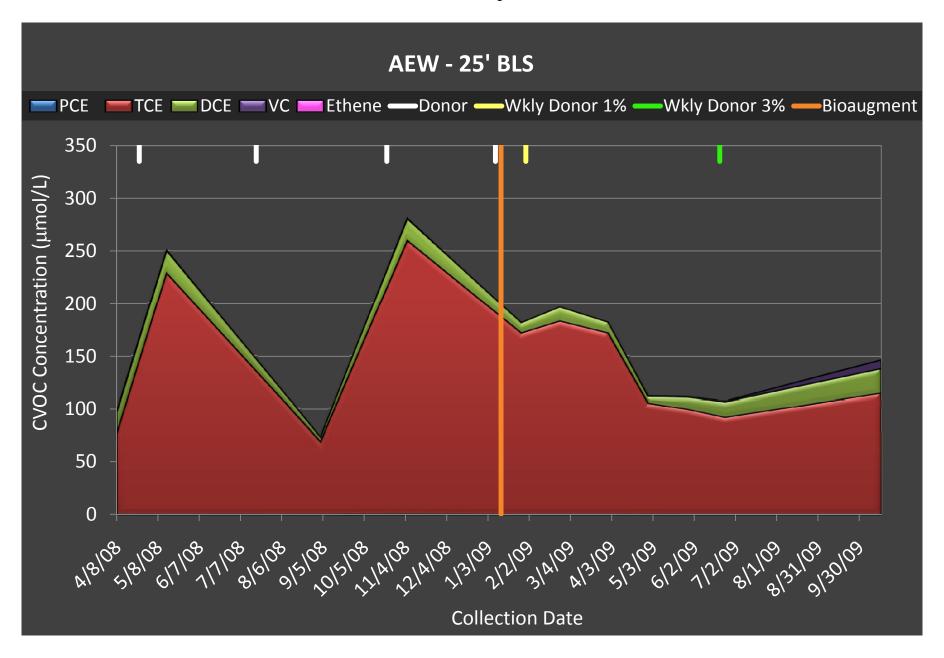
Seal Beach Groundwater Bioaugmentation



Seal Beach Groundwater Bioaugmentation

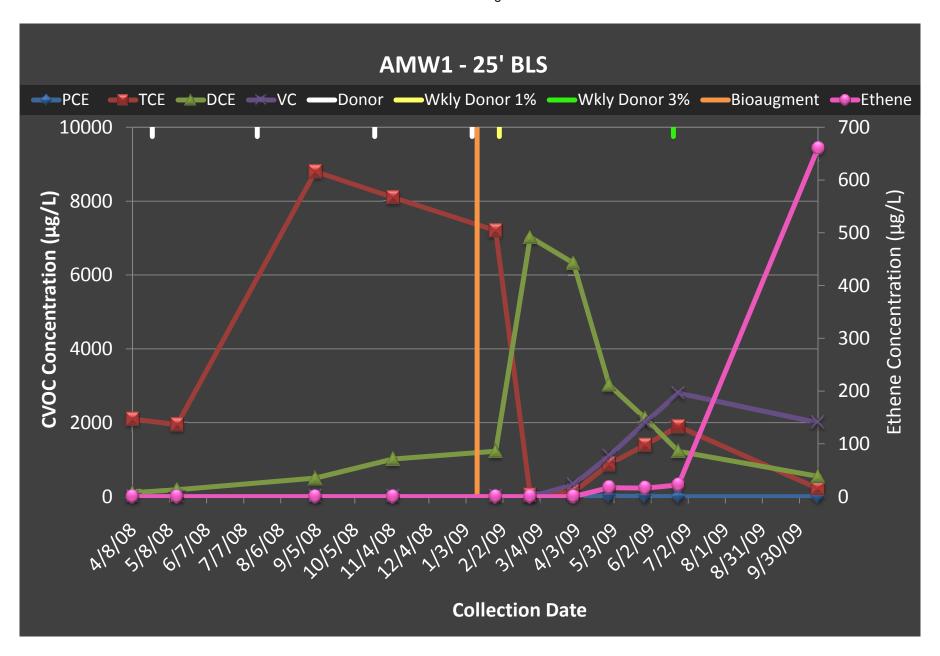


Seal Beach Groundwater Bioaugmentation

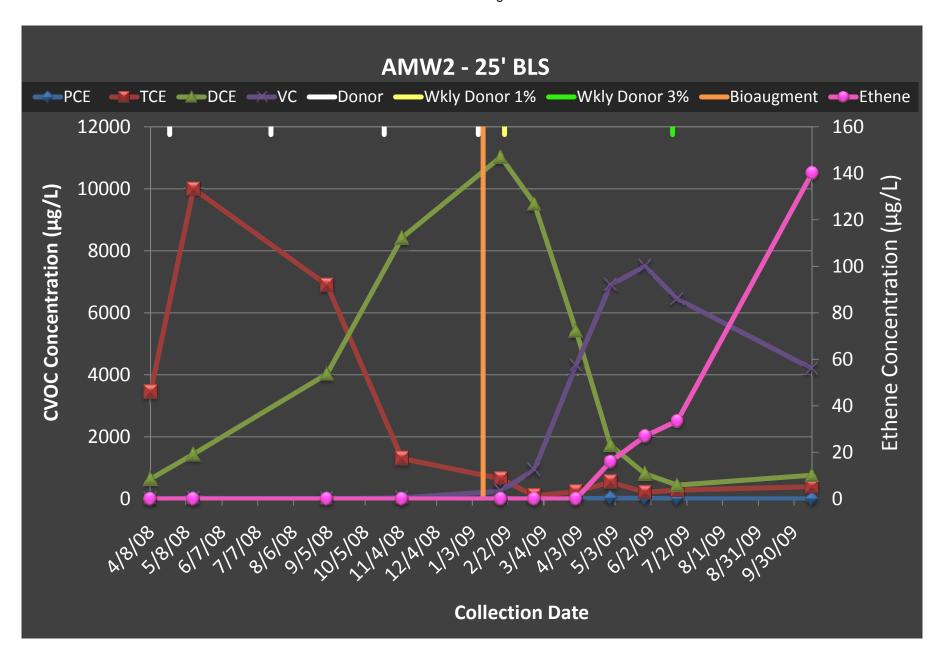


CVOCs Mass Concentrations

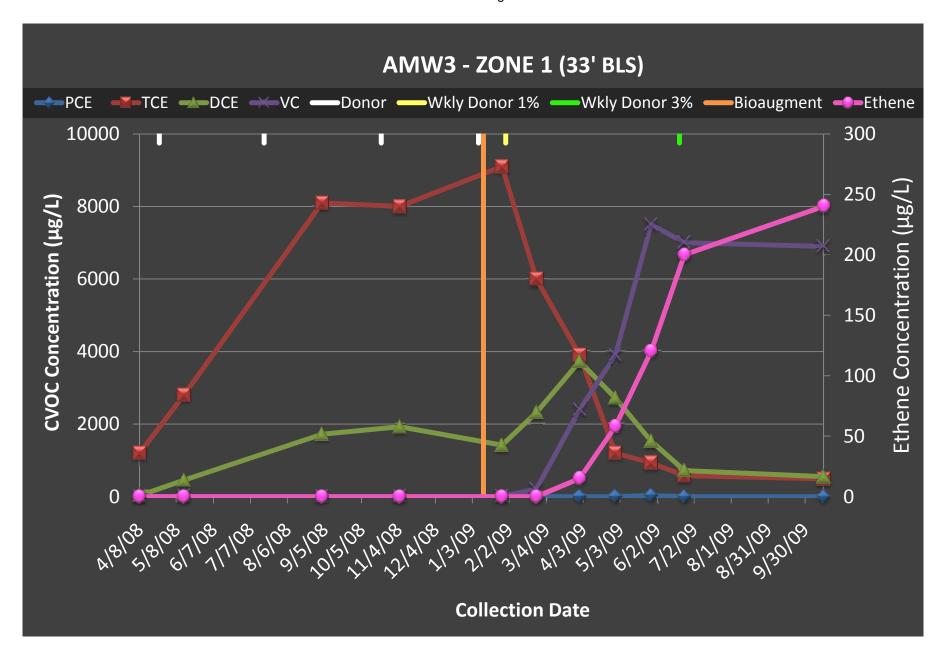
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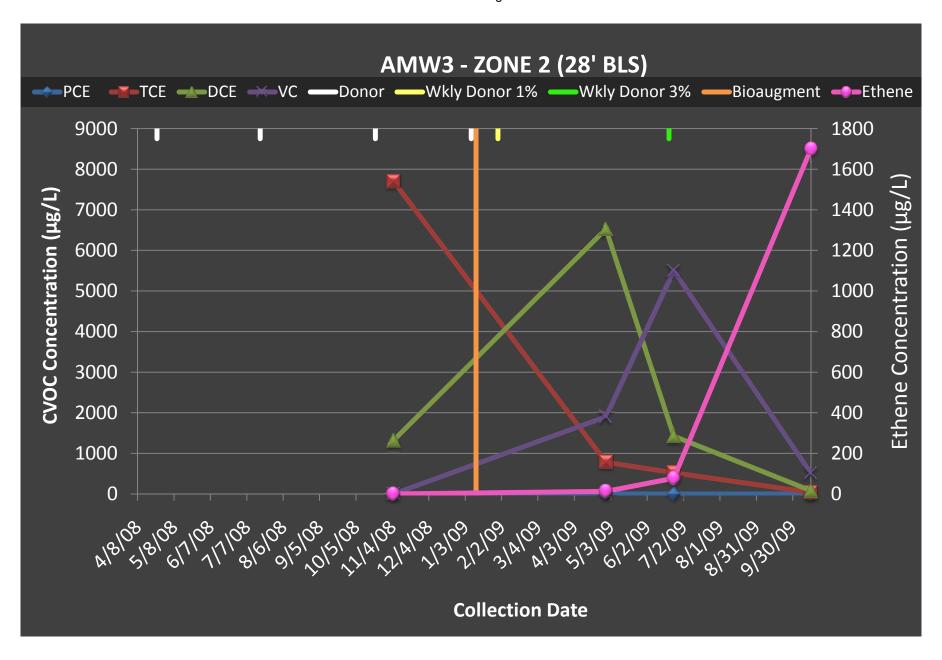
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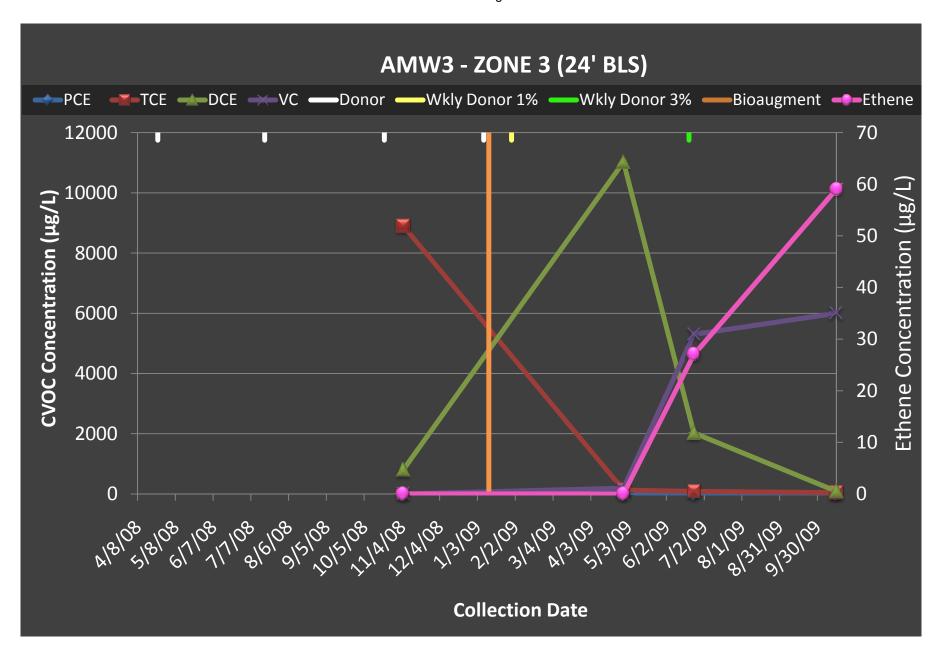
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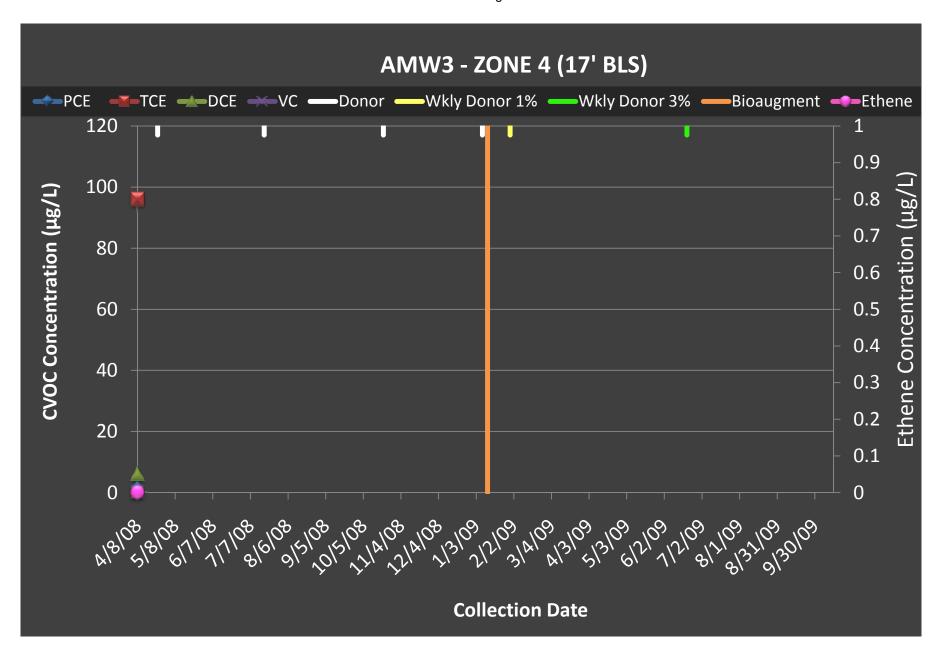
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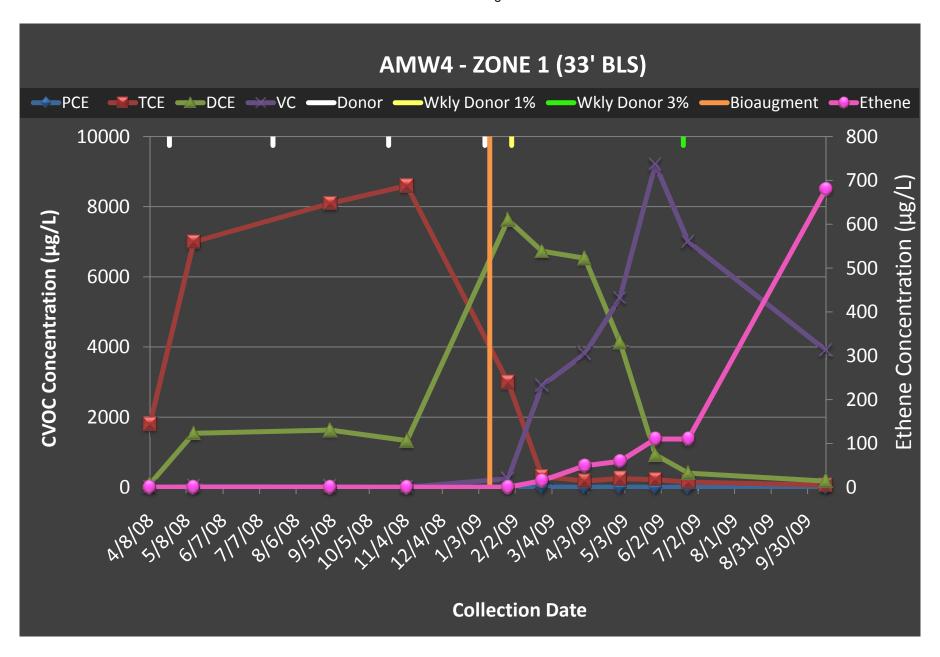
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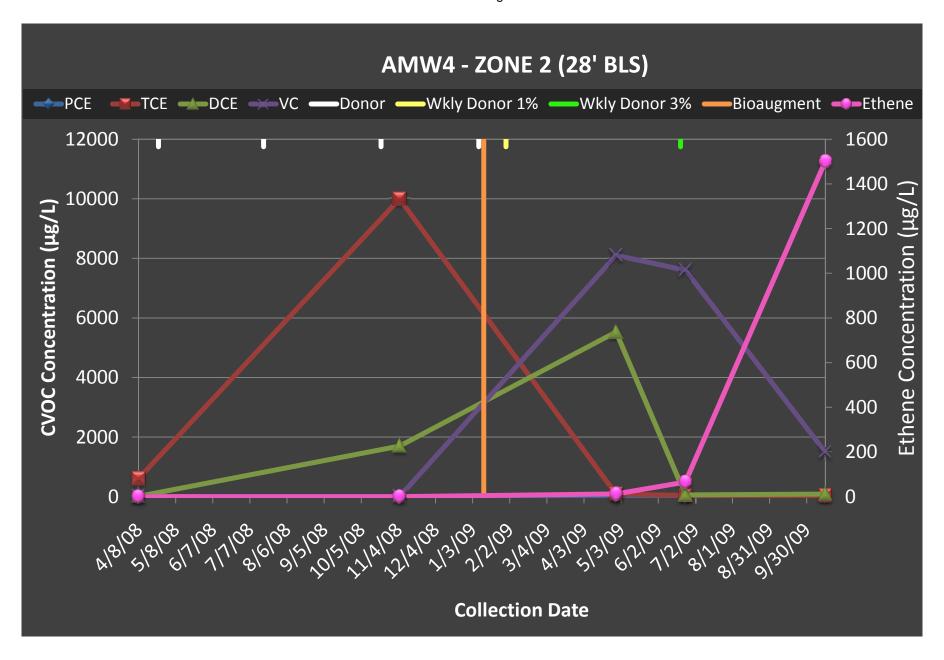
Seal Beach
Groundwater Bioaugmentation



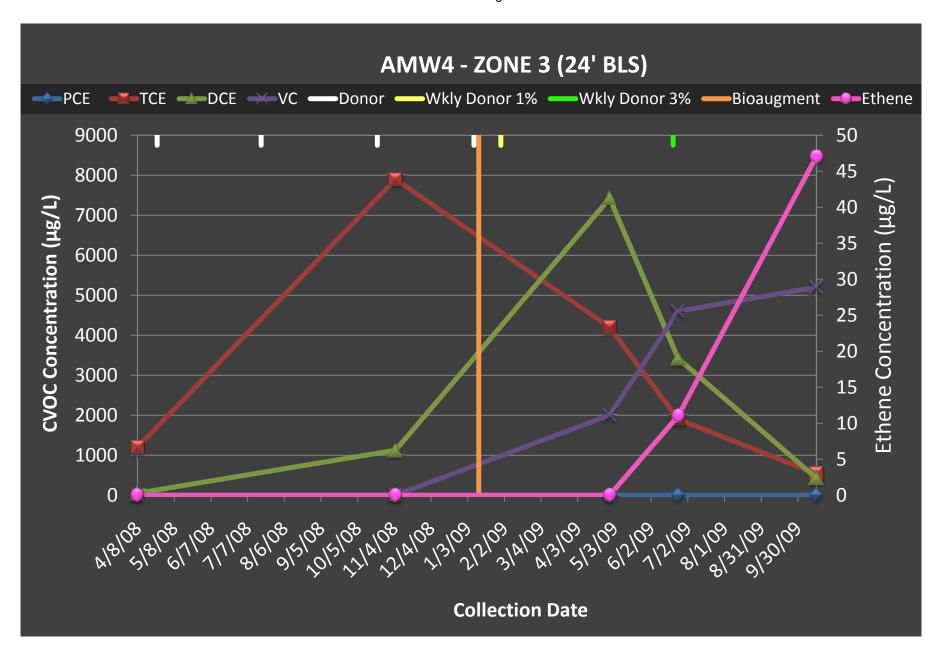
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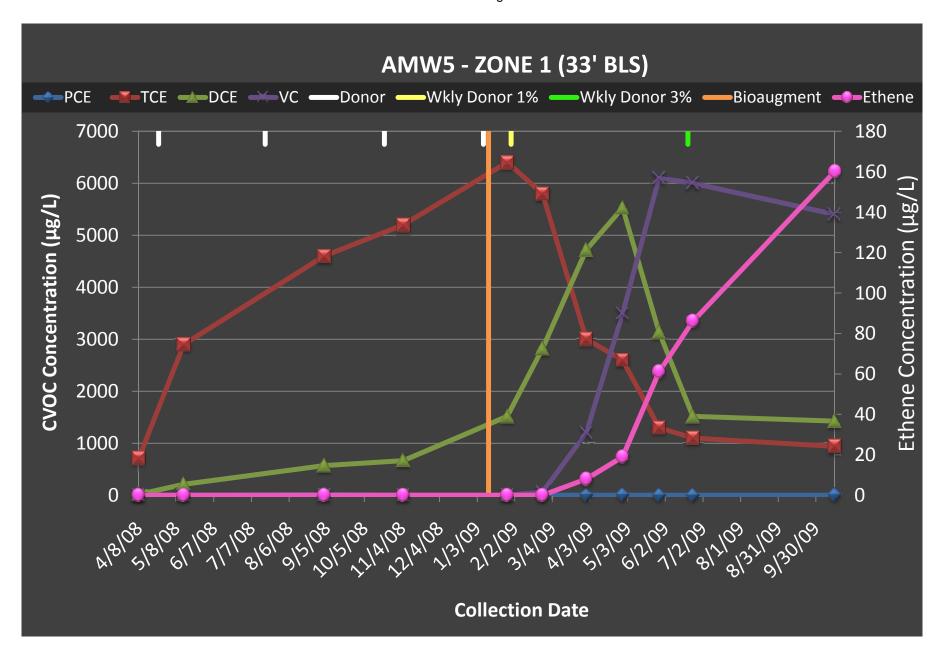
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Groundwater Bioaugmentation



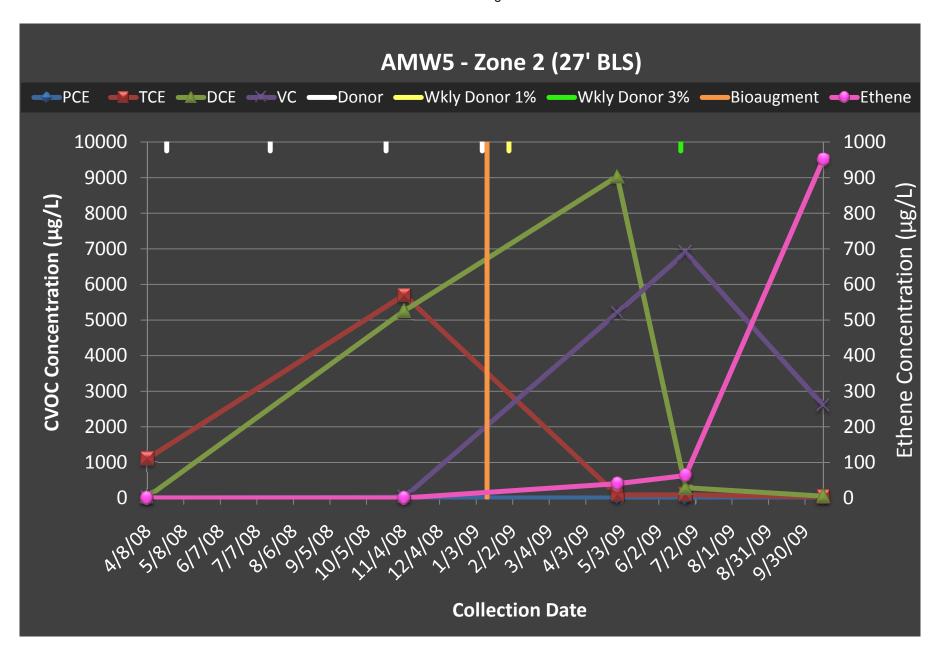
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Groundwater Bioaugmentation



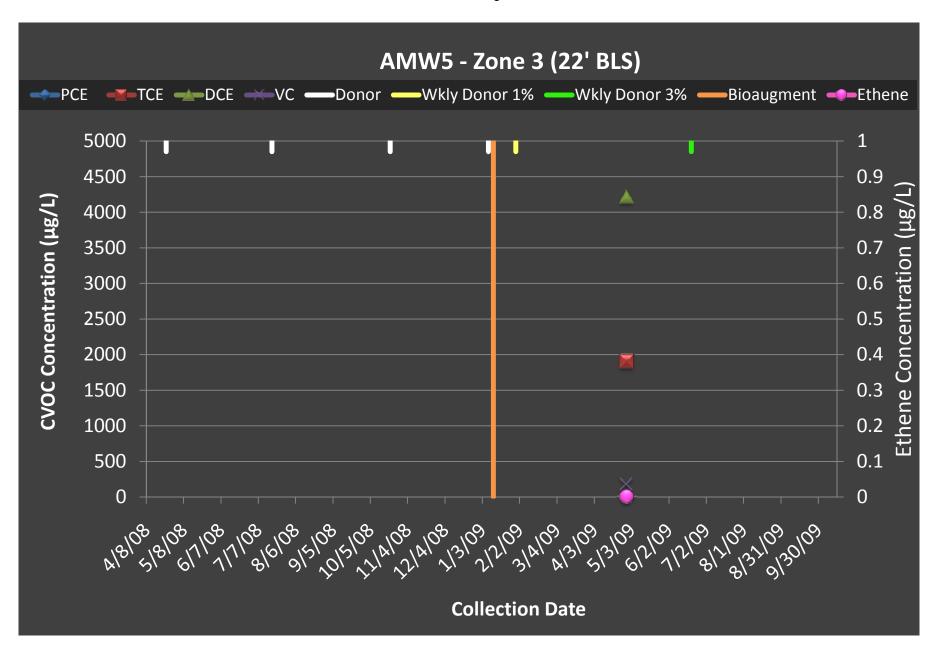
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Groundwater Bioaugmentation



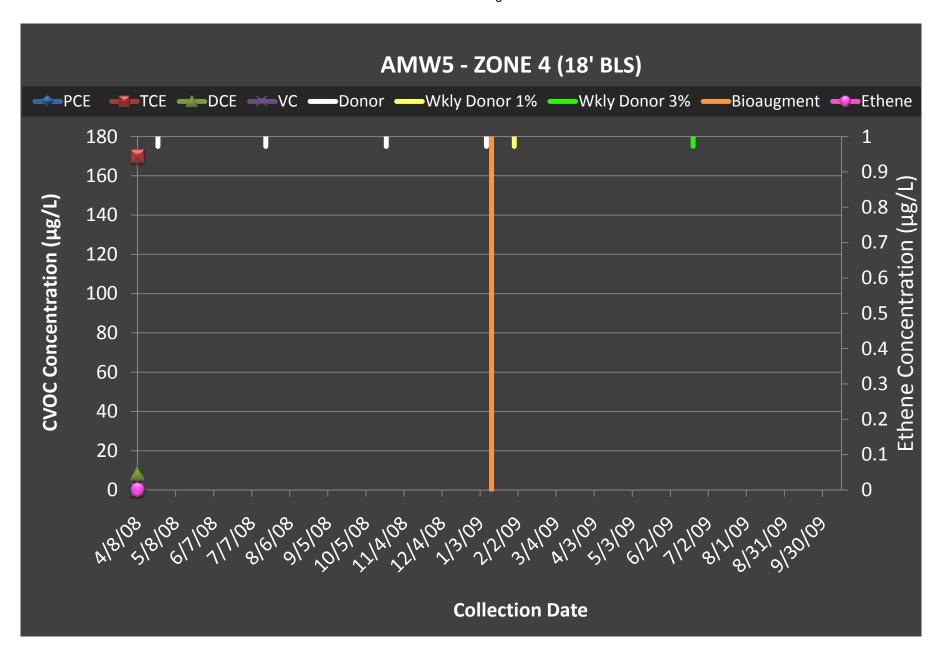
Seal Beach
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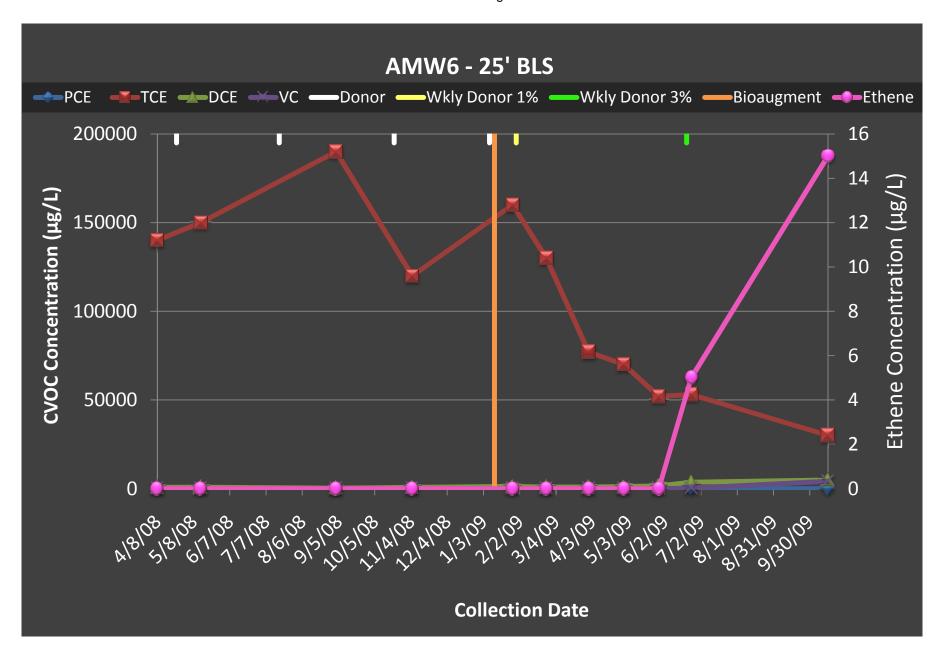
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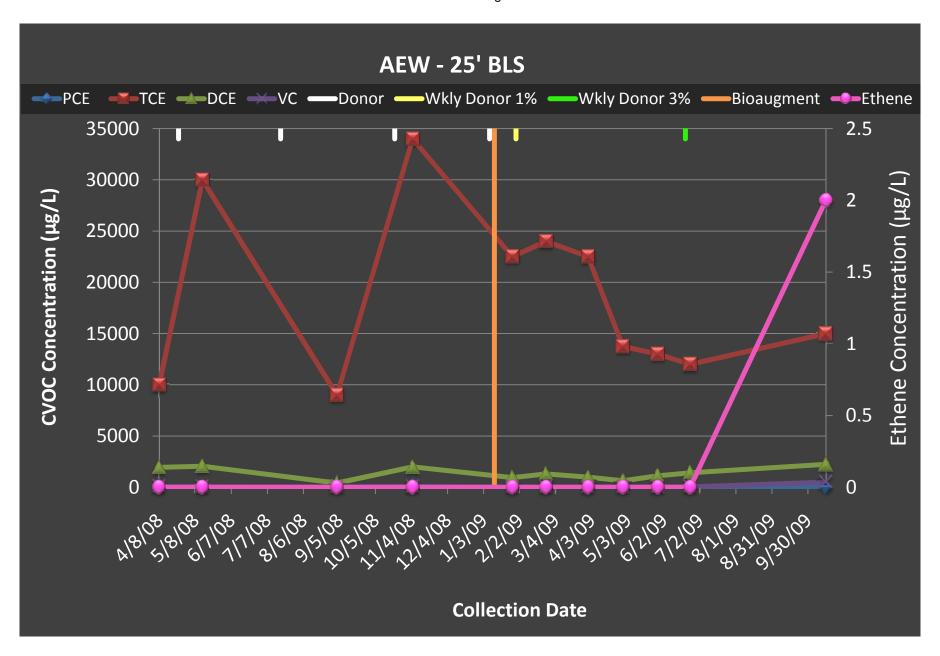
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Groundwater Bioaugmentation

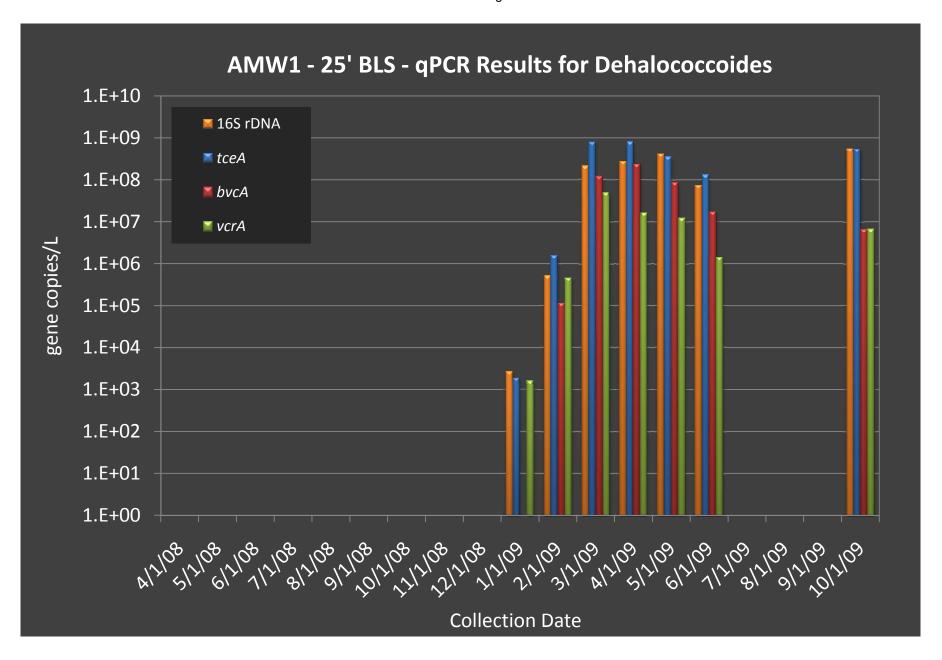


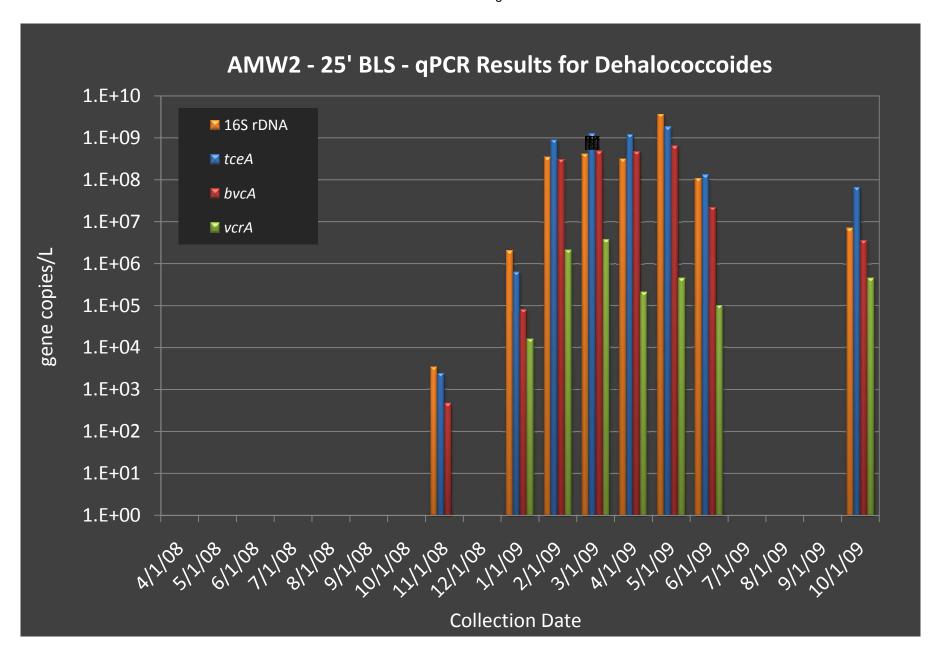
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Groundwater Bioaugmentation

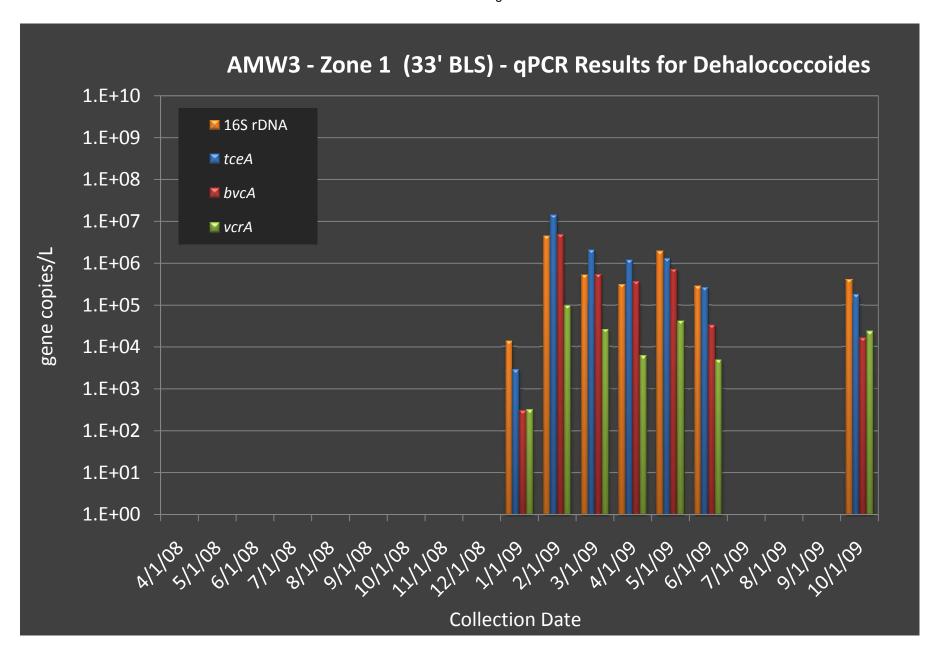


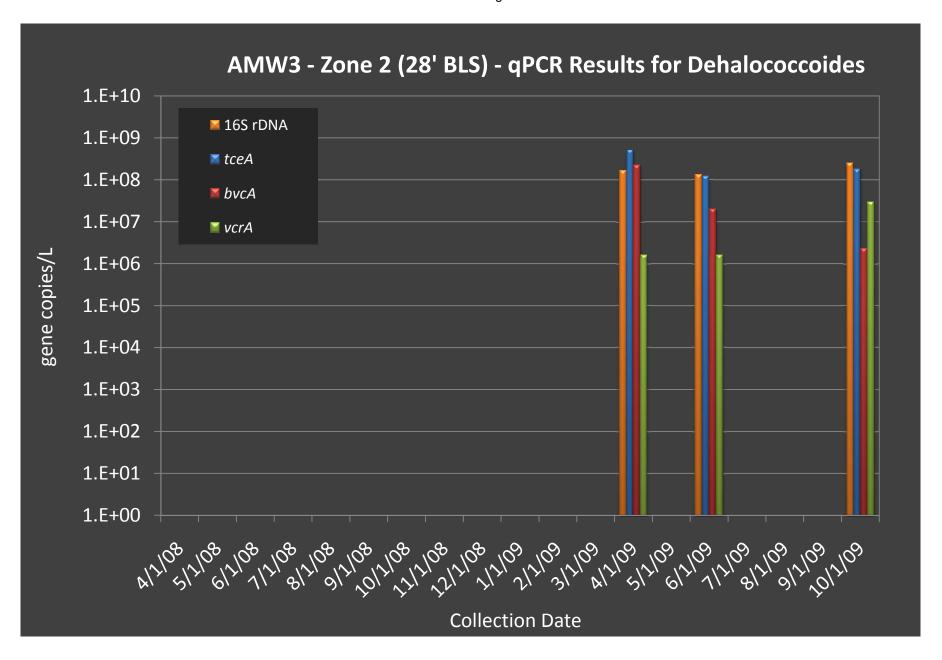
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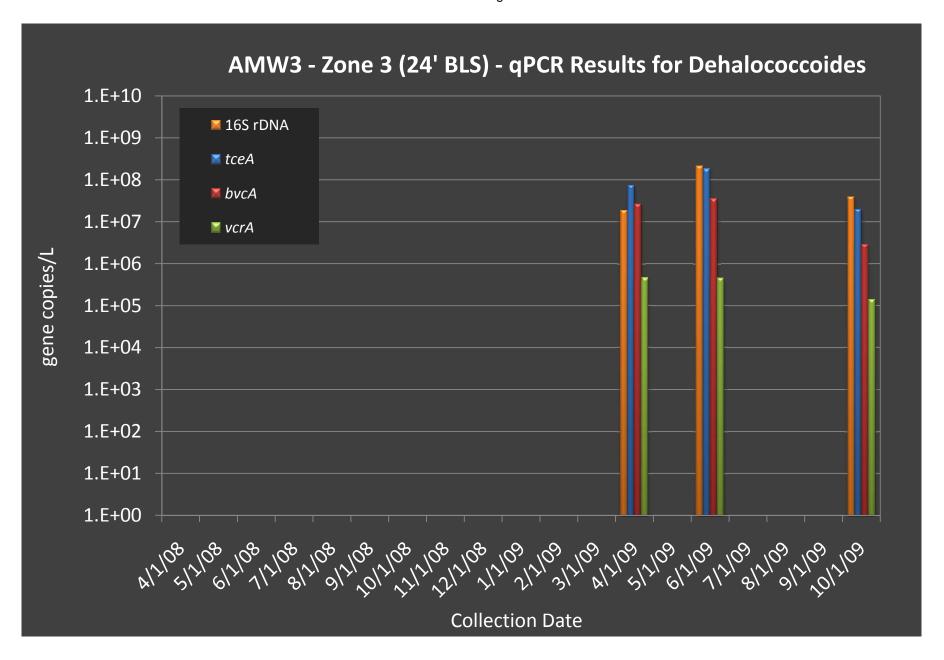


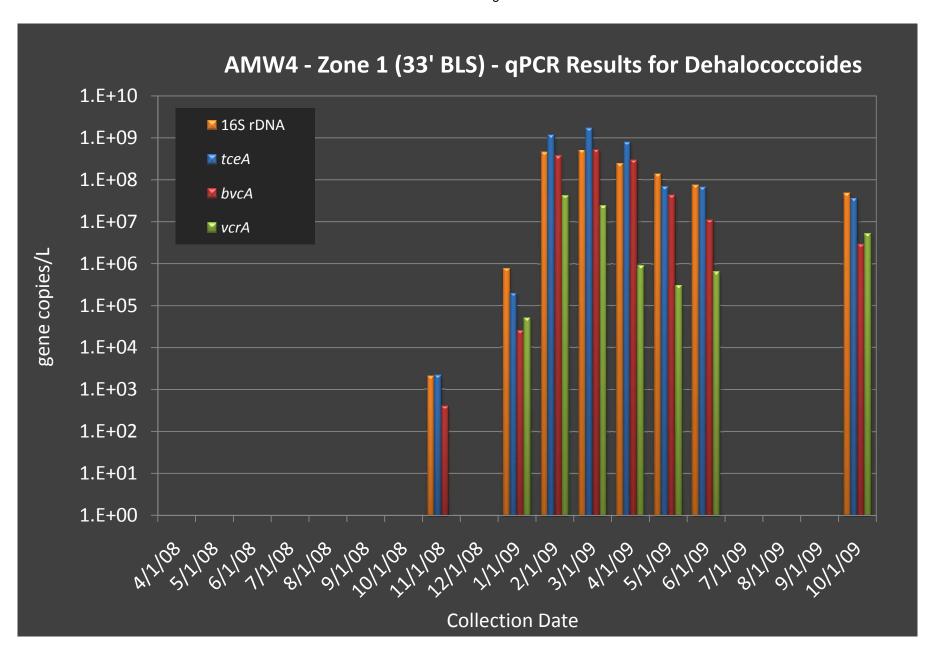


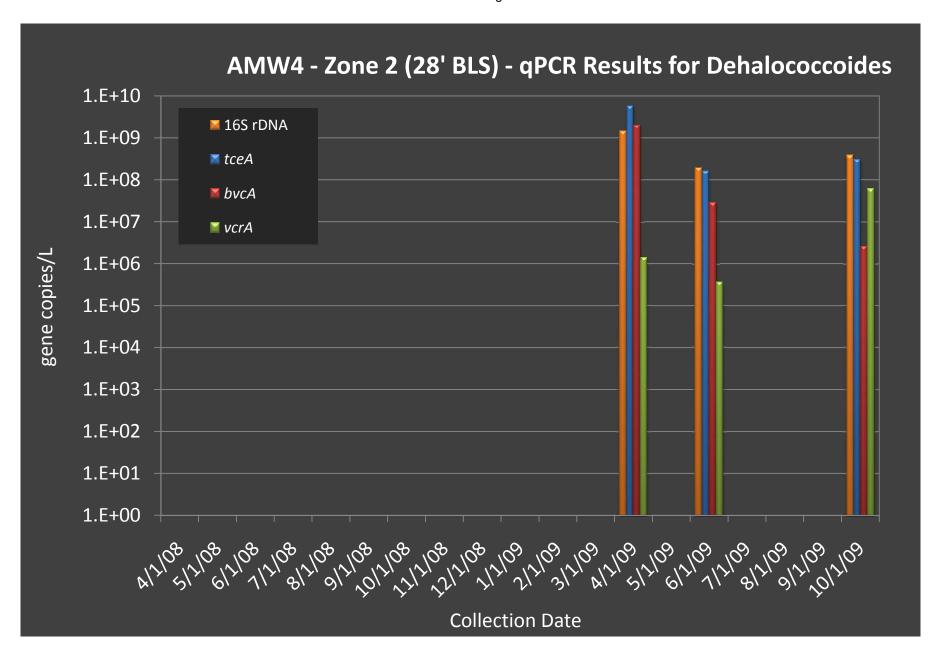


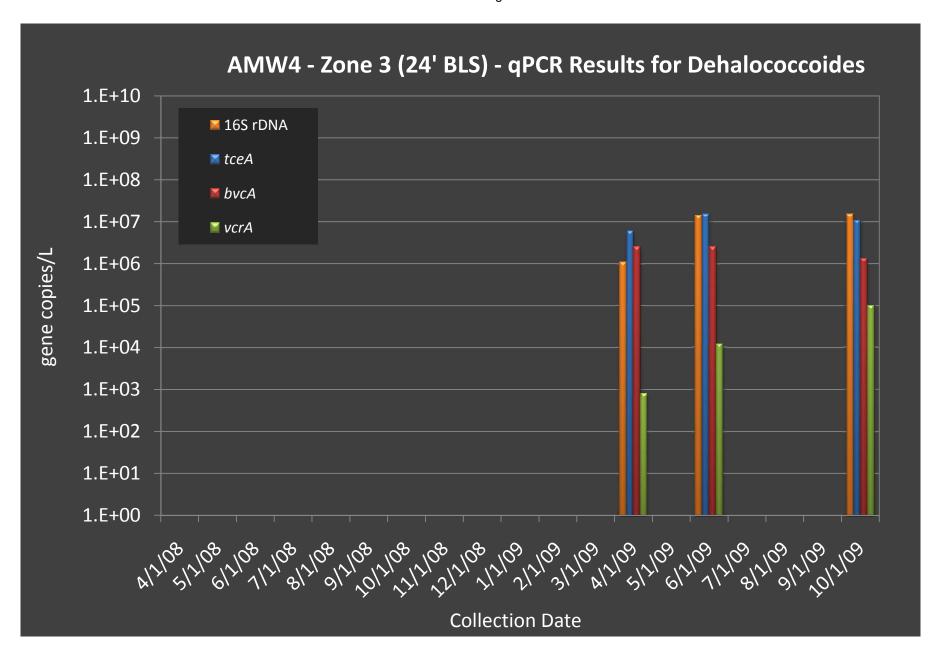


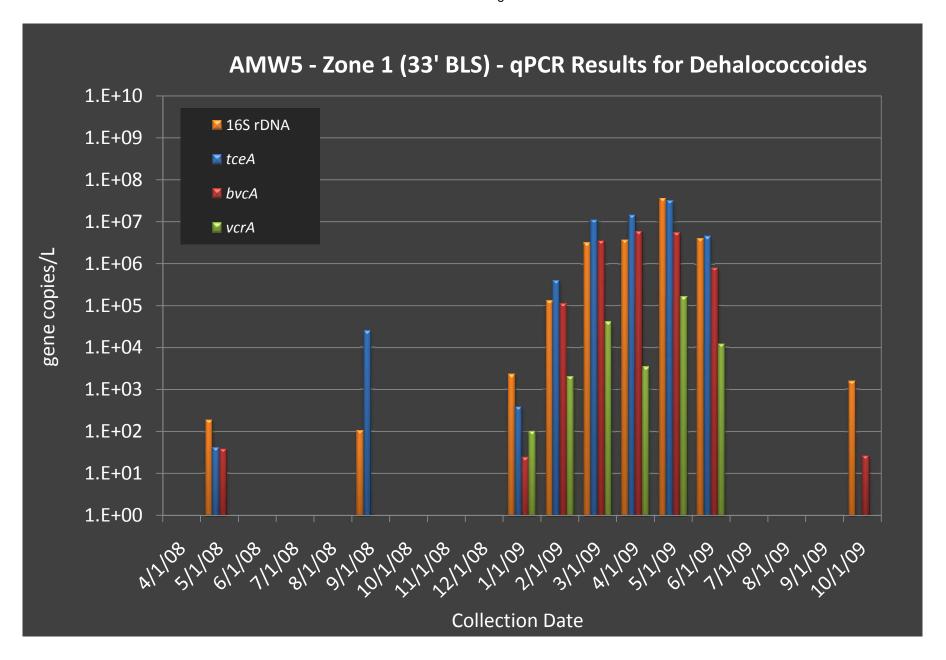


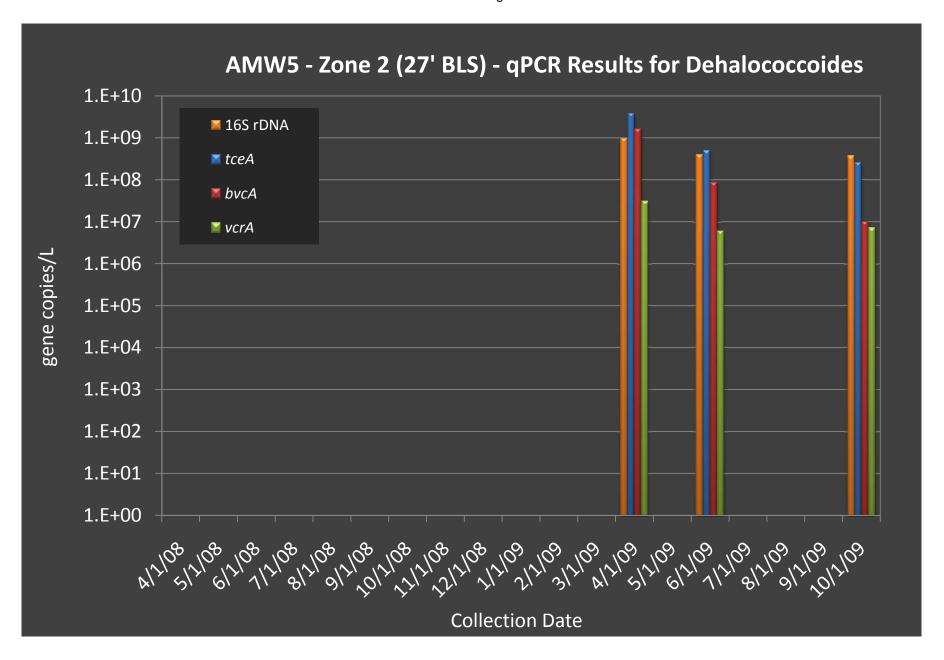


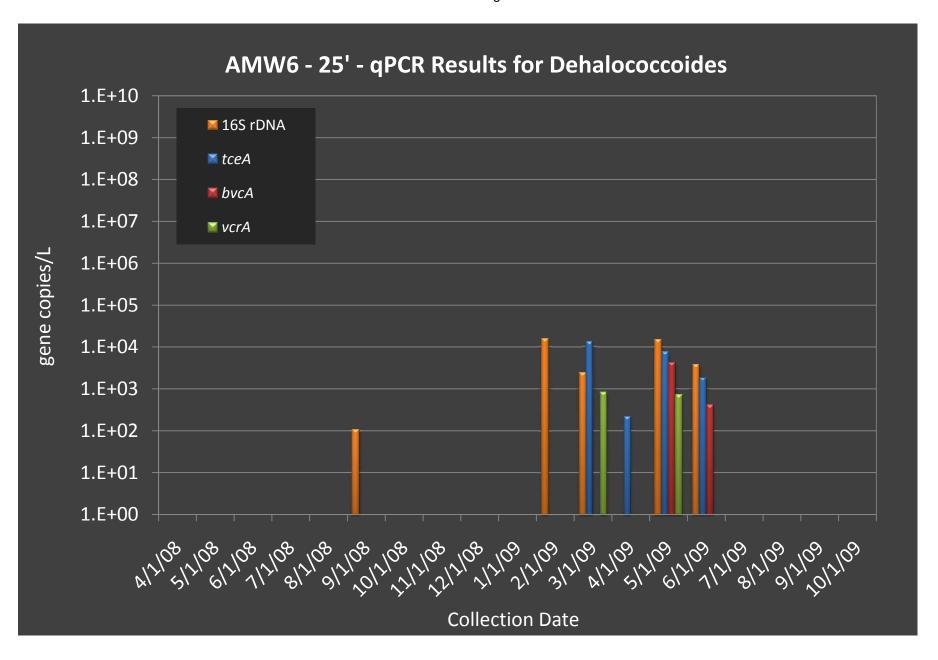


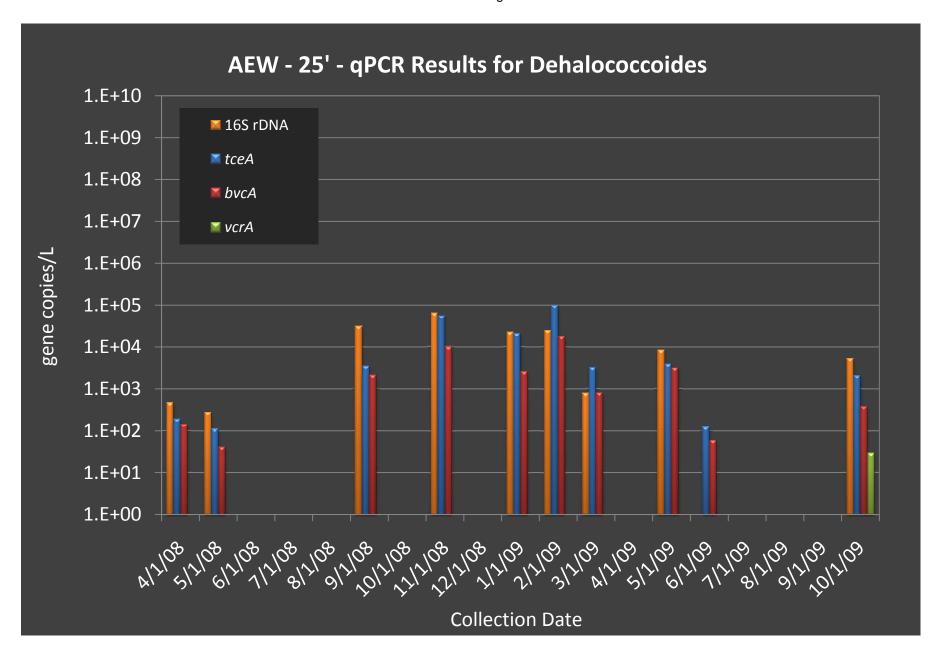


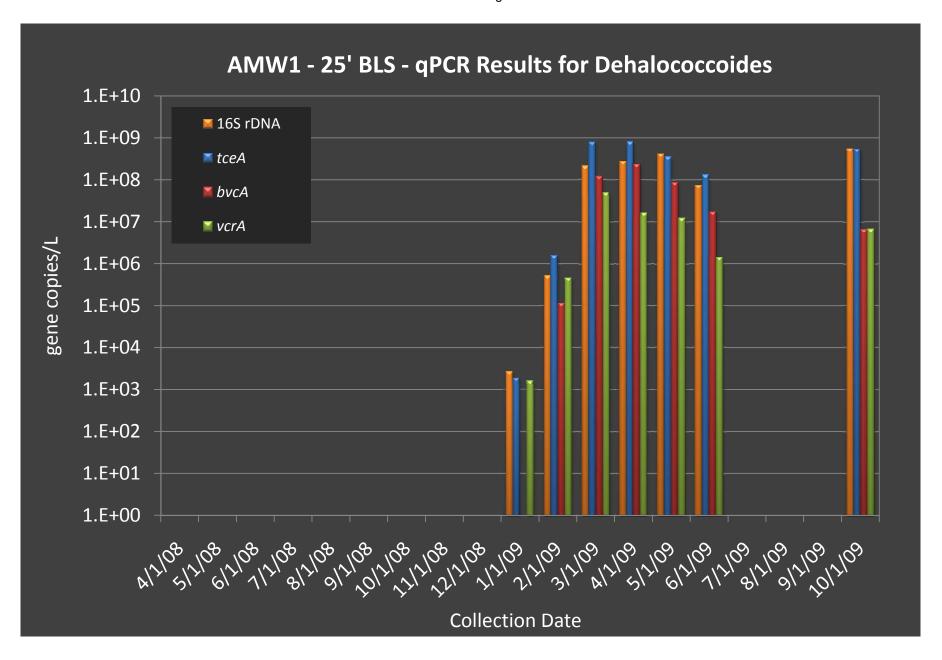


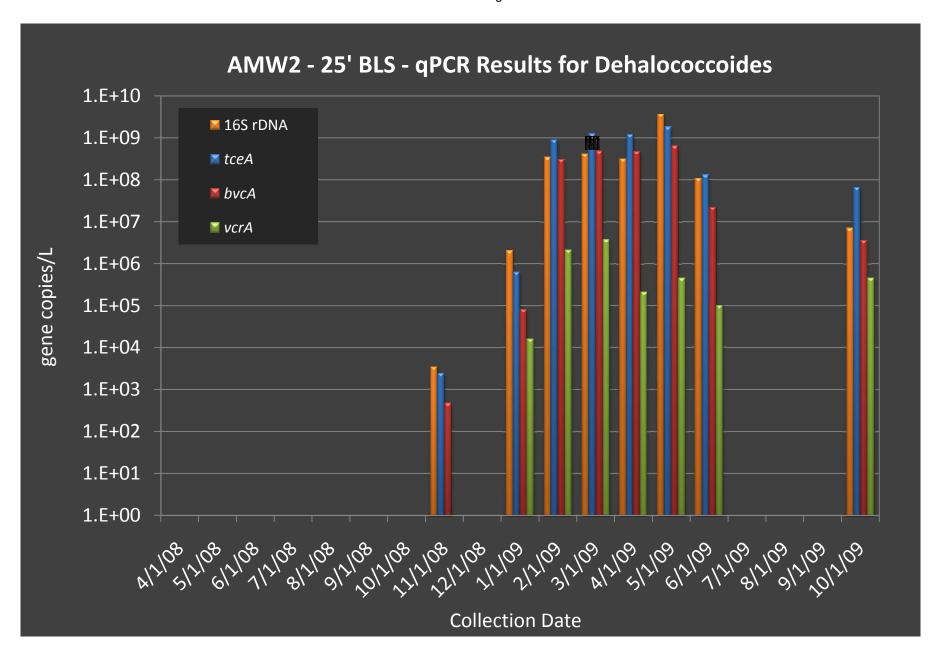


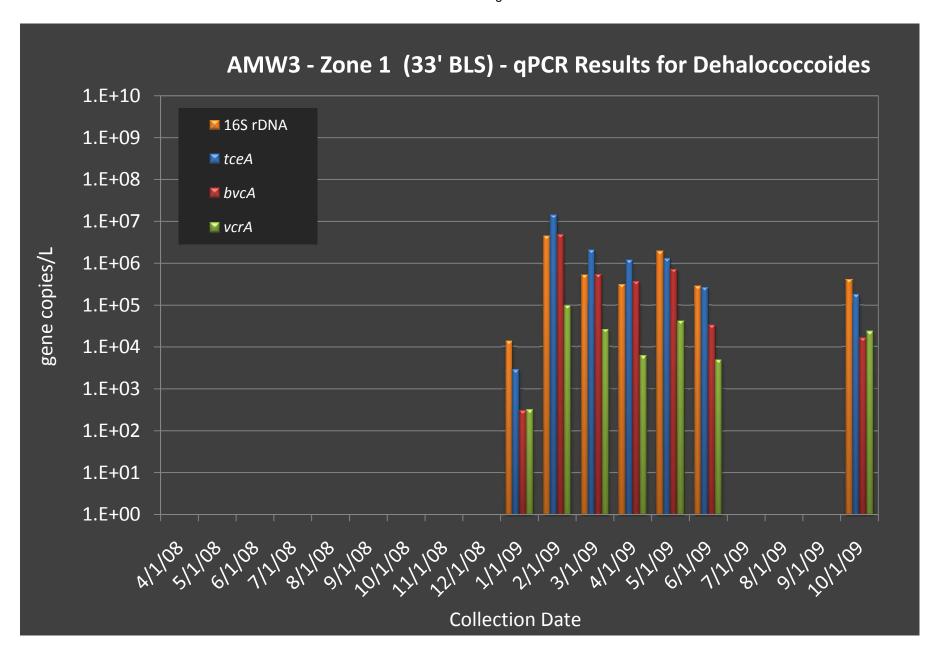


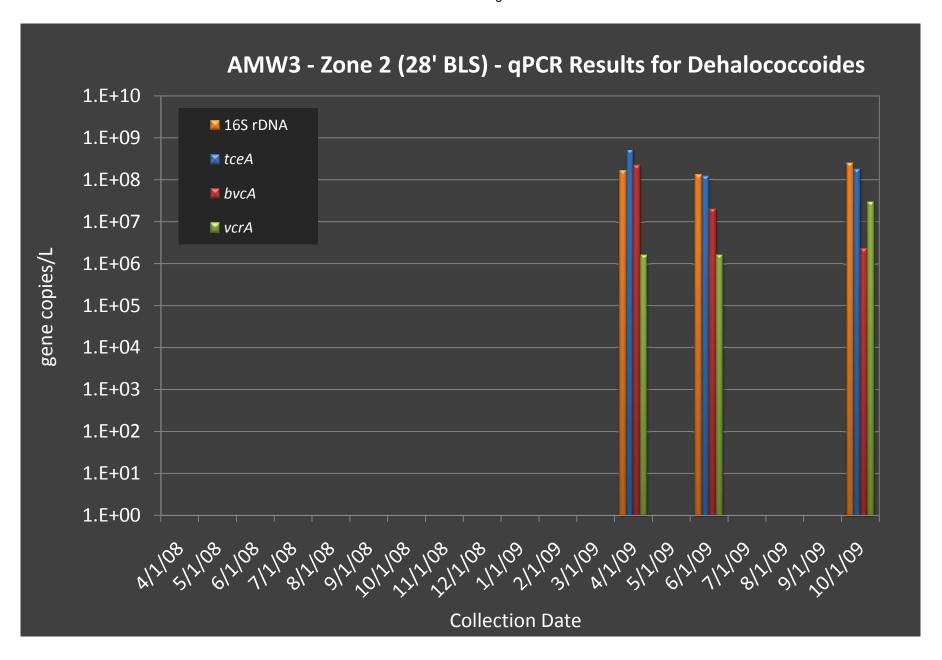


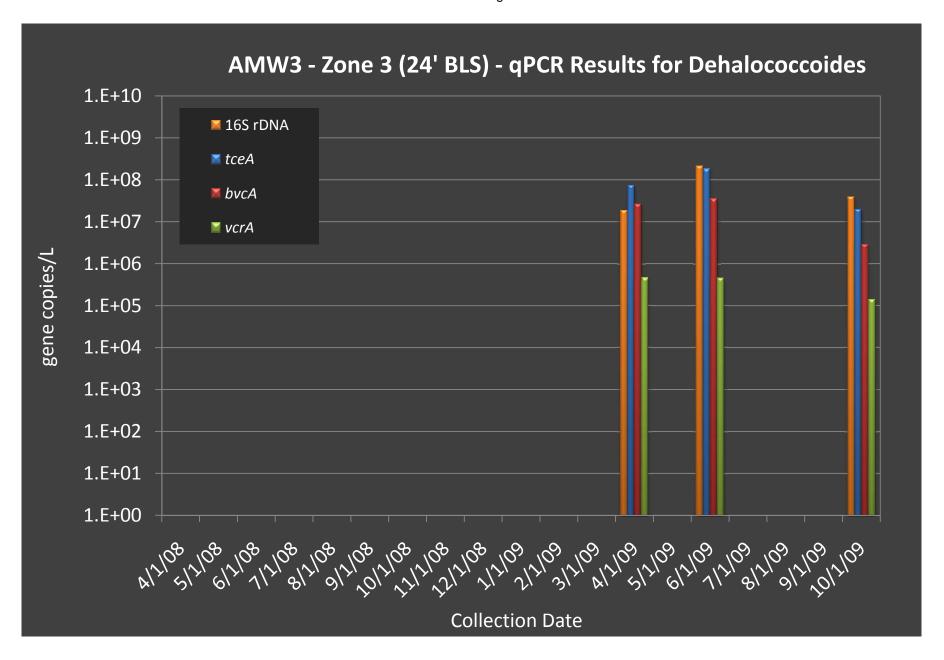


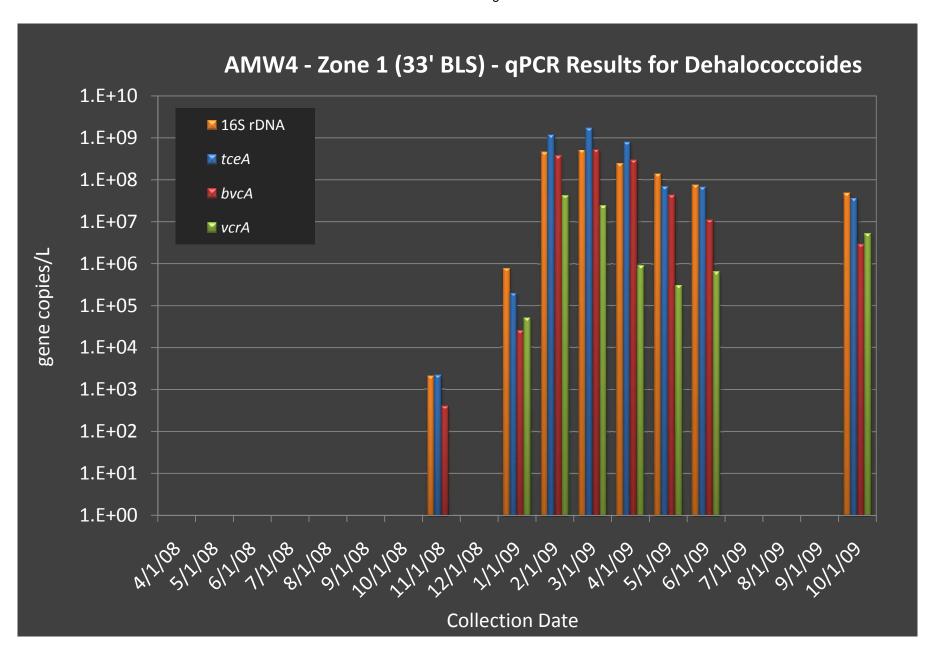


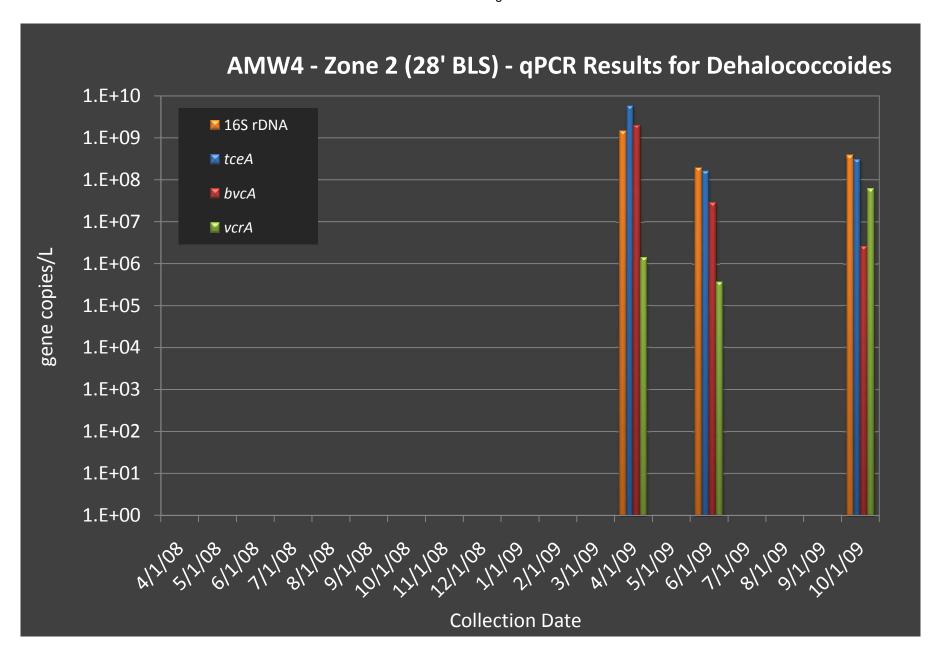


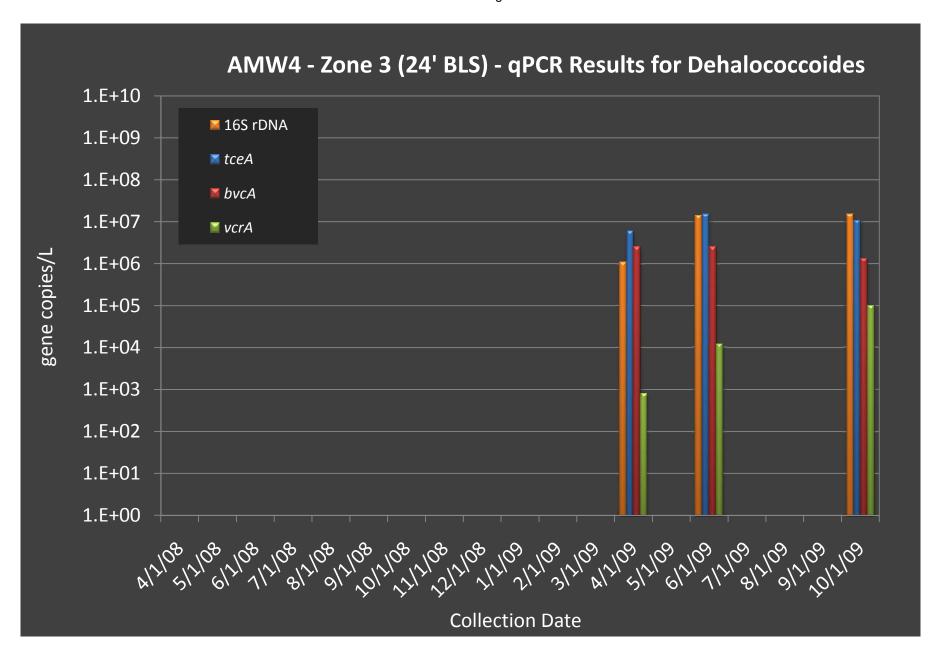


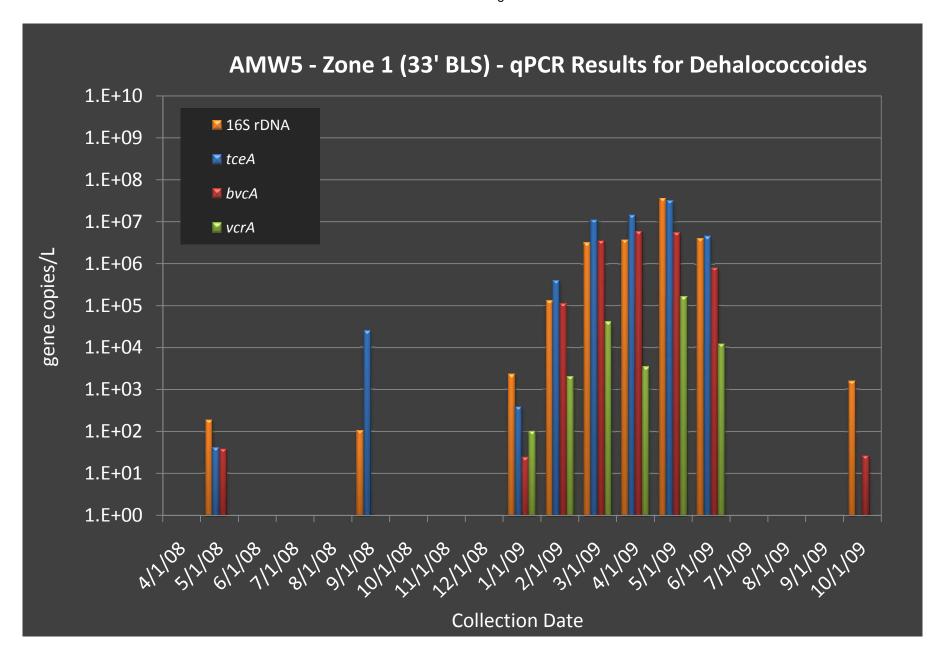


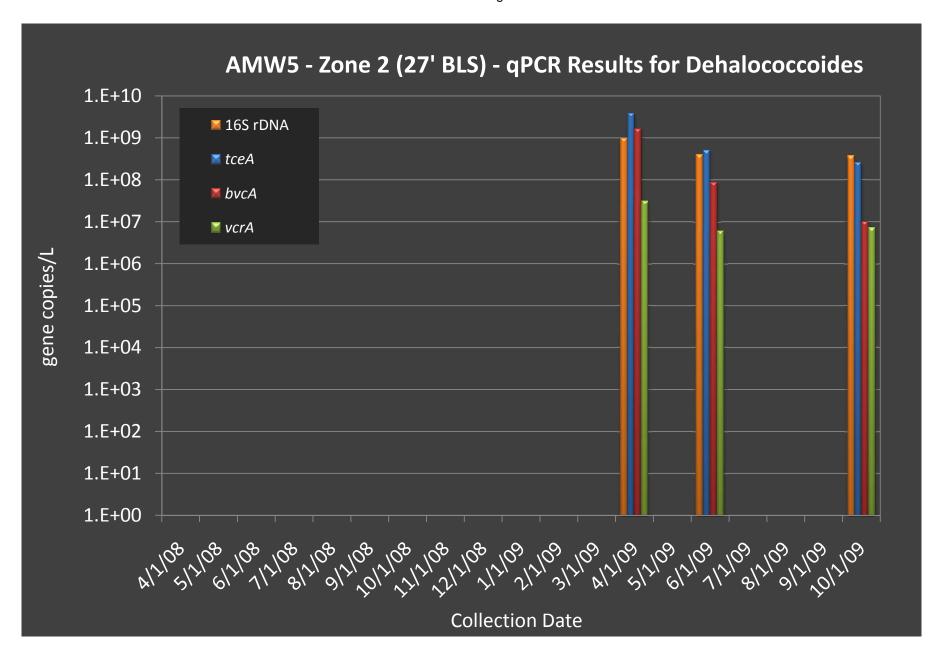


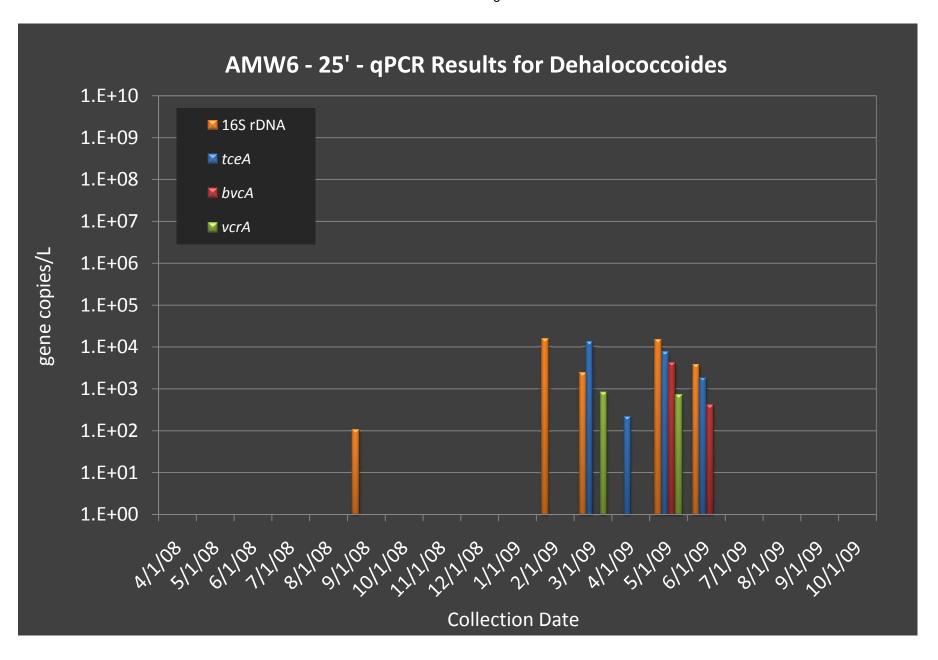


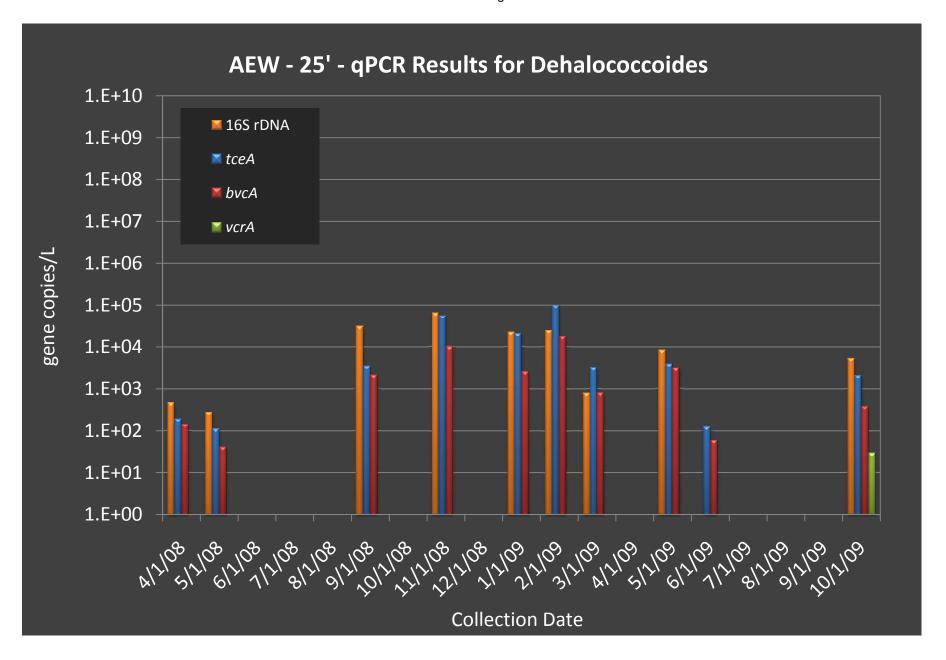




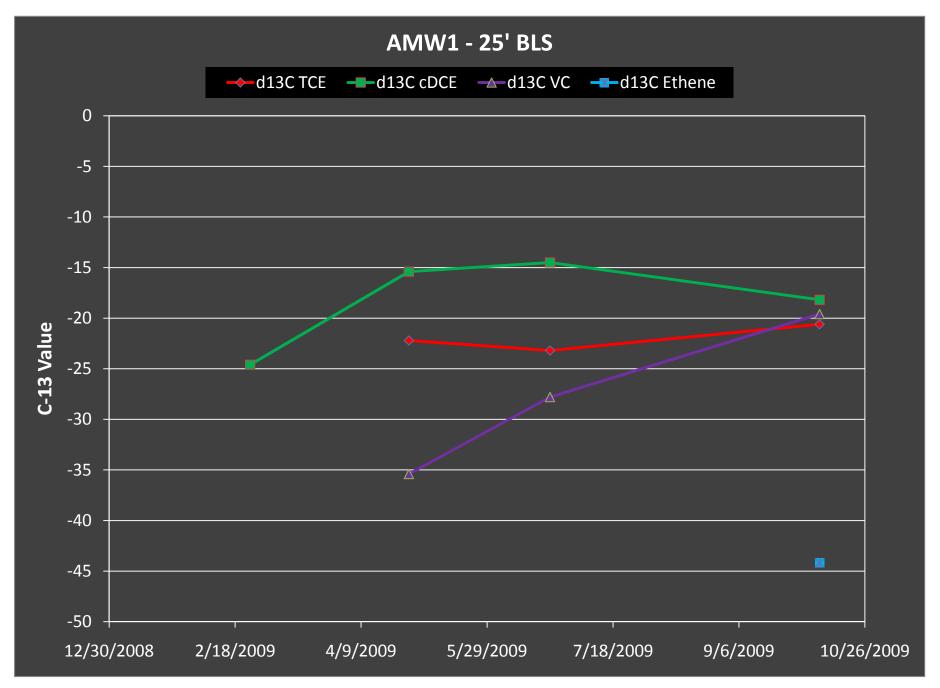


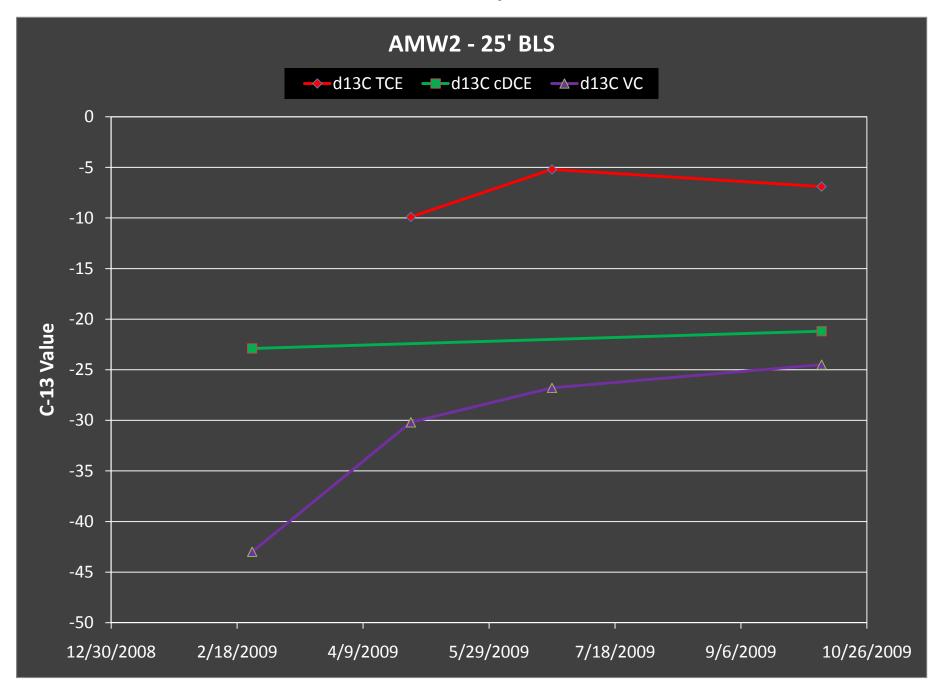


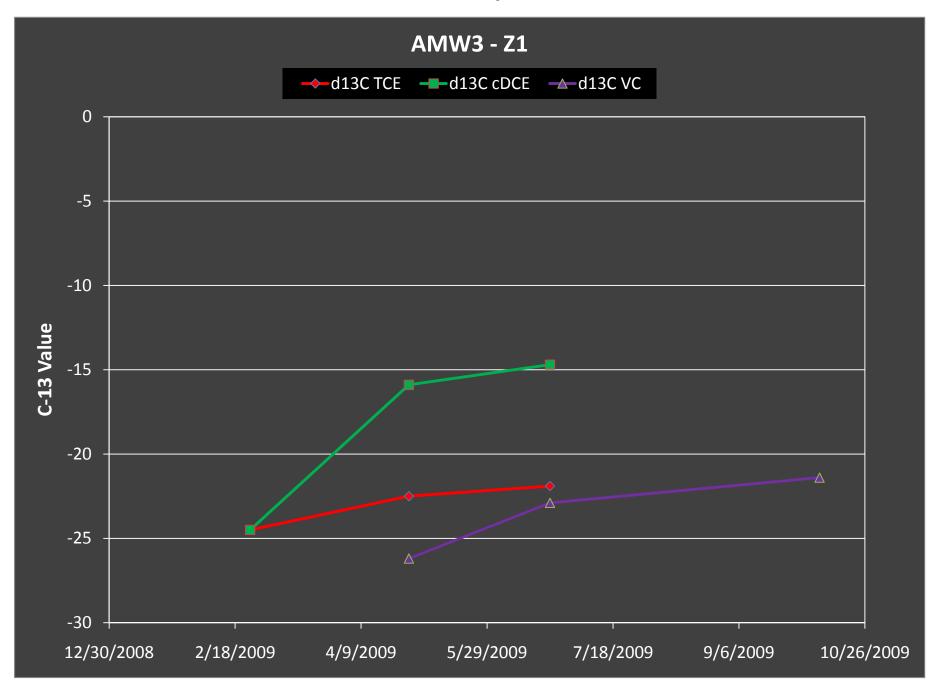




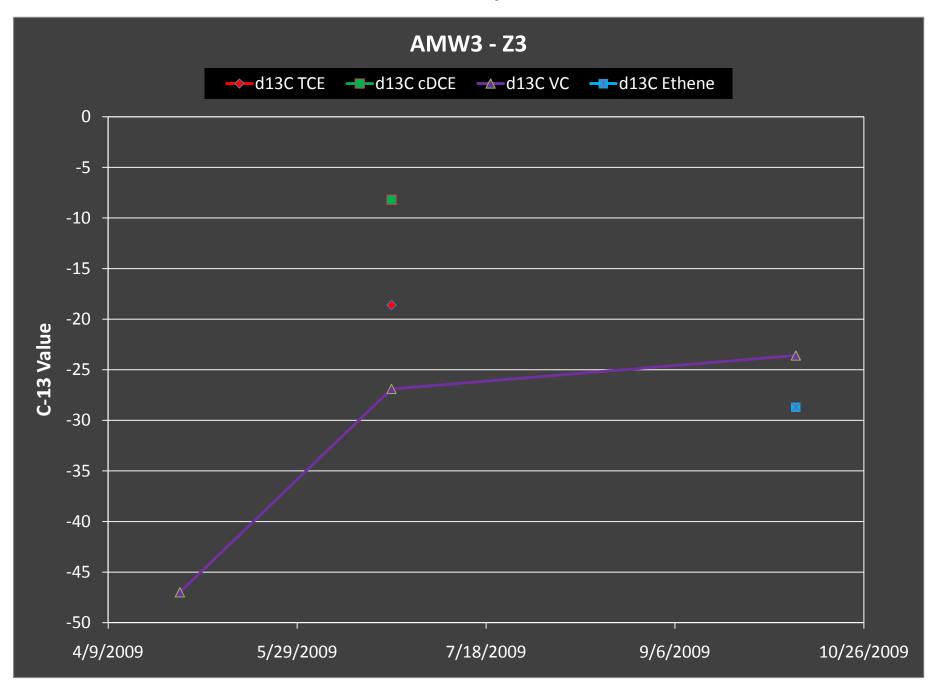
CSIA Results



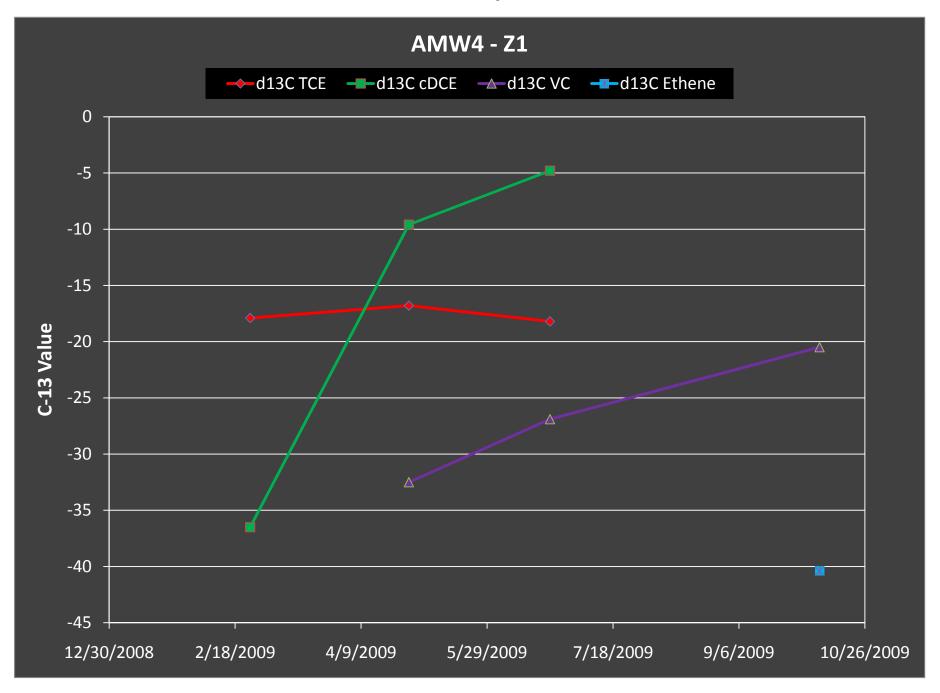


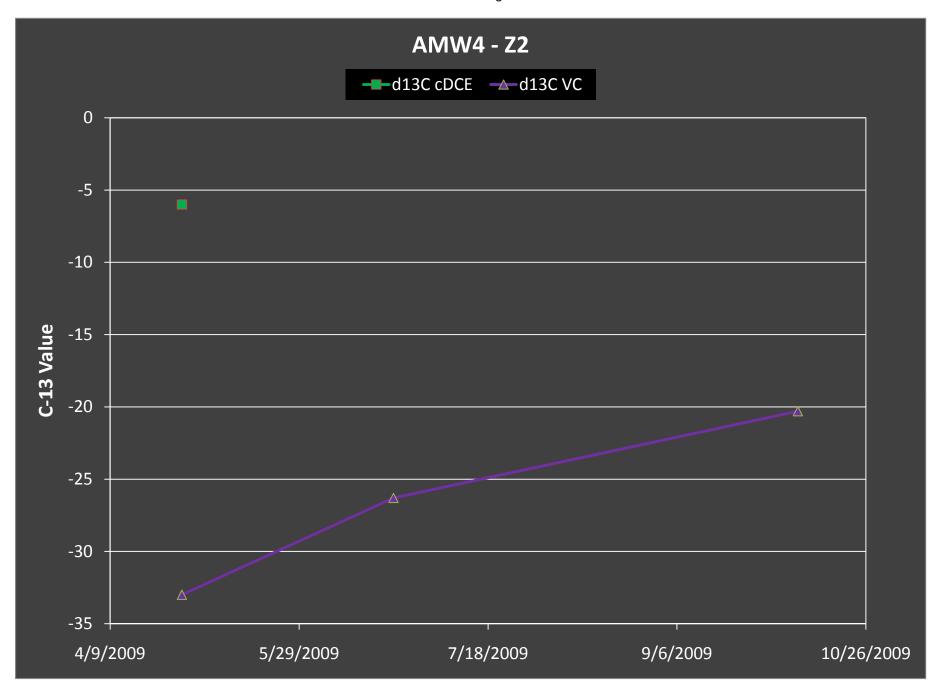






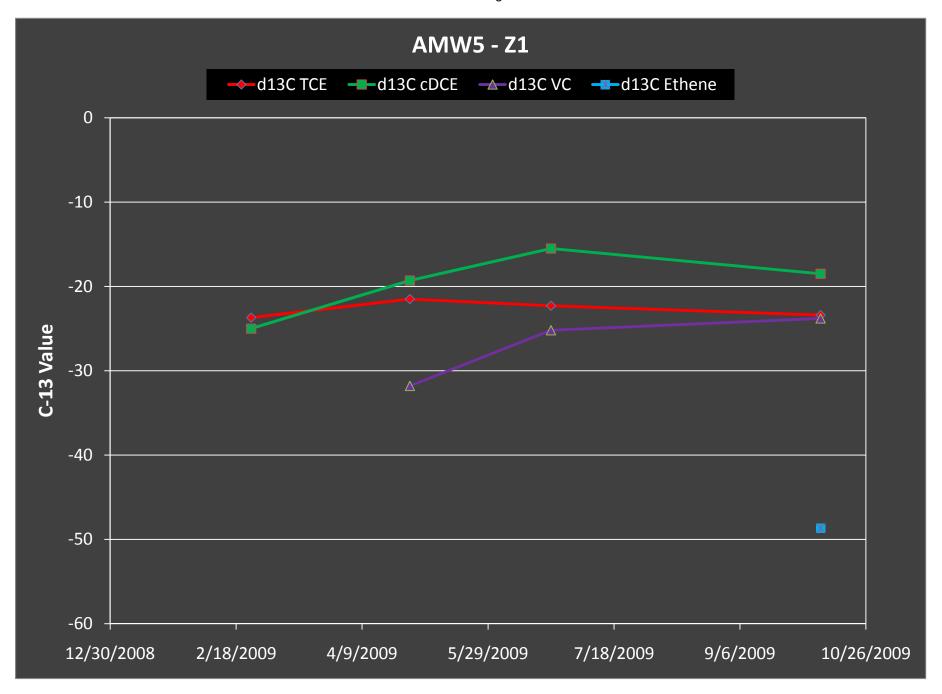
Seal Beach
Groundwater Bioaugmentation



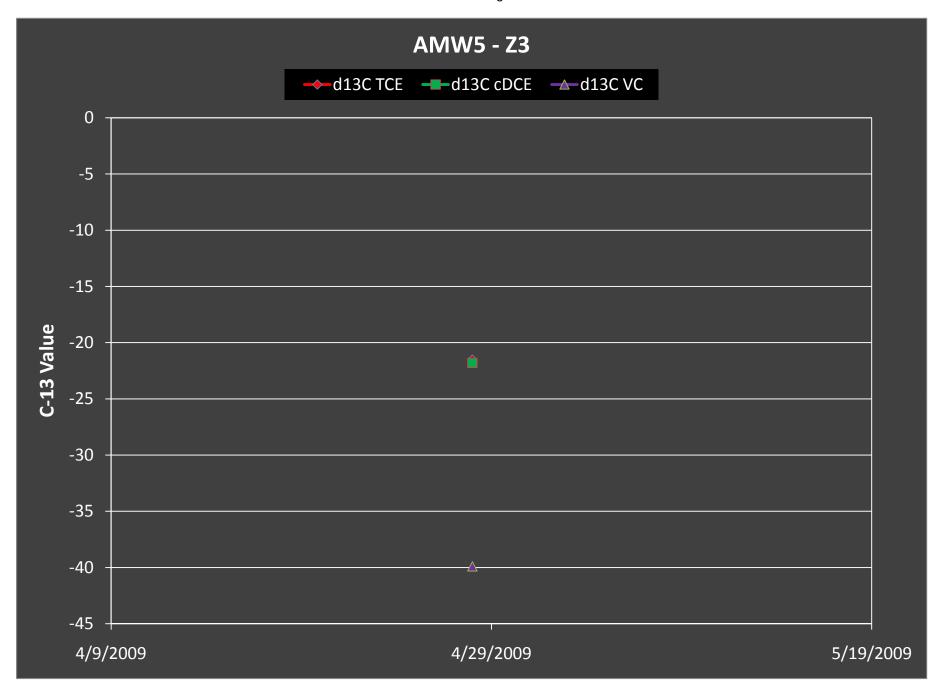


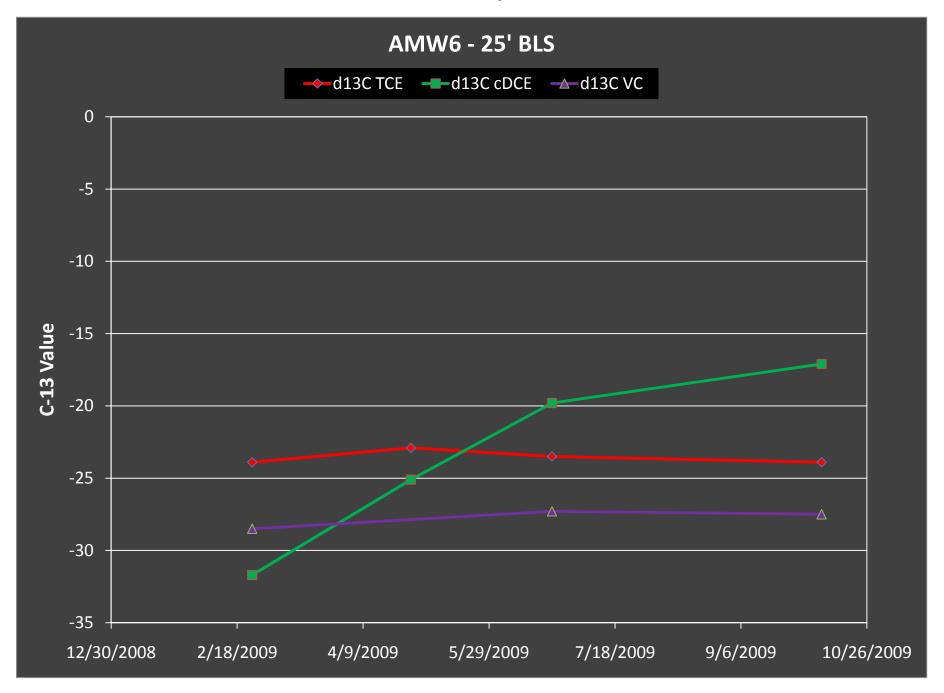


Seal Beach
Groundwater Bioaugmentation









Appendix I Passive Cell Concentration Trends

PASSIVE	CELL	_														4)	-	4)			I		
Monitoring		Tetrachloroether	Trichloroethene	cis-1,2- Dichloroethene	trans-1,2- Dichloroethene	Φ									Dehalococcoide s - 16S rRNA	Dehalococcoide s - tceA	Dehalococcoide s - bvcA	Dehalococcoide s - vcrA					
Summary	_	roe	the	the	the	Chloride									9	9	99	555			$\overline{}$	Conductivity (µmhos/cm)	O
NAVFAC I		hlo	roe	- roe	,2- roe	Ĭ		4	Je	ιŧζ	_	40	e	cal n	00 X	00	00	000		(Mv)	(mg/L)	ctiv s/c	s =
Weapons	· · · · · · · · · · · · · · · · · · ·	racl	old:	1,2 Ioli	ıs-1	0 /	l eue	ane	haı	äli	ate	fate	oric	emi rge nar	nalc rR	nalc	nalc A	nalc			E)	npı	1 00
Station - S	Site 70	Геtı	lrio	sis- Dict	ran Dich	Vinyl	Ethene	Ethane	Methane	Alkalinity	Nitrate	Sulfate	Chloride	Chemical Oxygen Demand) et 6S	Deha s - tceA	Deha s - bvcA	Deha s - vcrA	ЬH	ORP	20	S H	Ferrous Iron
Otation C	Units:	μg/L	μg/L	μg/L	μg/L	μg/L	μg/L	μg/L	μg/L	mg/L	mg/L	mg/L	mg/L	mg/L		gene c		0 0 >	ď	mV	mg/L	00	mg/L
	4/23/08	33	28	3.1	1.5 U	1.5 U	5 U	100	2300	1900	0.04 J	1600	350	18	ND	NS	NS	NS	7.39	190.9	0.65	6562	0
	9/5/08	49	85	5.9	0.7	0.2 J	5 U	57	1300	1600	0.25 U	2200	520	170	ND	ND	ND	ND	7.41	-26.8	0.68	6597	0
	10/16/08	51	71	4.6	0.4 J	0.5 U	5 U	47	1200	1600	0.25 U	2200	530	47	NS	NS	NS	NS	7.09	-59.8	1.18	7589	0.03
	11/3/08	38	54	3.6	1 U	1 U	5 U	5 U	94	1600	0.25 U	2200	530	99	ND	ND	ND	ND	7.37	-190.2	0.61	6903	0.02
_	1/28/09	33	35	4.2	1 U	9.4	5 U	56	2600	1500	0.25 U	2700	590	28	1.15E+07	5.51E+06	ND	1.56E+06	7.27	-67.6	2.36	8929	0.12
PIW-1	2/23/09	26	42	3.7	0.2 J	0.6 J	5 U	96	2800	1600	0.25 U	2700	600	28	2.75E+07	8.15E+07	ND	1.38E+07	7.29	-220.8	1.08	7590	0
P	3/30/09- 25 ft	21	48	10	0.3 J	0.7	5 U	120	3500	1600	0.1 U	2500	530	28	7.83E+07	3.30E+08	ND	2.42E+07	7.2	-183.6	3.31	8287	0.49
	3/30/09- 35 ft	NS	NS	NS	NS	NS	NS	NS	NS 4000 b	NS	NS	NS	NS	30	NS 5.555.07	NS 0.00E+00	NS	NS 4.055.07	7.09	-272.1	2.01	8295	0.61
	4/27/09 5/28/09	11 7.7	26 17	23 16	0.3 J	4.3	10 U	140 94	4000 b	1700 1700	0.1 U	2300 2100	490 430	25	5.55E+07 2.99E+06	2.63E+08 1.69E+06	ND 2.505+02	1.05E+07 3.00E+05	7.24 7.33	-280.4 -229	16.29 1.24	7467 6709	0.5
	6/23/09	5.7	13	8.8	0.3 J 0.3 J	11 19	2 J	59	5300 7300	1600	0.25 U 0.25 U	1900	400	32 44	5.70E+06	1.09E+06 1.00E+07	2.50E+03 ND	1.20E+06	7.48	-310.3	0.24	6757	0.1
	10/15/09	4.1	8.5	2.7	0.3 J 0.4 J	12	4 J	6	9500	1600	0.25 U	2100	460	46	2.92E+05	1.48E+05	ND	2.10E+04	7.46	-288.3	0.24	6483	0
	4/22/08	100 U	20000	73 J	100 U	100 U	5 U	8	230	600	0.23 J	3900	3600	71	ND	NS	NS	NS	6.68	403.6	1.17	15280	0
	9/2/08	22 J	6100	72	42 U	42 U	5 U	5 U	74	1400	0.25 U	3300	2300	1700	ND	ND	ND	ND	5.99	-256.8	2.02	1166	0
	10/15/08	6.3 J	4100	53	13 J	25 U	5 U	5 U	62	1900	0.25 U	3000	1900	1900	NS	NS	NS	NS	6.25	-168.3	0.78	12740	>3.3
	11/3/08	0.5 J	240	3.9	0.7 J	1.3 U	5 U	5 U	6	3100	0.1 U	600	210	4900	ND	ND	ND	ND	6.73	-236.1	0.43	6160	2.92
7	1/27/09	1.3 U	2.3	0.7 J	0.4 J	19	17	5 U	690	5000	0.1 U	1700	510	8900	2.15E+08	3.74E+08	ND	5.49E+07	6.36	-208.9	0.57	12.69	>3.3
PIW-2	2/23/09	0.4 J	12	2.5	1 U	23	5 U	5 U	1600	5300	0.1 U	990	490	8300	2.30E+09	6.98E+09	ND	1.69E+09	6.64	-345.9	0.57	10320	0
п.	3/30/09	0.5 J	41	5.5	1 U	10	5 U	5 U	1800	4400	0.5 U	1400	700	4100	6.81E+08	2.76E+09	ND	2.24E+08	6.7	-363.2	2.3	11540	0
	4/27/09	1 U	44	5.6	1 U	6		10 U	2900	4400	0.1 U	1100	570	3800	8.63E+08	3.64E+09	ND	1.14E+08	6.83	-372.1	20.71	9539	NM
	5/27/09	1.3 U	22	3.5	1.3 U	9.5 7.7	3 J 2 J	5 U	3200	5600 6000	0.25 U	1000	490	7200	1.21E+09 2.65E+08	1.05E+09	ND	1.70E+08 3.40E+07	6.34	-356.2	46.31	10940	0
	6/22/09 10/15/09	0.5 U 1 U	17 11	2.9 2.8	0.2 J 0.3 J	8.7	4 J	5 U	1800 6300	2800	0.25 U 0.1 U	1000 930	480 490	8600 920	2.65E+08 2.48E+07	1.60E+09 1.63E+07	ND ND	3.40E+07 1.90E+06	6.69 5.45	-351.8 -344.7	0.83	12000 6449	0
	4/23/08	17 U	11000	82	7.5 J	17 U	5 U	10	150	620	0.1 U	3100	1800	30	2.46E+07 ND	NS	NS	NS	6.57	101.6	0.63	10219	0
	9/5/08	83 U	11000	92	83 U	83 U	5 U	8	170	800	0.25 U	2700	1500	390	ND	ND	ND	ND	5.82	-139.2	1.05	7748	3.03
	10/15/08	31 U	12000	85	8.8 J	31 U	5 U	7	140	1200	0.25 U	2100	1500	740	NS	NS	NS	NS	6.44	-290.1	0.7	8362	>3.3
	11/3/08	2 U	270	260	2 U	2 U	5 U	5 U	14	3300	0.1 U	15	89	5700	ND	ND	ND	ND	6.64	-249.3	0.68	6052	>3.3
	1/27/09	0.5 U	1.4	1.1	0.3 J	30	24	5 U	2500	4800	0.1 U	8.9	65	8900	1.91E+09	9.89E+08	ND	4.37E+08	6.33	-209.2	0.5	9334	3.13
_	1/27/2009-K	1.3 U	1.4	0.5 J	1.3 U	25	26	5 U	2300	4800	0.1 U	11	67	10000	3.12E+09	1.80E+09	ND	8.19E+08					
PIW-3	2/23/09	1.3 U	2.1	1.4	1.3 U	4.7	5	5 U	3100	5300	0.1 U	1 U	14	12000	1.52E+09	5.27E+09	ND	1.17E+09	6.35	-211.7	2.43	8363	3.05
Ğ	3/30/09	1.3 U	6.5	2.4	1.3 U	4.2	5 U	5 U	3900	2800	0.1 U	0.77 J	190	4800	3.45E+08	1.17E+09	ND	8.20E+07	6.25	-292.4	2.01	4889	>3.3
	4/27/09	1.3 U	1.2 J	1.3	1.3 U	6.1		25 U	4600	3600	0.1 U	0.14 J	68	5900	1.11E+09	5.09E+09	ND	1.52E+08	6.32	-280.3	4.72	7098	3.3
	4/27/09-K	1.3 U	1.0 J	1.3	1.3 U	5		25 U	4200	3500	0.1 U	0.12 J	70	5900	1.36E+09	6.06E+09	ND	1.71E+08	0.0	054.0	44.00	0004	2.45
	5/27/09 6/22/09	0.5 U 0.5 U	3.4 5.4	7.1 2.2	0.3 J 0.2 J	3.8	35 6	4 J 5 U	2900 2700	4300 7000	0.05 U 0.1 U	2.2 8.1	98 100	7300 14000	2.89E+08 5.29E+07	1.69E+08 7.40E+07	ND ND	3.90E+07 8.60E+06	6.2 6.53	-251.6 -202.3	11.66 0.73	6934 11030	3.15 2.98
	10/13/09		0.7	2.7	0.2 J	4.2	19	3 J	10000	2700	0.10 0.05 U	0.61	130	2100	1.48E+07	5.45E+07	ND	1.90E+06	6.82	-94.3	0.75	4477	>3.3
	10/13/2009-K	0.5 U	1	3.4	0.4 J	5.2	18	2 J	10000	2700	0.05 U	2.3	130	2100	1.81E+07	1.16E+07	ND	2.50E+06	0.02	34.3	0.00	77//	70.0
	4/23/08	18	1100	48	6.5 J	10 U		5 U	35	1400	0.53	3800	1000	24	ND	NS	NS	NS	6.9	161.1	0.45	10673	0
	4/23/2008 - K	19	1200	49	6.4 J	8.3 U		5 U	37	1400	0.51	3800	1000	28	NS	NS	NS	NS					0
	9/5/08	15	2000	66	8.8 J	10 U	5 U		29	890	0.7	4300	1400	32	ND	1.03E+01	ND	ND	6.45	146.2	0.6	9118	0
	10/16/08	13	1800	55	9.8	3.1 U	5 U	5 U	27	880	0.52	4300	1300	23	NS	NS	NS	NS	6.45	102.7	1.2	10630	0
-	11/4/08	11 J	1600	64	10 J	17 U	5 U	5 U	14	880	0.72	4700	1400	25	ND	ND	ND	ND	6.69	159.8	0.83	9533	0
PMW-1	1/28/09	9.9 J	1500	54	9.3 J	10 U	5 U	5 U	31	850	0.67	4800	1300	28	3.48E+02	4.88E+01	ND 0.77F : 00	7.16E+01	6.66	-19.7	2.41	11560	0
₽	2/23/09	11	1700	79	14	10 U	5 U	12	320	870	0.76	5100	1400	25	ND	ND 6 20E+02	3.77E+02	ND	6.64	-266	1.05	10439	NM
	3/30/09 4/28/09	10 J 13	1400 1400	64 65	12 J 9.4 J	13 U	5 U	6	150 140	880 850	0.66	4600 4400	1300	23 30	1.58E+03 1.33E+02	6.28E+03 7.17E+02	ND ND	4.18E+02 4.90E+01	6.55 6.57	-85.9	2.77 0.58	10900 10280	0 NM
	5/28/09	12	1500	76	9.4 J 9.6 J	10 U 10 U	5 U	7 8	160	870	0.4 J 0.4 J	4300	1200 1100	17	6.46E+01	2.67E+02	ND	1.40E+02	6.44	91.6	0.56	9504	NM 0.02
	6/23/09	7.1 J	1400	69	8.9 J	13 U	5 U	5 J	85	830	1 U	4300	1100	57	8.83E+02*	8.50E+02	ND	1.40L+02 1.5E+02*	6.9	-185.4	0.72	9596	0.02
	10/15/09	6.2 J	1600	67	12	10 U	5 U	2 J	22	910	0.85	4900	1300	61	1.41E+02*	ND	1.3E+01*	2.8E+01*	6.68	-36.2	0.83	10190	0
	4/22/08	11 J	2600	61	18 J	25 U	5 U	5 U	15	1100	0.11 J	3400	1400	28	ND	NS	NS	NS	6.94	483.9	1.01	10928	0
	9/5/08	20	3400	74	16 J	20 U	5 U	5 U	39	900	0.5 U	5400	2200	42	ND	ND	ND	ND	6.66	85.2	0.3	1194	0
	10/16/08	15	2900	64	13	5 U	5 U	5 U	60	1100	0.25 U	4400	1800	53	NS	NS	NS	NS	6.77	-63.1	1.65	10610	1.01
	11/4/08	15 J	2600	64	10 J	25 U	5 U		7	1000	0.5 U	5000	2200	120	ND	ND	ND	ND	7.07	-74.4	2.67	10870	2.19
0.1	11/4/2008 - K	19	3000	65	15	10 U	5 U	5 U	71	970	0.5 U	5100	2300	110	ND	ND	ND	ND	7.0	40.0	4 70	4040	0.50
PMW-2	1/28/09	13 J	2300	41	7.9 J	17 U	5 U	5 U	59	1200	0.25 U	2700	1200	36	ND	ND ND	ND	ND ND	7.6	-49.9	1.73	4312	0.53
Ž	2/23/09	16 J	1800	43 7.8	7.3 J	17 U	5 U	5 U	70 5.11	200	0.05 U	550	240	53	ND ND	ND 1.07E+04	ND ND	7.23E+02	7.49	-155.9	0.87	5265	2.72
п.	3/30/09 4/28/09	1.9 3.8	88 280	7.8 6.5	0.4 J 0.5 J	0.5 U 1 U	5 U	5 U	5 U 5 J	1500 1500	0.05 J 0.1 U	390 440	51 87	13 19	2.23E+03 1.06E+02	1.07E+04 1.12E+03	ND ND	7.23E+02 4.21E+01	8 8.25	-110.6 36.9	6.14 1.46	3495 3339	0.01
	4/28/09-K	3.3	370	9.1	0.5 J	2.5 U	5 U	5 U	12	1699	0.1 U	1000	460	84	ND	ND	ND ND	4.21E+01	0.20	30.9	1.40	5558	
	5/28/09	13	1600	8.5	2.9	2.5 U	5 U	5 U	480	1600	0.1 U	420	63	11	4.16E+03	2.10E+03	1.40E+03	6.10E+01	7.23	-93.8	0.83	3353	0
	6/23/09	30 J	4400	14 J	42 U	42 U	5 U	5 J	3900	1800	0.05 U	400	72	25	ND	2.9E+02*	ND	6.9E+01*	6.79	53	0.46	3594	0.97
	10/15/09	18 J	2200	15 J	25 U	25 U	1 J	2 J	3200	2000	0.1 U	690	190	410	ND	ND	ND	ND	6.7	-121.6	0.33	3310	1.13
	4/23/08		49000	68 J	100 U	100 U	5 U		220	360	0.03 J	2000	2500	64	1.10E+03	8.60E+03	ND	ND	6.61	354.8	1.05	10070	0
																					-		-

PASSIVE		nen	e.	Φ	Φ	-									o o	o o	<u>ө</u>	g					
Monitoring Summary	_	Tetrachloroether	Trichloroethene	cis-1,2- Dichloroethene	trans-1,2- Dichloroethene	Chloride									Dehalococcoide s - 16S rRNA	Dehalococcoide s - tceA	Dehalococcoide s - bvcA	Dehalococcoide s - vcrA			(rity (E	on
NAVFAC I		힏	roe	, oe	,2- oe	矣		4.	<u>e</u>	iţ			<u> </u>	cal n	alococ rRNA	000	Ö	Ö		(Mv)	(mg/L)	ctiv s/cı	s Ir
Weapons	i va vai	acl	Plo	1,2 olor	s-1 nor	0/	ene	ane	har	ili	ate	ate	oric	imi gel nar	ialo rR	ialc	alc A	ialc		I) c	E)	ng qu	no.
Station - S	Site 70	letr	lric	sis- Oict	ran Dict	Vinyl	Ethene	Ethane	Methane	Alkalinity	Nitrate	Sulfate	Chloride	Chemical Oxygen Demand	Deha s - 16S	Deha s - tceA	Deha s - bvcA	Deha s - vcrA	hd	ORP	DO	Conductivity (µmhos/cm)	Ferrous Iron
<u> </u>	Units:	 μg/L	μg/L	µg/L	μg/L	μg/L	μg/L	μg/L	μg/L	mg/L	mg/L	mg/L	mg/L	mg/L	<u> </u>	gene c			<u>u</u>	mV	mg/L	00	mg/L
	9/2/08	360 U	61000	110 J	360 U	360 U	5 U	5	170	360	0.25 U	2000	2600	53	ND	ND	ND	ND	6.31	191.4	6.3	9122	0
	10/16/08	170 U	56000	90 J	170 U	170 U	5 U	7	200	370	0.25 U	1900	2500	95	NS	NS	NS	NS	6.35	15.1	2.17	9690	1.25
e 7	11/3/08	310 U	61000	310 U	310 U	310 U	5 U	7	220	430	0.25 U	2100	2600	170	ND	ND	ND	ND	6.51	-88.6	0.72	9050	1.92
Zone	1/27/09	360 U	46000	160 J	360 U	360 U	5 U	9	240	600	0.1 U	1900	1800	230	1.49E+07	6.28E+06	ND	4.52E+06	6.53	-200	1.13	9777	2.56
3.5	2/23/09	360 U	41000	170 J	360 U	360 U	5 U	12	330	680	0.25 U	1900	2300	290	3.03E+08	9.76E+08	ND	2.18E+08	6.47	-153	1.8	8384	>3.3
Š	3/30/09	20 J	44000	150	8.7 J	25 U	5 U	6	160	590	0.25 U	1900	2400	170	5.84E+07	2.36E+08	ND	2.20E+07	6.26	-183.6	7.49	9589	2.67
PMW-3	4/28/09	360 U	45000	260 J	360 U	360 U	1 J	7	180	500	0.1 U	2000	2600	150	8.04E+06	4.17E+07	ND	1.55E+06	6.38	-97.8	1.14	9089	3.3
	5/27/09	310 U	50000	230 J	310 U	310 U	1 J	7	200	450	0.25 U	2000	2700	74	1.47E+07	9.38E+06	ND	2.80E+06	6.24	-112	4.8	9282	3.26
	6/22/09	310 U	47000	190 J	310 U	310 U	5 U	6	190	410	0.25 U	2000	2600	290	3.71E+06	5.20E+06	ND	8.20E+05	6.71	-196.7	0.79	9432	2.86
(A	10/15/09	310 U	50000	110 J	310 U	310 U	2 J	6	240	360	0.5 U	2000	2700	76	1.71E+06	4.22E+05	ND	1.80E+05	6.47	-21.6	0.84	8876	>3.3
Zone	4/23/08	19 J	4700	90	20 J	42 U	5 U	5 U	81	730	0.04 J	4200	2500	67	ND	NS	NS	NS	6.63	208.7	0.79	13923	0
Zo	11/3/08	7.5	3100	69	16	7.1 U	5 U	5 U	86	650	0.1 U	3800	1700	30	ND	ND	ND	ND	6.68	-23.9	1.5	9867	0.18
PMW-3	4/28/09	10	1300	17	4.7 J	10 U	5 U	2 J	70	4000	0.1 U	1800	660	3100	2.23E+05	1.58E+06	ND	6.50E+04	6.27	-205.4	11.92	9727	3.3
\downarrow	6/22/09	9.3 J	1400	11	2.7 J	3.1 J	5 U	3 J	230	4100	0.25 U	1000	680	4000	3.78E+05	8.90E+05	ND	1.80E+05	6.56	-296.5	0.63	8791	>3.3
	10/15/09	15 J	3600	53	11 J	25 U	0.3 J	5 J	550	2100	0.5 U	2600	2000	1300	6.53E+04	1.44E+05	ND	8.4E+03*	6.25	-78	0.78	10460	>3.3
e 3	4/24/08	20	5400	74	24	3.1 U	5 U	5 U	160	750	0.5 U	3900	3400	100	ND	NS	NS	NS	6.68	117.8	0.9	13706	0
Zone	11/3/08	16	4800	51	12 J	13 U	5 U	5 U	98	880	0.5 U	4400	3100	68	ND	ND	ND	ND	6.76	-3.2	1.81	13140	1.18
-3.7	4/28/09	43	4400	32 J	42 U	42 U	5 U	4 J	150	1700	0.1 U	5900	2600	890	ND	ND	ND	ND	6.46	-216.4	5.98	15.53	2.5
PMW-3	6/22/09	29	2900	18 J	20 U	20 U	5 U	1 J	76	3000	1 U	5800	1600	3000	ND	ND	ND	ND	6.75	-294.1	0.51	15410	2.76
P	10/15/09	20 J	2300	42 U	42 U	42 U	5 U	0.9 J	71	3100	1 U	5800	1100	2300	2.92E+02*	1.54E+02*	2.6E+01*	7.0E+01*	6.57	-103.2	1.09	13680	>3.3
6-3	4/24/08	17 U	1600	10 J	17 U	17 U	5 U	5 U	88	NS	NS	NS	NS	NS	NS	NS	NS	NS	7.47	-38.7	3.68	14267	0
PMW-3 Zone 4																							
PN Zo																							
	4/24/08	630 U	63000	630 U	630 U	630 U	5 U	7	180	350	0.09 J	2000	2500	58	ND	NS	NS	NS	6.82	14.7	0.39	9571	0
	4/24/2008 - K	500 U	61000	500 U	500 U	500 U	5 U	7	180	300	0.08 J	2000	2500	58	NS	NS	NS	NS					
	9/2/08	420 U	63000	160 J	420 U	420 U	5 U	9	240	380	0.25 U	2000	2200	44	ND	ND	ND	ND	6.5	166.8	4.07	8224	0
-	10/16/08	71 U	56000	76	71 U	71 U	5 U	11	280	400	0.25 U	2000	2100	70	NS	NS	NS	NS	6.38	54	2.17	8577	0.09
Zone	11/4/08	360 U	50000	310 J	360 U	360 U	5 U	11	290	410	0.25 U	2000	2100	74	ND	ND	ND	ND	6.6	52.3	1.03	7992	0.62
† Zí	1/27/09	360 U	51000	360 U	360 U	360 U	5 U	9	230	460	0.1 U	2000	2000	99	6.95E+07	4.09E+07	ND	2.73E+07	6.52	-35.4	1.54	8982	2.8
MW-4	2/23/09	360 U	41000	170 J	360 U	360 U	5 U	13	310	530	0.25 U	2000	2000	140	1.67E+08	5.29E+08	ND	1.47E+08	6.54	-215.9	1.39	7726	2.31
	3/30/09	360 U	45000	140 J	360 U	360 U	5 U	8	180	540	0.25 U	1900	1900	140	9.13E+07	2.83E+08	ND	2.62E+07	6.27	-173.4	8.99	8547	2.4

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PASSIVE (Tetrachloroether	Je J	<u>e</u>	trans-1,2- Dichloroethene										Dehalococcoide s - 16S rRNA	de	de	0					
Monitoring	g Data	oet	Trichloroethene	cis-1,2- Dichloroethene	Jer	Chloride									.00	Dehalococcoide s - tceA	Dehalococcoide s - bvcA	Dehalococcoide s - vcrA				₹ `	<u>_</u>
Summary		lorc	Det	ett	2- eeth	Jo			ø)	>			40	a J	ე ≰	၁၀	၁၀	000		(Mv)	(T	tivi ′cn	<u> </u>
NAVFAC N	Naval	ch	lor	,2- orc	-1,- orc	$\dot{\circ}$	ЭE	e e	au	init	ā	te	iġ	nic Jen and	og K	olloc	oli	oll		≥	(mg/L)	luc os,	Sna
Weapons		etra	ig	chl 3:-	chl	Vinyl	Ethene	Ethane	Methane	Alkalinity	Nitrate	Sulfate	Chloride	Chemical Oxygen Demand	ehe	Deha s - tceA	Deha s - bvcA	Deha s - vcrA	_	ORP	0	Conductivity (µmhos/cm)	Ferrous Iron
Station - S		•		Di Sis	tra Di								Ċ	تَ ô تَ	S - 8 - 16	S - tce	S. bv	S - S	рН		8	ٽ ٽ ٽ	
	Units:	μg/L	μg/L	μg/L	μg/L	μg/L	μg/L	μg/L	μg/L	mg/L	mg/L	mg/L	mg/L	mg/L		gene c				mV	mg/L		mg/L
颪	4/28/09	360 U	42000	170 J	360 U	360 U	2 J	12	280	590	0.1 U	1900	1900	170	2.74E+07	1.33E+08	ND	5.19E+06	6.33	-154	1.73	8432	3.18
	5/27/09	250 U	41000	2100	250 U	250 U	3 J	10	250	630	0.25 U	1900	1900	170	6.69E+07	2.08E+08	ND	1.10E+07	6.15	-224.3	14.98	8026	0
1	6/23/09	9 J	30000	4500	11 J	20 J	4 J	11	310	640	0.25 U	1900	1900	190	8.56E+06	1.30E+07	ND	2.00E+06	6.51	-58.4	0.9	8030	2.34
	10/13/09	250 U	35000	10000	250 U	250 U	5	10	650	630	0.25 U	1800	2100	91	3.50E+07	2.51E+07	ND	3.70E+06	6.36	12	1.16	7.904	>3.3
4-7 2 2	4/24/08	130 U	17000	130 U	130 U	130 U	5 U	5 U	50	NS	NS	NS	NS	NS	NS	NS	NS	NS	7.05	41.5	0.94	13331	0
PMW- Zone																							_
	11/4/08	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	6.85	-32.7	1.48	15990	NM
ဗ	4/24/08	16 J	8500	85	43 J	63 U	5 U	5 U	89	820	0.5 U	5600	3100	79	ND	NS	NS	NS	6.72	42.1	0.64	17019	0
Zone	11/4/08	12 J	4600	96	40	13 U	5 U	5 U	75	730	0.5 U	5100	2200	53	ND	ND	ND	ND	6.75	50.8	0.8	14850	0.1
	4/28/09	33 J	3400	41 J	50 U	50 U	5 U	3 J	88	2100	0.1 U	5000	2100	2000	ND	ND	ND	ND	6.26	-37.7	3.17	15790	3.16
\	6/23/09	16	2000	22	8.8 J	13 U	5 U	1 J	51	3000	0.5 U	3900	1100	2000	2.70E+04*	2.4E+04*	ND	5.1E+03*	6.51	-144.9	0.4	13430	2.79
PMW-4	10/13/09	19	2600	29	8.7 J	5.3 J	5 U	4 J	270	2700	0.5 U	3900	1400	2400	1.03E+03*	7.12E+02*	ND	3.0E+03*	6.32	-62	0.74	12770	>3.3
4	4/24/08	16 J	8900	77	42 J	50 U	5 U	5 U	190	810	0.5 U	5000	3800	67	ND	NS	NS	NS	6.65	37.8	0.99	16058	0
Zone	11/4/08	21	6300	86	27	13 U	5 U	5 U	130	820	0.5 U	4400	3000	57	ND	ND	ND	ND ND	6.73	50.8	1.27	14180	0.12
Ζį																							
>	4/28/09	30 Jb	7900	57	12 J	50 U	5 U	4 J	120	850	0.1 U	4400	2900	55	ND	ND	ND	ND	6.44	-24.7	4.11	13840	3.3
PMW-4	6/23/09	18 J	6700	59	15 J	50 U	5 U	3 J	110	970	1 U	4300	2800	99	ND	ND	ND	ND	6.71	-81	0.89	13320	0.76
ļ .	10/15/09	14 J	7400	50	50 U	50 U	5 U	2 J	200	2000	0.5 U	3800	1900	910	ND	ND	ND	ND	6.39	-55.6	1.09	12.07	>3.3
V-4 e 5	4/24/08	1.9 J	990	6.7 J	1.6 J	7.1 U	5 U	5 U	110	NS	NS	NS	NS	NS	NS	NS	NS	NS	7.24	-61	1.02	13894	0.53
PMW- Zone																							
_	1/01/00	400.11		400.11	400.11	400.11			100	222		0.4.0.0	1000				110	110	0.00	10.5	0.00	70.40	
1	4/24/08	420 U	57000	420 U	420 U	420 U		5 U	130	220	0.57	2100	1900	38	ND	NS	NS	NS	6.99	42.5	0.33	7842	0
1	9/2/08	420 U	52000	420 U	420 U	420 U	5 U	6	170	280	0.79	2200	1900	32	ND	ND	ND	ND	6.71	125.4	9.15	7680	0
1	10/16/08	71 U	41000	45 J	71 U	71 U	5 U	8	200	330	0.25 U	2200	1700	42	NS	NS	NS	NS	6.67	28.5	1.28	7858	0.01
Φ 7	11/4/08	200 U	37000	91 J	200 U	200 U	5 U	10	270 250	360 350	0.25 U	2200	1700	44 57	ND 3.83E+05	ND 9.81E+04	ND ND	ND 4.44E+0E	6.94 6.79	387.4	1.46	7190 7899	0
Zone	1/27/09 2/23/09	13 J 360 U	37000 40000	45 J 360 U	50 U 360 U	50 U 360 U	5 U	9	260	370	0.1 U 0.25 U	2100 2200	1600		6.85E+05	9.81E+04 2.52E+06	ND ND	1.11E+05 5.57E+05	6.89		0.91	6740	0.23
	3/30/09	360 U	39000	360 U	360 U	360 U	5 U	10 7	180	390	0.25 U	2100	1600 1600	63 55	1.86E+05	8.31E+05	ND ND	6.32E+04	6.41	-179.1 -269	1.31 1.28	7779	0.08
≽	4/28/09	360 U	44000	360 U	360 U	360 U	5 U	7	170	410	0.25 U	2100	1600	68	1.00E+05 1.97E+05	9.35E+05	ND	3.51E+04	6.6	59.8	2.98	8366	0.08
PMW-5	5/27/09	250 U	35000	82 J	250 U	250 U	5 U	8	200	460	0.15 U	2100	1600	70	4.32E+06	2.49E+06	ND	8.40E+05	6.59	-158.5	0.9	7207	0.03
-	6/23/09	310 U	39000	380	310 U	310 U	5 U	8	210	440	0.25 U	2000	1500	93	1.55E+06	3.70E+06	ND	5.50E+05	6.87	-10.9	0.92	7258	0.03
1	10/7/09	200 U	23000	10000	200 U	450	10	8	830	640	0.25 U	2000	1500	100	1.43E+08	9.35E+07	ND	1.50E+07	6.27	-21.7	0.99	6.769	1.57
1	10/7/2009-K	200 U	24000	11000	200 U	490	10	7	800	650	0.25 U	2000	1500	100	2.46E+08	2.21E+08	ND	4.10E+07	0.2.		0.00	0.1.00	
(4	4/24/08	100 U	13000	75 J	54 J	100 U	5 U	5 U	62	610	0.5 U	5700	3000	100	ND	NS	NS	NS	6.96	52	0.61	15992	0
one	11/4/08	15	6100	86	43	13 U	5 U	5 U	57	950	0.5 U	6000	3100	95	ND	ND	ND	ND	6.89	127.5	1.67	16730	0.09
2 2	4/28/09	20 J b	7100	56	31 J	50 U	5 U	2 J	69	1400	0.1 U	5900	2600	1000	ND	ND	ND	ND	6.56	-37.5	1.94	12740	3.3
Š	6/23/09	12 J	4600	69	26 J	42 U	5 U		61	1900	1 U	5000	1900	2200	ND	ND	ND	ND	6.61	-82.6	0.94	16230	
PMW-5								2 J															1.56
(·)	10/13/09	42 U	6700	220	23 J	42 U	5 U	2 J	67	2500	1 U	4600	1500	2300	3.32E+03*	1.38E+03*	ND	8.2E+02*	6.33	-98.7	0.65	1373	>3.3
Zone	4/24/08	83 U	11000	90	32 J	83 U	5 U	5 U	74	750	0.5 U	5800	3100	87	ND	NS	NS	NS	6.82	21.1	0.53	16173	0.015
Zc	11/4/08	12 J	9100	72	34	25 U	5 U	5 U	83	890	0.5 U	5700	2900	83	ND	ND	ND	ND	6.82	66.8	1.02	15380	0.7
\-\-2	4/28/09	41 J	6900	67	20 J	50 U	5 U	1 J	49	1100	1 U	5400	2600	NA	ND	ND	ND	ND	6.57	-22.2	0.57	15700	2.21
PMW-5	6/23/09	42 U	5600	53	13 J	42 U	5 U	1 J	36	1200	1 U	5100	3000	430	ND	ND	ND	ND	6.7	-49.6	0.36	15010	3
	10/13/09	42 U	6900	72	42 U	42 U	5 U	2 J	71	1400	1 U	5300	2500	360	ND	ND	ND	ND	6.38	-45.9	1.46	14140	>3.3
7-5 ⊕ 4	4/24/08	31 U	5000	19 J	31 U	31 U	5 U	5 U	16	NS	NS	NS	NS	NS	NS	NS	NS	NS	7.41	-25.1	1.28	11794	0.76
PMW-5 Zone 4																							
_ N	4/00/00	04.11	44000	F.4	44 1	04.11		F 11	470	500	04 '	2000	2000		NID	NO	NO	NO	0.50	0.45	0.00	40070	
	4/22/08	31 U	11000	51	14 J	31 U		5 U	170	530	0.1 J	3000	3000	56	ND 0.65E+02	NS 6.77E±01	NS 2.71E+01	NS	6.53	345	0.98	12976	0
	9/5/08	68 J	12000	86	21 J	83 U	5 U	5 U	96 110	600 600	0.5 U 0.25 U	3200 3000	3100 2800	51 61	9.65E+02	6.77E+01	3.71E+01	ND NS	5.82	247.5	3.96	1106	0
	10/15/08 0/15/2008 - K	50 U 10 J	9500 10000	86 77	19 J 21	50 U 20 U	5 U	5 U	100	610	0.25 U	3100	2700	57	NS NS	NS NS	NS NS	NS NS	6.4	34	1.8	12530	0.06
	11/3/08	50 U	9700	80	14 J	50 U	5 U	5 U	130	660	0.25 U	3300	2900	78	ND ND	ND ND	ND	ND ND	6.64	-45.2	0.75	11060	0.99
ي ا	1/27/09	50 U	7400	600	50 U	50 U	5 U	5 U	68	740	0.1 U	2600	2900	120	ND	ND	ND	ND	6.61	-45.2	0.75	12770	2.86
PMW-6	2/23/09	17 U	2100	800	7.6 J	54	5 U	5 U	98	1300	0.1 U	1900	1100	1100	2.76E+07	1.14E+08	ND	5.79E+06	6.5	-328.4	0.44	7552	3.03
Σ	3/30/09	5 J	1000	350	4.9 J	350	380	5 U	55	2400	0.1 U	2100	2000	1600	1.71E+09	5.90E+09	ND	6.11E+08	6.47	-320.4	2.92	11900	0.67
"	4/27/09	4.2 J	740	360	8.2	410	310	2 J	72	3100	0.035	1500	1700	2300	5.62E+08	3.61E+09	ND	1.72E+08	6.46	-336.4	5.98	7721	0.07
	5/27/09	2.7 J	820	310	4.9 J	340	290	2 J	260	3100	0.035	790	1300	3400	1.00E+09	9.54E+08	ND	2.60E+08	6.23	-330.4	12.58	8591	0.35
	6/22/09	3 J	790	460	6.4	120	190	5 U	1100	3200	0.23 0.06 J	600	1400	3400	1.59E+09	3.80E+09	ND	4.60E+08	6.72	-335.3	0.54	9043	0.33
	6/22/2009-K	3.1 J	910	490	7	130	170	5 U	900	3100	0.13	760	1500	3200	9.64E+08	2.00E+09	ND	2.40E+08	0.12	0.00.0	0.04	0010	
	10/15/09	5.7	760	690	11	470	220	1 J	2300	1900	0.25 U	2000	1800	400	4.22E+08	2.44E+08	ND	4.90E+07	5.54	-328.7	0.98	9763	NM
	4/22/08	31 U	17000	76	21 J	31 U	5 U	9	210	540	0.03 J	3000	3200	50	ND	NS	NS	NS	6.66	301	0.96	13572	0
- L									-										-				

PASSIVE Monitorin Summary NAVFAC Weapons Station -	ng Data / Naval s Site 70	Tetrachloroethen	Trichloroethene	cis-1,2- Dichloroethene	trans-1,2- Dichloroethene	Vinyl Chloride	Ethene	Ethane	Methane	Alkalinity	Nitrate	Sulfate		Chemical Oxygen Demand	Dehalococcoide s - 16S rRNA	Dehalococcoide s - tceA	Dehalococcoide s - bvcA	Dehalococcoide s - vcrA	Hd	ORP (Mv)	DO (mg/L)	Conductivity (µmhos/cm)	Ferrous Iron
	Units:	μg/L	μg/L	μg/L	μg/L	μg/L	μg/L	μg/L	μg/L	mg/L	mg/L	mg/L	mg/L	mg/L		gene c	•	_		mV	mg/L		mg/L
	9/5/08	27 J	15000	78 J	28 J	83 U	5 U	5 U	150	600	0.5 U	3500	3200	53	1.51E+03	1.75E+02	2.51E+01	ND	6.11	310.2	1.23	1274	0
	10/15/08	9.6 J	10000	57	13 J	25 U	5 U	5 U	110	1200	0.25 U	3000	2400	1300	NS	NS	NS	NS	6.26	-221.9	0.81	12010	NM
	11/3/08	50 U	9600	59	50 U	50 U	5 U	5 U	140	1100	0.1 U	2600	2600	1100	ND	ND	ND	ND	6.49	-240.9	0.5	11990	1.94
!	1/27/09	50 U	5800	33 J	50 U	50 U	5 U	5 U	96	2300	0.1 U	2800	2200	1700	6.15E+03	1.02E+03	ND	2.19E+03	6.6	-287.9	0.44	13.3	0.4
՛≽	2/23/09	50 U	6500	44 J	50 U	50 U	5 U	5 U	110	2800	0.1 U	2800	1900	2700	ND	ND	ND	ND	6.46	-327.4	0.52	12848	0.66
PMW-7	2/23/2009 - K	31 U	6700	46	31 U	31 U	5 U	5 U	110	2600	0.1 U	2900	2000	2600	ND	ND	ND	ND					
1 "	3/30/09	50 U	8600	58	18 J	50 U	5 U	5 U	100	2300	0.1 U	2900	2400	1500	5.78E+03	2.29E+04	ND	1.88E+03	6.42	-324.8	6.58	13520	1.56
	4/27/09	50 U	5800	1700	16 J	50 U	5 U	4 J	120	2100	0.1 U	3300	2500	1500	2.04E+03	1.43E+04	ND	3.83E+02	6.31	-222.4	0.47	8383	3.3
	5/27/09	25 U	2900	1300	9.2 J	120	5 U	3 J	96	3300	0.5 U	2200	1600	3200	3.83E+08	3.04E+08	ND	2.50E+06	6.28	-322.5	12.58	11720	0.57
	6/22/09	4.2 U	190	96	2.4 J	590	28	1 J	47	4400	0.1 U	650	820	5400	1.06E+09	2.20E+09	ND	7.10E+07	6.75	-320.8	0.66	9510	0
	10/15/09	5 U	120	41	4.2 J	67		0.5 J	2600	5200	0.1 U	320	710	8400	2.63E+08	2.22E+08	ND	3.40E+07	7.01	-340.7	1.49	9644	0
	4/23/08	8 J	15000	120	10 J	17 U	5 U	34	430	590	0.5 U	2400	2400	46	ND	NS	NS	NS	6.64	189.5	0.96	10876	0
	9/3/08	71 U	12000	100	71 U	71 U	5 U	6	140	780	0.25 U	2300	1800	660	ND	ND	ND	ND	6.01	-141.7	1.92	7889	0
	10/15/08	17 U	8300	58	8.8 J	17 U	5 U	6	120	1400	0.25 U	2000	1600	1600	NS	NS	NS	NS	6.22	-243	0.68	8568	>3.3
	11/3/08	83 U	8500	69 J	83 U	83 U	5 U	7	150	1400	0.1 U	1900	1600	1400	ND	ND	ND	ND	6.42	-297.8	0.31	8123	3.07
PMW-8	1/27/09	17 U	1300	2700	3.7 J	17 U	5 U	8	160	2000	0.1 U	1500	1500	1300	ND	ND	ND	ND	6.55	-257	0.36	9157	0.85
¥	2/23/09	17 U	1100	4500	4.4 J	17 U	5 U	9	190	1500	0.1 U	1900	1700	420	7.05E+04	1.50E+05	ND	3.28E+04	6.35	-268.4	2.22	8445	1.49
₫	3/30/09	10 U	1700	1400	3 J	140	150	8	170	1900	0.1 U	1500	1500	1400	1.40E+08	6.59E+08	ND	6.09E+07	6.45	-241.7	0.24	8802	1.03
	4/27/09	6.1 J	1500	470	2.7 J	420	330	11	350	2000	0.1 U	1400	1500	1600	1.61E+08	1.08E+09	ND	5.11E+07	6.32	-313.9	15.41	8485	0.36
	5/27/09	10 U	1300	330	3.3 J	260	410	12	1100	2300	0.25 U	1100	1300	2400	4.49E+08	7.42E+08	ND	1.80E+08	6.2	-323	17.02	7700	1.38
	6/22/09	5 U	710	180	2.9 J	250	350	10	1300	2700	0.1 U	760	1100	3500	2.16E+09	5.50E+09	ND	6.90E+08	6.6	-353.1	0.56	7496	1.46
	10/7/09	1.8	72	76	6.3	100	250	9	2300	3200	0.1 U	33	860	3200	7.06E+07	5.39E+07	ND	1.20E+07	6.52	-313.8	0.2	6622	1
	4/23/08	6.3 U	840	18	6.3 U	6.3 U	5 U	48	2800	990	0.01 J	1100	160	16	ND	NS	NS	NS	6.75	35.2	0.57	4114	0
	9/3/08	3.5 J	2000	35	6.3 U	6.3 U	5 U	6	460	810	0.1 U	2700	400	13	ND	ND	ND	ND	6.11	-88.4	1.85	5543	0
	10/15/08	4.2 U	1800	34	1.3 J	4.2 U	5 U	5	360	810	0.25 U	2800	400	13	NS	NS	NS	NS	6.35	47.7	0.98	6034	0
	11/3/08	17 U	1500	31	17 U	17 U	5 U	5	370	800	0.1 U	3000	440	13	ND	ND	ND	ND	6.52	115.7	0.84	6066	0
6-	1/27/09	0.7 J	350	7	2 U	2 U	5 U	27	770	1000	0.1 U	1700	220	19	ND	ND	ND	ND	6.71	-96.4	0.63	4993	0.48
	2/23/09	0.3 J	96	5.8	0.1 J	0.5 U	5 U	48	1500	1100	0.1 U	950	120	17	ND	ND	ND	1.31E+03	6.72	-222.1	4.65	3098	0.25
PMW-9	3/30/09	0.8 J	240	160	0.4 J	0.2 J	5 U	20	640	1100	0.33	1200	120	15	ND	ND	ND	ND	6.53	-102	0.82	3562	2.88
	4/27/09	1.4 J	110	270	0.6 J	11	4 J	20	640	1500	0.1 U	1600	290	320	9.79E+05	5.18E+06	ND	1.79E+05	6.43	-295.2	10.54	5846	0.12
	5/27/09	0.8 J	100	200	1.2 J	80	100	39	2200	2200	0.25 U	1300	480	720	2.48E+08	1.86E+08	ND	5.20E+07	6.3	-305.8	2.63	5772	0.62
	6/22/09	0.2 J	19	19	1.1	37	110	26	3000	2100	0.1 J	960	580	400	2.80E+07	5.60E+07	ND	8.00E+06	6.8	-290.2	0.6	5921	0
	10/7/09	0.2 J	45	14	1.2	25	66	16	6600	1700	0.25 U	2500	1100	120	1.12E+07	9.83E+06	ND	1.90E+06	6.67	-282.4	0.56	7605	0.05

Notes:

K - Duplicate sample

J - estimated value

U - nondetect (detection limit is indicated)

NA - not analyzed ND - not detected

NS - not sampled

* - indicates that the value presented is below the reporting limit.

µmhos/cm - micromhos per centimeter

µg/L - mcrograms per liter mg/L - milligrams per liter

mV - millivolts
ORP - oxidation reduction potential

DO - dissolved oxygen

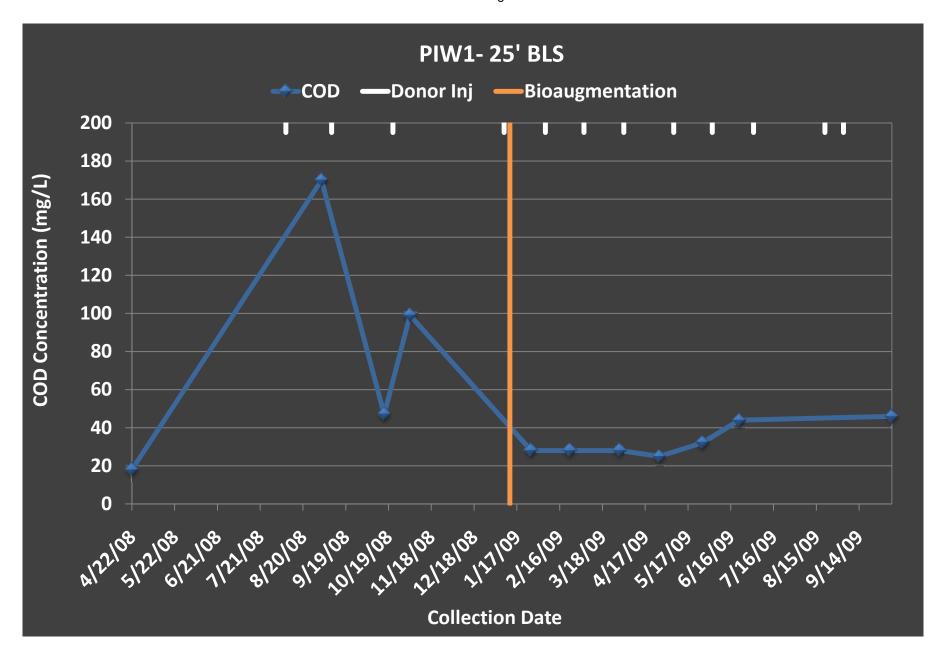
> - greater than NM - not measured

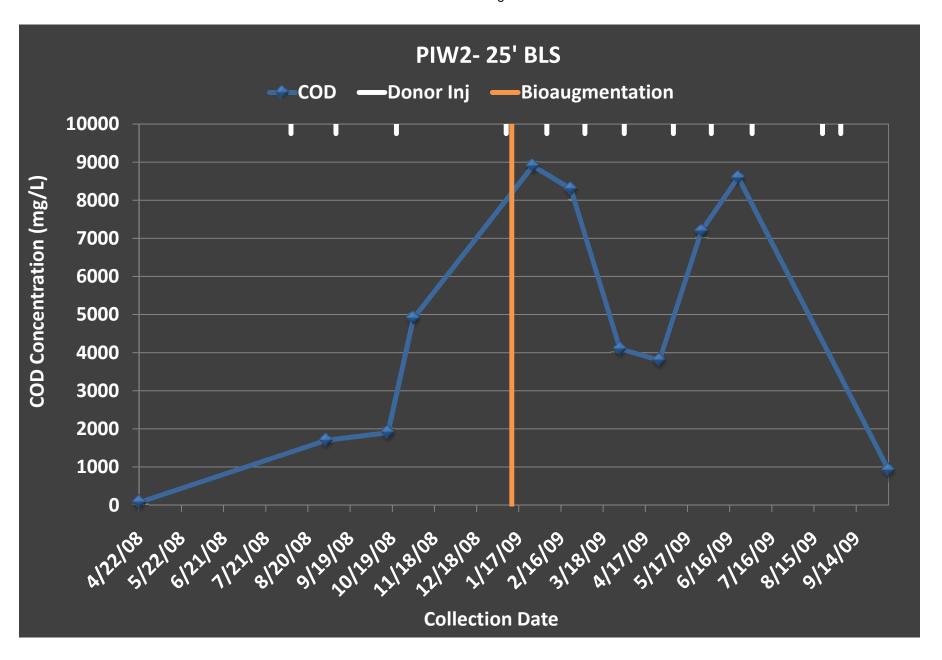
February 2009	TCE (ppb)	$\delta^{13}C$	cDCE	$\delta^{13}C$	VC	$\delta^{13}C$	Ethene	δ^{13} C eth	2-butanone	chloroform	chlor/DCE
SB-AEW	21000	-24.3	1100	-26.8							
SB-AMW1-25'	27000		7000	-24.6							
SB-AMW2-25'	97		9500	-22.9	940	-43.0				75	0.01
SB-AMW3-Z1	6000	-24.5	2300	-24.5	210					55	0.02
SB-AMW4-Z1	310	-17.9	6700	-36.5	2900		59				
SB-AMW5-Z1	5800	-23.7	2800	-25.0	57					100	0.04
SB-AMW6-25'	130000	-23.9		-31.7		-28.5					
SB-PIW1-25'	42		4	-27.0						120	32.43
SB-PIW2-25'	12	-26.1	3	-27.9	23				110	3	1.00
SB-PIW3-25'	2	-23.5	1	-25.4	5				160	-	
SB-PMW1-25'	1700	-28.0	79		-					74	0.94
SB-PMW2-25'	1800	20.0	43								0.0 .
SB-PMW3-Z1	41000	-23.3									
SB-PMW4-Z1	41000	-23.3									
SB-PMW5-Z1	40000	-23.5									
			900	10.2	E /					200	0.25
SB-PMW6-25'	2100	-23.3	800	-19.2	54					280	0.35
SB-PMW7-25'	6500	20.0	4500	25.2						690	0.01
SB-PMW8-25'	1100	-20.0	4500	-25.2						30	0.01
SB-PMW9-25'	96	-22.6	6								
Average of all		-23.5		-26.4		25.0					
Average of all		-23.5 -22.9		-20.4 -27.4		-35.8 -35.8					
AEW/AMW Ave						-33.6					
PIW/PMW Ave		-23.9		-24.9							
April 2009	TCE (ppb)	$\delta^{13}C$	cDCE	δ ¹³ C	VC	$\delta^{13}C$	Ethene	δ^{13} C eth	2-butanone	chloroform	chlor/DCE
SB-AEW	6500	-22.7	330	-24.9							
SB-AMW1-25'	880	-22.2	3000	-15.4	1100	-35.4	17				
SB-AMW2-25'	540	-9.9	1700		6900	-30.2	16				
					0000	00.0					
SB-AMW3-Z1	1200	-22.5	2700	-15.9	3900	-26.2	58				
SB-AMW3-Z1 SB-AMW3-Z2	1200 780	-22.5 -22.6	2700 6500	-15.9 -13.9	1900	-26.2 -40.2	58 13			99	0.02
										99	0.02
SB-AMW3-Z2	780		6500		1900	-40.2				99	0.02
SB-AMW3-Z2 SB-AMW3-Z3	780 130	-22.6	6500 11000	-13.9	1900 190	-40.2 -47.0	13			99	0.02
SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1	780 130 250	-22.6	6500 11000 4100	-13.9 -9.6	1900 190 5400	-40.2 -47.0 -32.5	13 14			99	0.02
SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z2	780 130 250 81	-22.6 -16.8	6500 11000 4100 5500	-13.9 -9.6 -6.0	1900 190 5400 8100	-40.2 -47.0 -32.5 -33.0	13 14			99 72	0.02
SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z2 SB-AMW4-Z3	780 130 250 81 4200	-22.6 -16.8 -21.9	6500 11000 4100 5500 7400	-13.9 -9.6 -6.0 -19.4	1900 190 5400 8100 2000	-40.2 -47.0 -32.5 -33.0 -37.8	13 14 12				
SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z2 SB-AMW4-Z3 SB-AMW5-Z1	780 130 250 81 4200 2600	-22.6 -16.8 -21.9	6500 11000 4100 5500 7400 5500	-9.6 -6.0 -19.4 -19.3	1900 190 5400 8100 2000 3500	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8	13 14 12 19				
SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z2 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z2	780 130 250 81 4200 2600 81	-22.6 -16.8 -21.9 -21.5	6500 11000 4100 5500 7400 5500 9000	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7	1900 190 5400 8100 2000 3500 5200	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0	13 14 12 19			72	0.01
SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z2 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z3	780 130 250 81 4200 2600 81 1900	-22.6 -16.8 -21.9 -21.5	6500 11000 4100 5500 7400 5500 9000 4200	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8	1900 190 5400 8100 2000 3500 5200	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0	13 14 12 19			72 77	0.01
SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z2 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z3	780 130 250 81 4200 2600 81 1900	-22.6 -16.8 -21.9 -21.5	6500 11000 4100 5500 7400 5500 9000 4200	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8	1900 190 5400 8100 2000 3500 5200	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0	13 14 12 19	-17.2		72 77	0.01
SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW6-25'	780 130 250 81 4200 2600 81 1900 70000	-22.6 -16.8 -21.9 -21.5	6500 11000 4100 5500 7400 5500 9000 4200 1100	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8	1900 190 5400 8100 2000 3500 5200 170	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0	13 14 12 19	-17.2	180	72 77 120	0.01 0.02 0.11
SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z2 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z3 SB-AMW6-25' SB-PIW1-25'	780 130 250 81 4200 2600 81 1900 70000	-22.6 -16.8 -21.9 -21.5 -21.5 -22.9	6500 11000 4100 5500 7400 5500 9000 4200 1100	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8	1900 190 5400 8100 2000 3500 5200 170	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0	13 14 12 19	-17.2	180 170	72 77 120	0.01 0.02 0.11
SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z3 SB-AMW6-25' SB-PIW1-25' SB-PIW2-25'	780 130 250 81 4200 2600 81 1900 70000	-22.6 -16.8 -21.9 -21.5 -21.5 -22.9	6500 11000 4100 5500 7400 5500 9000 4200 1100	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8	1900 190 5400 8100 2000 3500 5200 170	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0 -39.9	13 14 12 19	-17.2		72 77 120	0.01 0.02 0.11
SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW6-25' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PMW1-25' SB-PMW1-25' SB-PMW1-25' SB-PMW1-25'	780 130 250 81 4200 2600 81 1900 70000	-22.6 -16.8 -21.9 -21.5 -21.5 -22.9	6500 11000 4100 5500 7400 5500 9000 4200 1100 23	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8	1900 190 5400 8100 2000 3500 5200 170	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0 -39.9	13 14 12 19	-17.2		72 77 120 11	0.01 0.02 0.11 0.48
SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW6-25' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PMW1-25' SB-PMW1-25'	780 130 250 81 4200 2600 81 1900 70000 26 44	-22.6 -16.8 -21.9 -21.5 -21.5 -22.9 -23.1 -28.3	6500 11000 4100 5500 7400 5500 9000 4200 1100 23	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8	1900 190 5400 8100 2000 3500 5200 170	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0 -39.9	13 14 12 19	-17.2		72 77 120 11	0.01 0.02 0.11 0.48
SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW6-25' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PMW1-25' SB-PMW1-25' SB-PMW1-25' SB-PMW1-25'	780 130 250 81 4200 2600 81 1900 70000 26 44	-22.6 -16.8 -21.9 -21.5 -21.5 -22.9 -23.1 -28.3 -21.7	6500 11000 4100 5500 7400 5500 9000 4200 1100 23 1 65 7	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8	1900 190 5400 8100 2000 3500 5200 170	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0 -39.9	13 14 12 19	-17.2		72 77 120 11 58 140	0.01 0.02 0.11 0.48
SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z2 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW6-25' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PMW1-25' SB-PMW1-25' SB-PMW1-25' SB-PMW1-25' SB-PMW1-25' SB-PMW1-25' SB-PMW1-25'	780 130 250 81 4200 2600 81 1900 70000 26 44 1400 280 45000	-22.6 -16.8 -21.9 -21.5 -21.5 -22.9 -23.1 -28.3 -21.7 -21.9	6500 11000 4100 5500 7400 5500 9000 4200 1100 23 1 65 7 260	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8	1900 190 5400 8100 2000 3500 5200 170	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0 -39.9	13 14 12 19	-17.2		72 77 120 11 58 140 48	0.01 0.02 0.11 0.48 0.89 21.54 0.18
SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW6-25' SB-PIW1-25' SB-PIW2-25' SB-PIW3-Z5' SB-PIW3-Z5' SB-PMW1-25' SB-PMW1-Z5' SB-PMW1-Z5' SB-PMW1-Z5' SB-PMW1-Z5' SB-PMW2-Z5' SB-PMW3-Z1 SB-PMW3-Z1	780 130 250 81 4200 2600 81 1900 70000 26 44 1400 280 45000 42000	-22.6 -16.8 -21.9 -21.5 -21.5 -22.9 -23.1 -28.3 -21.7 -21.9 -22.0	6500 11000 4100 5500 7400 5500 9000 4200 1100 23 1 65 7 260 520	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8	1900 190 5400 8100 2000 3500 5200 170	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0 -39.9	13 14 12 19	-17.2		72 77 120 11 58 140 48 220	0.01 0.02 0.11 0.48 0.89 21.54 0.18
SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z3 SB-AMW6-25' SB-PIW1-25' SB-PIW2-25' SB-PIW3-Z5 SB-PMW1-Z5' SB-PMW1-Z5' SB-PMW1-Z5' SB-PMW1-Z5' SB-PMW3-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z1	780 130 250 81 4200 2600 81 1900 70000 26 44 1400 280 45000 42000 3400	-22.6 -16.8 -21.9 -21.5 -21.5 -22.9 -23.1 -28.3 -21.7 -21.9 -22.0 -23.4	6500 11000 4100 5500 7400 5500 9000 4200 1100 23 1 65 7 260 520 41	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8	1900 190 5400 8100 2000 3500 5200 170	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0 -39.9	13 14 12 19	-17.2		72 77 120 11 58 140 48 220 5600	0.01 0.02 0.11 0.48 0.89 21.54 0.18 0.42 136.59
SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW6-25' SB-PIW1-25' SB-PIW2-25' SB-PIW2-25' SB-PIW3-Z5' SB-PMW3-Z5' SB-PMW3-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z3 SB-PMW4-Z3 SB-PMW4-Z4	780 130 250 81 4200 2600 81 1900 70000 26 44 1400 280 45000 42000 3400 7900	-22.6 -16.8 -21.9 -21.5 -21.5 -22.9 -23.1 -28.3 -21.7 -21.9 -22.0 -23.4 -23.5	6500 11000 4100 5500 7400 5500 9000 4200 1100 23 1 65 7 260 520 41	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8	1900 190 5400 8100 2000 3500 5200 170	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0 -39.9	13 14 12 19	-17.2		72 77 120 11 58 140 48 220 5600 3300	0.01 0.02 0.11 0.48 0.89 21.54 0.18 0.42 136.59
SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW6-25' SB-PIW1-25' SB-PIW2-25' SB-PIW2-25' SB-PIW3-25' SB-PMW3-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z4 SB-PMW5-Z1	780 130 250 81 4200 2600 81 1900 70000 26 44 1400 280 45000 42000 3400 7900 44000	-22.6 -16.8 -21.9 -21.5 -22.9 -23.1 -28.3 -21.7 -21.9 -22.0 -23.4 -23.5 -24.1 -24.0	6500 11000 4100 5500 7400 5500 9000 4200 1100 23 1 65 7 260 520 41 57	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8	1900 190 5400 8100 2000 3500 5200 170	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0 -39.9	13 14 12 19	-17.2		72 77 120 11 58 140 48 220 5600 3300 130 2900	0.01 0.02 0.11 0.48 0.89 21.54 0.18 0.42 136.59 57.89
SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z2 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW6-25' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PMW1-25' SB-PMW1-Z3 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z2	780 130 250 81 4200 2600 81 1900 70000 26 44 1400 280 45000 42000 3400 7900 44000 7100	-22.6 -16.8 -21.9 -21.5 -21.5 -22.9 -23.1 -28.3 -21.7 -21.9 -22.0 -23.4 -23.5 -24.1	6500 11000 4100 5500 7400 5500 9000 4200 1100 23 1 65 7 260 520 41 57	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8	1900 190 5400 8100 2000 3500 5200 170	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0 -39.9	13 14 12 19	-17.2		72 77 120 11 58 140 48 220 5600 3300 130	0.01 0.02 0.11 0.48 0.89 21.54 0.18 0.42 136.59 57.89
SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z3 SB-PIW1-25' SB-PIW2-25' SB-PIW2-25' SB-PIW3-Z1 SB-PMW1-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z3 SB-PMW4-Z3 SB-PMW5-Z2 SB-PMW5-Z2 SB-PMW5-Z2 SB-PMW5-Z2	780 130 250 81 4200 2600 81 1900 70000 26 44 1400 280 45000 42000 3400 7900 44000 7100 6900	-22.6 -16.8 -21.9 -21.5 -22.9 -23.1 -28.3 -21.7 -21.9 -22.0 -23.4 -23.5 -24.1 -24.0 -23.5	6500 11000 4100 5500 7400 5500 9000 4200 1100 23 1 65 7 260 520 41 57	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8	1900 190 5400 8100 2000 3500 5200 170 4 6 6	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0 -39.9	13 14 12 19 40		170	72 77 120 11 58 140 48 220 5600 3300 130 2900 3800	0.01 0.02 0.11 0.48 0.89 21.54 0.18 0.42 136.59 57.89 51.79 56.72
SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-PIW1-25' SB-PIW1-25' SB-PIW2-25' SB-PIW2-25' SB-PIW3-25' SB-PMW3-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z3 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW6-Z5'	780 130 250 81 4200 2600 81 1900 70000 26 44 1400 280 45000 42000 3400 7900 44000 7100 6900 740	-22.6 -16.8 -21.9 -21.5 -21.5 -22.9 -23.1 -28.3 -21.7 -21.9 -22.0 -23.4 -23.5 -24.1 -24.0 -23.5 -24.1	6500 11000 4100 5500 7400 5500 9000 4200 1100 23 1 65 7 260 520 41 57	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8 -25.1	1900 190 5400 8100 2000 3500 5200 170 4 6 6	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0 -39.9	13 14 12 19 40		170	72 77 120 11 58 140 48 220 5600 3300 130 2900 3800 230	0.01 0.02 0.11 0.48 0.89 21.54 0.18 0.42 136.59 57.89 51.79 56.72 0.64 0.34
SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW6-25' SB-PIW1-25' SB-PIW2-25' SB-PIW2-25' SB-PIW3-25' SB-PMW1-25' SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW6-Z5' SB-PMW6-Z5' SB-PMW6-Z5' SB-PMW6-Z5' SB-PMW6-Z5' SB-PMW6-Z5' SB-PMW6-Z5' SB-PMW6-Z5' SB-PMW6-Z5' SB-PMW6-Z5'	780 130 250 81 4200 2600 81 1900 70000 26 44 1400 280 45000 42000 3400 7900 44000 7100 6900 740 5800	-22.6 -16.8 -21.9 -21.5 -21.5 -22.9 -23.1 -28.3 -21.7 -21.9 -22.0 -23.4 -23.5 -24.1 -24.0 -23.5 -17.9 -22.0	6500 11000 4100 5500 7400 5500 9000 4200 1100 23 1 65 7 260 520 41 57	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8 -25.1	1900 1900 5400 8100 2000 3500 5200 170 4 6 6	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0 -39.9	13 14 12 19 40	-27.5	170	72 77 120 11 58 140 48 220 5600 3300 130 2900 3800 230 570	0.01 0.02 0.11 0.48 0.89 21.54 0.18 0.42 136.59 57.89 51.79 56.72 0.64
SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z3 SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PMW1-25' SB-PMW1-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW6-25' SB-PMW7-25' SB-PMW7-25' SB-PMW7-25' SB-PMW7-25' SB-PMW7-25' SB-PMW7-25'	780 130 250 81 4200 2600 81 1900 70000 26 44 1400 280 45000 42000 3400 7900 44000 7100 6900 740 5800 1500	-22.6 -16.8 -21.9 -21.5 -22.9 -23.1 -28.3 -21.7 -21.9 -22.0 -23.5 -24.1 -24.0 -23.5 -17.9 -22.0 -15.5	6500 11000 4100 5500 7400 5500 9000 4200 1100 23 1 65 7 260 520 41 57 56 67 360 1700 470	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8 -25.1	1900 190 5400 8100 2000 3500 5200 170 4 6 6	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0 -39.9 -28.0	13 14 12 19 40	-27.5	170	72 77 120 11 58 140 48 220 5600 3300 130 2900 3800 230 570	0.01 0.02 0.11 0.48 0.89 21.54 0.18 0.42 136.59 57.89 51.79 56.72 0.64 0.34
SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-PW5-Z3 SB-PW1-25' SB-PIW1-25' SB-PIW2-25' SB-PIW2-25' SB-PMW3-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z3 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW6-25' SB-PMW6-25' SB-PMW6-25' SB-PMW6-25' SB-PMW9-25'	780 130 250 81 4200 2600 81 1900 70000 26 44 1400 280 45000 42000 3400 7900 44000 7100 6900 740 5800 1500 110	-22.6 -16.8 -21.9 -21.5 -21.5 -22.9 -23.1 -28.3 -21.7 -21.9 -22.0 -23.4 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -22.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.0 -23.5 -24.0 -23.5 -24.0 -23.5 -24.0 -23.5 -24.0 -24.0 -25.0 -27.0 -2	6500 11000 4100 5500 7400 5500 9000 4200 1100 23 1 65 7 260 520 41 57 56 67 360 1700 470	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8 -25.1 -23.8 -9.5 -22.8	1900 190 5400 8100 2000 3500 5200 170 4 6 6	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0 -39.9 -28.0	13 14 12 19 40	-27.5	170	72 77 120 11 58 140 48 220 5600 3300 130 2900 3800 230 570	0.01 0.02 0.11 0.48 0.89 21.54 0.18 0.42 136.59 57.89 51.79 56.72 0.64 0.34
SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z2 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-PW5-Z3 SB-PW2-25 SB-PIW2-25 SB-PIW3-25 SB-PIW3-25 SB-PMW1-25 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z5 SB-PMW5-Z5 SB-PMW5-Z5 SB-PMW8-Z5 SB-PMW8-25	780 130 250 81 4200 2600 81 1900 70000 26 44 1400 280 45000 42000 3400 7900 44000 7100 6900 740 5800 1500 110	-22.6 -16.8 -21.9 -21.5 -22.9 -23.1 -28.3 -21.7 -21.9 -22.0 -23.5 -24.1 -24.0 -23.5 -17.9 -22.0 -23.5 -17.9 -22.0	6500 11000 4100 5500 7400 5500 9000 4200 1100 23 1 65 7 260 520 41 57 56 67 360 1700 470	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8 -25.1 -23.8 -9.5 -22.8 -17.1 -16.6	1900 190 5400 8100 2000 3500 5200 170 4 6 6	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0 -39.9 -28.0	13 14 12 19 40	-27.5	170	72 77 120 11 58 140 48 220 5600 3300 130 2900 3800 230 570	0.01 0.02 0.11 0.48 0.89 21.54 0.18 0.42 136.59 57.89 51.79 56.72 0.64 0.34
SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-PW5-Z3 SB-PW1-25' SB-PIW1-25' SB-PIW2-25' SB-PIW2-25' SB-PMW3-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z3 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW6-25' SB-PMW6-25' SB-PMW6-25' SB-PMW6-25' SB-PMW9-25'	780 130 250 81 4200 2600 81 1900 70000 26 44 1400 280 45000 42000 3400 7900 44000 7100 6900 740 5800 1500 110	-22.6 -16.8 -21.9 -21.5 -21.5 -22.9 -23.1 -28.3 -21.7 -21.9 -22.0 -23.4 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -22.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.1 -24.0 -23.5 -24.0 -23.5 -24.0 -23.5 -24.0 -23.5 -24.0 -23.5 -24.0 -24.0 -25.0 -27.0 -2	6500 11000 4100 5500 7400 5500 9000 4200 1100 23 1 65 7 260 520 41 57 56 67 360 1700 470	-13.9 -9.6 -6.0 -19.4 -19.3 -11.7 -21.8 -25.1 -23.8 -9.5 -22.8	1900 190 5400 8100 2000 3500 5200 170 4 6 6	-40.2 -47.0 -32.5 -33.0 -37.8 -31.8 -40.0 -39.9 -28.0	13 14 12 19 40	-27.5	170	72 77 120 11 58 140 48 220 5600 3300 130 2900 3800 230 570	0.01 0.02 0.11 0.48 0.89 21.54 0.18 0.42 136.59 57.89 51.79 56.72 0.64 0.34

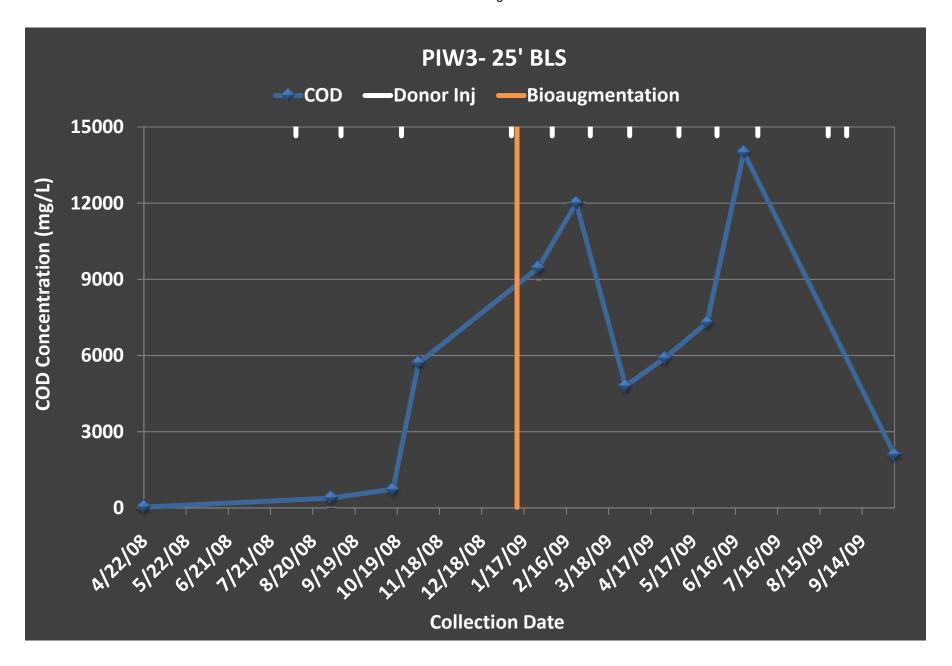
luno 2000	TCE (nnh)	δ ¹³ C	•DCE	δ^{13} C	VC	$\delta^{13}C$	Ethene	S ¹³ C oth	2-butanone chloroform	oblor/DCE
June 2009	TCE (ppb)		cDCE		VC		Ethene	o Celli	2-butanone chlorolom	CHIOI/DCE
SB-AEW	8000	-24.4	920	-25.6		-37.2				
SB-AEW K	16000	-23.9	1900	-23.0						
SB-AMW1-25'	1900	-23.2	1200	-14.5	2800	-27.8	22			
SB-AMW2-25'	280	-5.2	440	6.8	6400	-26.8	33			
SB-AMW2-25' K	290	-4.3	430	6.6	6500	-26.1	34			
SB-AMW3-Z1	580	-21.9	690	-14.7	7000	-22.9	200			
SB-AMW3-Z2	520	-20.0	1400	-11.7	5500	-26.4	77			
SB-AMW3-Z3	87	-18.6	2000	-8.2	5300	-26.9	27			
SB-AMW4-Z1	150	-18.2	380	-4.8	7000	-26.9	110			
SB-AMW4-Z2					7600	-26.3	65			
SB-AMW4-Z3	1900	-22.2	3400	-16.4	4600	-27.4	11			
SB-AMW5-Z1	1100	-22.3	1500	-15.5	6000	-25.2	86			
SB-AMW5-Z2	91	-8.6	290	4.0	6900	-27.0	63			
			3600		310	-27.3	00			
SB-AMW6-25'	53000	-23.5	3000	-19.8	310	-27.5				
SB-PIW1-25'	13		9		19	-19.3		-12.6		
						-19.3		-12.0		
SB-PIW2-25'	17		3		8					
SB-PIW3-25'	54	00.0	2	00.5	4					
SB-PMW1-25'	1400	-28.0	69	-30.5			6			
SB-PMW2-25'	4400	-23.4								
SB-PMW3-Z1	47000	-24.3	190		310					
SB-PMW3-Z2	1400	-24.1	11							
SB-PMW4-Z1	30000	-23.3	4500	-25.2						
SB-PMW4-Z3	2000	-24.2	22							
SB-PMW4-Z4	6700	-24.3	59							
SB-PMW5-Z1	39000	-24.6	380		310					
SB-PMW5-Z2	4600	-23.9	69							
SB-PMW5-Z3	5600	-24.6	53							
SB-PMW6-25'	790	-16.2	460	-20.5	120	-18.5	190	-21.0		
SB-PMW6-25' K		-16.9	490	-20.3	130	-20.6	170	-21.7		
SB-PMW7-25'	190	-8.7	96	1.3	590	-18.3	28	-34.1		
SB-PMW8-25'	710	-12.6	180	-7.3	250	-8.8	350	-26.5		
SB-PMW9-25'		-12.8		-1.3	37	-6.6 -4.2		-23.4		
3D-F101009-23	19	-10.0	19		31	-4.2	110	-23.4		
Average of all		10.0		12.6		22.4		22.2		
Average of all		-19.8		-12.6		-23.4		-23.2		
AEW/AMW Ave		-18.2		-10.5		-27.2				
PIW/PMW Ave		-21.2		-17.1		-15.0		-23.2		
		12		10		12		10		
October 2009	TCE (ppb)	$\delta^{13}C$	cDCE	$\delta^{13}C$	VC	$\delta^{13}C$	Ethene	δ^{13} C eth	2-butanone chloroform	chlor/DCE
October 2009 SB-AEW	TCE (ppb)	δ ¹³ C -24.2	cDCE	δ ¹³ C -19.8	VC	δ ¹³ C -30.0	Ethene	δ ¹³ C eth	2-butanone chloroform	chlor/DCE
	TCE (ppb)		cDCE		VC		Ethene	δ ¹³ C eth	2-butanone chloroform	chlor/DCE
SB-AEW	TCE (ppb)	-24.2 -20.6	cDCE	-19.8 -18.2	VC	-30.0 -19.6	Ethene		2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25'	TCE (ppb)	-24.2	cDCE	-19.8	VC	-30.0	Ethene		2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1	TCE (ppb)	-24.2 -20.6	cDCE	-19.8 -18.2	VC	-30.0 -19.6 -24.5	Ethene		2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2	TCE (ppb)	-24.2 -20.6	cDCE	-19.8 -18.2	VC	-30.0 -19.6 -24.5 -21.4	Ethene	-44.2	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW3-Z3	TCE (ppb)	-24.2 -20.6	cDCE	-19.8 -18.2	VC	-30.0 -19.6 -24.5 -21.4	Ethene	-44.2 -28.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1	TCE (ppb)	-24.2 -20.6	cDCE	-19.8 -18.2	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5	Ethene	-44.2	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1	TCE (ppb)	-24.2 -20.6	cDCE	-19.8 -18.2	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3	Ethene	-44.2 -28.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW4-Z3 SB-AMW4-Z2 SB-AMW4-Z2 SB-AMW4-Z3	TCE (ppb)	-24.2 -20.6 -6.9	cDCE	-19.8 -18.2 -21.2	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1	Ethene	-44.2 -28.7 -40.4	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW4-Z3 SB-AMW4-Z3 SB-AMW5-Z1	TCE (ppb)	-24.2 -20.6	cDCE	-19.8 -18.2	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3	Ethene	-44.2 -28.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2	TCE (ppb)	-24.2 -20.6 -6.9	cDCE	-19.8 -18.2 -21.2	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1	Ethene	-44.2 -28.7 -40.4	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z3	TCE (ppb)	-24.2 -20.6 -6.9	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2	TCE (ppb)	-24.2 -20.6 -6.9	cDCE	-19.8 -18.2 -21.2	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1	Ethene	-44.2 -28.7 -40.4	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW3-21 SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW6-25'	TCE (ppb)	-24.2 -20.6 -6.9 -23.4	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z2 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW6-25' SB-PIW1-25'	TCE (ppb)	-24.2 -20.6 -6.9	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z5' SB-PIW1-25' SB-PIW1-25' SB-PIW2-25'	TCE (ppb)	-24.2 -20.6 -6.9 -23.4	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z5' SB-PIW1-25' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25'	TCE (ppb)	-24.2 -20.6 -6.9 -23.4	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z5' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25'	TCE (ppb)	-24.2 -20.6 -6.9 -23.4 -23.9 -22.9	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z5' SB-PIW1-25' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25'	TCE (ppb)	-24.2 -20.6 -6.9 -23.4 -23.9 -22.9	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z5' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25'	TCE (ppb)	-24.2 -20.6 -6.9 -23.4 -23.9 -22.9	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW3-Z1 SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z3 SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PMW1-25' SB-PMW1-25' SB-PMW1-25' SB-PIW3-25' SB-PIW3-25' SB-PMW1-25'	TCE (ppb)	-24.2 -20.6 -6.9 -23.4 -23.9 -22.9	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z2 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z5' SB-PIW1-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PHW3-25' SB-PHW3-25' SB-PHW3-25' SB-PHW3-25' SB-PHW3-25' SB-PHW3-25'	TCE (ppb)	-24.2 -20.6 -6.9 -23.4 -23.9 -22.9	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z5' SB-PIW1-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z5'	TCE (ppb)	-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW3-Z1 SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z5 SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z5'	TCE (ppb)	-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.2 -24.4	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z5' SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z3 SB-PMW3-Z3 SB-PMW3-Z3 SB-PMW3-Z3 SB-PMW3-Z3	TCE (ppb)	-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.2 -24.4 -23.3	cDCE	-19.8 -18.2 -21.2 -18.5	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z5' SB-PIW3-Z5' SB-PIW3-Z5' SB-PIW3-Z5' SB-PIW3-Z5' SB-PIW3-Z5' SB-PIW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z3 SB-PMW4-Z1 SB-PMW4-Z3	TCE (ppb)	-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.2 -24.4 -23.3 -24.3	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z5' SB-PIW1-25' SB-PIW2-25' SB-PIW3-Z5' SB-PIW3-Z1 SB-PMW3-Z1 SB-PMW4-Z1 SB-PMW4-Z3 SB-PMW4-Z3 SB-PMW4-Z3	TCE (ppb)	-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.2 -24.4 -23.3 -24.3 -24.3	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW3-21 SB-AMW3-21 SB-AMW3-23 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z5' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25'		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -24.4 -23.2 -24.4 -23.3 -24.3 -24.7 -23.3	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8 -27.5	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z2 SB-AMW5-Z5' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z5' SB-PIW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.3 -24.3 -24.3 -23.3 -24.3 -23.3 -23.5	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z3 SB-AMW5-Z5' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW4-Z4 SB-PMW4-Z4 SB-PMW5-Z1		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.3 -24.3 -24.3 -24.3 -23.3 -24.3 -23.5 -23.5 -23.1	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8 -27.5	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z3 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z5' SB-PIW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW4-Z4 SB-PMW4-Z4 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.2 -24.4 -23.3 -24.7 -23.3 -24.7 -23.5 -23.1 -23.2	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8 -22.8 -22.2 -23.1	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8 -27.5	Ethene	-44.2 -28.7 -40.4 -48.7 -42.7 -45.8 -53.6	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-Z5' SB-PIW3-Z5' SB-PIW3-Z5' SB-PIW3-Z1 SB-PMW3-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW4-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.3 -24.3 -24.3 -24.3 -23.3 -24.3 -23.5 -23.5 -23.1	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8 -27.5	Ethene	-44.2 -28.7 -40.4 -48.7	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW3-Z1 SB-AMW3-Z1 SB-AMW3-Z1 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z3 SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW6-Z5' SB-PMW6-Z5' SB-PMW6-Z5'		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.3 -24.7 -23.3 -24.7 -23.3 -23.5 -23.1 -23.5 -23.1	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8 -22.8 -22.2 -23.1	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8 -27.5	Ethene	-44.2 -28.7 -40.4 -48.7 -42.7 -45.8 -53.6	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z5' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW4-Z1 SB-PMW4-Z3 SB-PMW4-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z5' SB-PMW7-25' SB-PMW7-25'		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.2 -24.4 -23.3 -24.7 -23.3 -24.7 -23.5 -23.1 -23.2	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8 -22.8 -22.2 -23.1	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8 -27.5	Ethene	-44.2 -28.7 -40.4 -48.7 -42.7 -45.8 -53.6	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW3-Z1 SB-AMW3-Z1 SB-AMW3-Z1 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z3 SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW6-Z5' SB-PMW6-Z5' SB-PMW6-Z5'		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.3 -24.7 -23.3 -24.7 -23.3 -23.5 -23.1 -23.5 -23.1	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8 -22.8 -22.2 -23.1	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8 -27.5	Ethene	-44.2 -28.7 -40.4 -48.7 -42.7 -45.8 -53.6	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-Z5' SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW4-Z4 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z5' SB-PMW5-Z5'		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.2 -24.4 -23.3 -24.7 -23.3 -24.7 -23.3 -23.5 -24.7 -23.3 -24.7 -23.3 -23.5 -24.7 -23.9 -24.7 -25.7 -	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8 -22.8 -22.2 -23.1	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8 -27.5	Ethene	-44.2 -28.7 -40.4 -48.7 -42.7 -45.8 -53.6	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z5' SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW4-Z1 SB-PMW4-Z3 SB-PMW4-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z3 SB-PMW5-Z5' SB-PMW7-25' SB-PMW7-25'		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.3 -24.7 -23.3 -24.7 -23.3 -23.5 -23.1 -23.5 -23.1	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8 -22.8 -22.2 -23.1	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8 -27.5	Ethene	-44.2 -28.7 -40.4 -48.7 -42.7 -45.8 -53.6	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW2-25' SB-AMW3-Z1 SB-AMW3-Z2 SB-AMW3-Z3 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z3 SB-AMW5-Z3 SB-AMW5-Z3 SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-Z5' SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW4-Z4 SB-PMW5-Z1 SB-PMW5-Z3 SB-PMW5-Z5' SB-PMW5-Z5'		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.2 -24.4 -23.3 -24.7 -23.3 -24.7 -23.3 -23.5 -24.7 -23.3 -24.7 -23.3 -23.5 -24.7 -23.9 -24.7 -25.7 -	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8 -22.8 -22.2 -23.1	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8 -27.5	Ethene	-44.2 -28.7 -40.4 -48.7 -42.7 -45.8 -53.6	2-butanone chloroform	chlor/DCE
SB-AEW SB-AMW1-25' SB-AMW3-21 SB-AMW3-21 SB-AMW3-21 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW4-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z1 SB-AMW5-Z2 SB-AMW5-Z1 SB-PIW1-25' SB-PIW2-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-25' SB-PIW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW3-Z1 SB-PMW4-Z1 SB-PMW5-Z1 SB-PMW5-Z5' SB-PMW6-25' SB-PMW9-25'		-24.2 -20.6 -6.9 -23.4 -23.9 -22.9 -28.2 -23.5 -24.4 -23.3 -24.7 -23.3 -24.7 -23.3 -24.7 -23.3 -24.7 -23.3 -24.7 -23.9 -24.7 -23.9 -24.7 -23.9 -24.7 -23.9 -24.7 -23.9 -24.7 -25.7 -25.7 -26.7 -27.7 -	cDCE	-19.8 -18.2 -21.2 -18.5 -17.1 -21.8 -22.8 -22.2 -23.1 -18.4	VC	-30.0 -19.6 -24.5 -21.4 -23.6 -20.5 -20.3 -22.1 -23.8 -27.5	Ethene	-44.2 -28.7 -40.4 -48.7 -42.7 -45.8 -53.6 -43.3 -36.2 -40.3 -42.4	2-butanone chloroform	chlor/DCE

COD

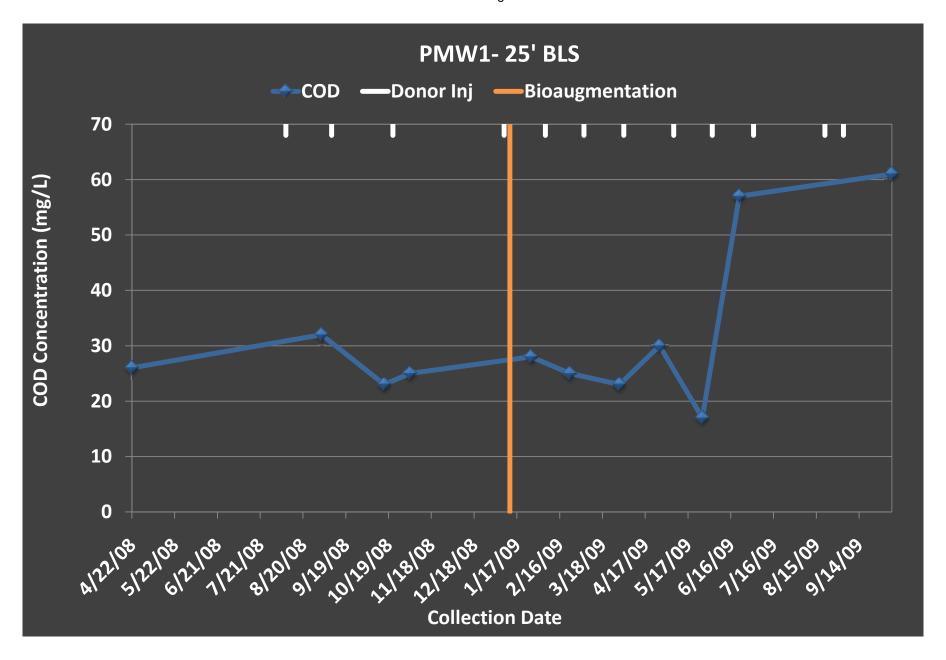
Seal Beach
Groundwater Bioaugmentation

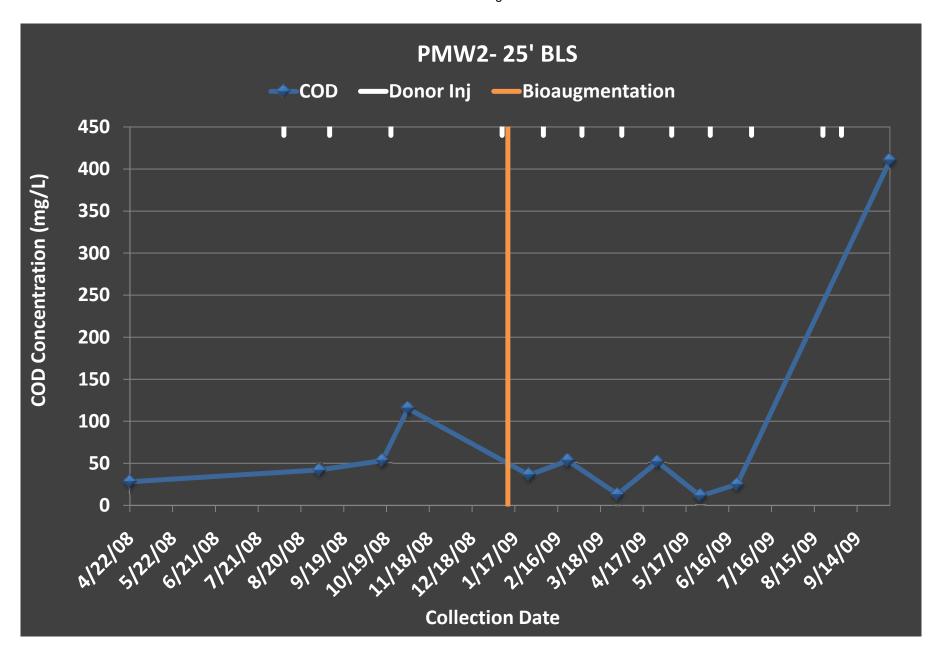


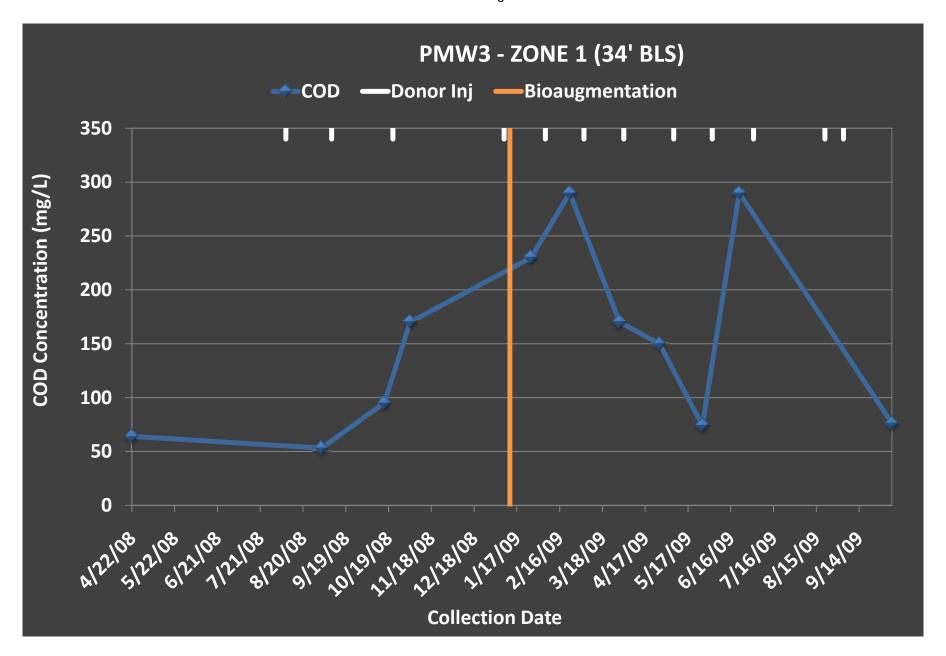


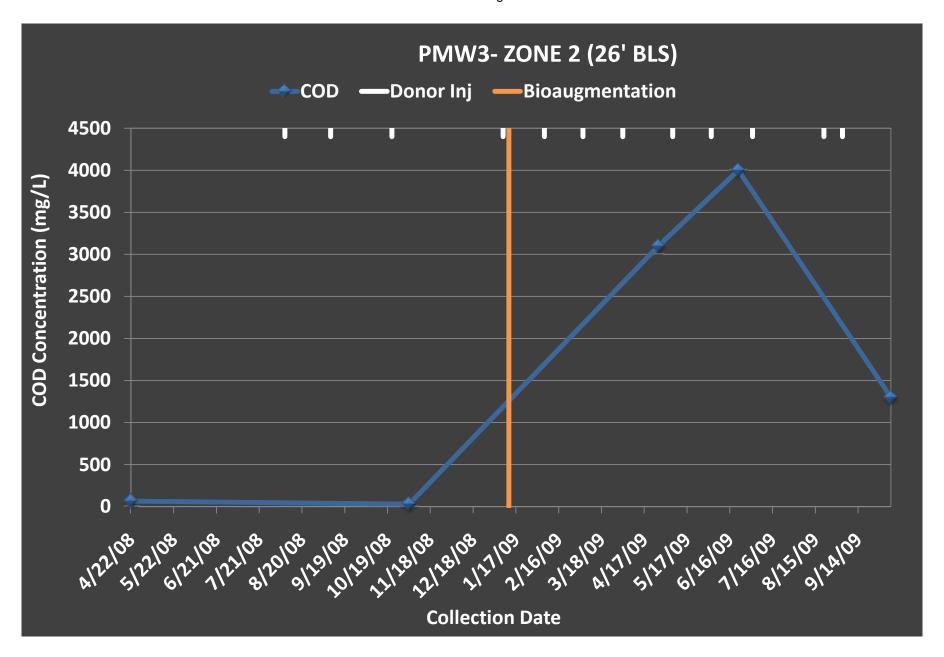


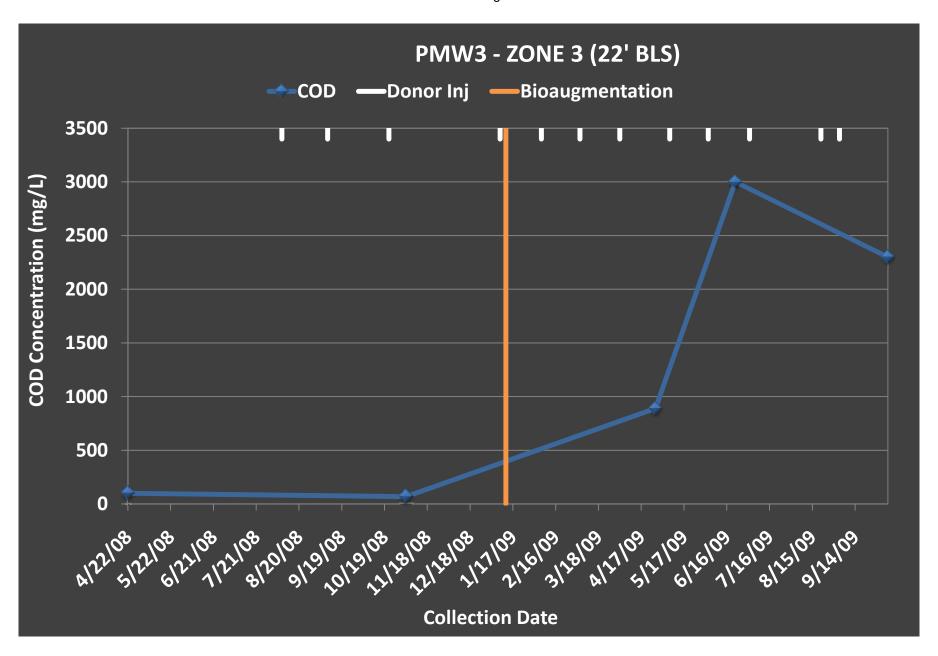
Seal Beach
Groundwater Bioaugmentation

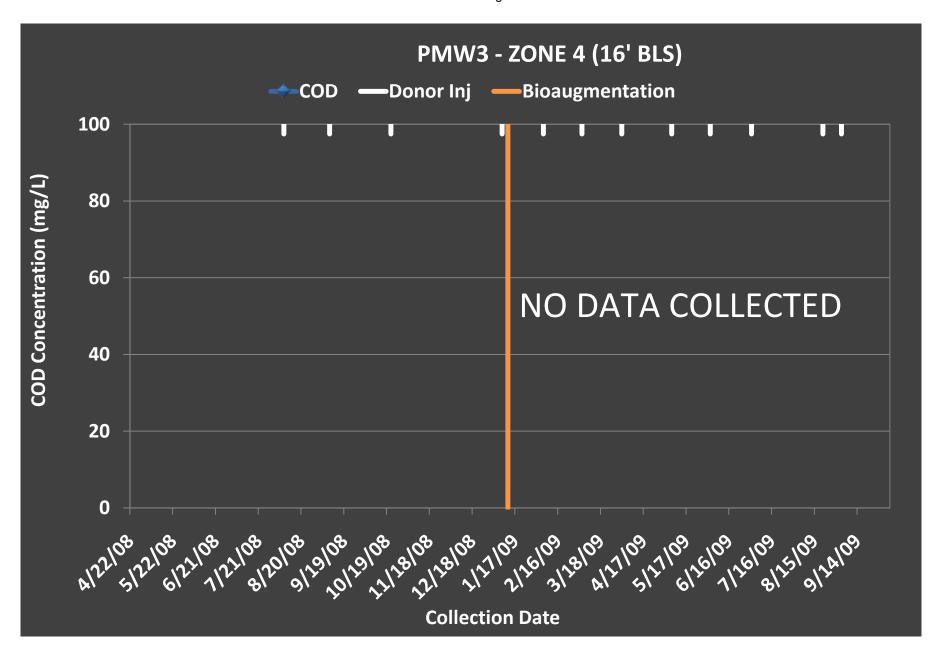


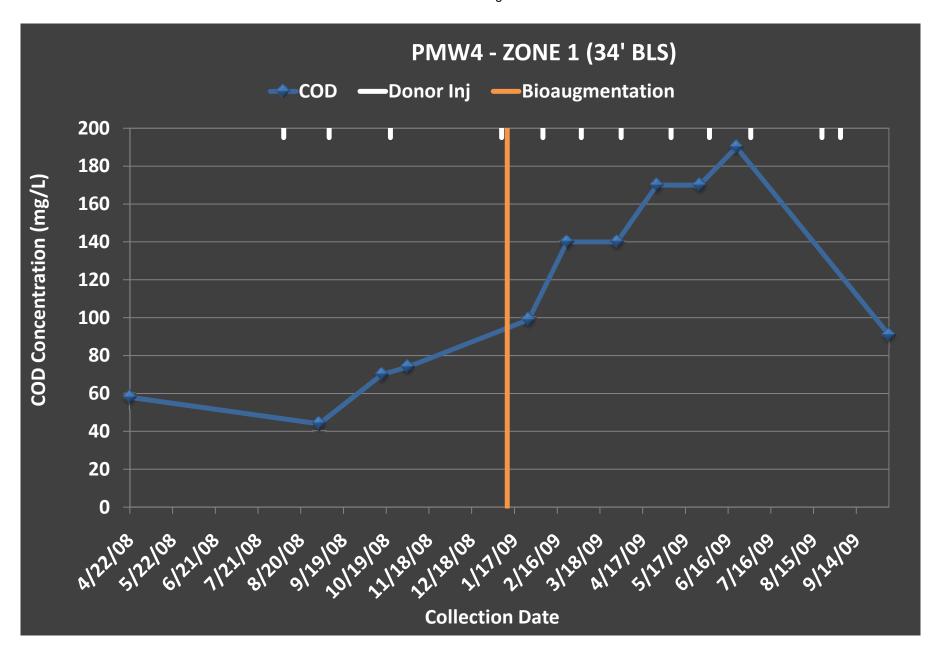


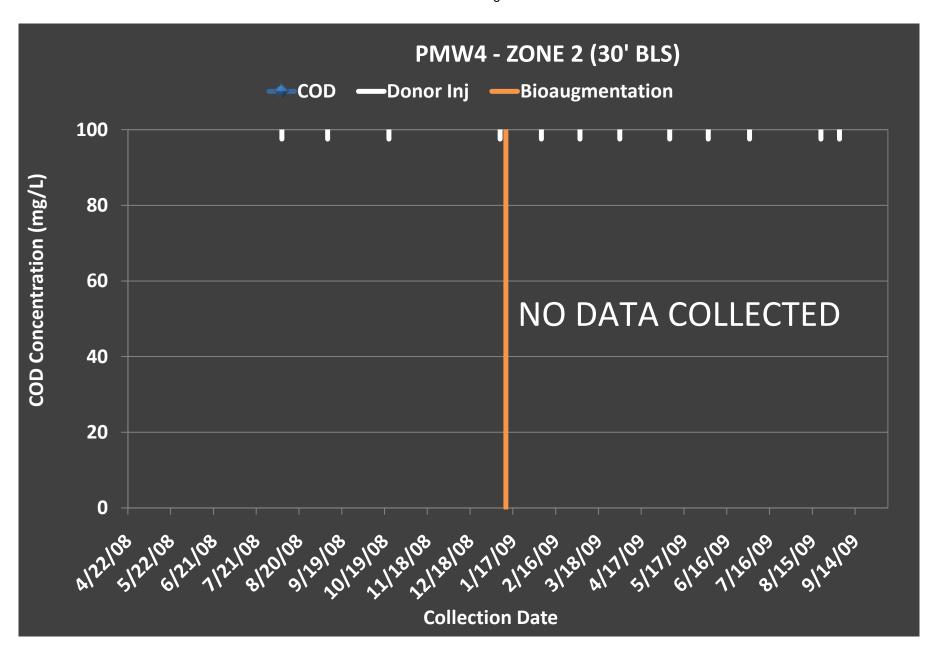


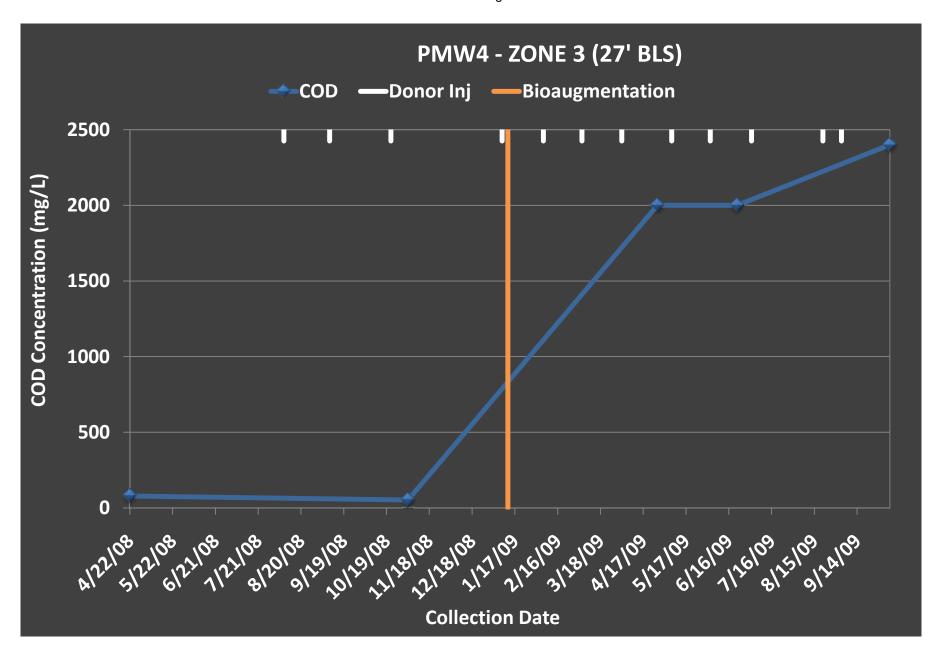


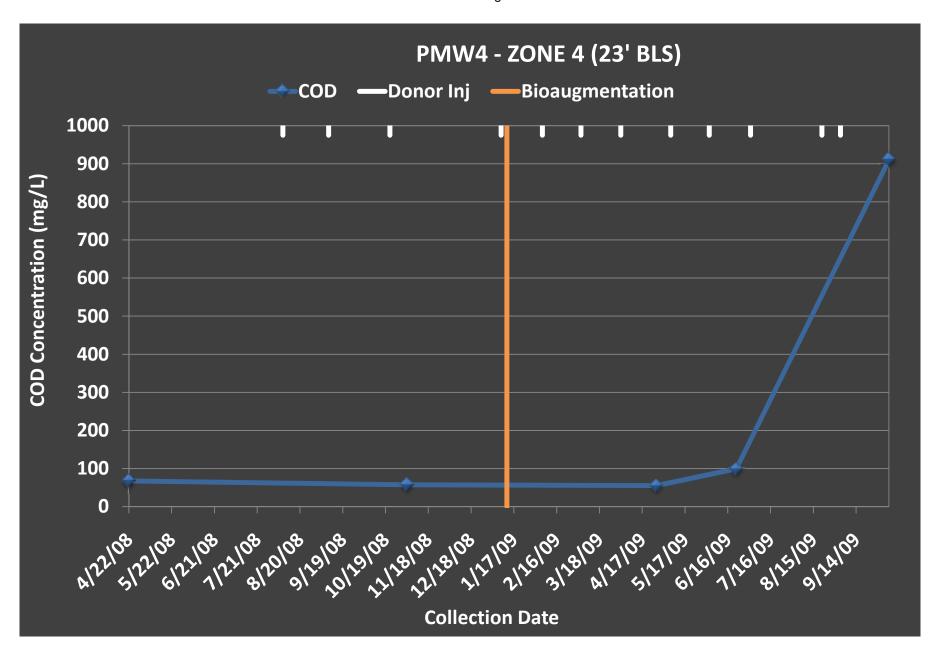


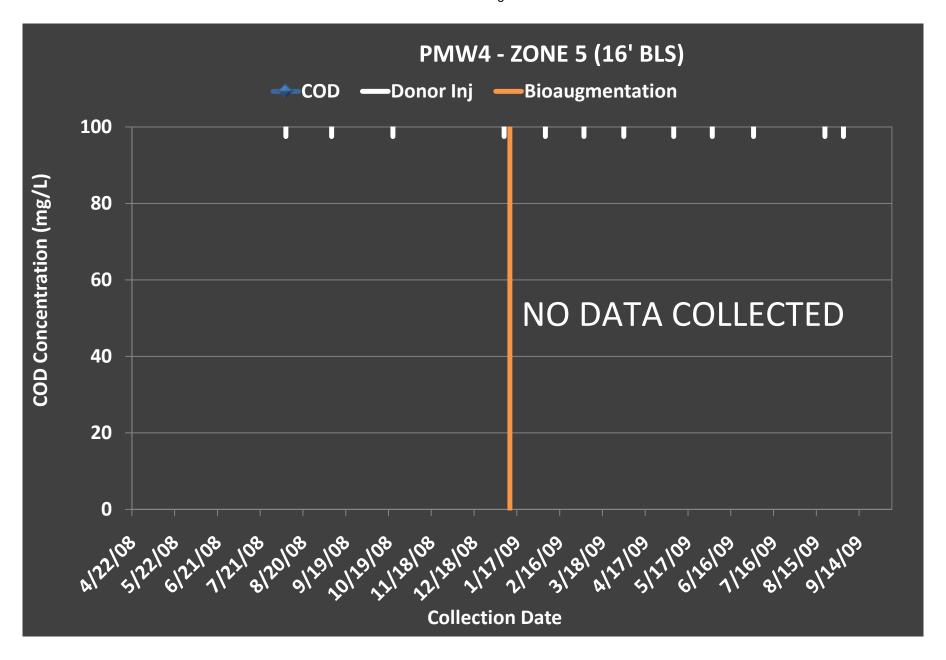


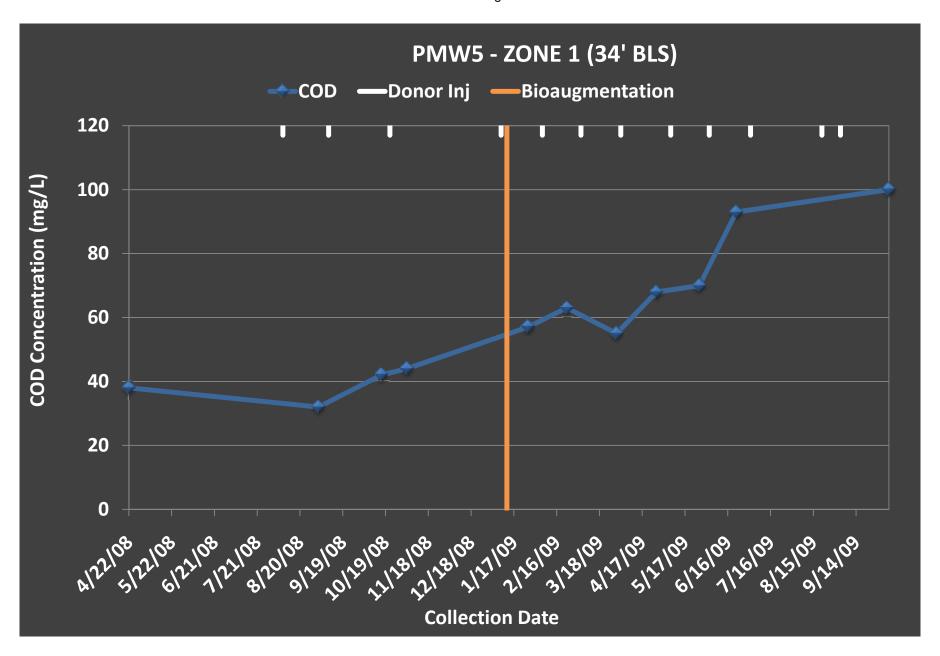


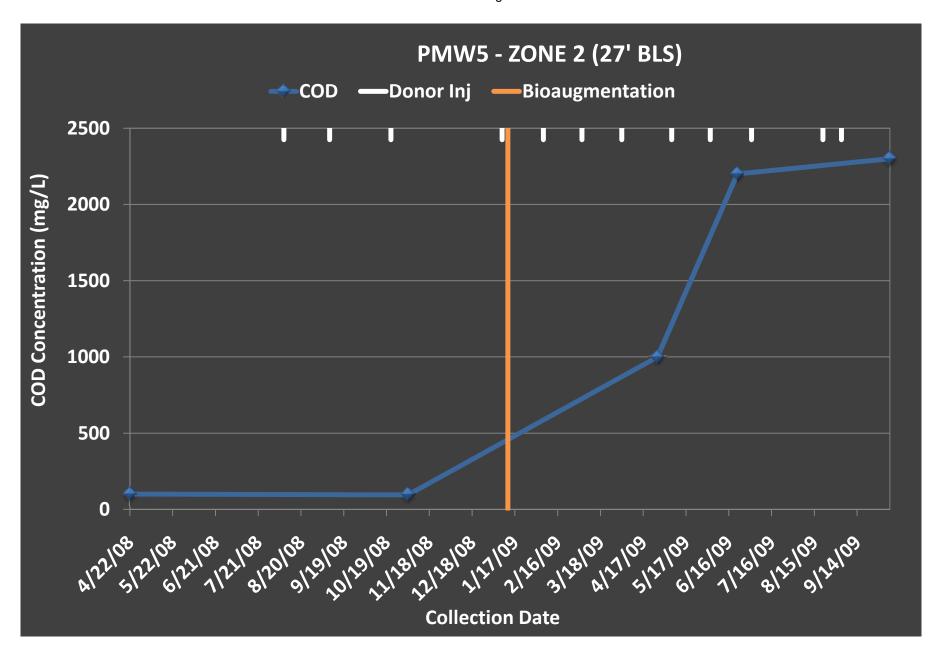




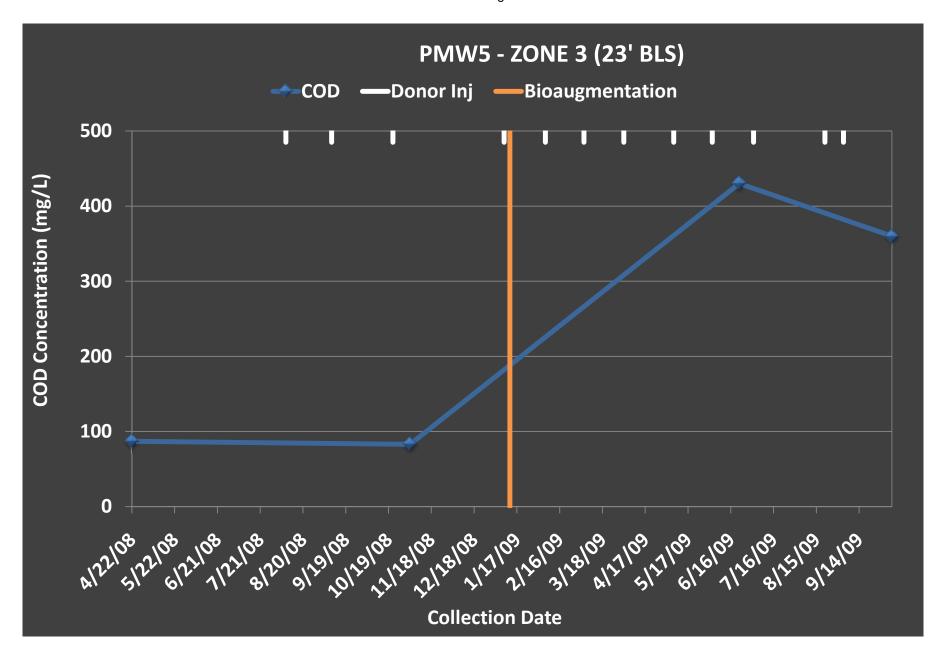


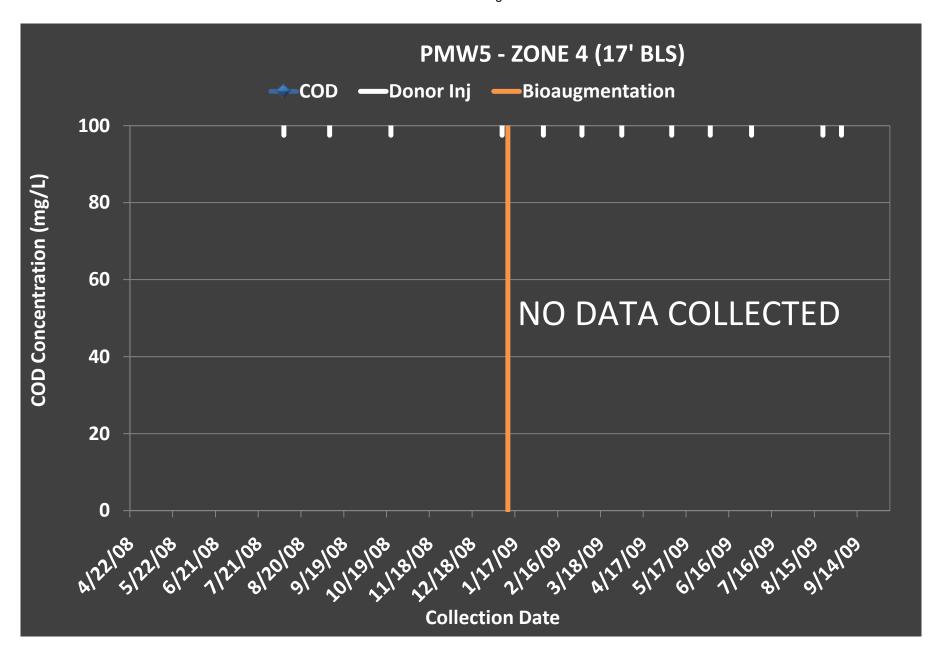


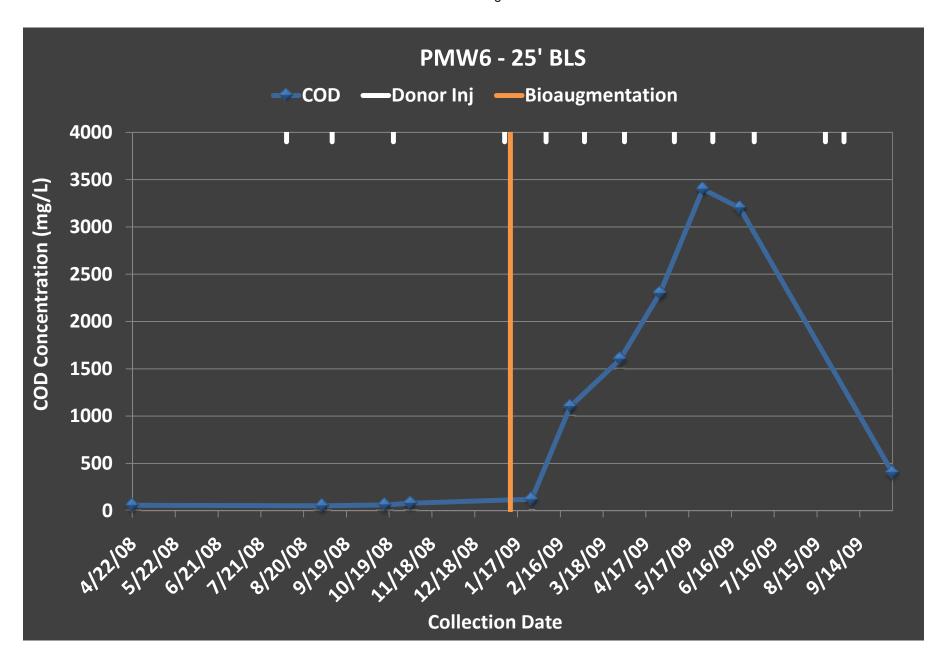




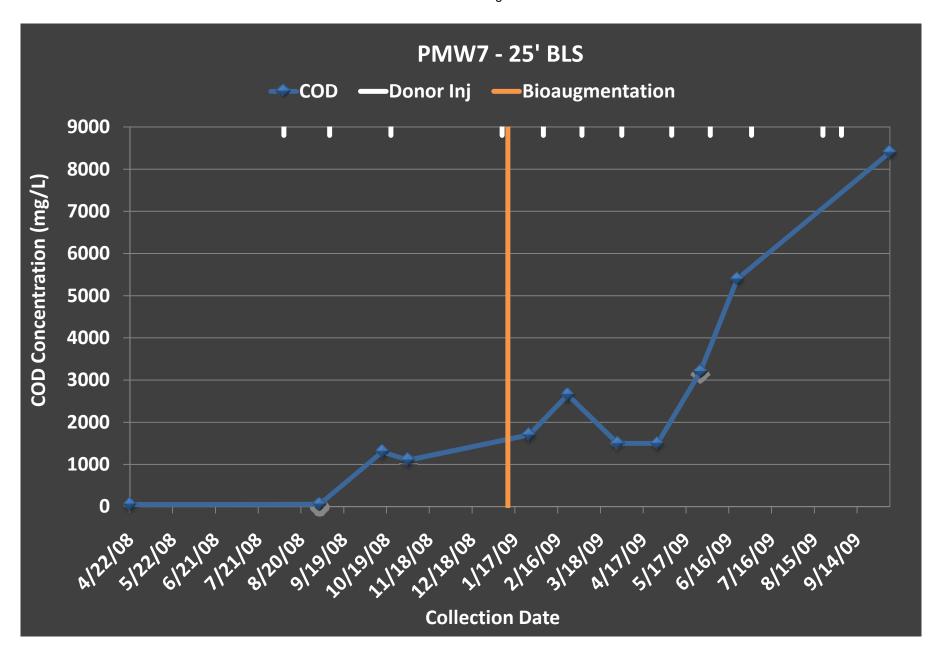
Seal Beach
Groundwater Bioaugmentation



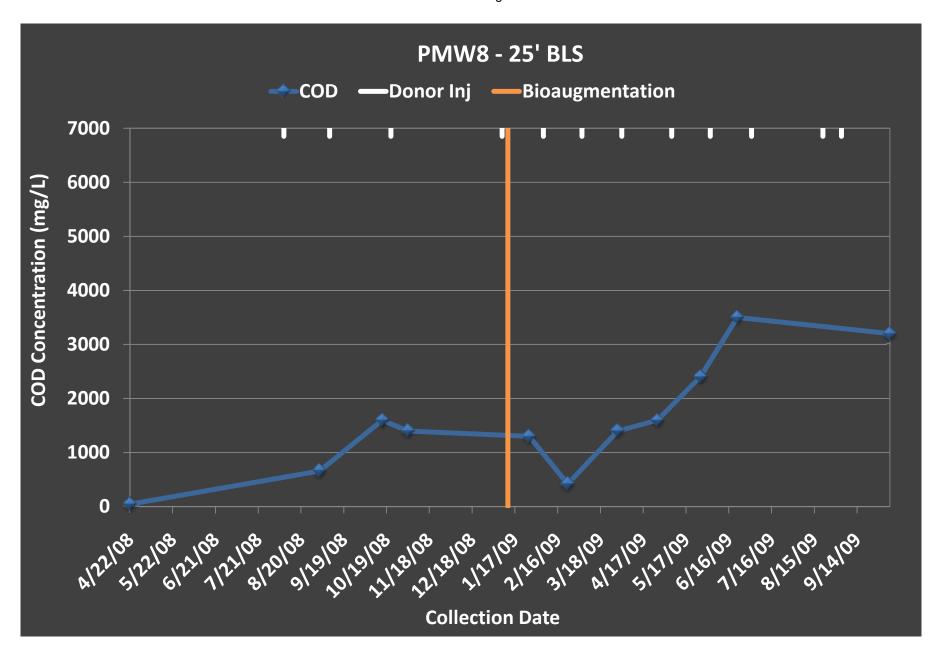




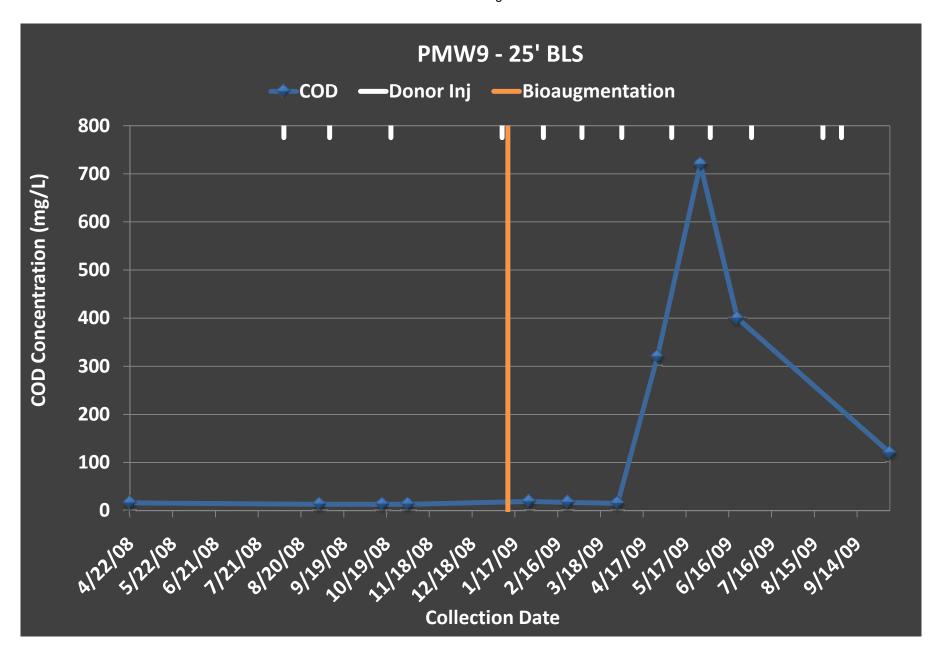
Seal Beach
Groundwater Bioaugmentation



Seal Beach
Groundwater Bioaugmentation

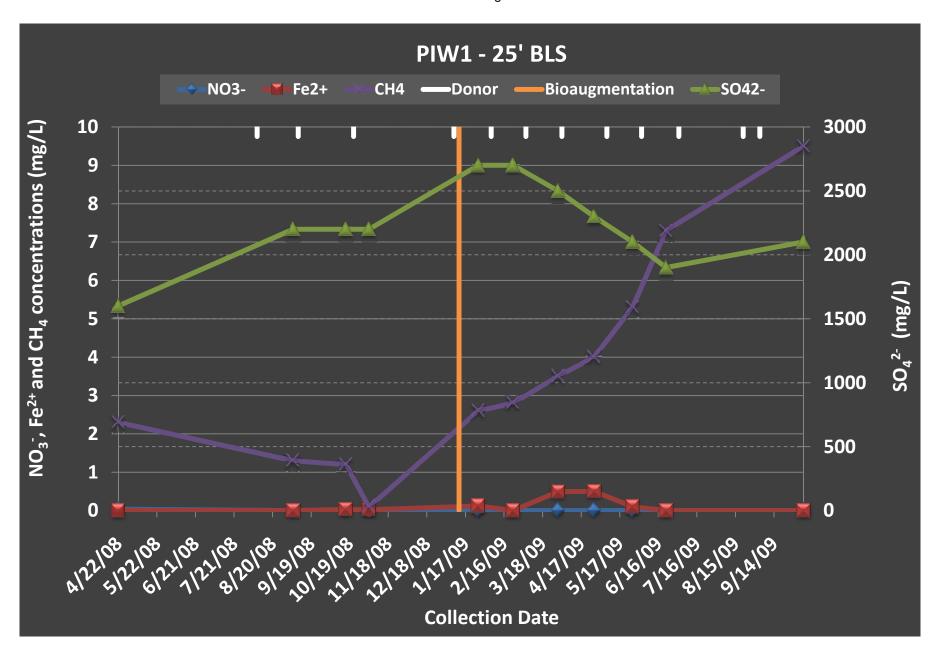


Seal Beach
Groundwater Bioaugmentation

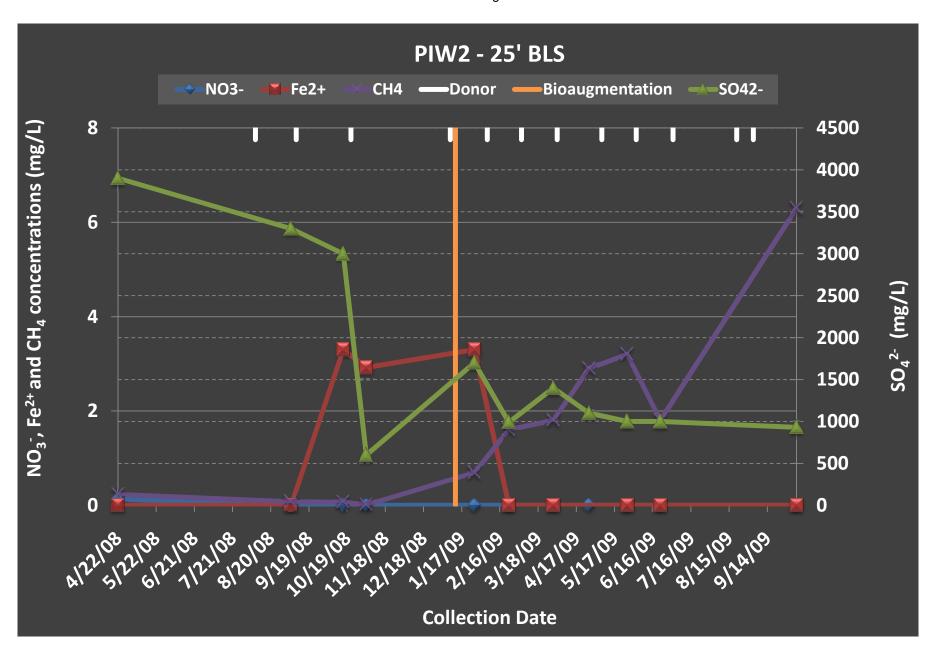


Electron Acceptors

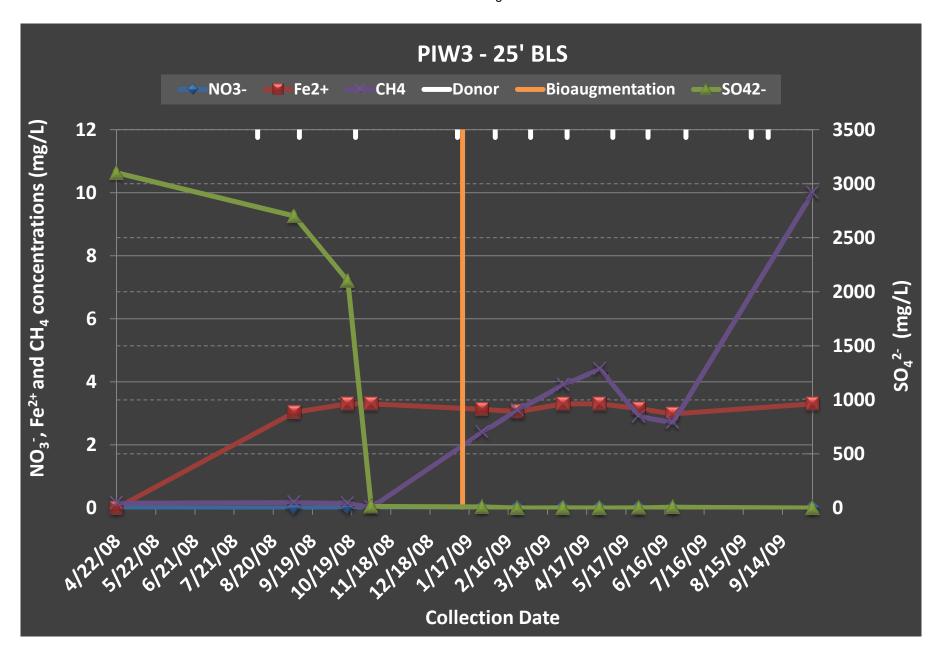
Seal Beach
Groundwater Bioaugmentation



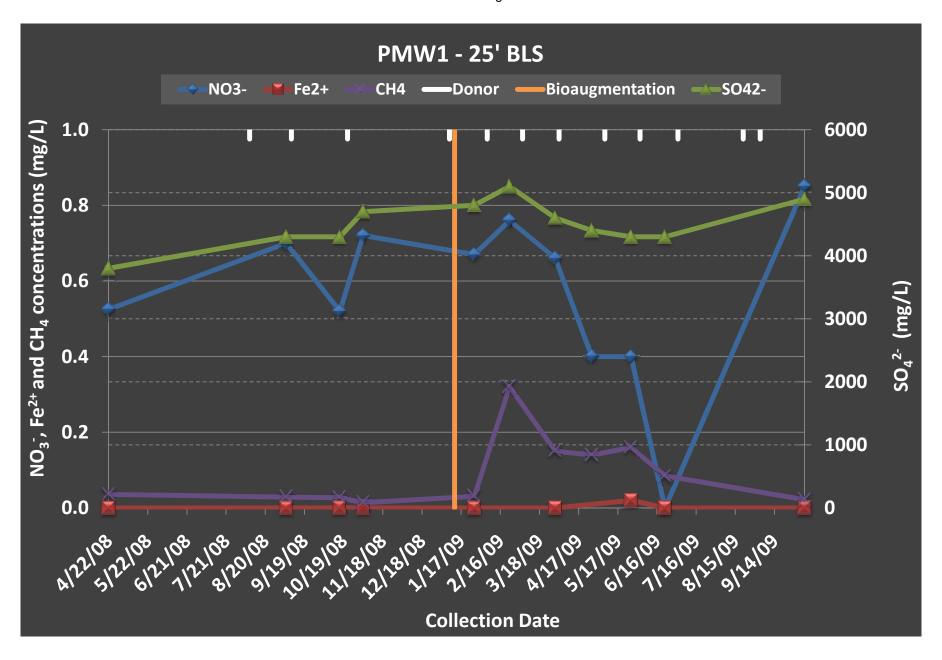
Seal Beach
Groundwater Bioaugmentation



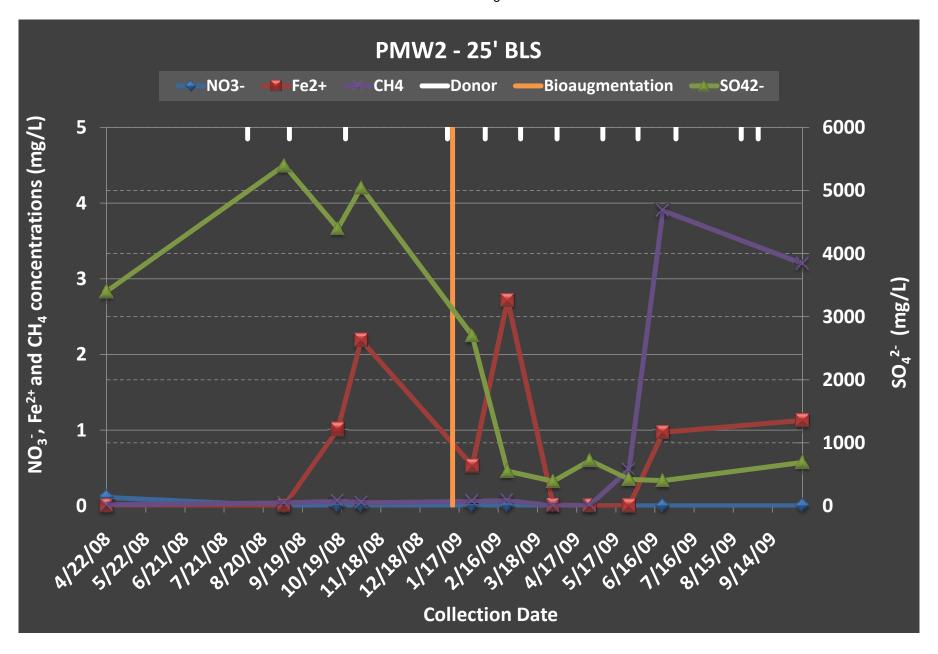
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Groundwater Bioaugmentation



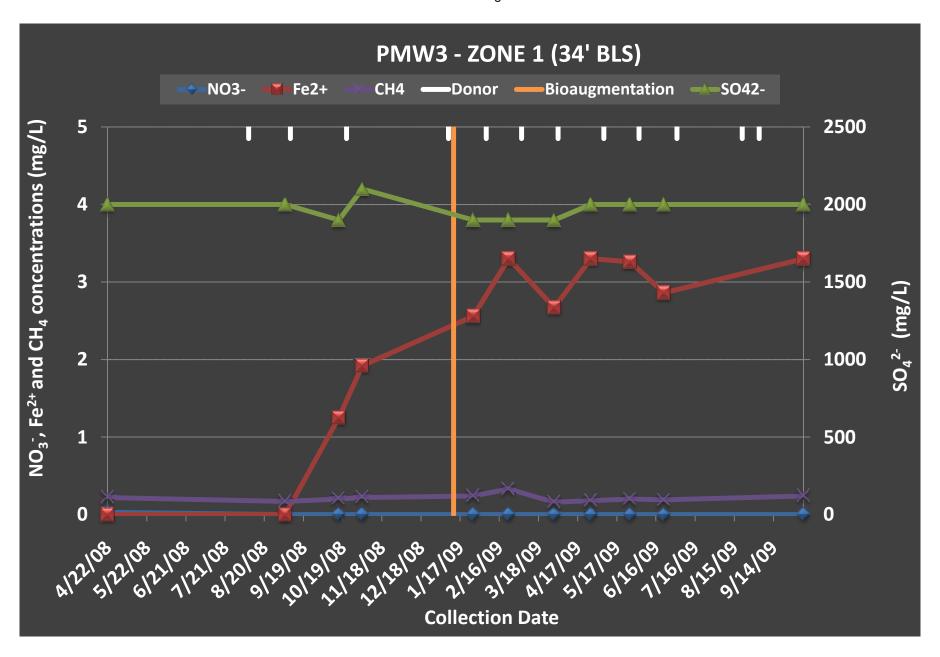
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Groundwater Bioaugmentation



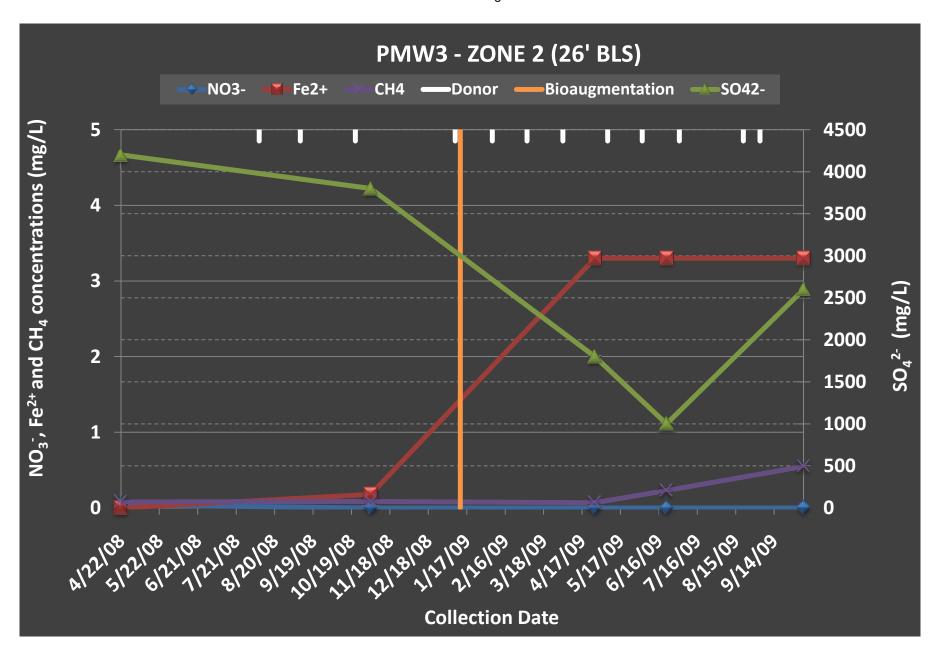
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Groundwater Bioaugmentation



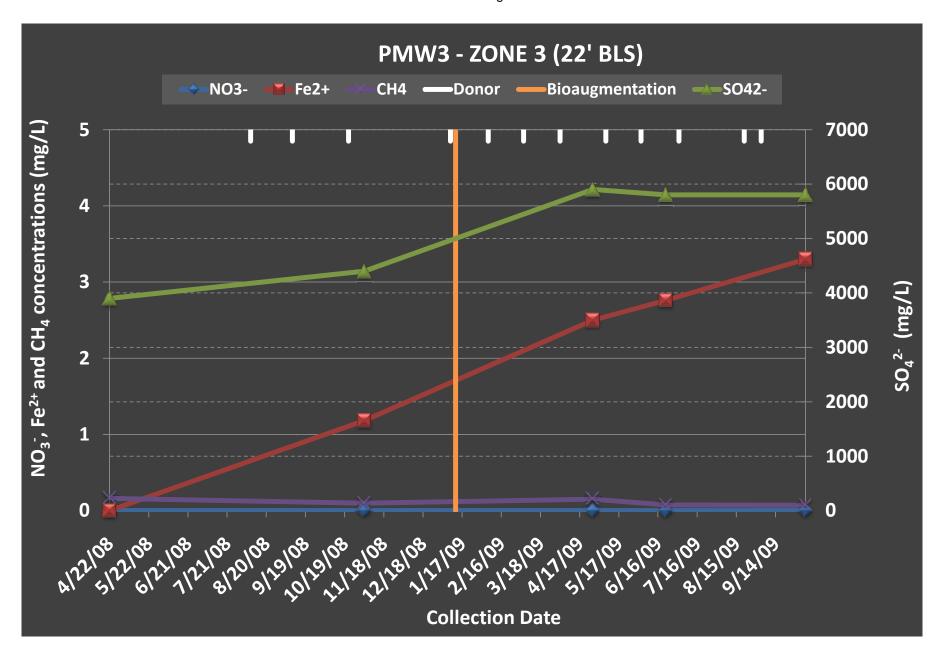
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Groundwater Bioaugmentation



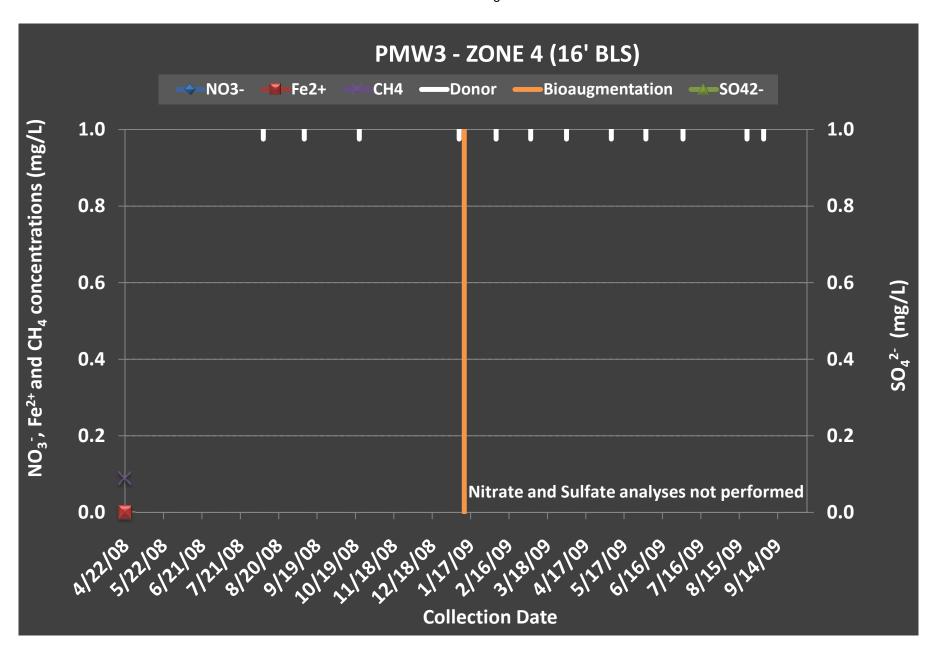
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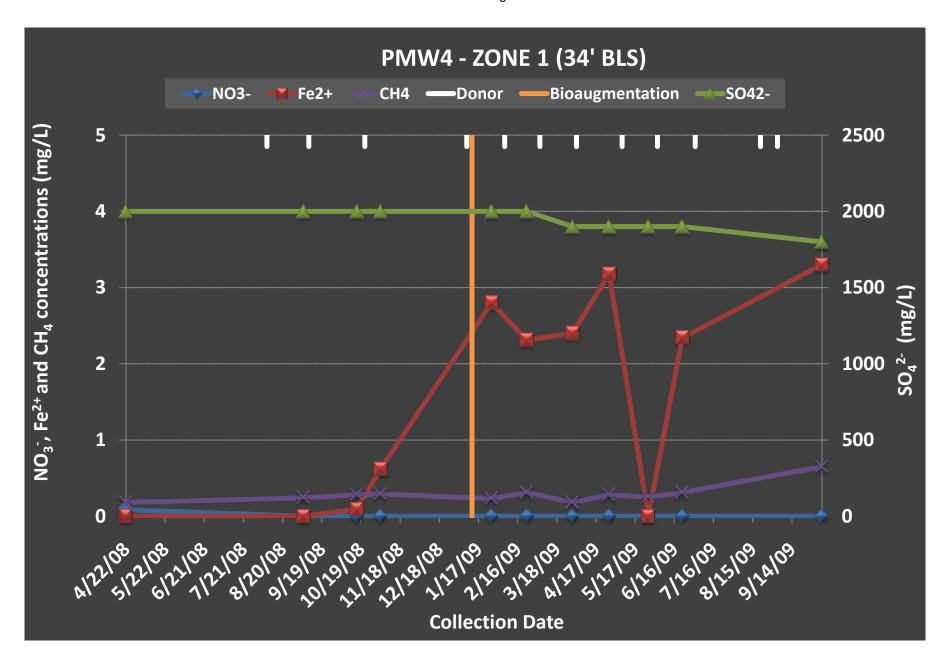
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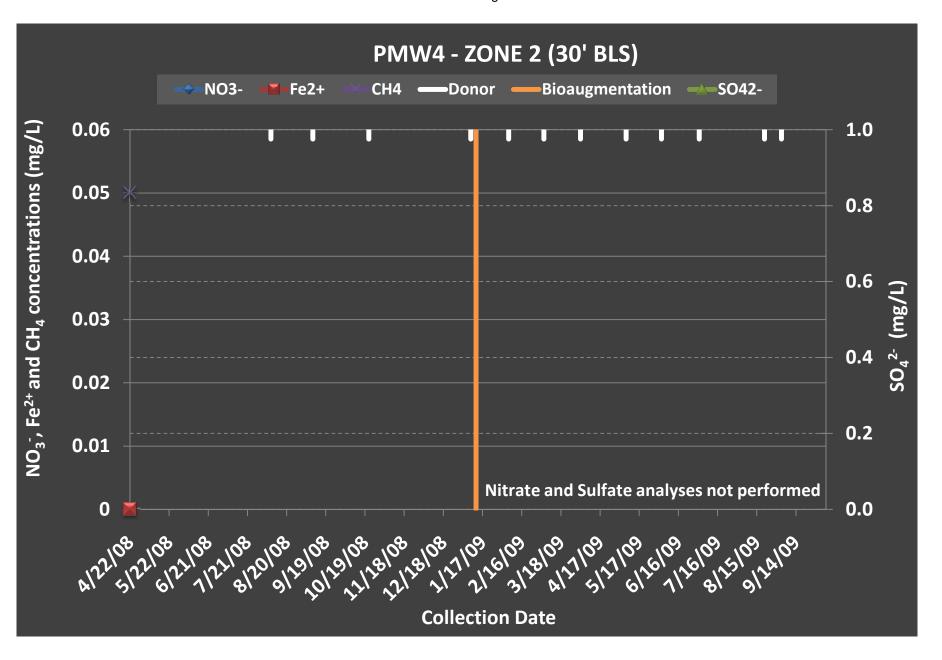


Seal Beach
Groundwater Bioaugmentation

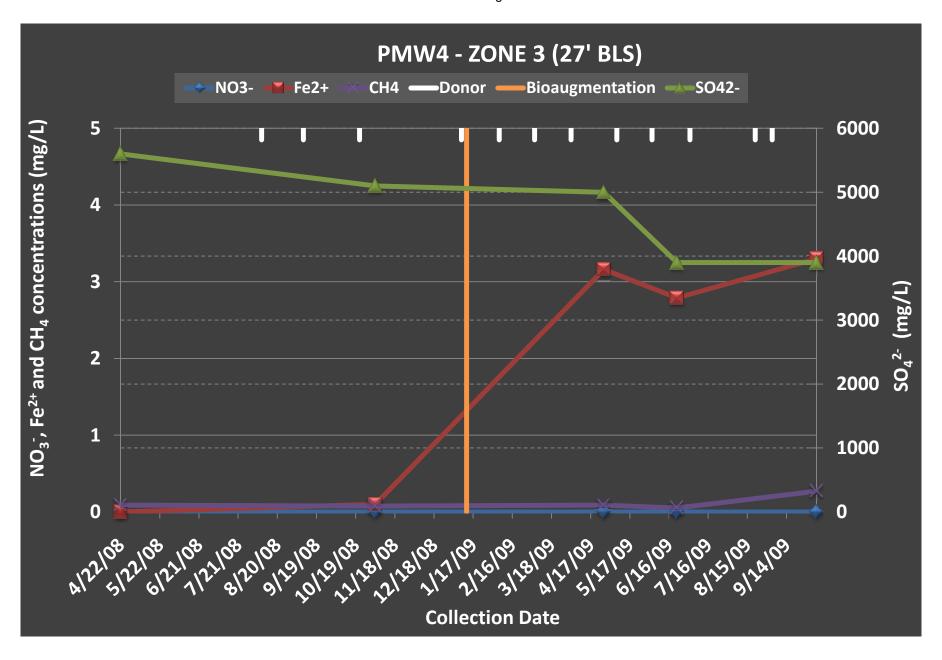


Seal Beach
Groundwater Bioaugmentation

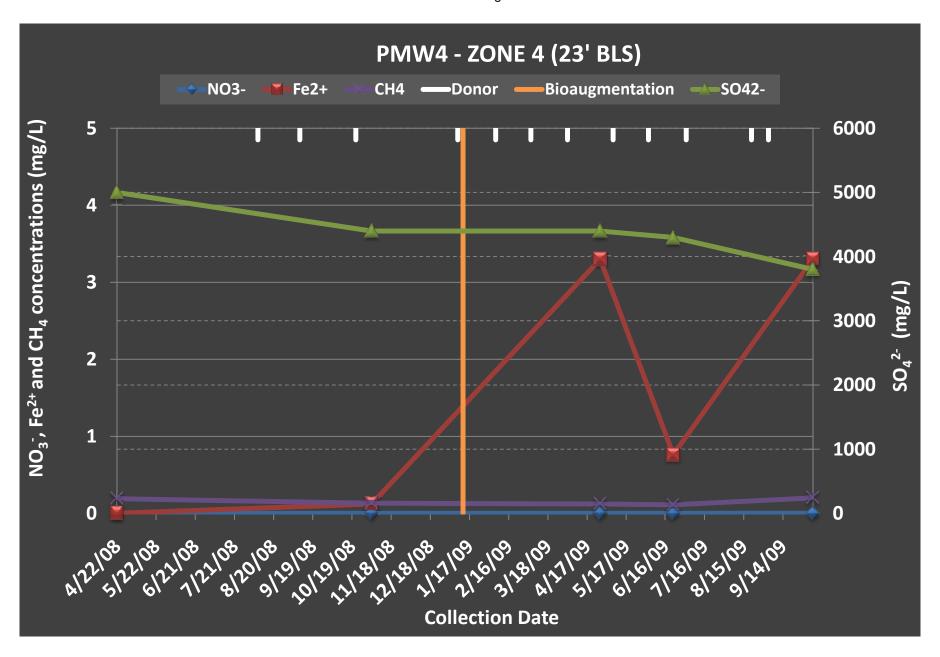




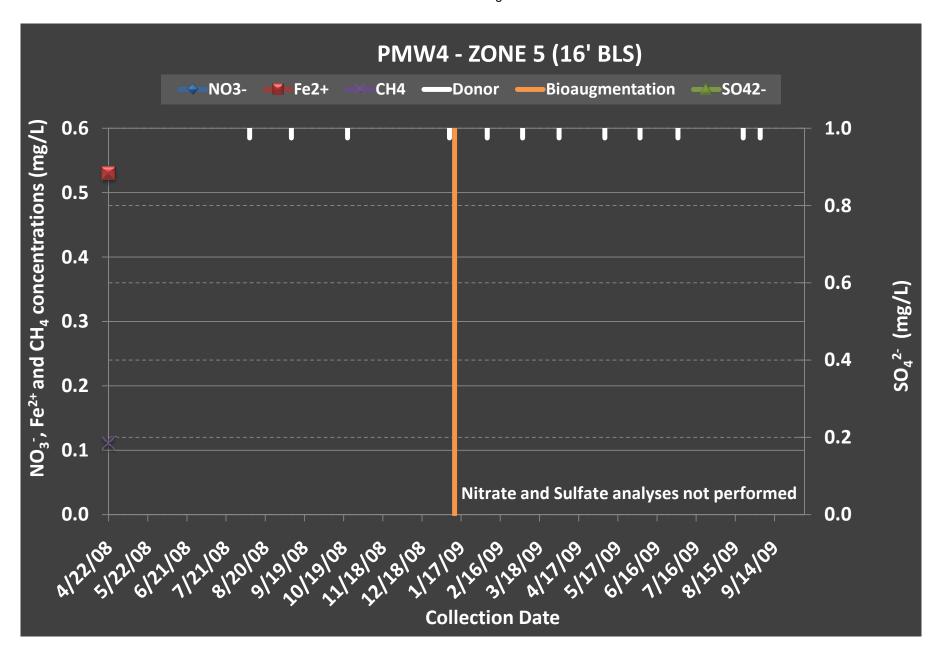
Seal Beach
Groundwater Bioaugmentation



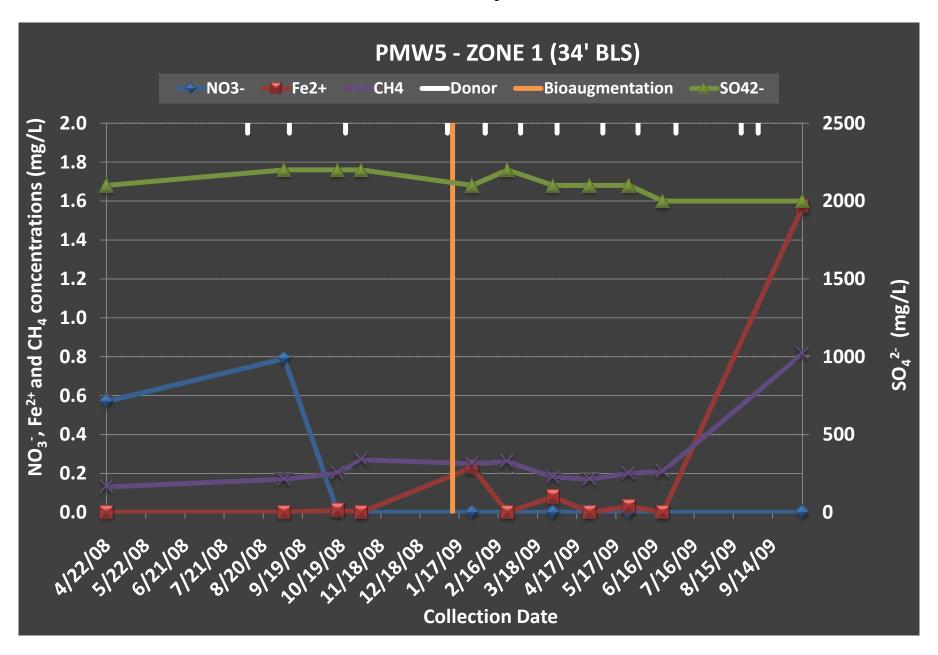
Seal Beach
Groundwater Bioaugmentation



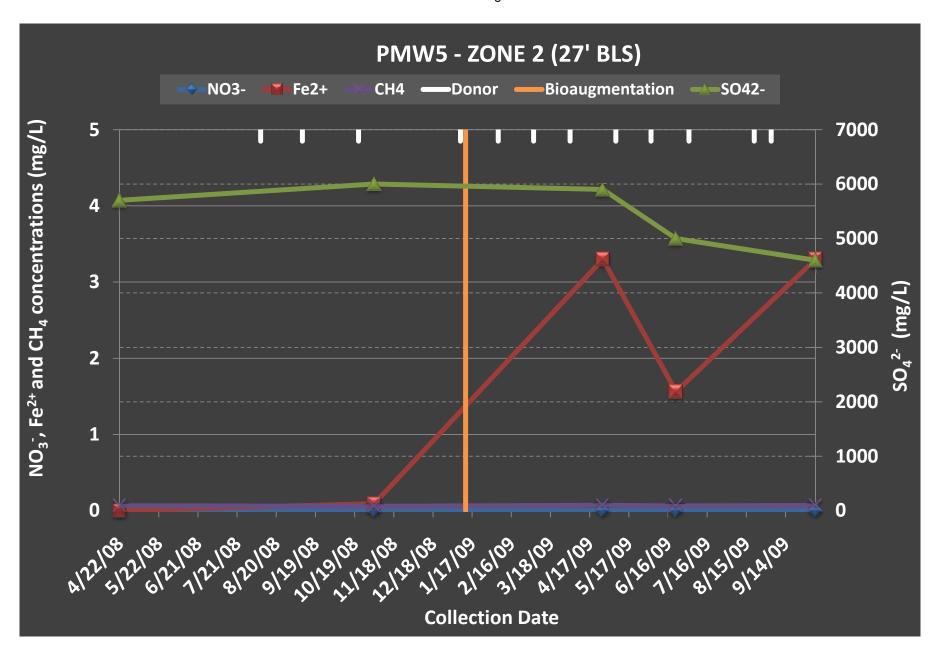
Seal Beach
Groundwater Bioaugmentation



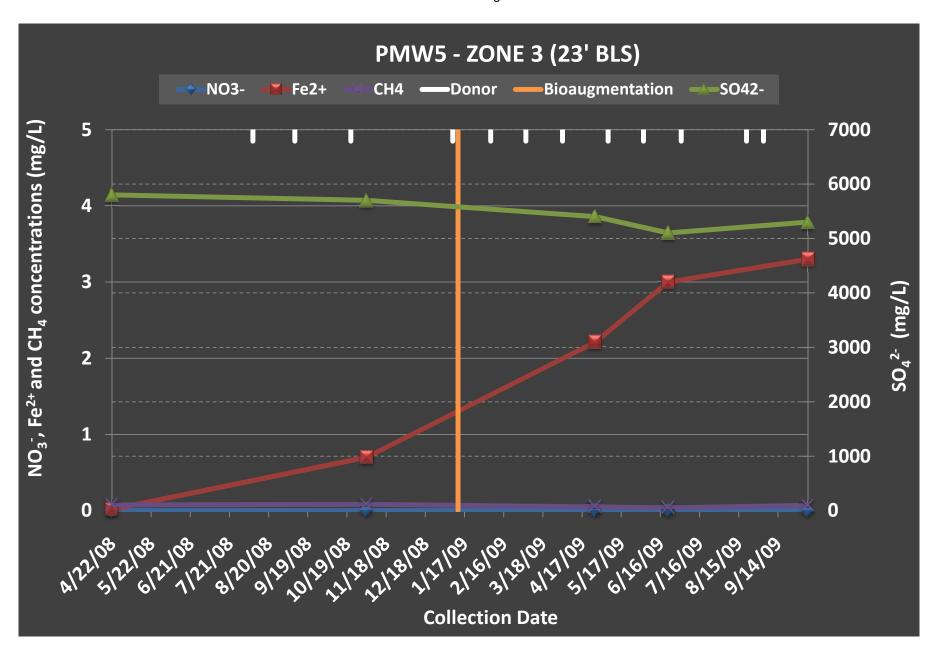
Seal Beach Groundwater Bioaugmentation



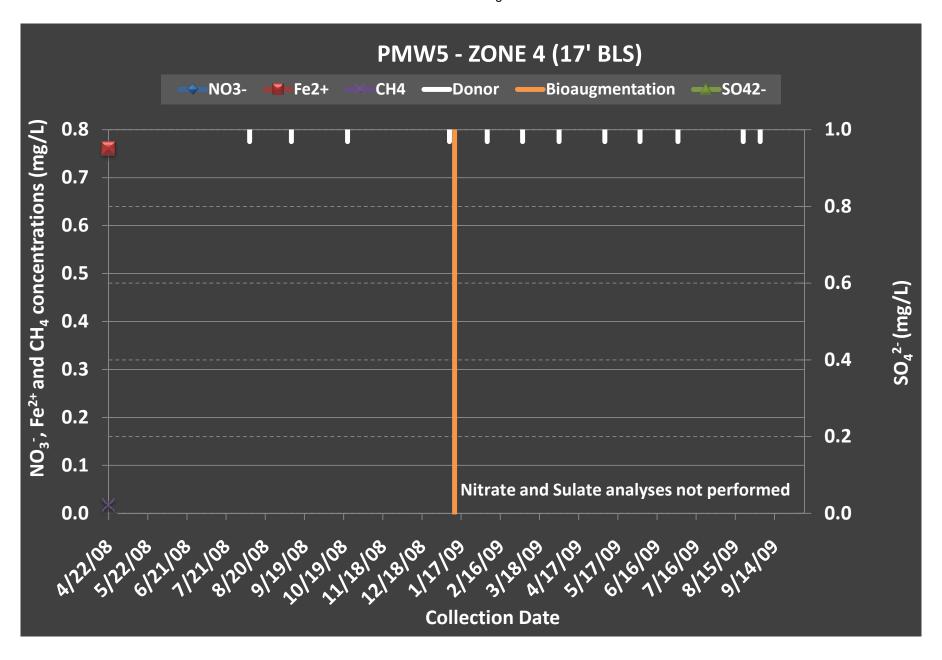
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Groundwater Bioaugmentation



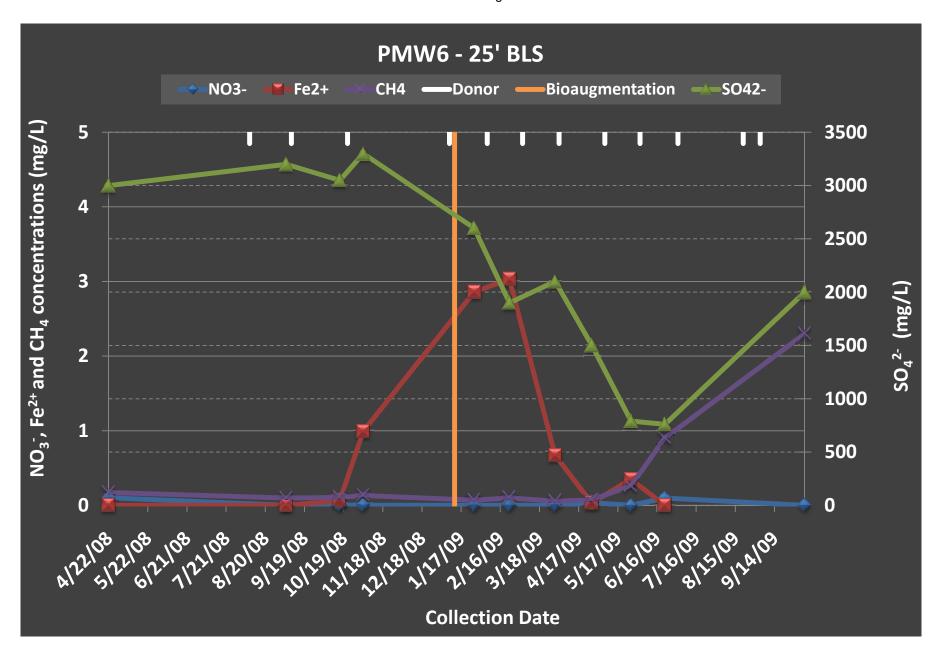
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Groundwater Bioaugmentation



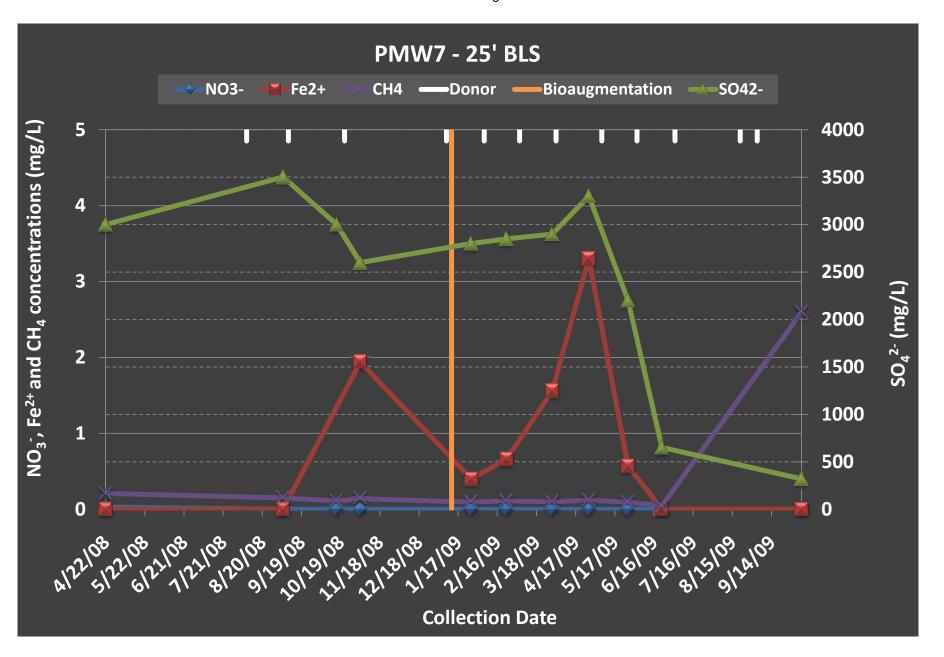
Seal Beach
Groundwater Bioaugmentation



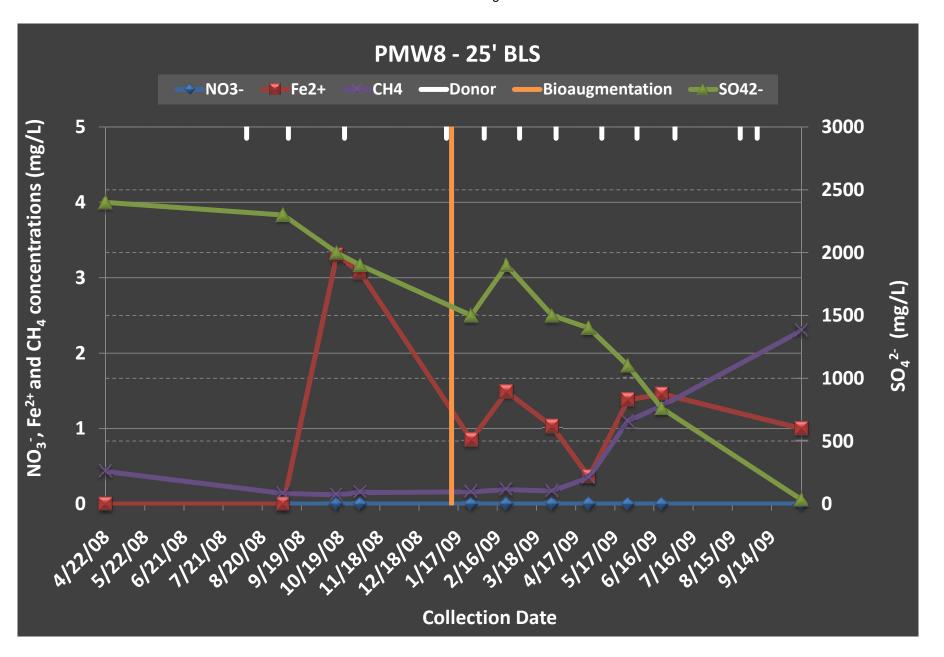
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Groundwater Bioaugmentation



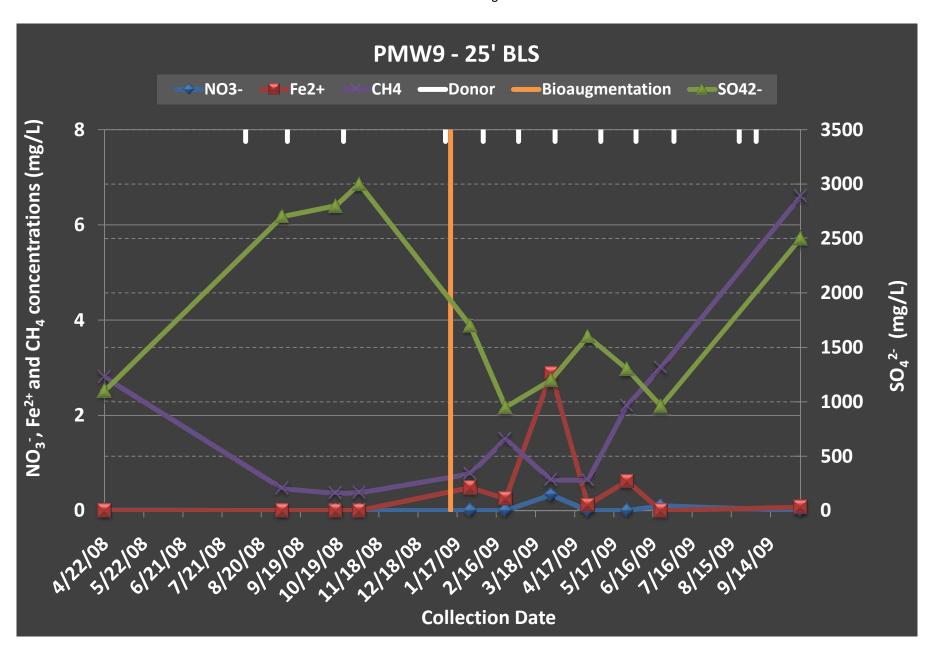
Seal Beach
Groundwater Bioaugmentation



Seal Beach
Groundwater Bioaugmentation

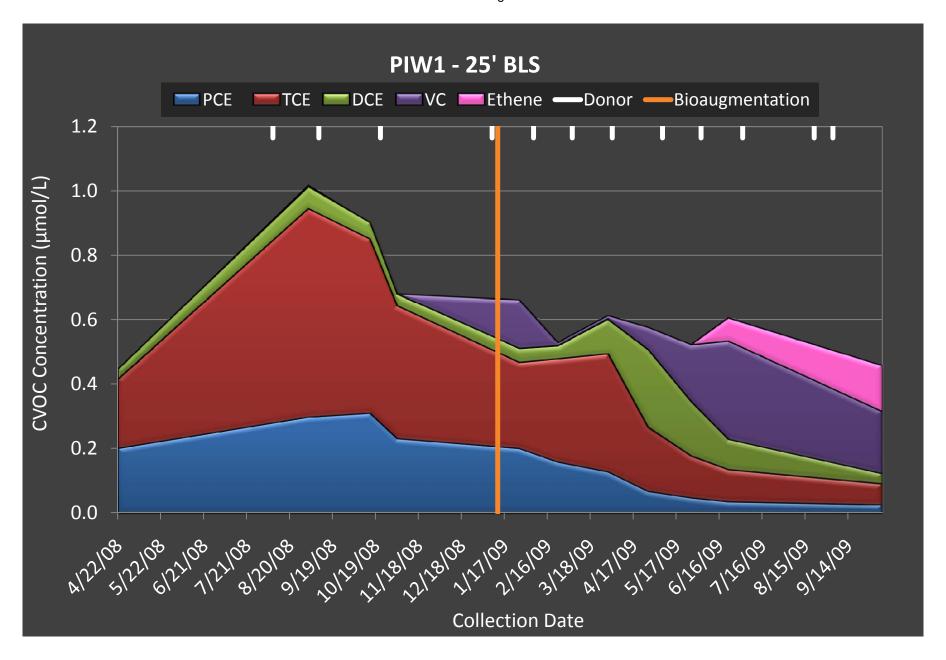


Seal Beach
Groundwater Bioaugmentation

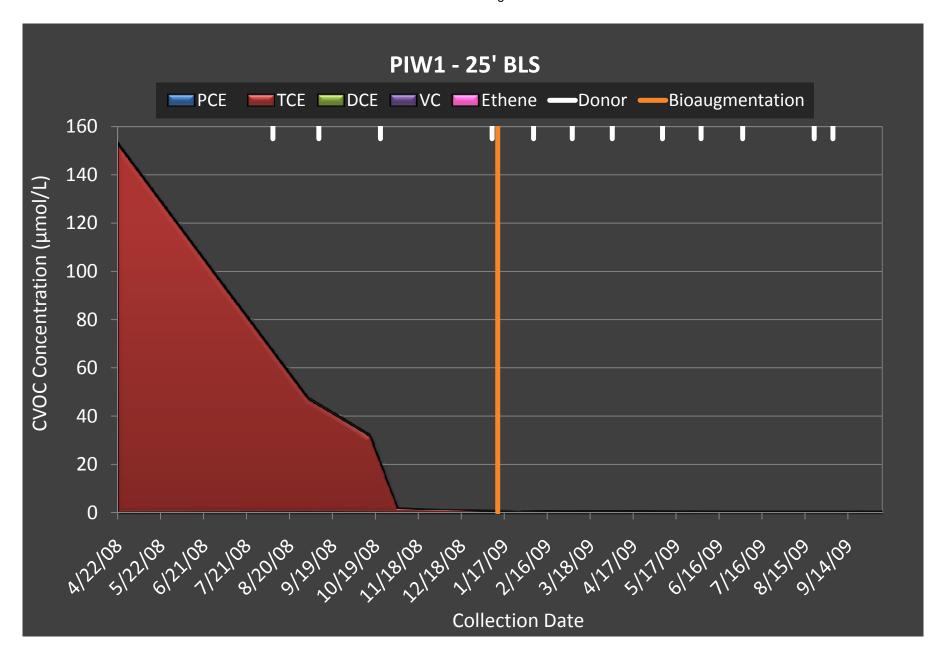


CVOCs Molar Concentrations

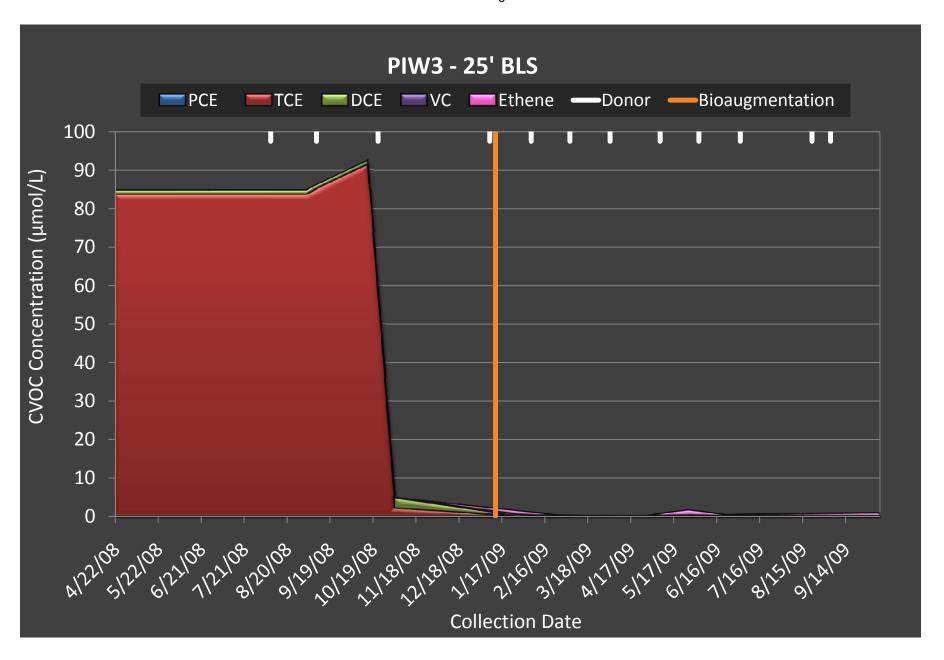
Seal Beach
Groundwater Bioaugmentation



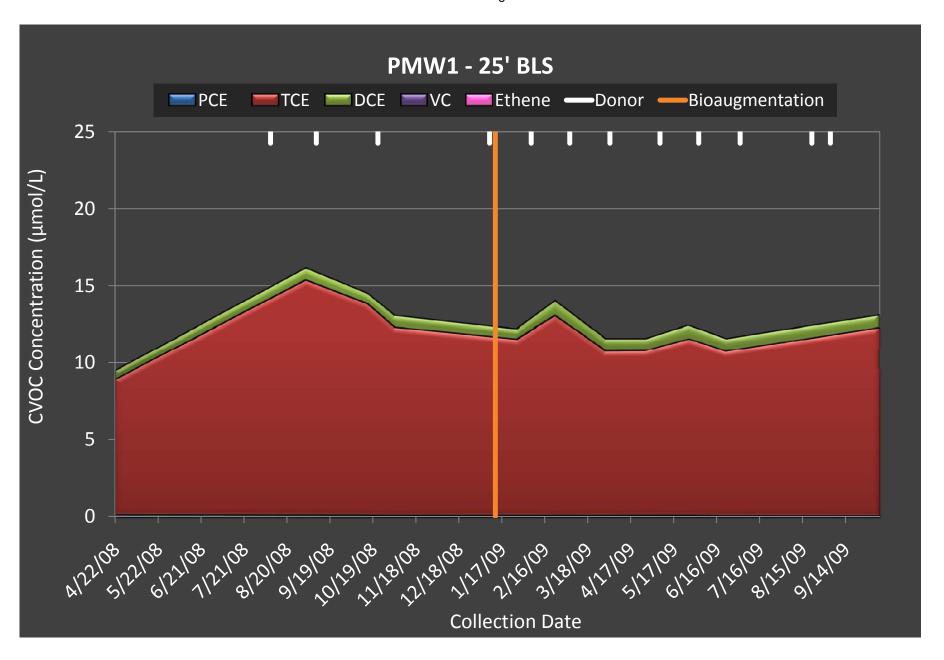
Seal Beach
Groundwater Bioaugmentation



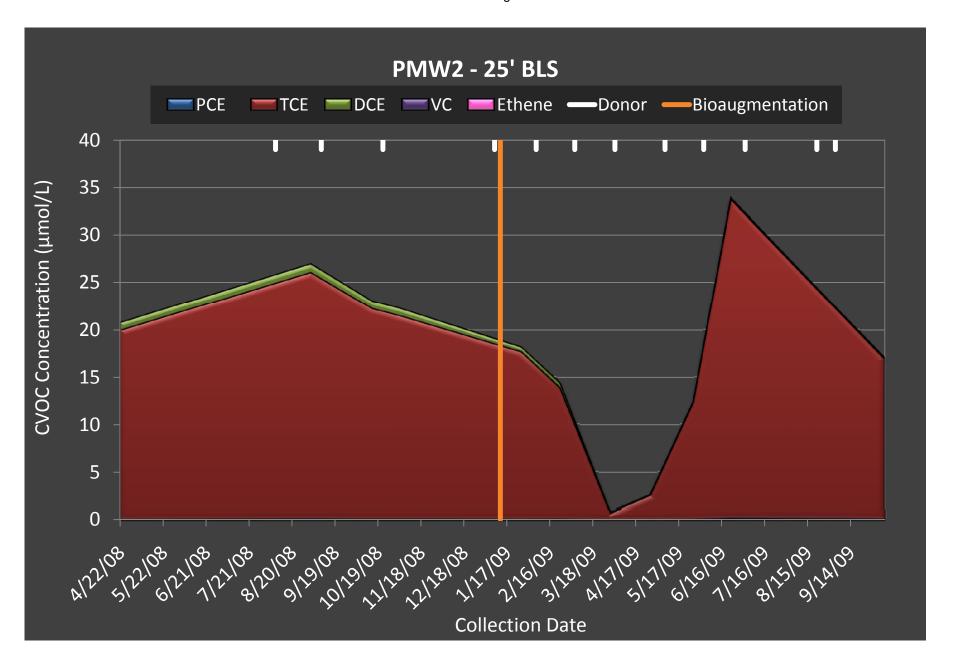
Seal Beach
Groundwater Bioaugmentation



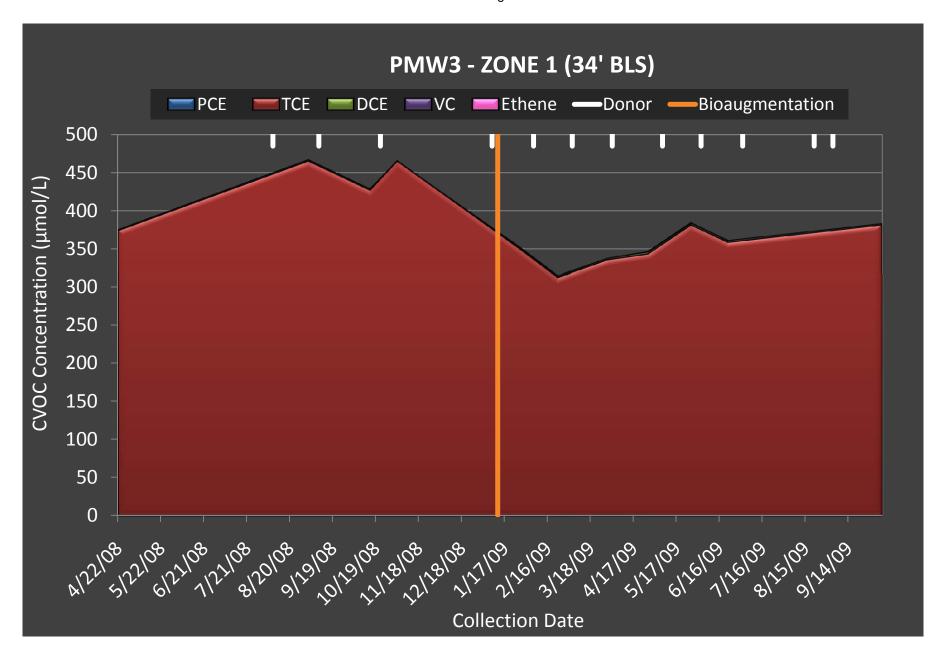
Seal Beach
Groundwater Bioaugmentation



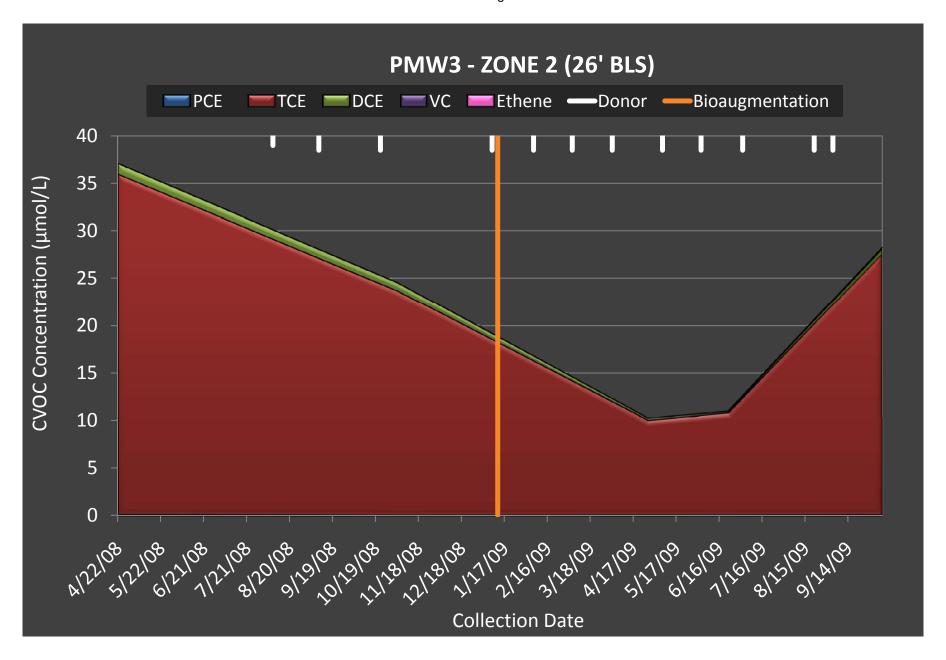
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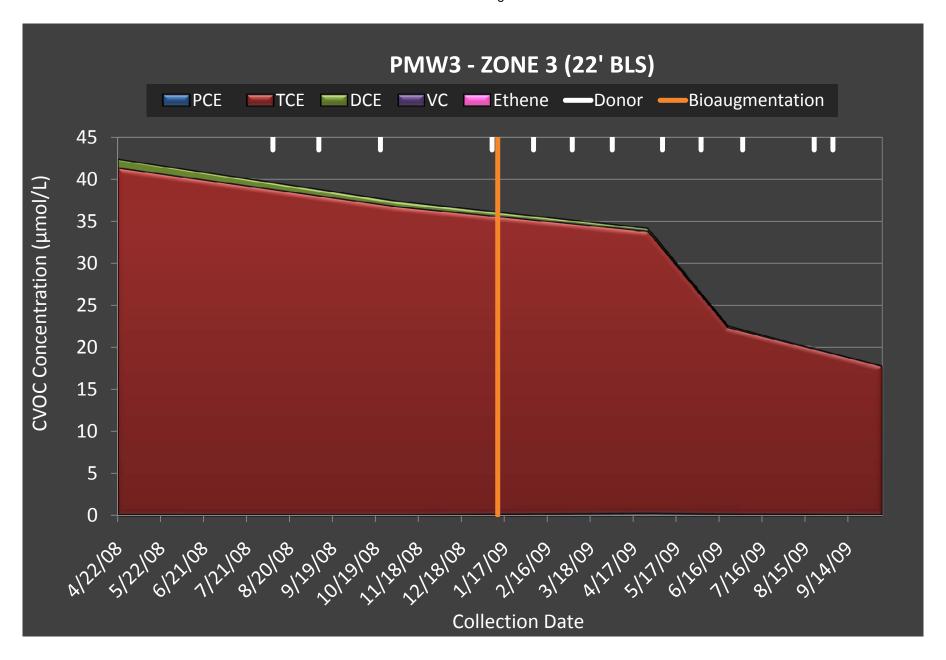
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Groundwater Bioaugmentation

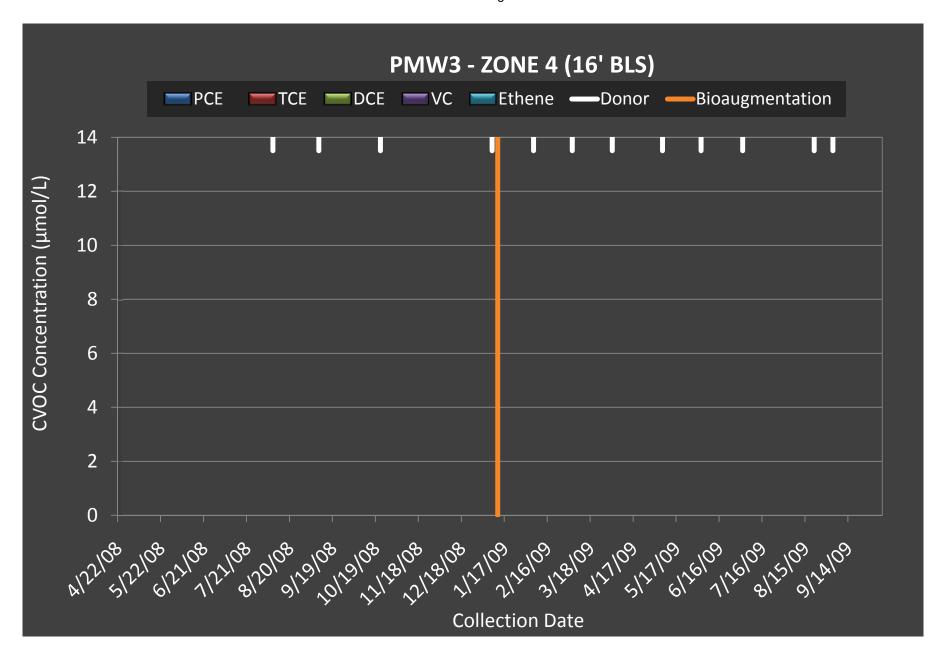


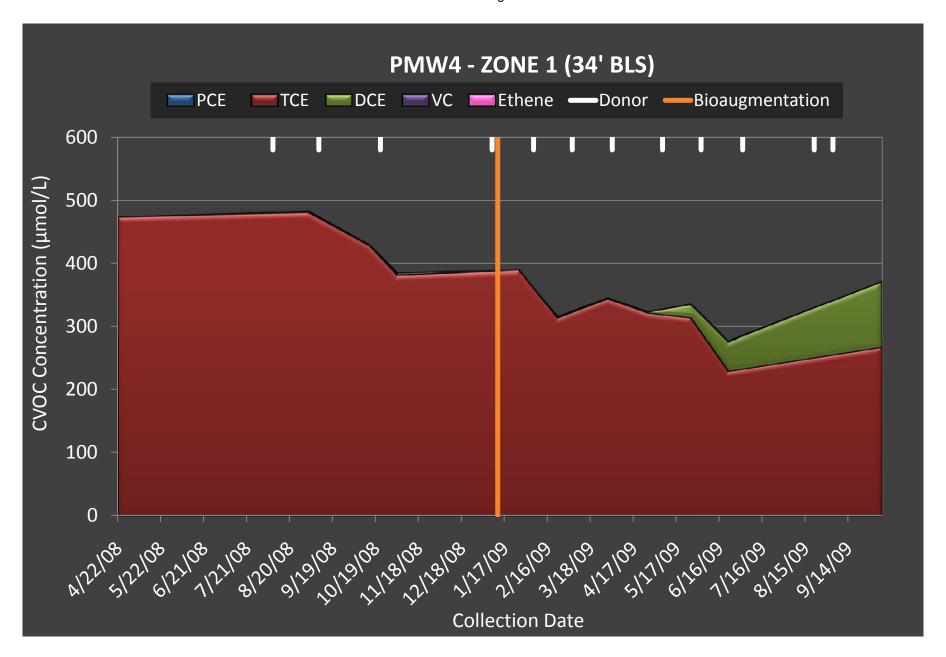
Seal Beach
Groundwater Bioaugmentation

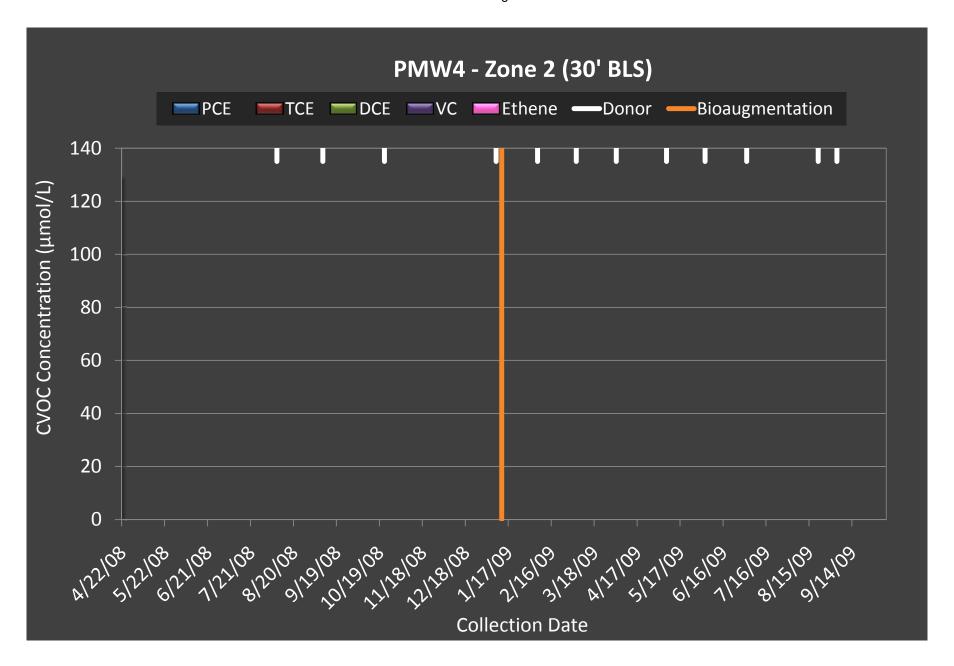


Seal Beach
Groundwater Bioaugmentation

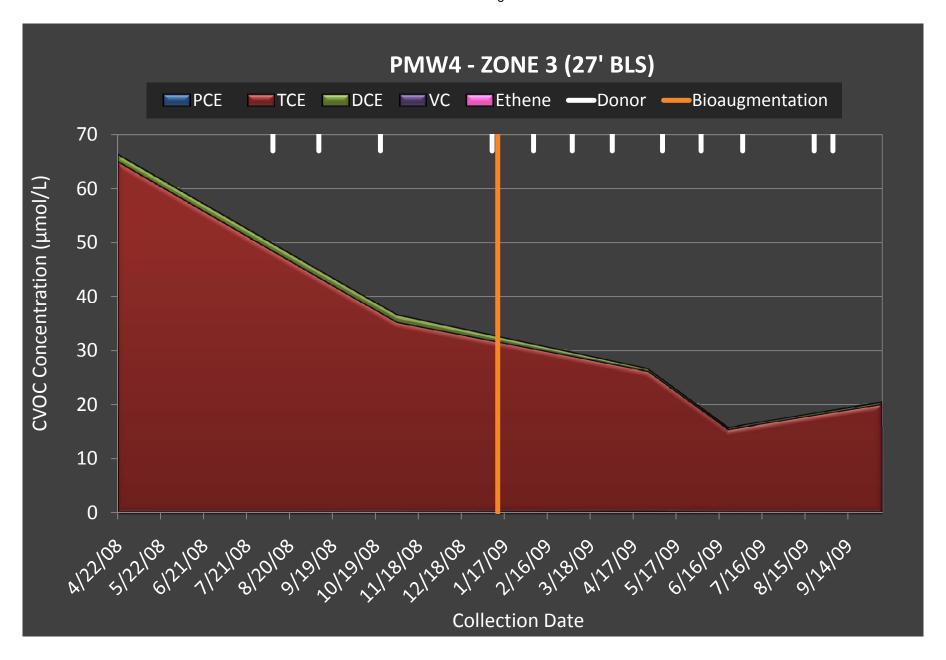




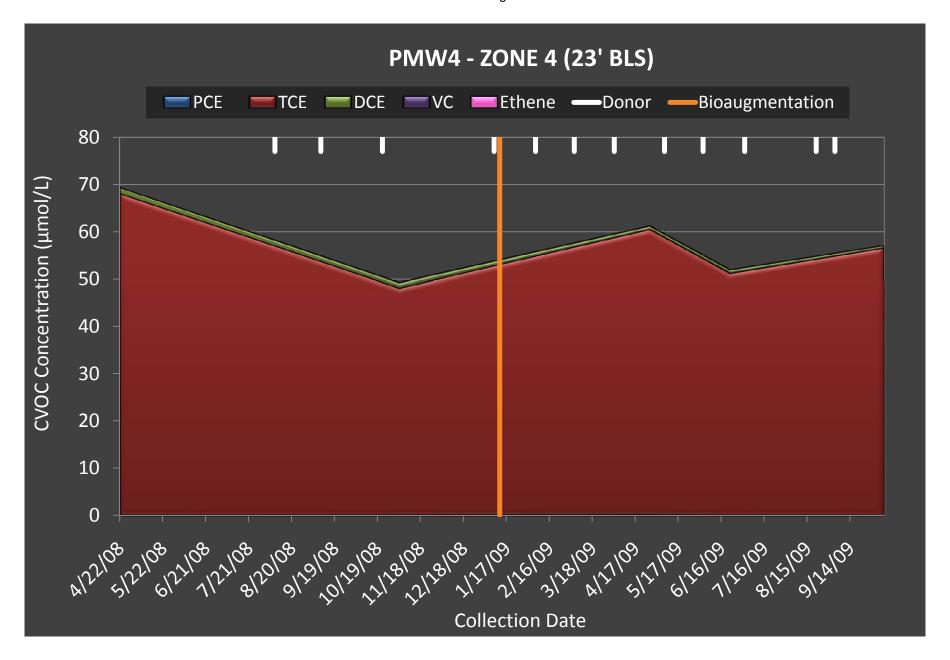


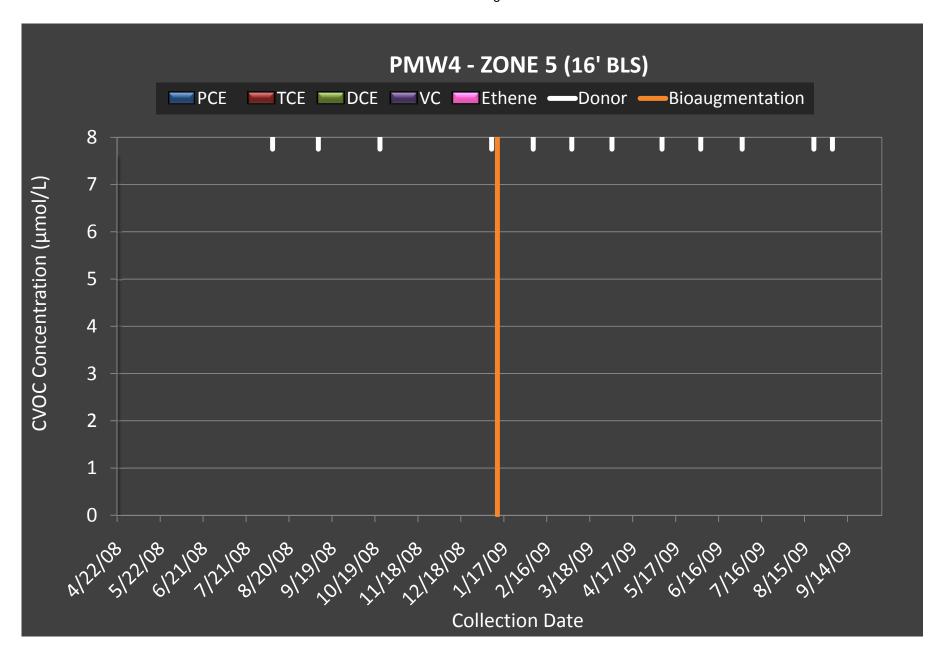


Seal Beach
Groundwater Bioaugmentation

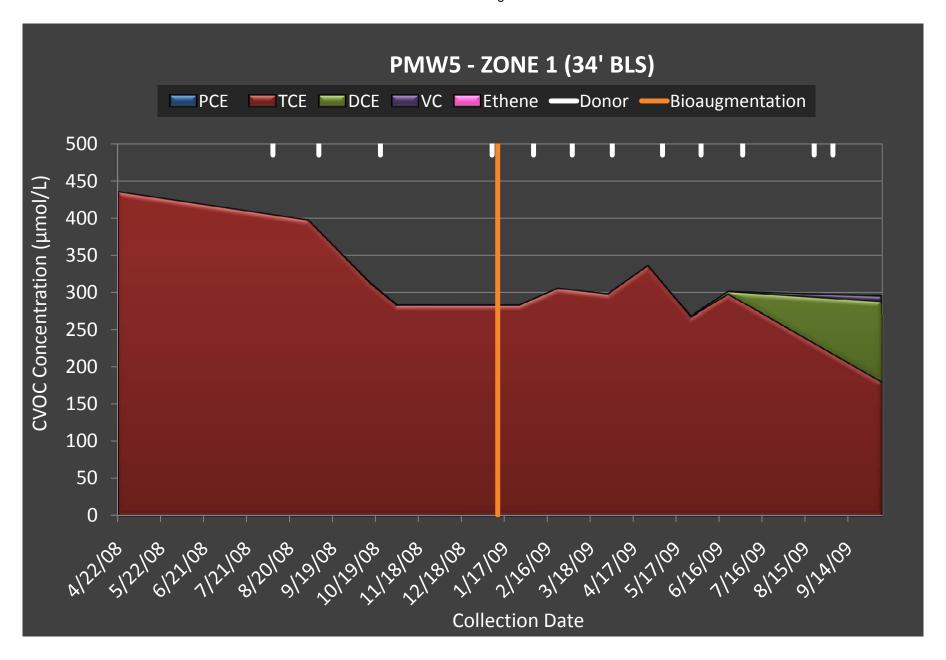


Seal Beach
Groundwater Bioaugmentation

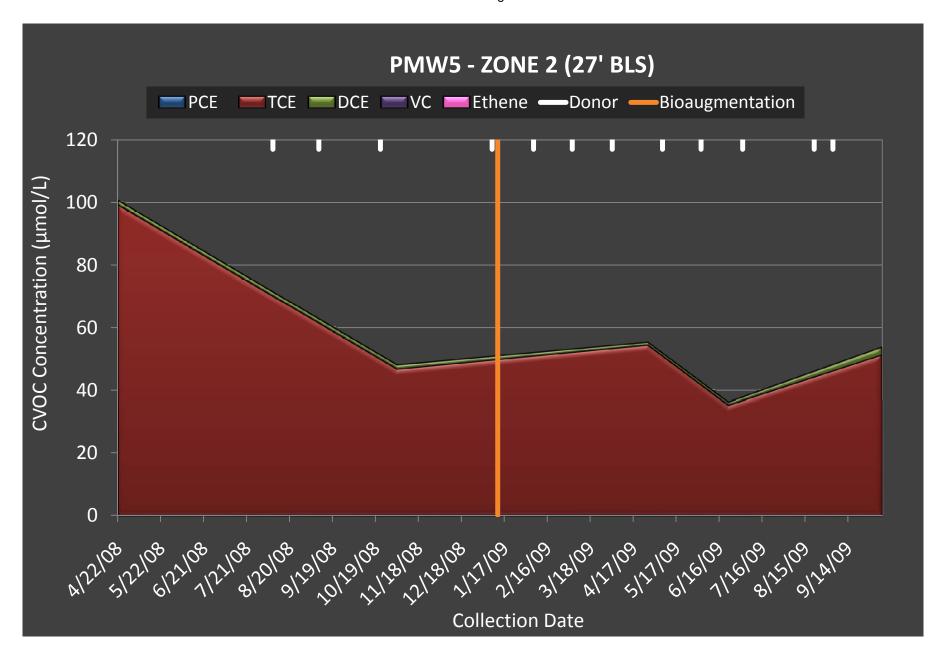




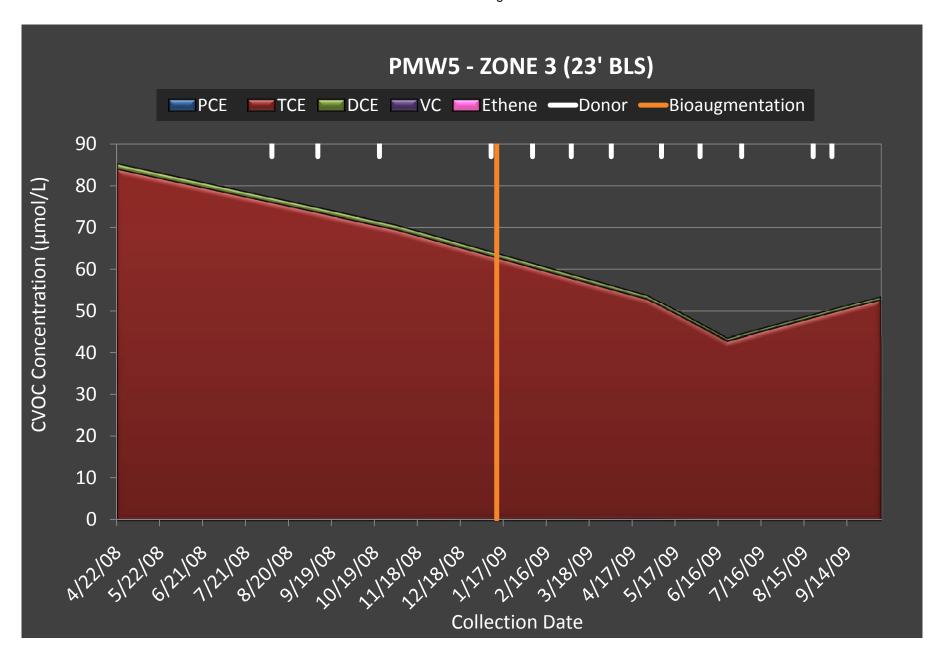
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Groundwater Bioaugmentation



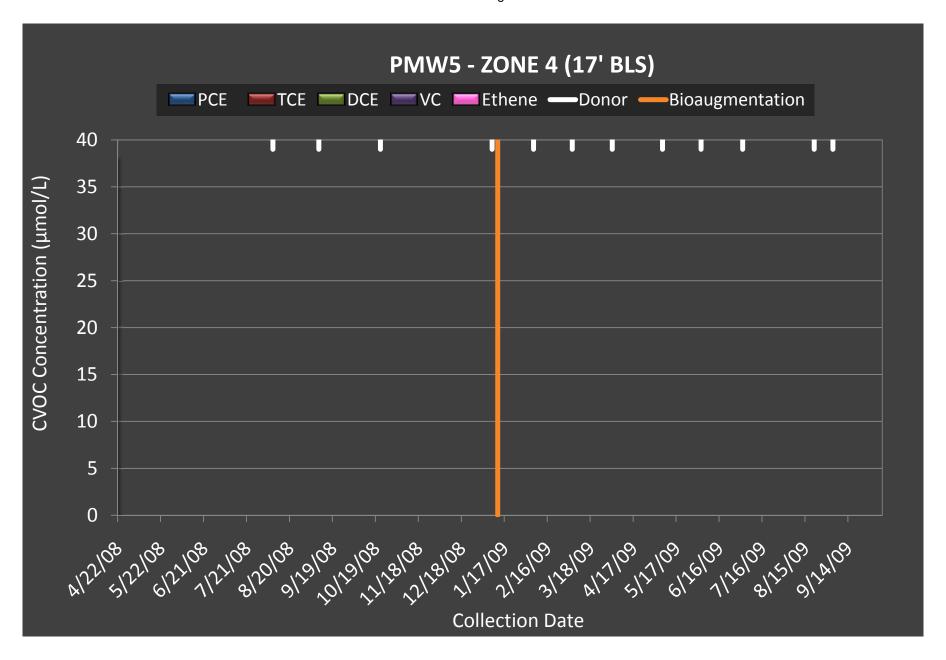
Seal Beach
Groundwater Bioaugmentation



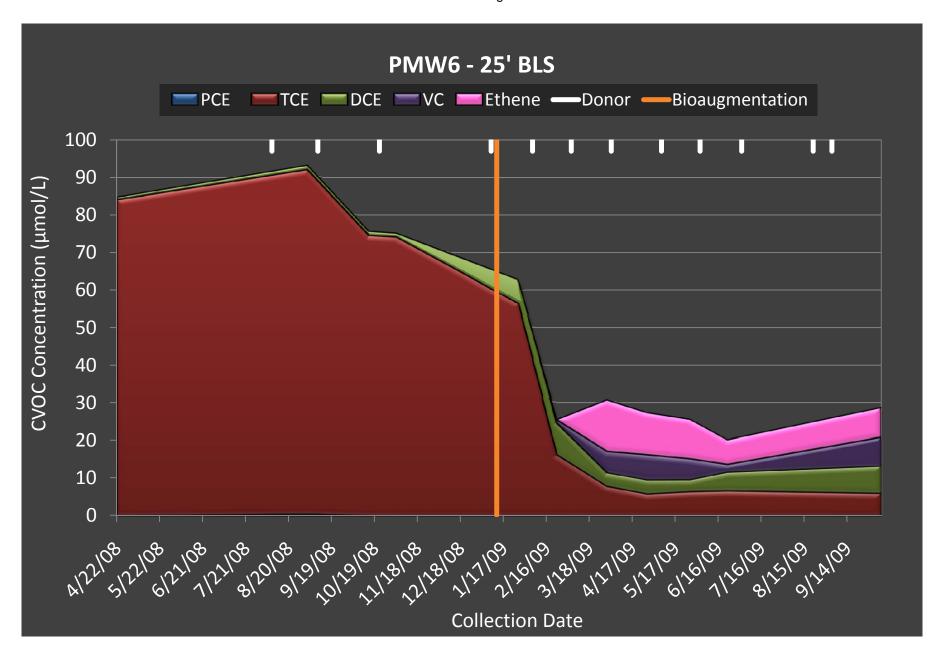
Seal Beach
Groundwater Bioaugmentation



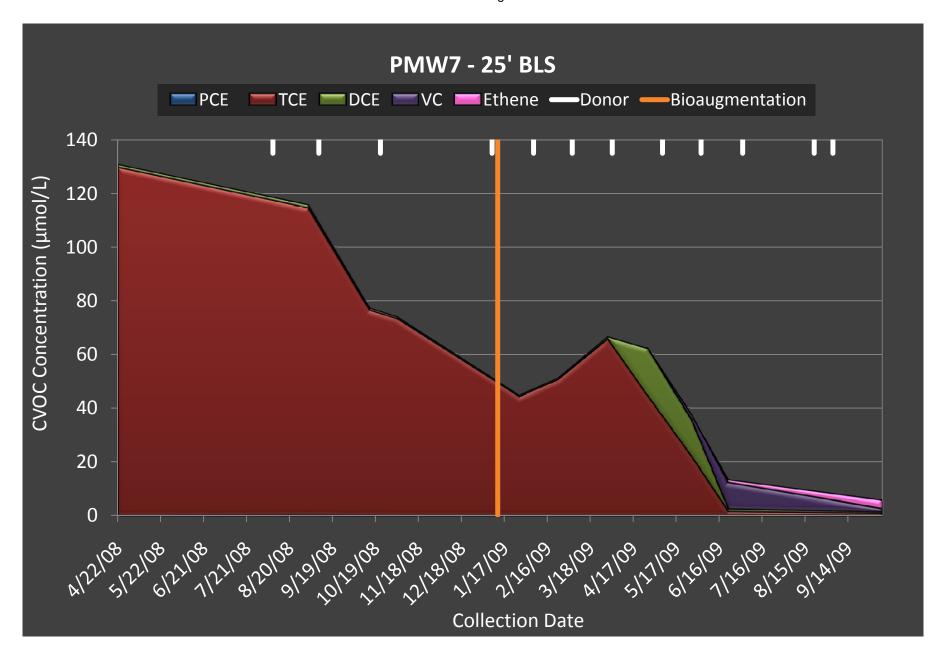
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Groundwater Bioaugmentation



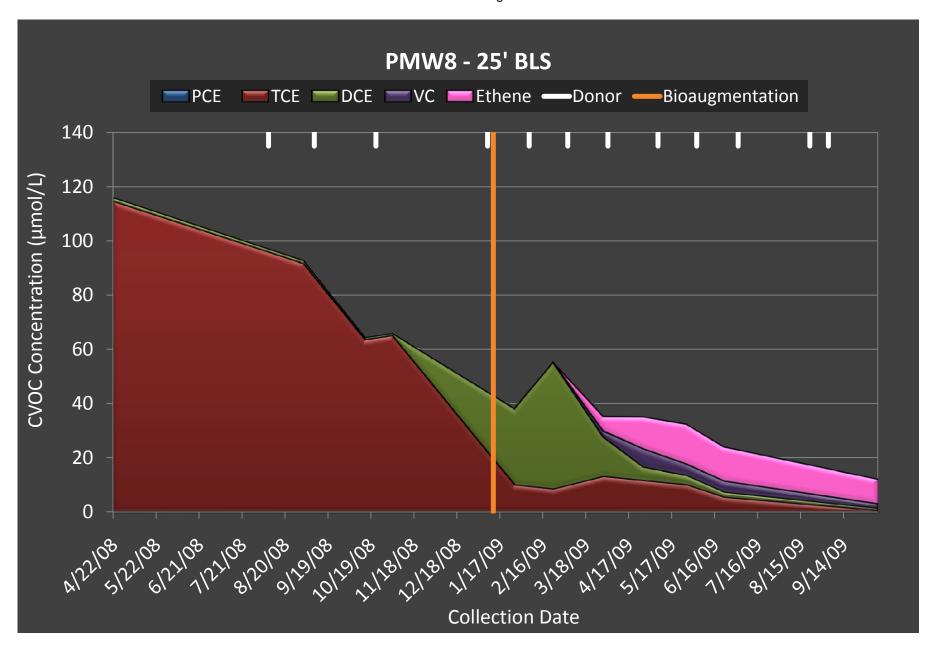
Seal Beach
Groundwater Bioaugmentation



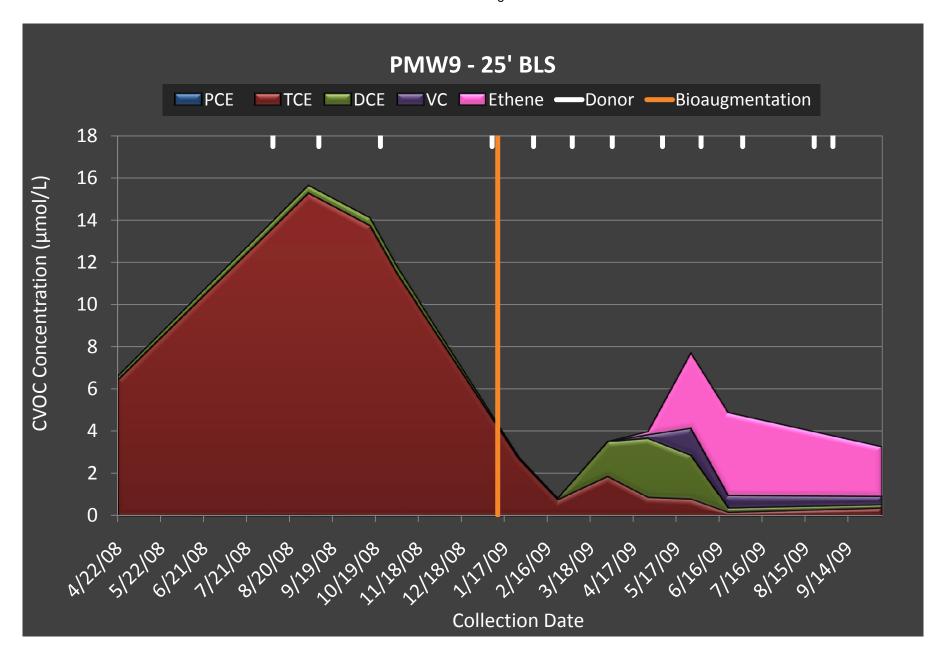
Seal Beach
Groundwater Bioaugmentation



Seal Beach
Groundwater Bioaugmentation

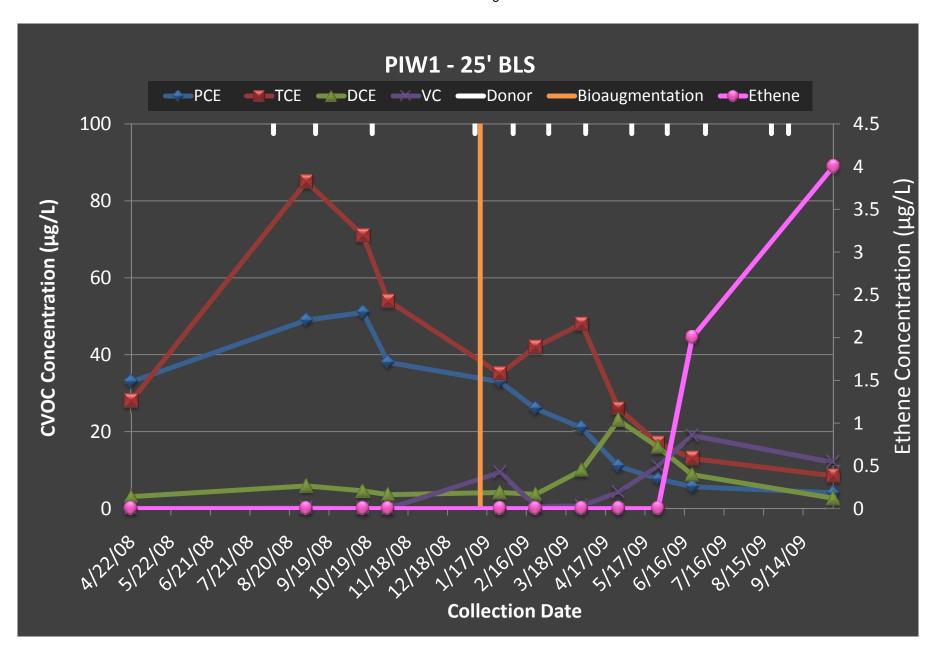


Seal Beach
Groundwater Bioaugmentation

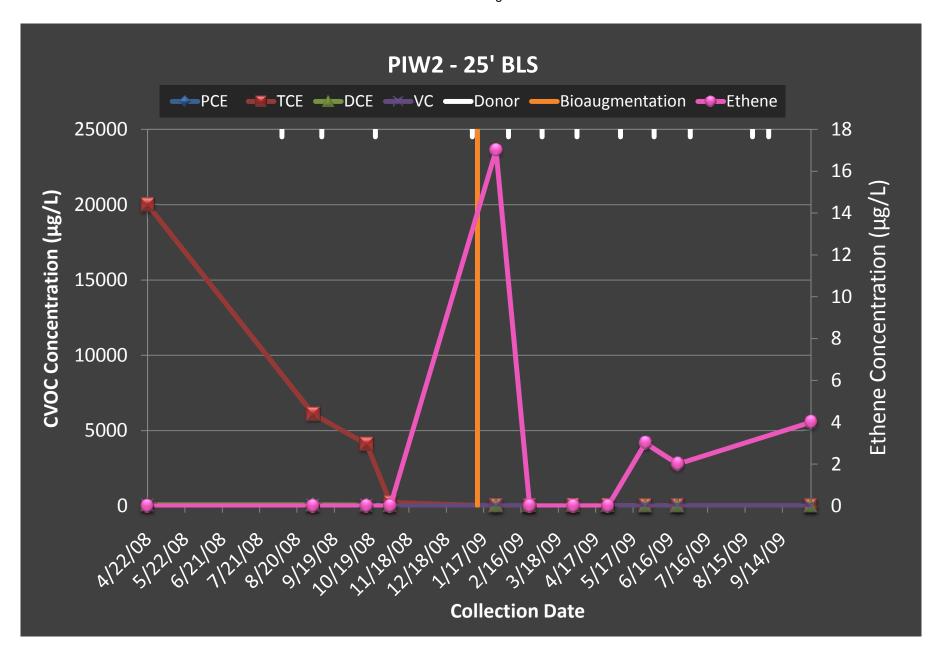


CVOCs Mass Concentrations

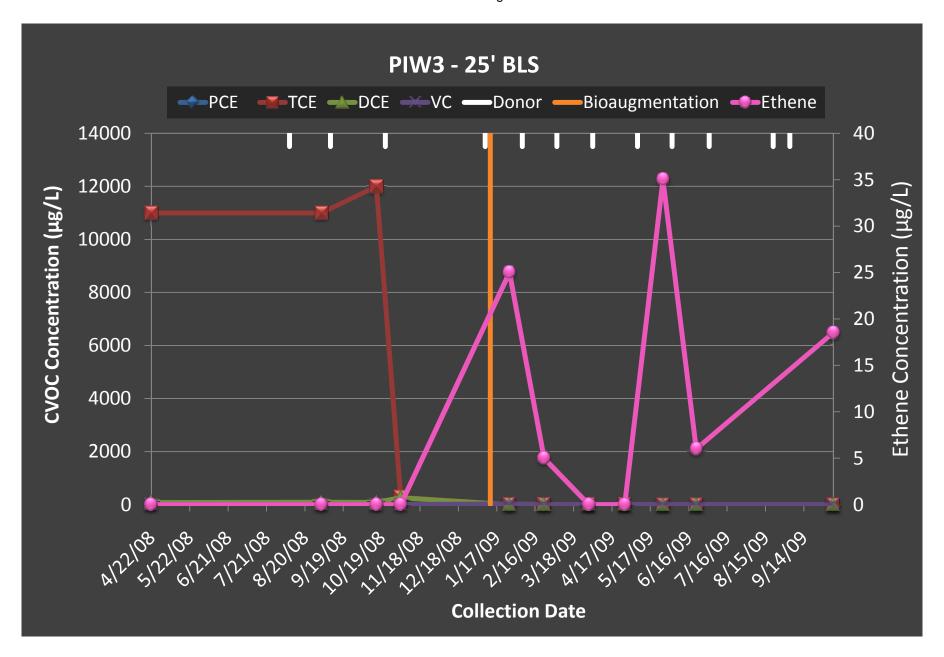
Seal Beach
Groundwater Bioaugmentation



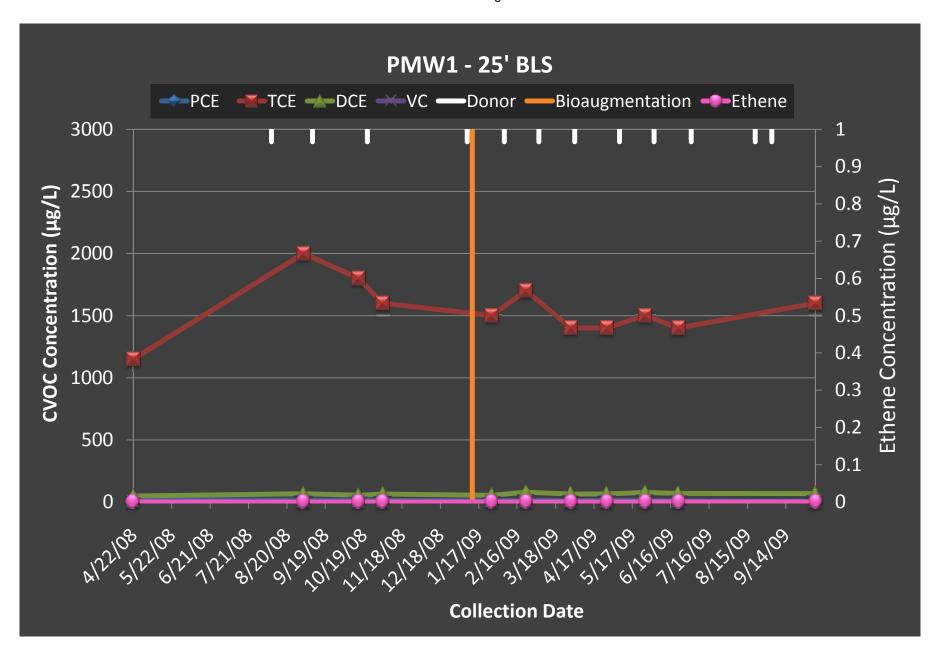
Seal Beach
Groundwater Bioaugmentation

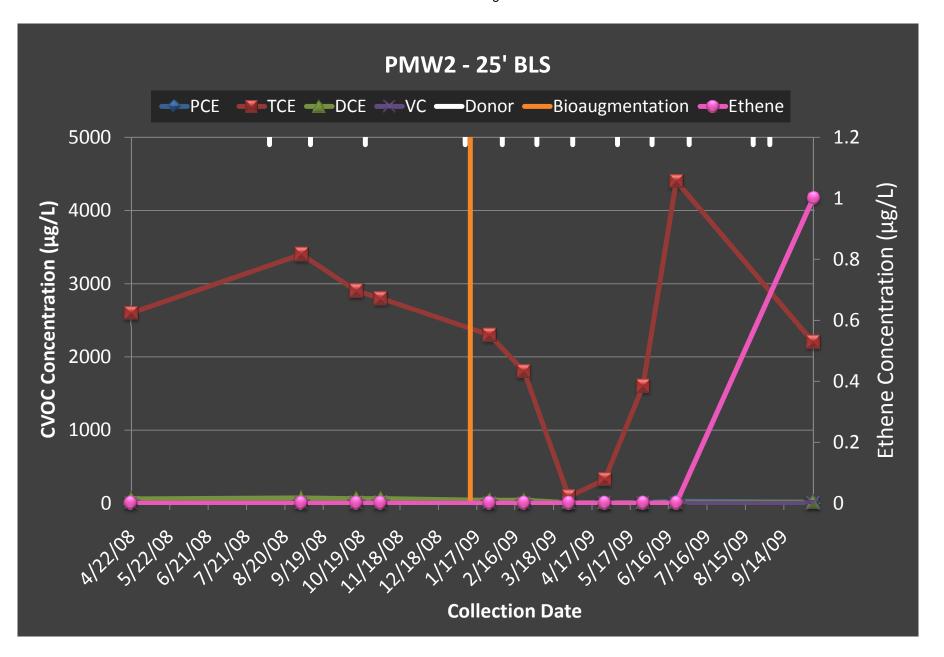


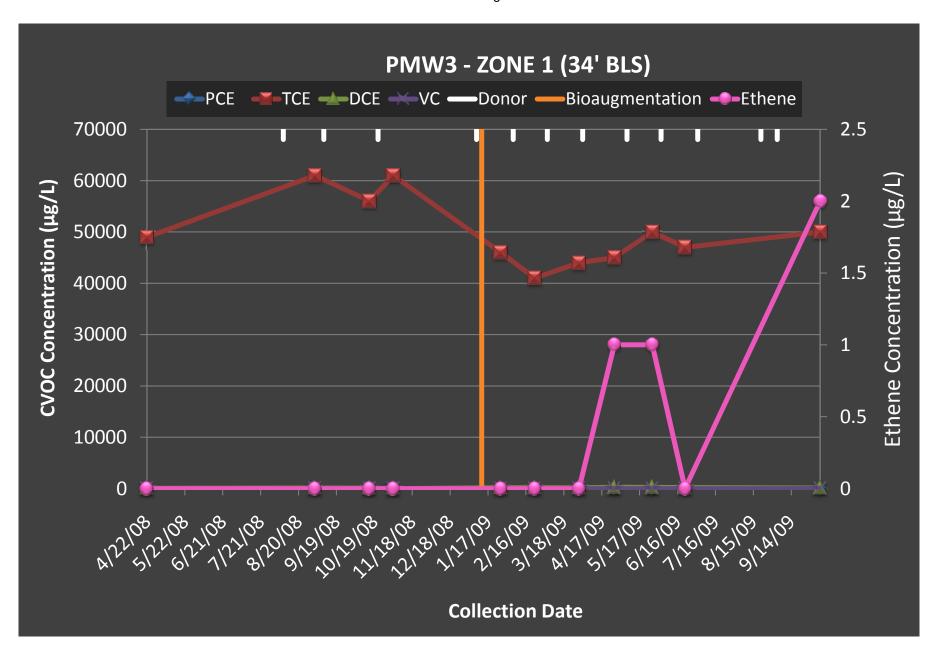
Seal Beach
Groundwater Bioaugmentation



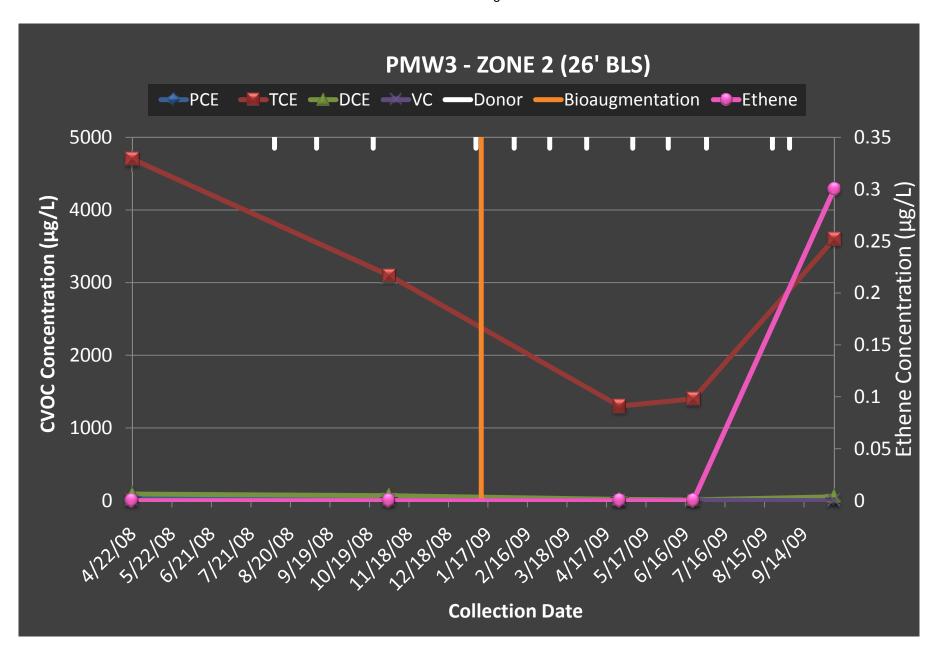
Seal Beach
Groundwater Bioaugmentation



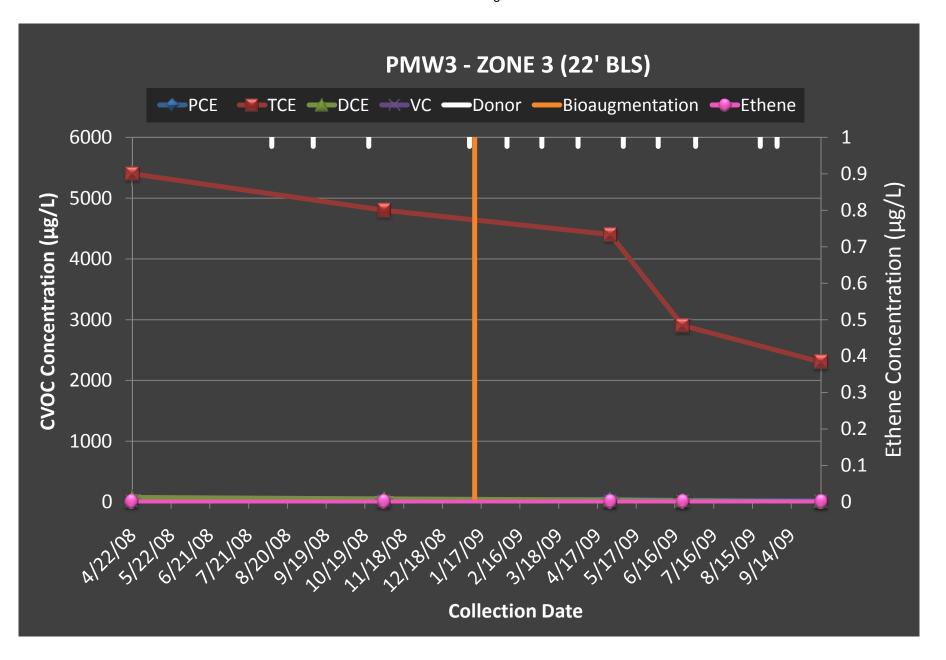




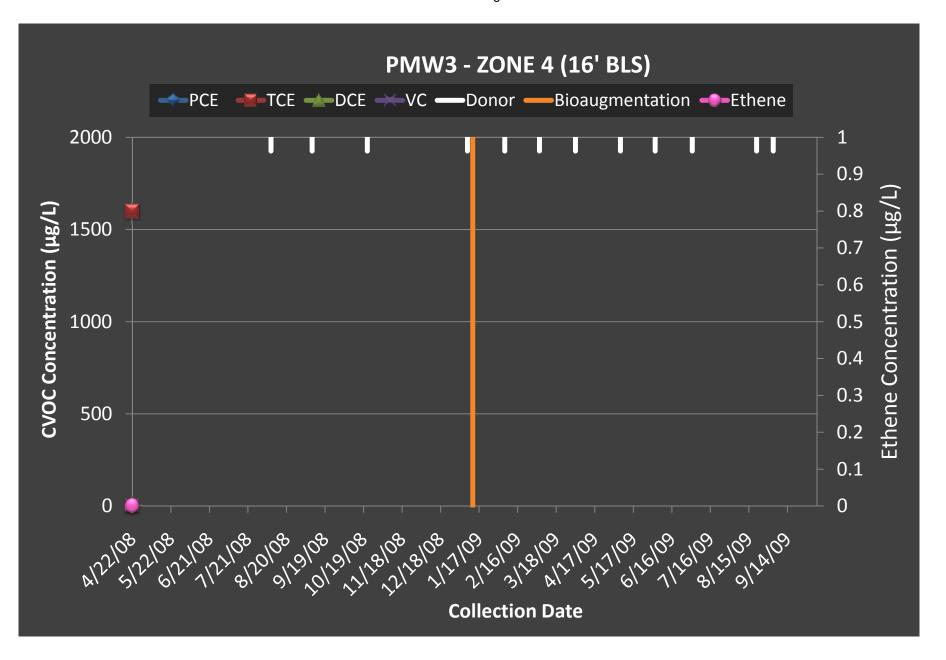
Seal Beach
Groundwater Bioaugmentation

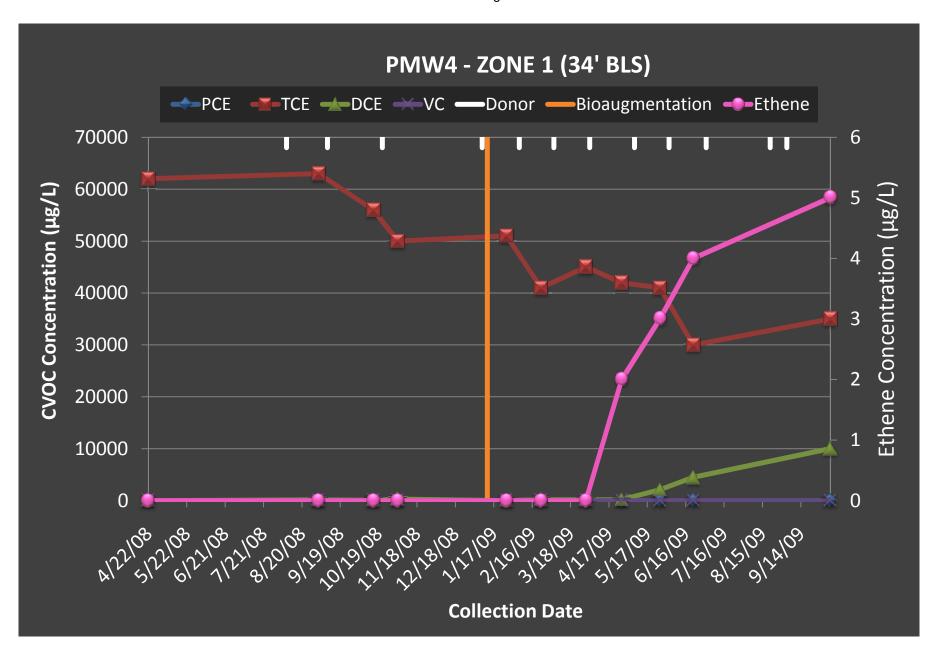


Seal Beach
Groundwater Bioaugmentation

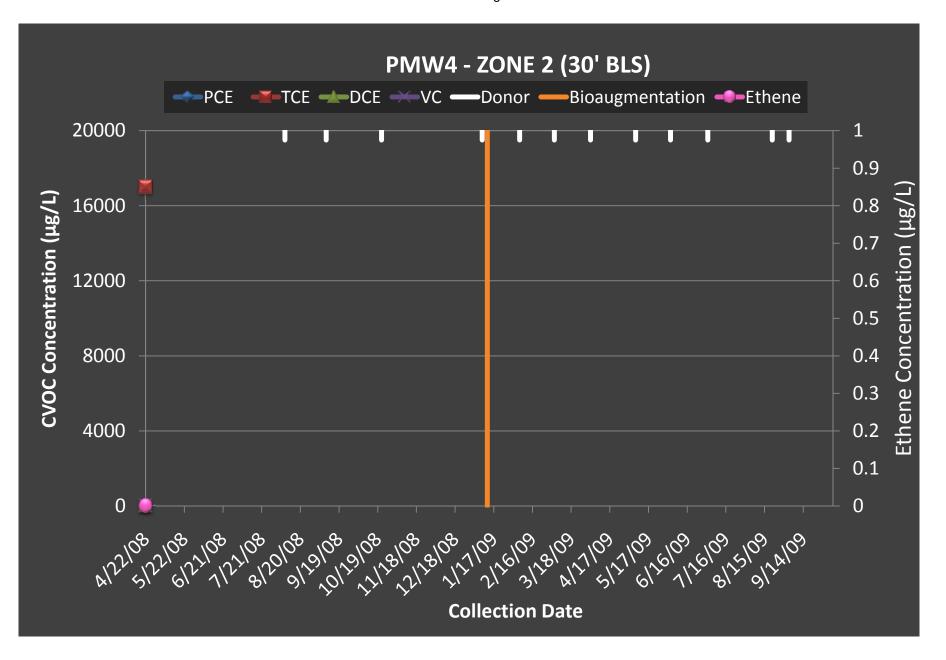


Seal Beach
Groundwater Bioaugmentation

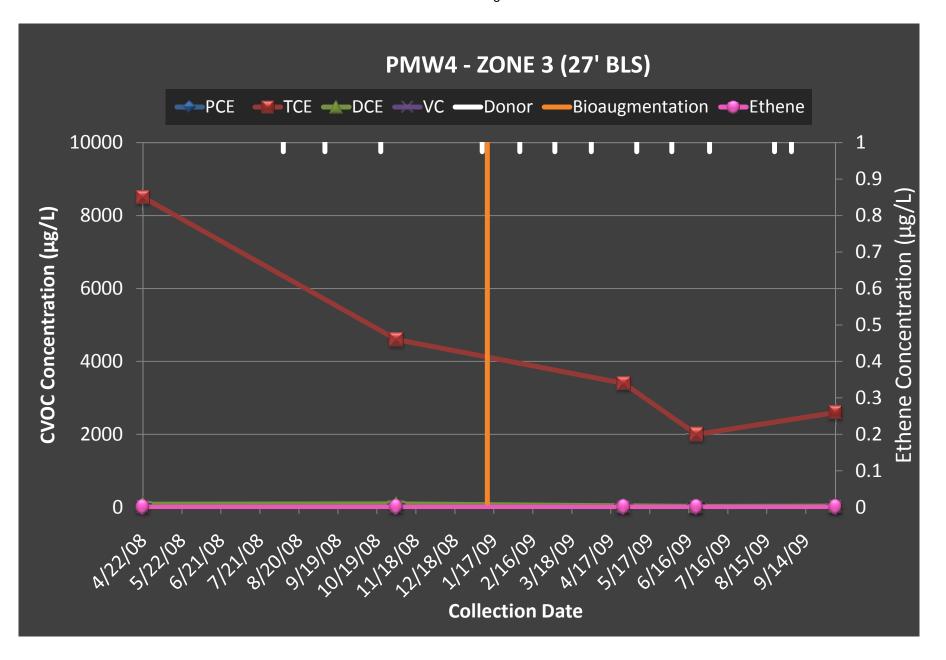


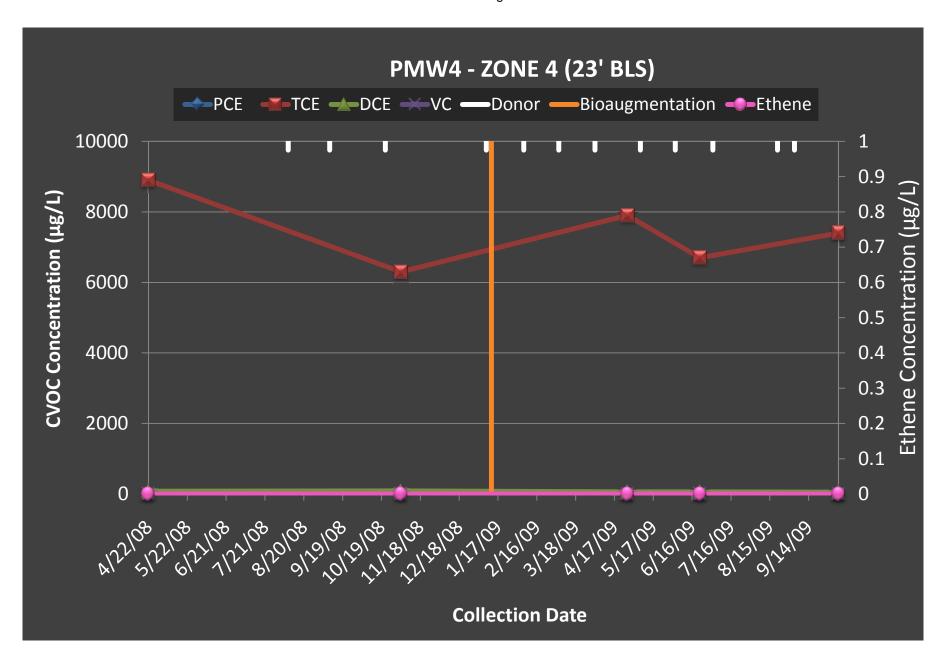


Seal Beach
Groundwater Bioaugmentation

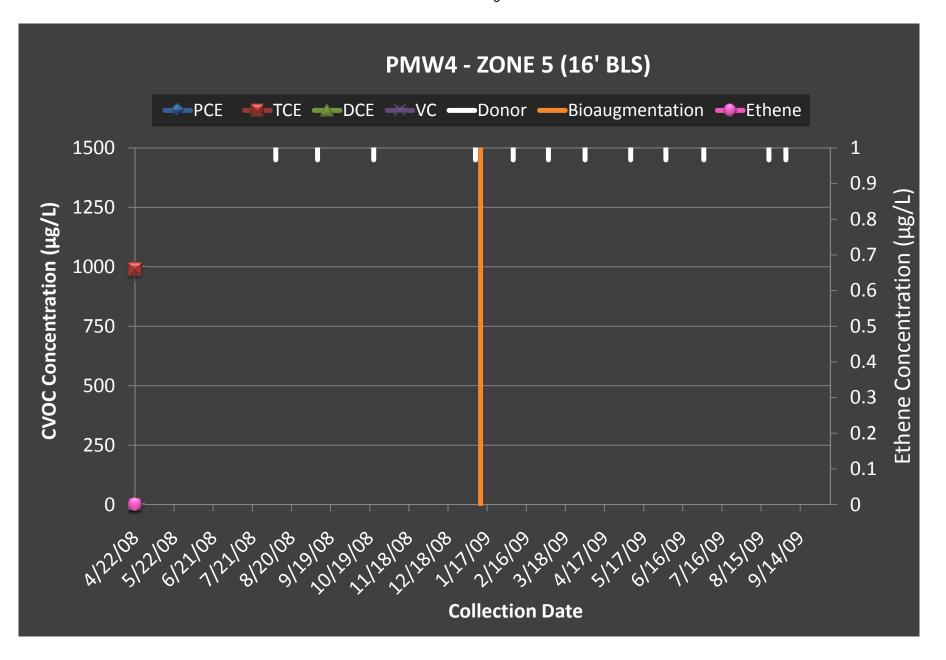


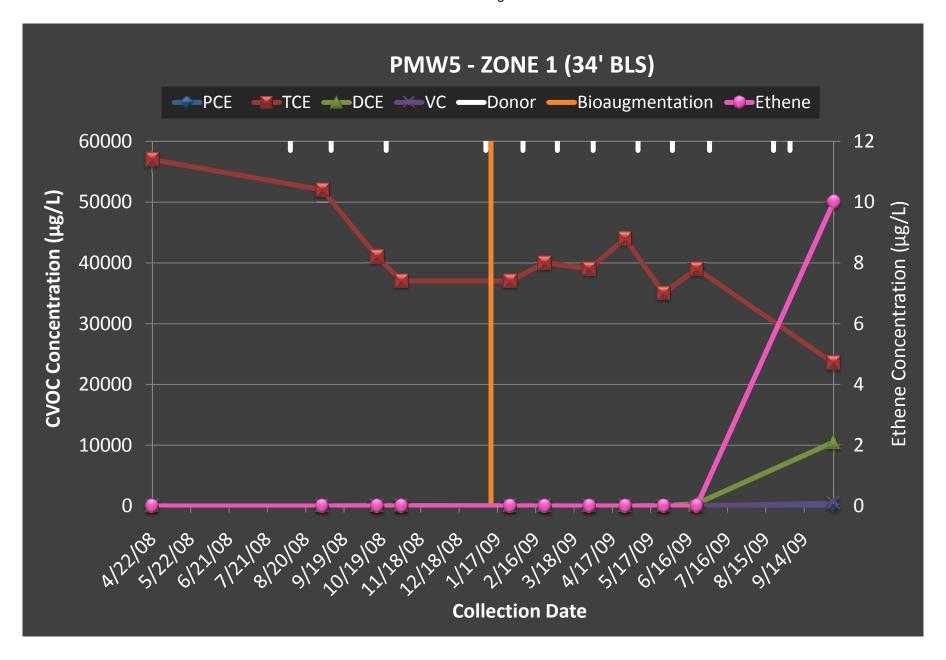
Seal Beach
Groundwater Bioaugmentation

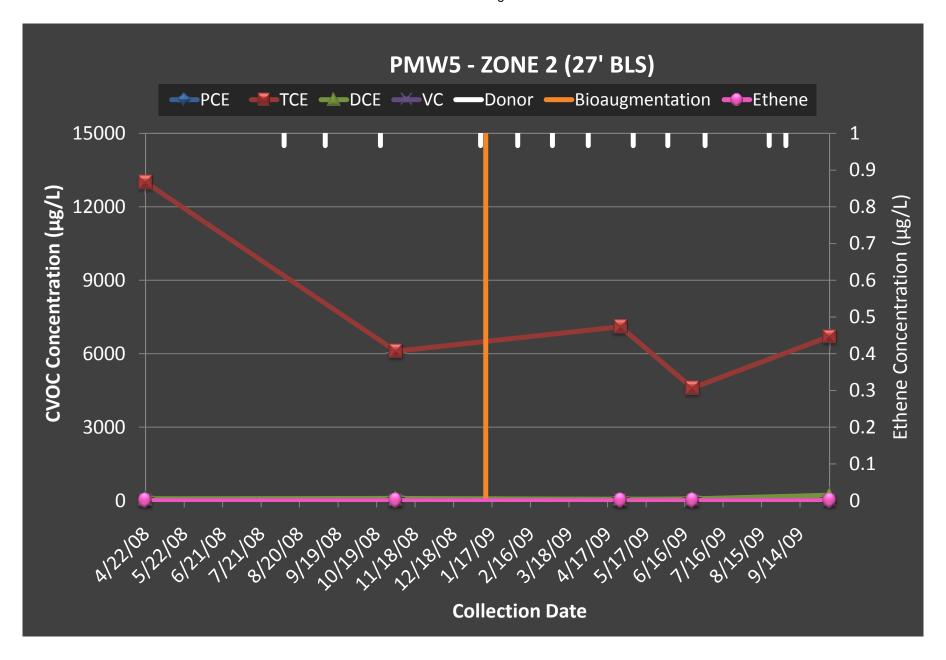




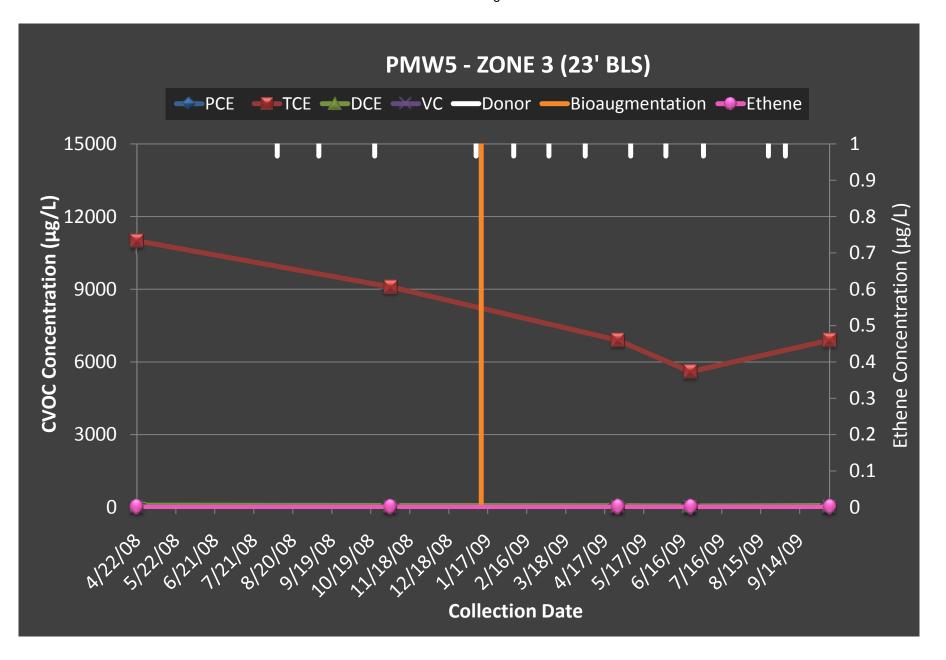
Seal Beach
Groundwater Bioaugmentation

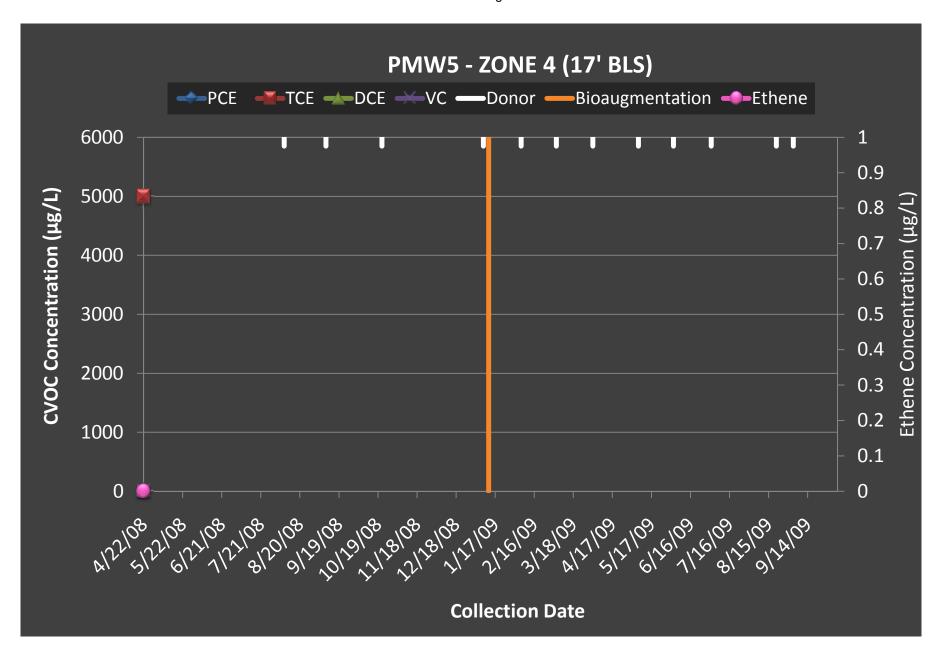




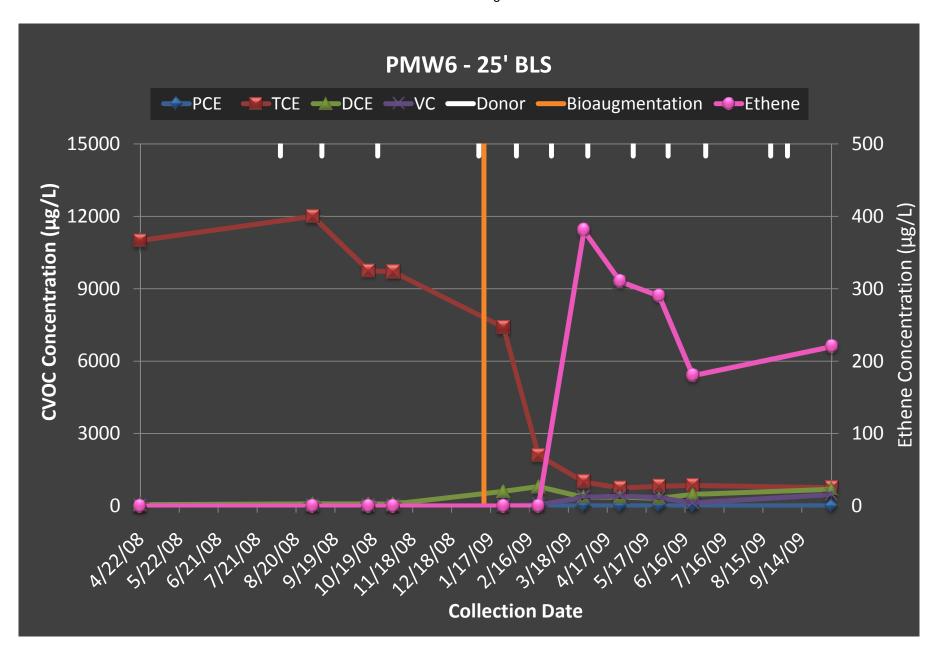


Seal Beach
Groundwater Bioaugmentation

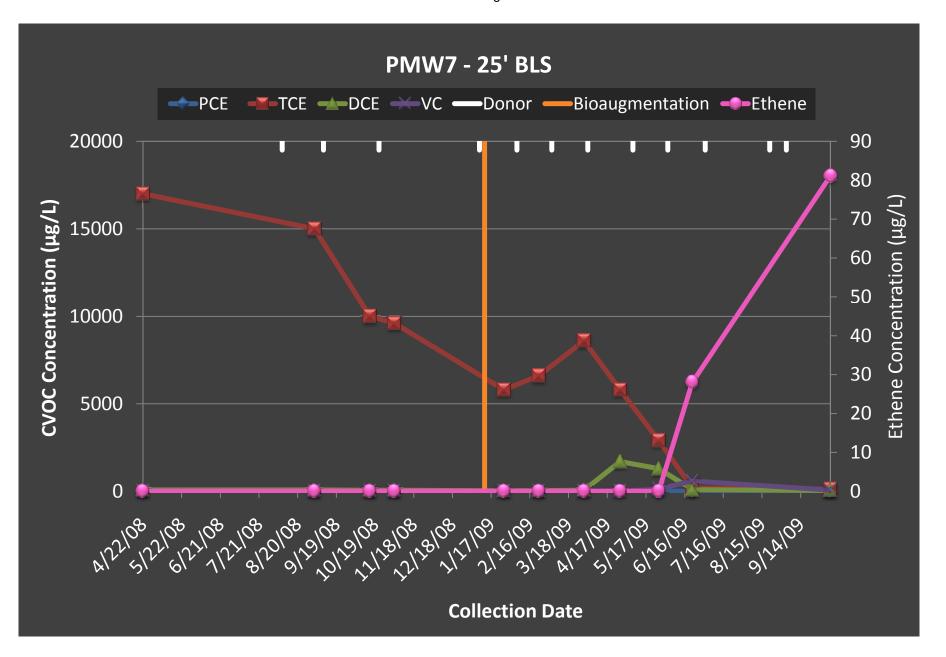




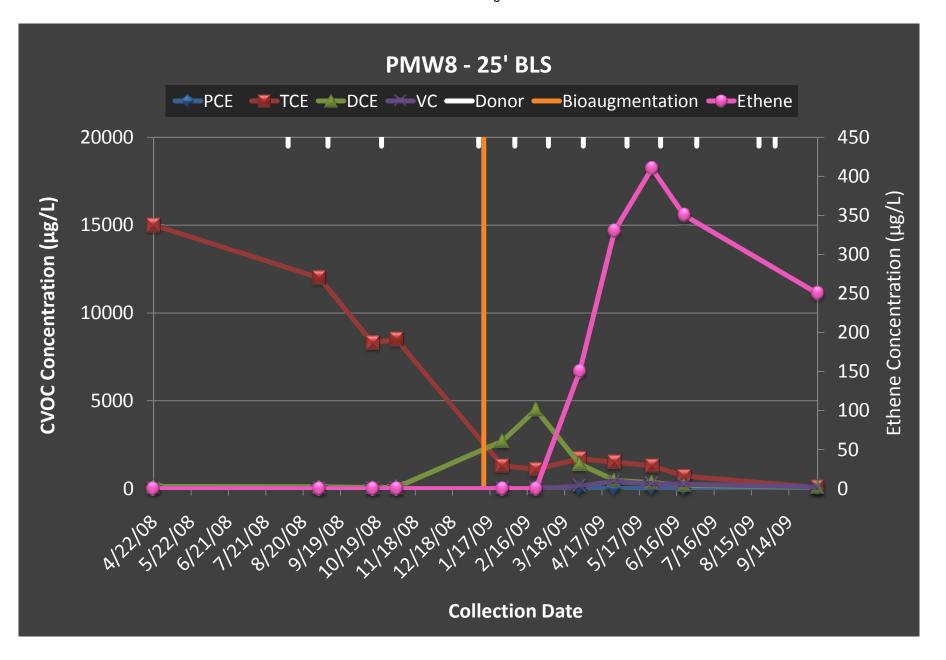
Seal Beach
Groundwater Bioaugmentation



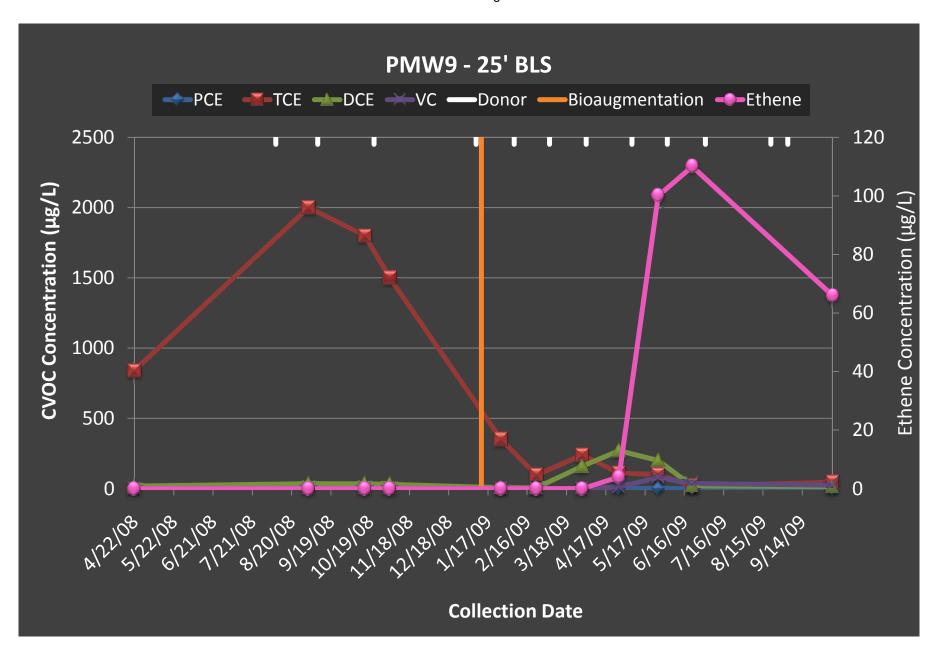
Seal Beach
Groundwater Bioaugmentation



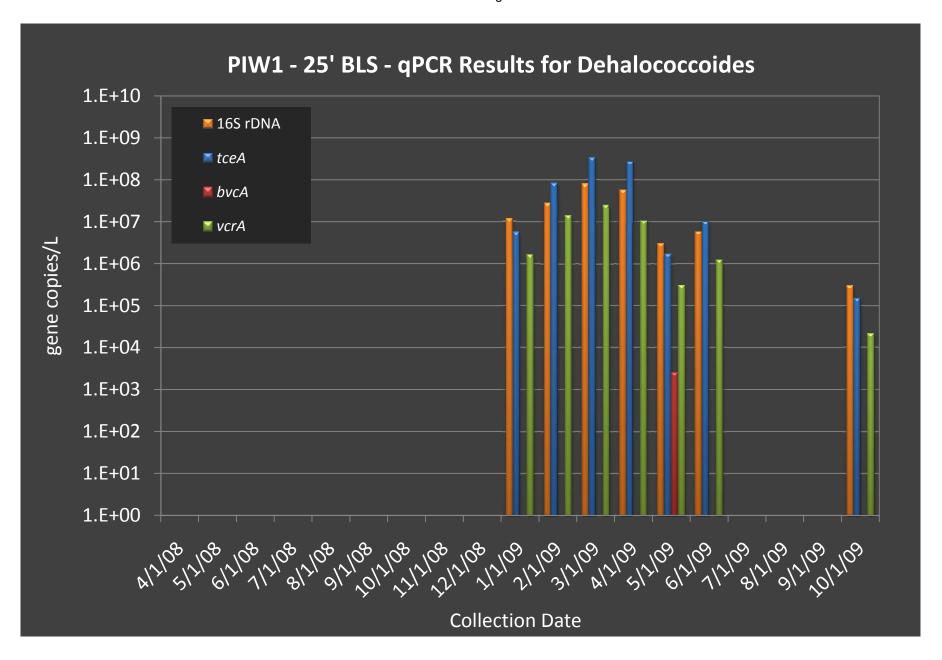
Seal Beach
Groundwater Bioaugmentation

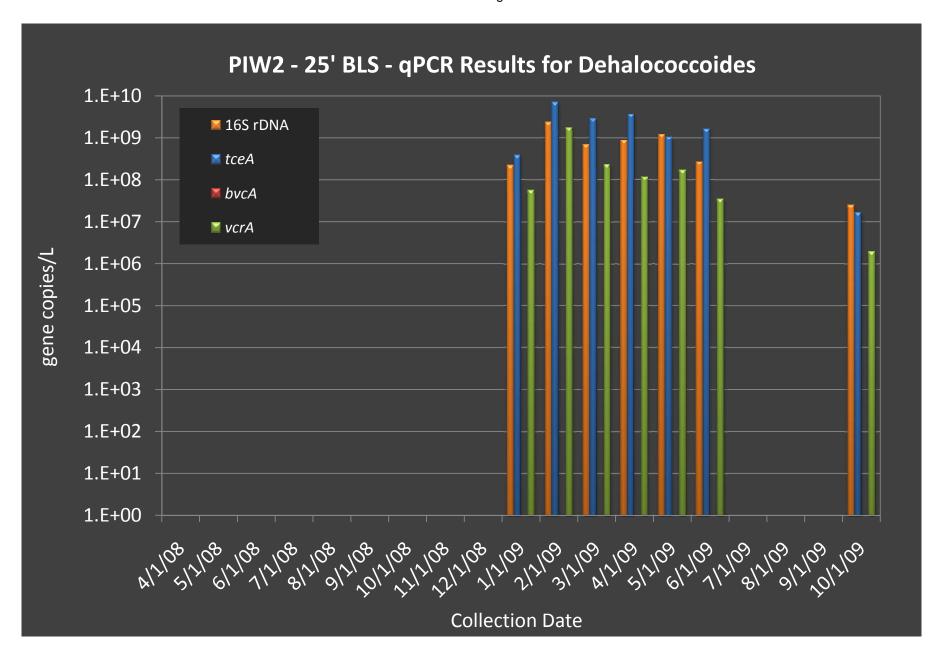


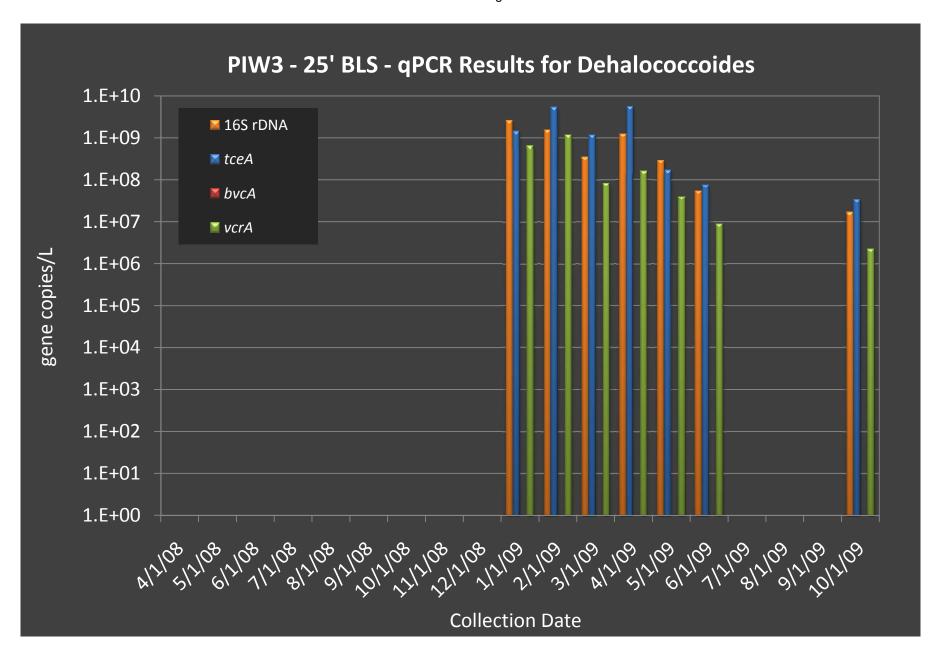
Seal Beach
Groundwater Bioaugmentation

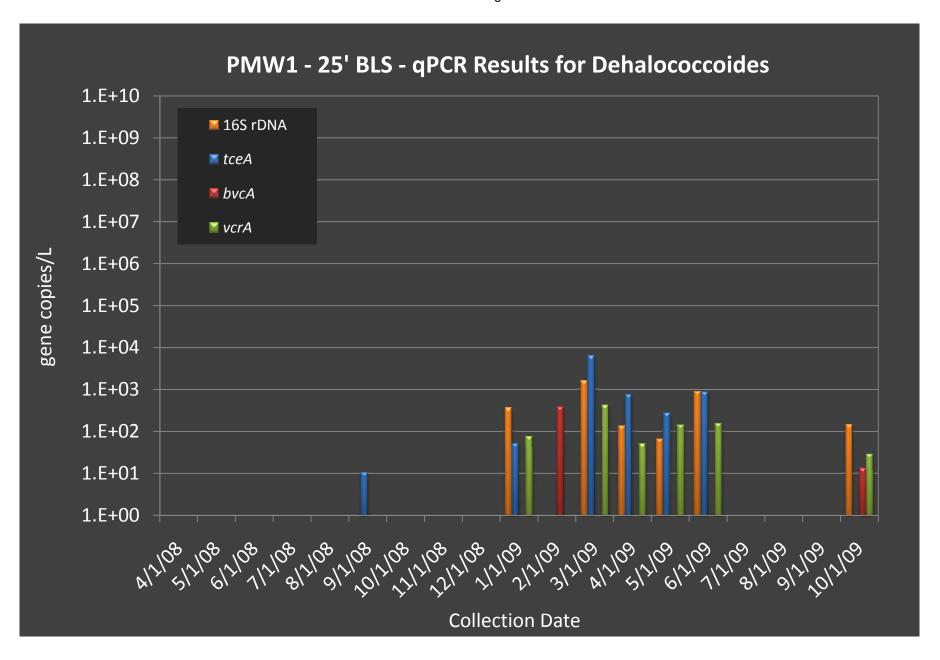


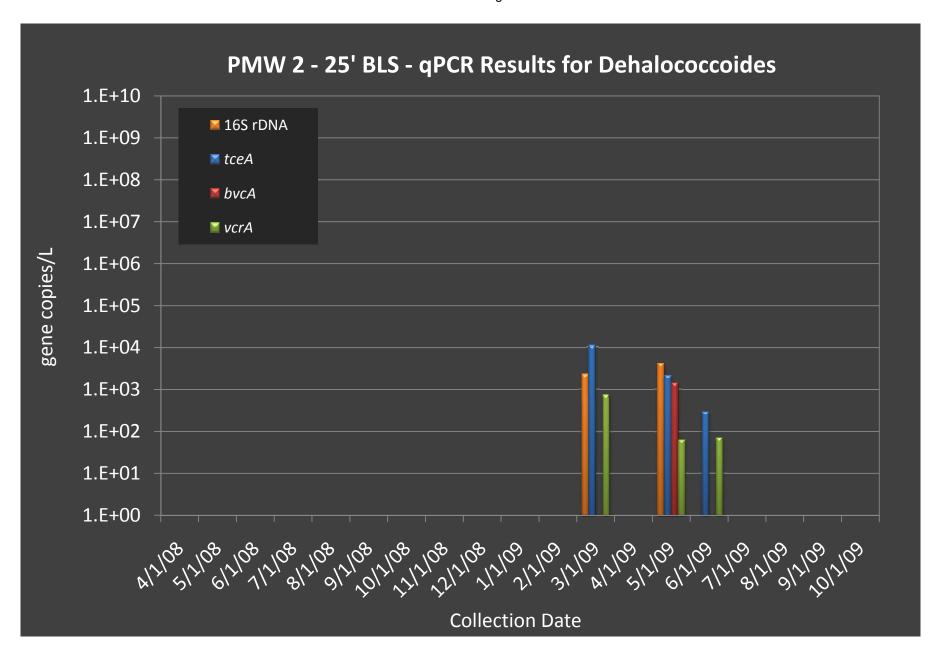
Dechlorinating Bacteria

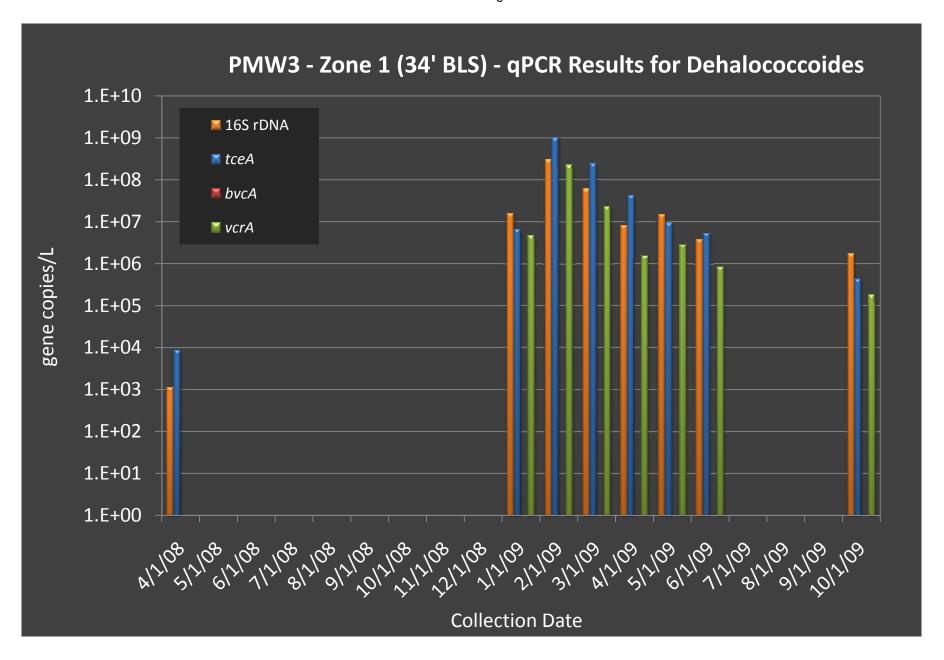


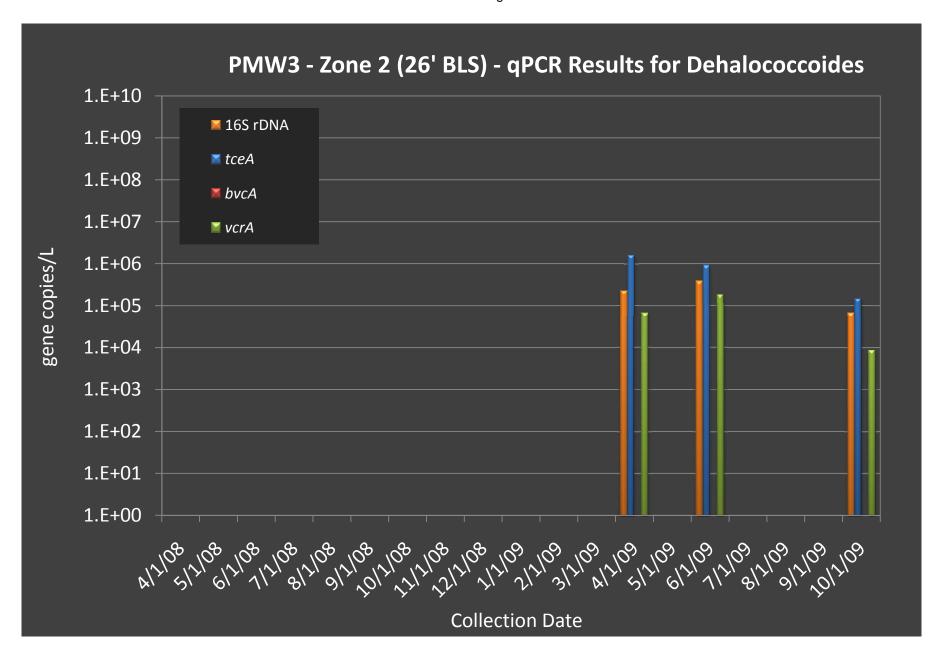


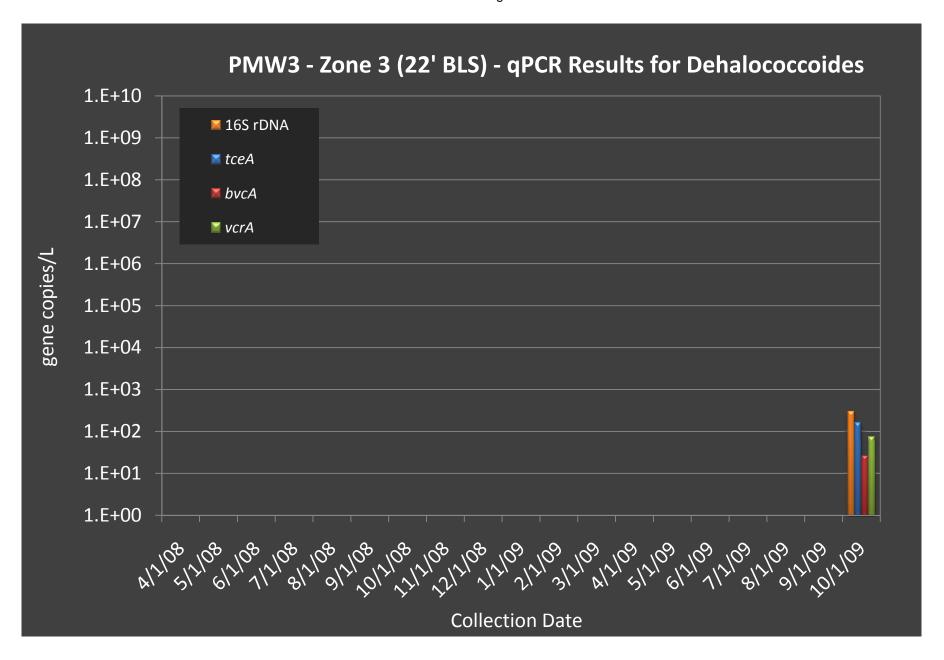


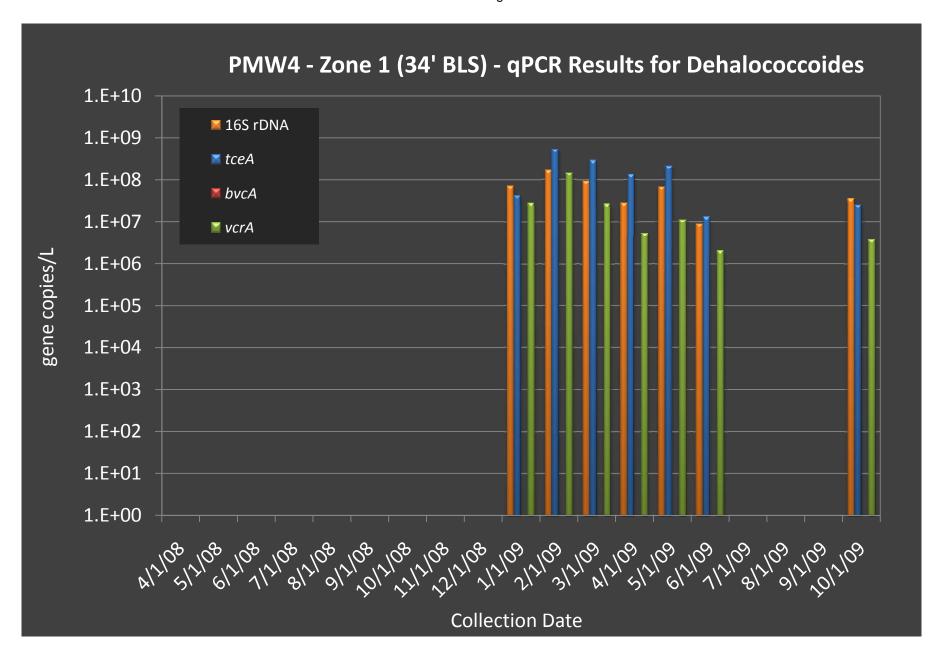


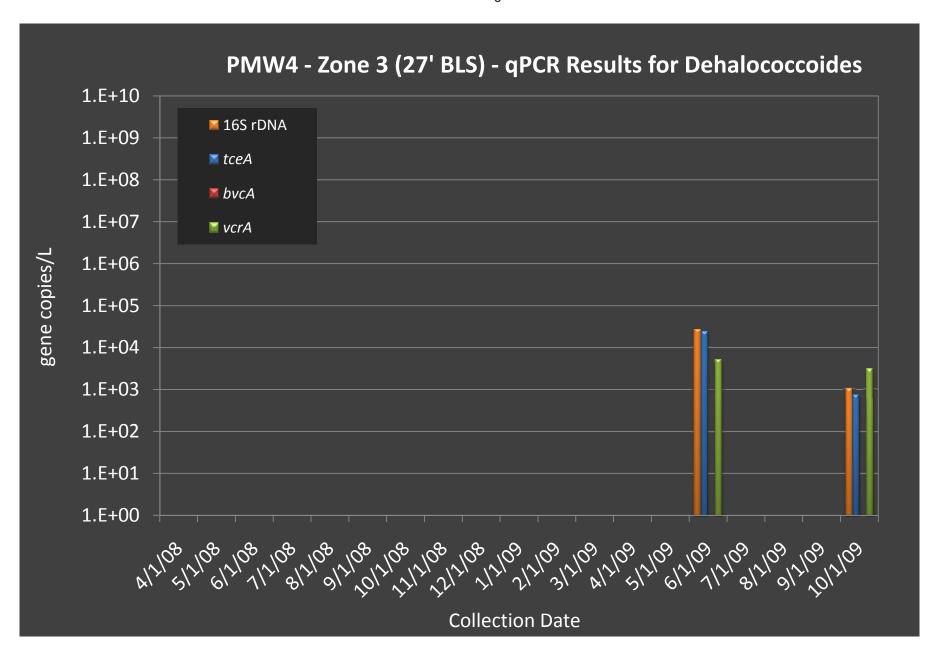


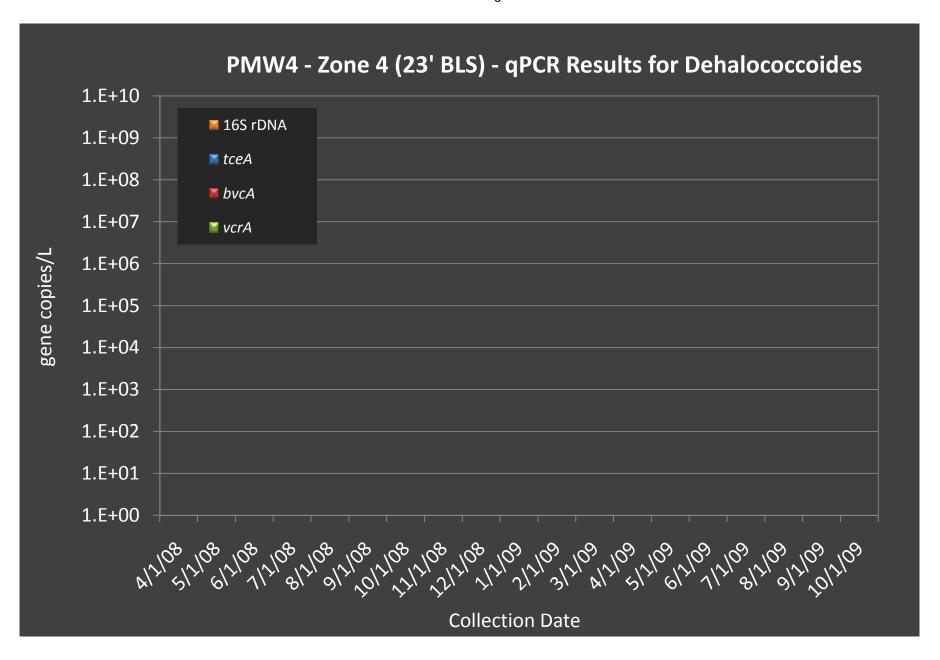


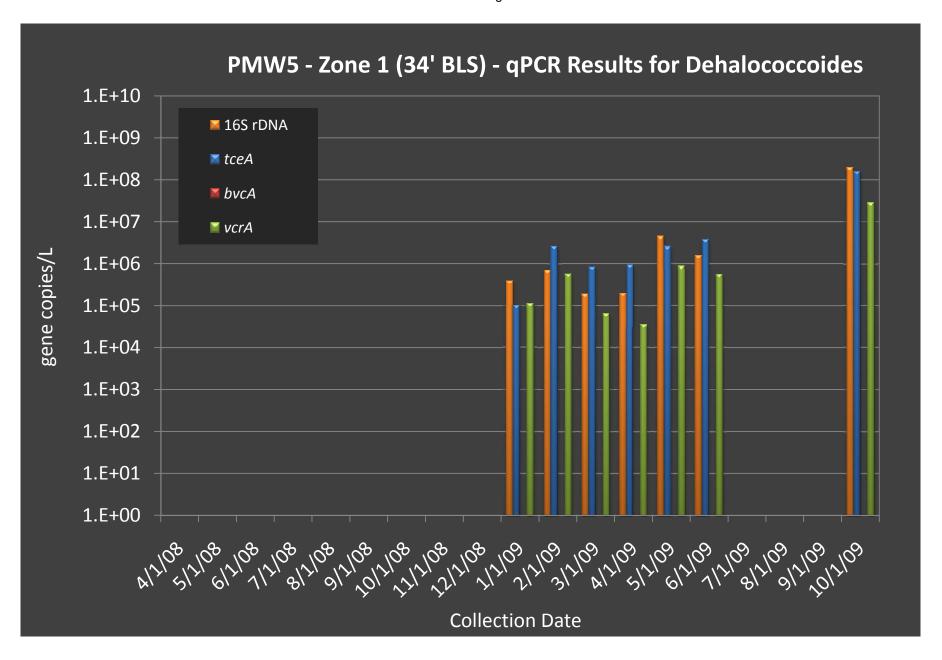


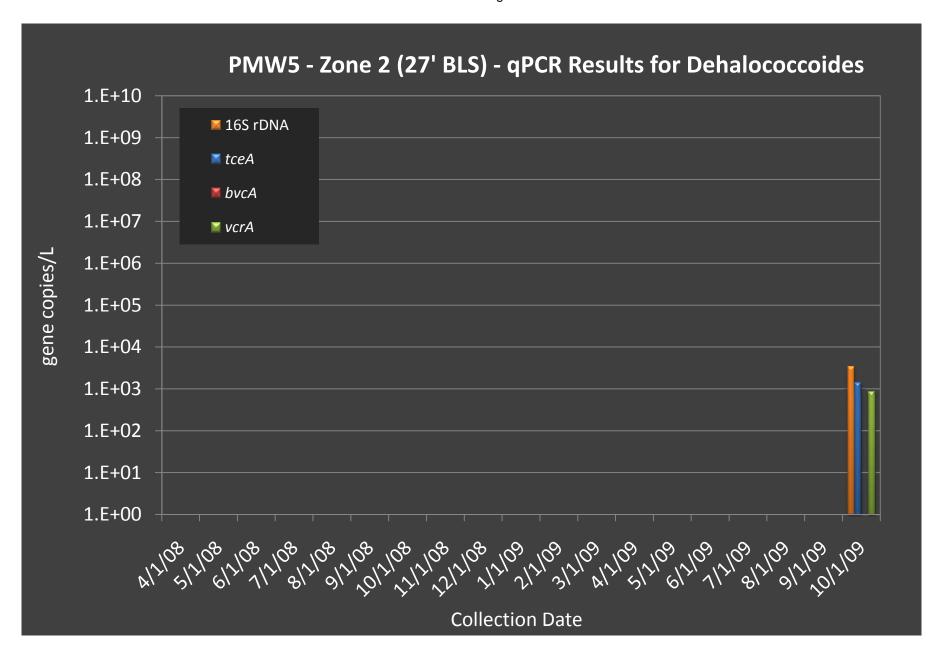


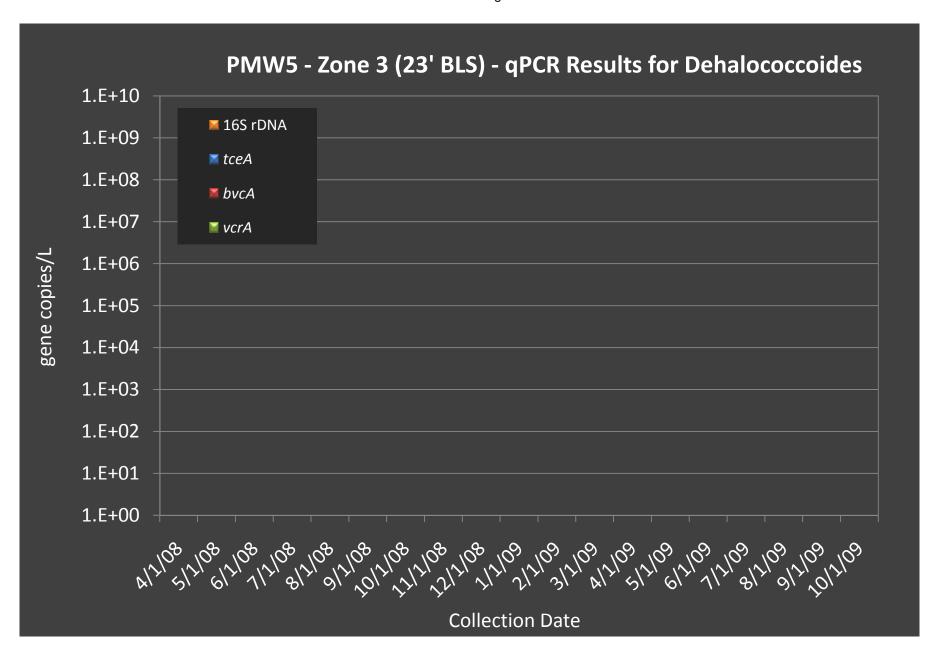


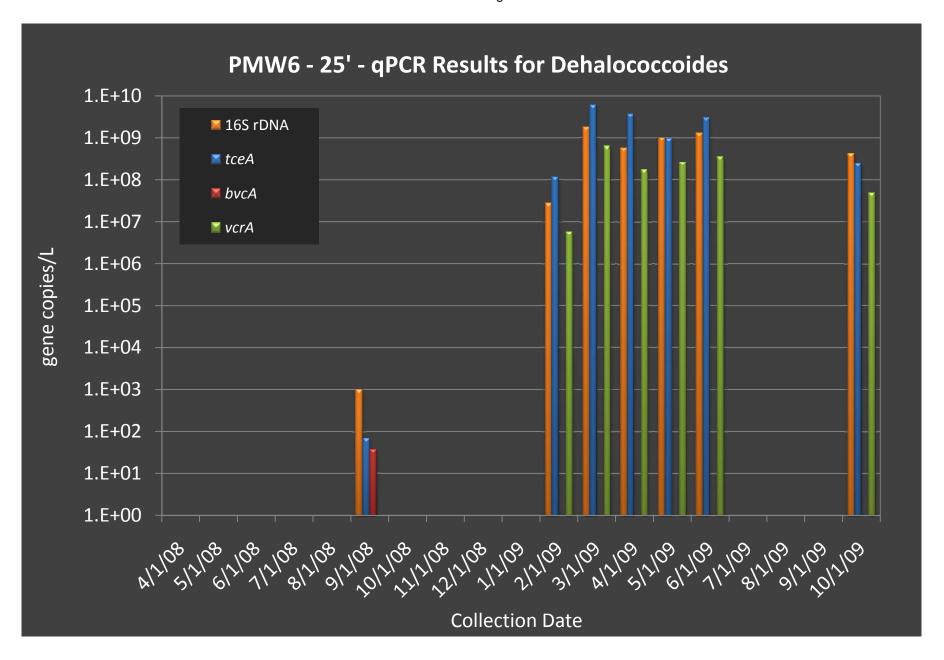


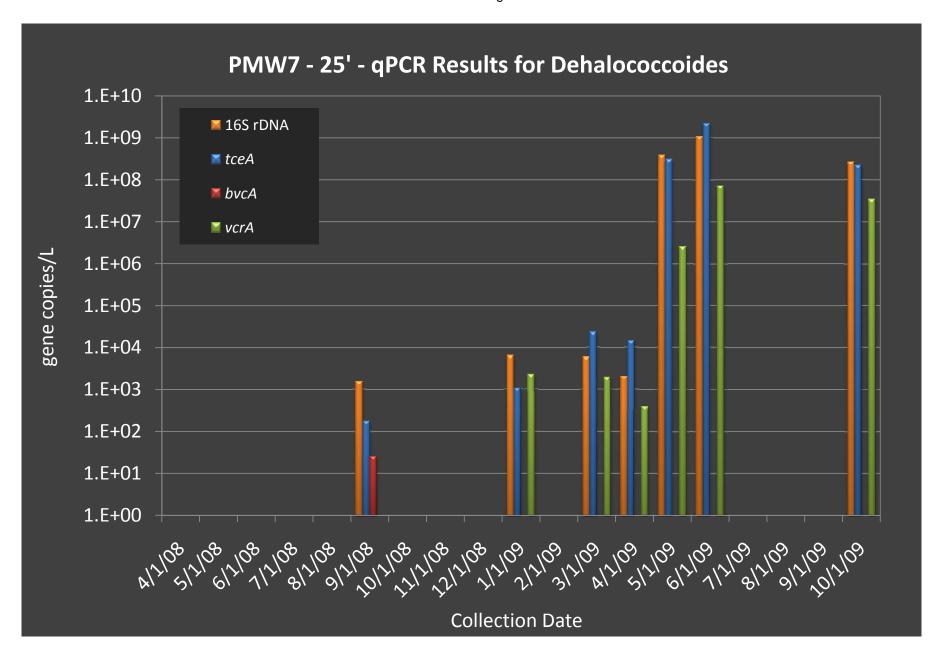


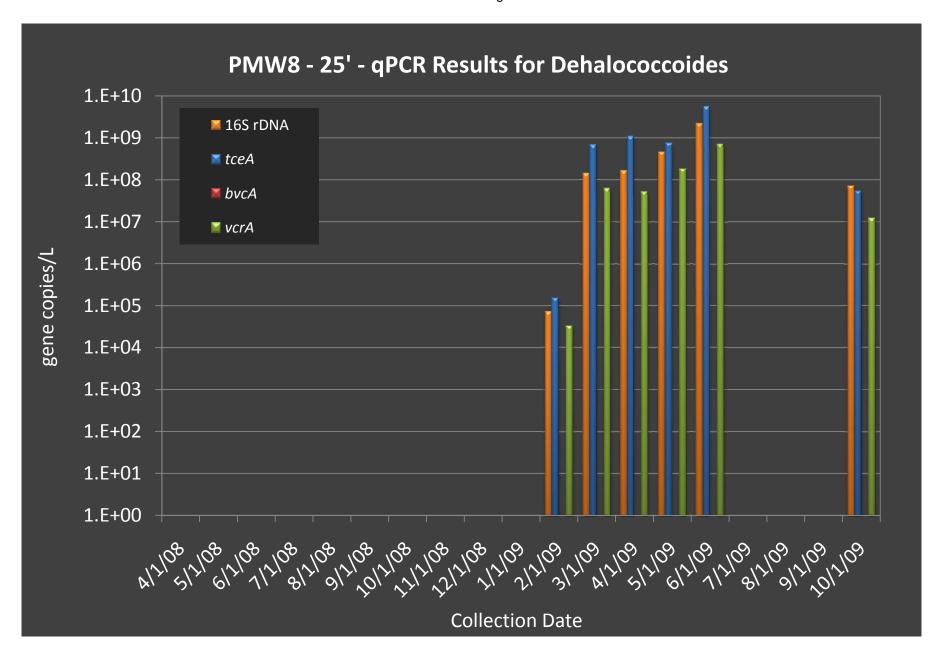


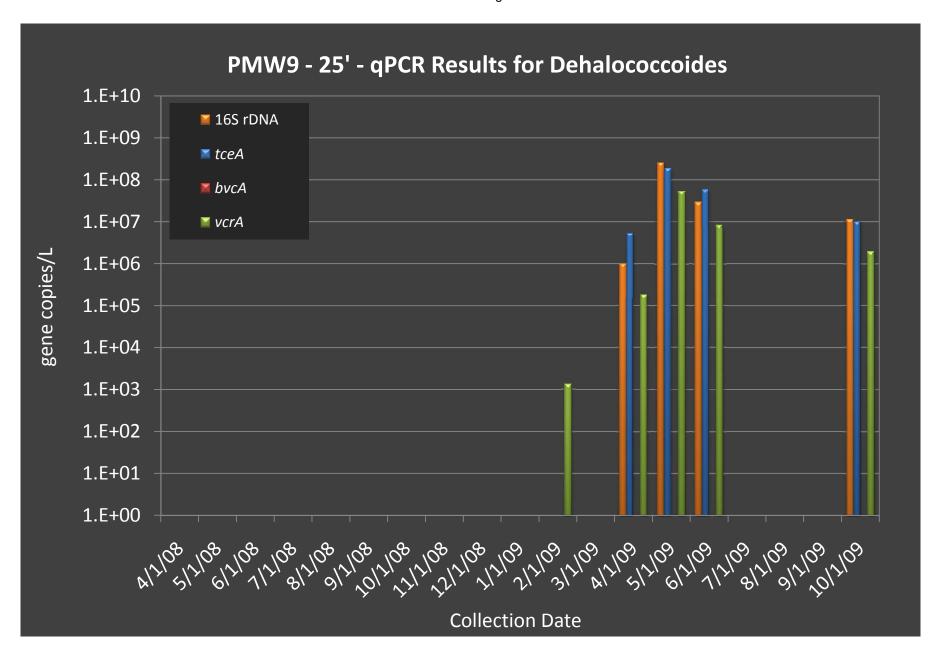












CSIA Results

